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# A mass balance model of the Newfoundland-Labrador Shelf 

by

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Abstract<br>Bundy, Alida, George R. Lilly, and Peter A. Shelton. 2000. A mass balance model of the Newfoundland-Labrador Shelf. Can. Tech. Rep. Fish. Aquat. Sci. 2310:xiv + 157 p..

A mass balance model using the Ecopath approach was constructed for the southern Labrador Shelf, Northeast Newfoundland Shelf and Grand Bank in Northwest Atlantic Fisheries Organisation (NAFO) Divisions 2J+3KLNO for the period 1985-1987. The exercise effected a synthesis of information on biomass, consumption, production and diet of major species or species groups. The paucity of data on biomass and diet of many groups was emphasised. Information on productivity of lower trophic levels was especially weak. Major imbalances were found in the original model. Biomass estimates of some prey species were increased substantially to meet the food requirements of predators. Hooded seals were at the top of the food web. Other high level predators included Greenland halibut, harp seals, Atlantic cod and skates. The major predators in the area were harp seals and Atlantic cod. The dominant planktivorous fish was capelin, but Arctic cod and sand lance were also important.

## Résumé

Bundy, Alida, George R. Lilly, and Peter A. Shelton. 2000. A mass balance model of the Newfoundland-Labrador Shelf. Can. Tech. Rep. Fish. Aquat. Sci. 2310:xiv + 157 p.

Un modèle de bilan massique fondé sur l'approache Ecopath a été établi pour le sud du plateau continental du Labrador, le nord-est du plateau continental de Terre-Neuve et la partie des grands bancs située dans les divisions $2 \mathrm{~J}+3 \mathrm{KLNO}$ de l'Organisation des pêches de l'Atlantique Nord-Ouest (OPANO) pour la période 1985-1987. L’Exercise comportait une synthèse de renseignements sur la biomasse, la consommation, la production et l'alimentation des principals espéces ou des principaux groupes d'espèces. Il a mis en evidence le manque du donées sur la biomasse et sur l'alimentation de nombreaux groupes. On a relevé d'importants déséquilibres dans le modèle original. Les estimations de biomasse de certains espéces proies ont été augmentées notablement, en fonction des besions alimentaire des prédateurs. Les phoques à capuchon occupaient le sommet du réseau trophique. Parmi les autres prédateurs de haut niveau, citons le flétan du Groënland, le phoque du Groënland, la morue et les raies. Dans la région, les grands prédateurs étaient le phoque Groënland et la morue. Le principal possion planctivore était le capelan, mais la morue arctique et le lançon occupaient aussie une place importante.

## Introduction

Mass-balance biomass models are being used globally as an efficient and useful method to systematise ecosystems and explore their properties (Christensen and Pauly 1993; Christensen 1995; Pauly and Christensen 1996). They represent the complexity of an ecosystem in a relatively simple way and are useful as a diagnostic tool. Christensen and Pauly (1993) describe three main benefits of the models: (i) they require researchers to review and standardise all available data on a given system; (ii) they identify states and rates which are mutually incompatible; and (iii) they bring together species specialists and modellers in a collaborative exercise. A fourth benefit is that the approach can highlight knowledge gaps in systems that are considered relatively well studied.

The model presented in this paper originated in a workshop that was held at the Northwest Atlantic Fisheries Centre, St. John's, NF, Canada 14-22 October 1993. Several experts from DFO and the Memorial University of Newfoundland gathered at this meeting with the aim of constructing a steady-state box model for the shelf system off Labrador and eastern Newfoundland. The workshop, structured around the ECOPATH approach of Polovina (1985, ECOPATH I) and Christensen and Pauly (1992, ECOPATH II), benefited from the advice of Dr. V. Christensen (ICLARM, Manila, Philippines), whose participation was supported in part by the Northern Cod Science Program.

The choice of study area involved compromise. For biological reasons, it may have been appropriate to choose areas corresponding to physiographic regions such as the Labrador Shelf, the Northeast Newfoundland Shelf and Grand Bank (Fig 1a), because species composition within the regions is likely to be more homogeneous than among regions. Groundfish assemblages, such as those described by Gomes et al. $(1992,1995)$, may have provided guidance. However, it was also convenient to choose boundaries coinciding with Northwest Atlantic Fisheries Organisation (NAFO) Divisions (Fig. 1b) because commercial fishery statistics are of great importance to the model and are often not available at spatial scales smaller than Division. The first problem with choosing boundaries is that physiographic boundaries and Division boundaries do not coincide. This is most notable in Division 3L, which straddles the boundary between the Northeast Newfoundland Shelf and Grand Bank.

A second consideration in choice of study area was finding contiguous Divisions which corresponded to stock boundaries for the majority of important commercial species. This was considered important because much of the data for commercial species, including population estimates from sequential population analyses, are organised by stock. This consideration also proved difficult to meet (Fig. 2). The largest and most important stock in the area was $2 \mathrm{~J}+3 \mathrm{KL}$ cod, which has its southern boundary across the centre of Grand Bank. However, several other important stocks, most notably 3LNO American plaice, have their northern stock boundary just north of Grand Bank. It was awkward, then, to place a boundary at either the $3 \mathrm{~K} / 3 \mathrm{~L}$ line or the $3 \mathrm{~L} / 3 \mathrm{NO}$ line. With respect to the northern boundary of the study area, several stocks extend to the tip of Labrador, but for some of these stocks the bulk of the biomass occurs south of the $2 \mathrm{H} / 2 \mathrm{~J}$ boundary. Furthermore, there have been few research surveys north of Division 2J. It seemed reasonable to place the northern boundary at the $2 \mathrm{H} / 2 \mathrm{~J}$ line. At the southern end of the range, none of the important species have stocks that
overlap with Subdivision 3Ps, so the 3O/3Ps boundary was taken as the southwestern limit. Flemish Cap (Division 3M) was readily eliminated, because most species occurring on the Cap have stocks that are clearly distinct from those on the adjacent Grand Bank. Greenland halibut, a deepwater species, is an important exception. For the above reasons, it was felt that the study area should be Divisions $2 \mathrm{~J}+3 \mathrm{KLNO}$ (the southern third of the Labrador Shelf, the Northeast Newfoundland Shelf and Grand Bank) from the coast out to the $1,000 \mathrm{~m}$ isobath, a total area of about $495,000 \mathrm{~km}^{2}$.

Because the study area is large and spread across three physiographic regions, there are numerous difficulties with assessing the degree of interaction among the various species. Some species (e.g. haddock, white hake) tend to be associated with the relatively warm water of the southwestern Grand Bank, and tend to interact very little with species north of the 3L/3NO boundary. Other species, such as Greenland halibut and sand lance, tend not to interact with one another because they have different depth requirements. Seasonal migrants are important to the model. Harp and hooded seals and certain birds come into the area from the north in late autumn and remain until spring whereas most whales, many birds, some fish (e.g. mackerel) and short-finned squid come into the area from the south in late spring or early summer and remain until autumn.

The period chosen for the model was 1985 to 1987. This was a time of relatively constant biomass for the major commercial species, in particular those groundfish whose biomasses were low prior to extension of jurisdiction in 1977, increased in the late 1970s and early 1980s, and decreased again during the mid to late 1980s and early 1990s. It is also the period for which there are most data for the diets of demersal fish, mainly because of seasonal sampling in Division 3L in the period 1984-1986. An additional consideration was partial avoidance of the very cold period centred on 1983-1984.

In the present paper, the mass balance model of the Newfoundland-Labrador Shelf that was initiated at the 1993 workshop is greatly revised, updated and presented at an advanced state. The structure of the model has been altered and parameter estimates either updated or made anew. The time period and area of the model remain the same, except that the seaward boundary has been extended from the 500 m isobath to the 1000 m isobath. Many colleagues have contributed data, analyses and advice. These people are listed in Appendix 1, and are identified at the start of those sections to which they contributed.

## Description of approach

The necessary condition in the ECOPATH approach for mass balance is:

$$
\begin{equation*}
P_{i}-M 0_{i}-M 2_{i}-C_{i}=0 \tag{1}
\end{equation*}
$$

where
$P_{i}=$ the annual biomass produced by species $i$
$M O_{i}=$ annual biomass loss of species $i$ not accounted for by any predator group included in the model nor by catch or export, that is, other mortality
$M 2_{i}=$ annual biomass of species $i$ consumed by all predator groups included in the model, that is, predation mortality
$C_{i}=$ annual catch of species $i$.
$M 2_{i}$ is obtained by summing the annual consumption of species group $i$ by all $j$ predator groups included in the model, so that

$$
\begin{equation*}
\mathrm{M} 2_{\mathrm{i}}=\sum \mathrm{p}_{\mathrm{ij}} \mathrm{Q}_{\mathrm{j}} \tag{2}
\end{equation*}
$$

where
$p_{i j}=$ proportion by mass of predator $j$ 's diet that is comprised of prey $i$
$Q_{j}=$ the annual consumption of biomass by predator $j$.
M0, other mortality, is expressed in Ecopath as $1-E E_{i}$ where $E E_{i}$ is the "ecotrophic efficiency". In these terms $1-E E_{i}$ is the proportion of $P_{i}$ that is not consumed by predators included in the model, nor caught by the fishery, so that,

$$
\begin{equation*}
E E_{i}=\left(M 2_{i}+C_{i}\right) / P_{i} \tag{3}
\end{equation*}
$$

The ecosystem is modelled using a set of simultaneous linear equations derived from the above relationships. Equation (1) is scaled by biomass. Thus production and consumption are expressed as biomass ratios ( $\mathrm{P} / \mathrm{B}, \mathrm{Q} / \mathrm{B}$ ). M0 and M 2 are instantaneous mortality rates. Each group in the model is represented by one balanced equation and requires six input parameters. Diet composition and catch (export) must be entered and three of the other four parameters (B, $\mathrm{P} / \mathrm{B}, \mathrm{Q} / \mathrm{B}$, and EE ) must be entered. The linear equations are solved and any missing parameters are estimated. ( EE is estimated if all other parameters have been entered.) Polovina's (1984) original Ecopath may be used to estimate the biomass. However, it is more robust to enter $\mathrm{B}, \mathrm{P} / \mathrm{B}$ and $\mathrm{Q} / \mathrm{B}$ where possible and allow the model to estimate EE . This provides an immediate check for mass-balance because EE cannot be greater than 1. Sometimes it is necessary to enter EE in which case a default value of 0.95 is used (Christensen and Pauly 1992).

In most cases, when all the information to run a model is assembled, the model will not balance. In such cases, the values of one or more of the terms can be altered iteratively until a balance is obtained. This may be done using the Ecoranger routine of Ecopath 3.0. This procedure enables the user to enter a range of values for each parameter, and assign a probability to them. The model is then fitted on a user-based goodness of fit test ${ }^{1}$. Thus, parameters with robust data can be used to constrain the model, while for less certain parameters a range of values may be entered.

There is more than one way to construct an Ecopath model and there is more than one solution to any model. However, if there are areas of the model which are well known and on which the modeller can place some certainty, then the number of plausible solutions is reduced. For the less certain parameters, sensitivity analysis can be used to examine their effects on the model.

[^0]
## Parameter Estimation

A total of 31 groups were defined in the model. Where possible, scientists with specific knowledge of a group were consulted in the estimation of the parameters. In addition, these specialists co-wrote the appropriate parameter estimation sections below. A full list of specialists and their contact information is given in Appendix 1.

## Biomass

Biomass was estimated from sequential population analysis, trawl survey data or hydroacoustics data for most groups in the model. There are two types of trawl survey data, Engels trawl data, which dates up to spring 1995, and Campelen trawl data, which dates from autumn 1995 onwards. The Campelen gear is more efficient at catching small fish and as such, is considered to give a better estimate of true biomass than the Engels data. Engels data have been converted to Campelen units for some species. The Campelen converted data have been used here when available. However, even with the Campelen converted data, biomass estimates are likely to apply only to a certain proportion of biomass, that which is recruited to the sample gear. To allow for the biomass of pre-recruits, two methods were used. The preferred method was to back-calculate numbers at age in the pre-recruits from a known number of first year recruits, assuming a natural mortality rate, and then to multiply by mean weight-at-age to get mean biomass-at-age. The alternative method was to raise the biomass estimate by a "bump-up factor". For those groups that are only present in the model area for part of the year, their biomass was pro-rated by the proportion of the year that they were present in the model area.

## Production

Christensen and Pauly (1992) define production as the total amount of tissue elaborated in the population or community under study during a given time period. Measurement of this production over the time period requires knowledge of the biomass of the population at the beginning and the end of the period and the mass of living components that have been lost by death or emigration during the period. In the Labrador-Newfoundland model it is assumed that there is no year to year change in biomass over the 1985-87 time period and that emigration is zero. Thus production in this model is simply the biomass that is lost to natural mortality and fishing mortality. In many Ecopath models, production, or more specifically, the production to biomass ratio ( $\mathrm{P} / \mathrm{B}$ ), is estimated by assuming that $\mathrm{P} / \mathrm{B}$ is equal to total mortality (Allen 1971). An alternative method is to estimate production from virtual population analysis, or MSVPA, as was done for some species in the North Sea (Christensen 1995). Both methods carry assumptions, as discussed below, and neither explicitly considers gonadic production.

Ricker (1946) provided a method for calculating production from instantaneous body growth rate (in terms of weight) and instantaneous total mortality rate. This approach assumes that the two rates remain constant over the lifespan of the animal (Allen 1971). Allen (1971) examined several models for which mortality and growth vary over the lifespan. While for some models the ratio of $\mathrm{P} / \mathrm{B}$ is a complex function of growth rate and mortality rate, Allen
(1971) found that $\mathrm{P} / \mathrm{B}=\mathrm{Z}$ for models in which total mortality is a constant exponential rate (i.e. $N_{t}=N_{0} e^{-Z t}$ ), irrespective of the form of the body growth rate function. Allen (1971) also provides relationships between $\mathrm{P} / \mathrm{B}$ and mean age or mean life span for different mortality and growth functions. Mertz and Myers (1998) have also examined the relationship between production, biomass and mortality and shown that $\mathrm{P} / \mathrm{B}=\mathrm{Z}$, the biomass averaged total mortality.

Production calculated from virtual population analysis is simply calculated as the difference between the biomass of the population at the beginning and the end of the year. This difference is assumed to be the production, that is, the biomass lost to fishing mortality and natural mortality (predation, disease and other natural causes of death). However, using this approach entails assuming a value for natural mortality. This parameter is not well known, and is estimated by Ecopath as predation mortality, plus other mortality. Assuming a value of M is equivalent to setting a prior natural mortality, assuming a component of production and forcing the model to accommodate this M.

An additional problem with both approaches is that neither deals specifically with the annual elaboration of tissue into gonads. This may be a considerable proportion of annual production for mature fish. Some of it is consumed directly through predation on fish with developed gonads prior to spawning. Some of the gonadic production is consumed or dies immediately after spawning, before fertilisation, while some ends up as icthyoplankton with production, mortality and growth rates similar to that of other components of the zooplankton.
Icthyoplankton are subject to a similar suite of predators as other members of the zooplankton that are in the same size range. Some of the icthyoplankton may be lost through transport in ocean currents out of the system.

In order to investigate methods of estimating production in this model, and to determine whether gonadic production is important, production was estimated for cod and American plaice using 3 approaches: (i) Z was estimated from catch curve analysis of RV survey data, (ii) somatic production was estimated using VPA-based methods and (iii) somatic plus gonadic production was estimated using VPA-based methods.

For the third method, gonadic production was calculated for all mature fish that survived until the end of the year and for $50 \%$ of the mature fish that died. This assumes that they all die half way through the year. Gonadic production was estimated, for each age group, as the weight of the ripe gonad multiplied by the proportion mature at age for each sex, times the number of fish. A 1:1 sex ratio was assumed. Gonad weights were estimated for cod and American plaice. The results are shown in Table 1.

Table 1 Comparison of $\mathrm{P} / \mathrm{B}$ estimates using different methods.

| Method of P/B Estimation | Atlantic cod <br> $($ Ages 3+) | American plaice <br> (Ages 8+) |
| :--- | :---: | :---: |
| Catch Curve | 0.65 | 0.52 |
| VPA-based | 0.38 | 0.47 |
| VPA-based + Gonadic | 0.46 | 0.51 |
| Production |  |  |

The inclusion of gonadic production increases the $\mathrm{P} / \mathrm{B}$ ratio by $17 \%$ for Atlantic cod and $7 \%$ for American plaice. All of the American plaice $\mathrm{P} / \mathrm{B}$ estimates are similar, regardless of which method is used. However, the catch curve estimate of $\mathrm{P} / \mathrm{B}(=\mathrm{Z})$ for Atlantic cod is higher than the two VPA-based methods. Thus for Atlantic cod, including gonadic production does not give the highest estimate of production. In all likelihood, the lower $\mathrm{P} / \mathrm{B}$ estimates from the VPA-based methods reflect the inaccuracy of assuming that natural mortality $=0.2 \mathrm{yr}^{-1}$. On the basis of these results, and to avoid making assumptions about natural mortality, where possible catch curve analysis was used to estimate $\mathrm{P} / \mathrm{B}$, assuming that $\mathrm{P} / \mathrm{B}=\mathrm{Z}$ (Allen 1971). It was not possible to estimate $Z$ this way for all groups. For such groups, other estimates of $Z$ or $\mathrm{P} / \mathrm{B}$ values from the literature were used, or, for commercial species, production was estimated as the sum of the biomass times an assumed natural mortality plus the catch.

The question about the fate of gonadic productivity remains. Here it is assumed that the uncertainty around the catch curve (or other) estimate of $Z$ will allow for gonadic productivity. This is reasonable for species such as American plaice where gonadic production only increased production by $7 \%$. It may also hold for cod-like species. Icthyoplankton, the surviving gonadic products, are included in the large zooplankton group.

## Consumption

Consumption was estimated from field studies where possible, either from the region, or for species from similar areas reported in the literature. In addition, models were used to estimate consumption by marine mammals and seabirds. Where it was not possible to do either, Q/B values were taken from the literature and/or estimated, assuming a P to Q ratio of 0.15 (after Christensen, 1995).

## Diet

The diet of many species varies temporally and spatially. For fish, there may also be considerable variation with size. Thus, to obtain an adequate representation of the annual diet, one would have to obtain stomach content data for the full size range of each species throughout its distribution in the study area with a temporal frequency sufficient to capture any seasonal variability. In addition, there should be appropriate adjustment for variability in population numbers (or biomass) at length and for variability in digestive rate associated with prey type and temperature. For some species (e.g. harp seals, Atlantic cod, Greenland halibut and American plaice) the participants had access to stomach content data collected from the study area during 1985-1987. For none of these species have the stomach collections covered the whole geographic area throughout the year. For the fish species, the average diets are calculated from the available stomach content data without adjustments for population size structure and variability in digestion rates. For many species or groups, there were no stomach content data available to the authors. In some of these cases diet information was available from literature reports from the study area, although often from time periods other than 19851987. In the worst cases (eg. small demersal-feeding fish), it was necessary to make educated guesses based solely on general diet descriptions from outside the study area.

## Species groups

Species groups were defined on the basis of their commercial significance and importance as predators or prey.

## Size Groupings

Atlantic cod, Greenland halibut and American plaice were each separated into two groups based on diet, age/size of first capture and age/size of maturity. Smaller animals prey mainly on invertebrates and larger animals prey mainly on fish. These changes tend to occur gradually with increasing length, but for this model it was assumed that the change occurs at 35 cm for Atlantic cod (Lilly and Fleming 1981; Lilly 1991) and American plaice (Pitt 1973) and 40 cm for Greenland halibut (Bowering and Lilly 1992).

## Input data by Ecopath group

The input parameters (Biomass, $\mathrm{P} / \mathrm{B}$ and $\mathrm{Q} / \mathrm{B}$ ) are given in Table 2a. Also given are the EEs and GEs. Some of these are input parameters, whilst others are estimated by the model (in italics). The diet matrix is given in Table 2b.

Table 2a. Initial input parameters for the Ecopath model. The values in italics were estimated by the model. The members of each group are described in the relevant section of the text. L.Dem.Feeders are large demersal feeders. S.Dem.Feeders are small demersal feeders. L.Pel.Feeders are large pelagic feeders. Pisc. SPF and Plankt. SPF are piscivorous and planktivorous small pelagic feeders. O.Benthic Inver are other benthic invertebrates.

| Group Name | Biomass $t \cdot k m^{-2}$ | $\begin{aligned} & P / B \\ & y r^{-1} \end{aligned}$ | $\begin{aligned} & \text { Q/B } \\ & y r^{-1} \end{aligned}$ | EE | GE |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1. Whales | 0.251 | 0.100 | 11.794 | 0.000 | 0.008 |
| 2. Harp Seals | 0.184 | 0.102 | 17.412 | 0.053 | 0.006 |
| 3. Hooded Seals | 0.034 | 0.109 | 13.100 | 0.000 | 0.008 |
| 4. Seabirds | 0.012 | 0.250 | 54.750 | 0.333 | 0.005 |
| 5. Cod $>35 \mathrm{~cm}$ | 2.044 | 0.651 | 3.240 | 0.516 | 0.201 |
| 6. $\mathrm{Cod}<=35 \mathrm{~cm}$ | 0.094 | 0.600 | 6.090 | 7.744 | 0.099 |
| 7. G. halibut $>40 \mathrm{~cm}$ | 0.348 | 0.508 | 1.478 | 0.361 | 0.344 |
| 8. G. halibut $<=40 \mathrm{~cm}$ | 0.165 | 0.247 | 3.401 | 9.641 | 0.073 |
| 9. Aplaice $>35 \mathrm{~cm}$ | 0.972 | 0.538 | 1.262 | 0.191 | 0.426 |
| 10. Aplaice < $=35 \mathrm{~cm}$ | 0.784 | 0.625 | 3.736 | 0.530 | 0.167 |
| 11. Flounders | 0.868 | 0.394 | 3.600 | 1.150 | 0.109 |
| 12. Skates | 0.517 | 0.286 | 2.878 | 0.264 | 0.099 |
| 13. Redfish | 0.975 | 0.330 | 2.000 | 1.789 | 0.165 |
| 14. L.Dem.Feeders | 0.845 | 0.262 | 1.747 | 0.676 | 0.150 |
| 15. S.Dem.Feeders | 0.227 | 0.300 | 2.000 | 15.139 | 0.150 |
| 16. Capelin | 13.289 | 1.145 | 6.400 | 0.740 | 0.179 |
| 17. Sand lance | 2.103 | 1.150 | 7.667 | 1.196 | 0.150 |
| 18. Arctic cod | 2.729 | 0.395 | 2.633 | 0.913 | 0.150 |
| 19. L.Pel.Feeders | 0.029 | 0.400 | 3.333 | 0.950 | 0.120 |
| 20. Pisc. SPF | 0.453 | 0.265 | 1.767 | 6.805 | 0.150 |
| 21. Plankt. SPF | 0.950 | 0.240 | 1.600 | 4.378 | 0.150 |
| 22. Shrimp | 0.202 | 1.450 | 9.667 | 2.756 | 0.150 |
| 23. Large Crustacea | 0.185 | 0.282 | 5.850 | 20.874 | 0.048 |
| 24. Echinoderms | 112.300 | 0.600 | 6.667 | 0.022 | 0.090 |
| 25. Molluscs | 42.100 | 0.570 | 6.333 | 0.035 | 0.090 |
| 26. Polychaetes | 10.500 | 2.000 | 22.222 | 0.116 | 0.090 |
| 27. O.Benthic Inver | 7.800 | 2.500 | 12.500 | 0.115 | 0.200 |
| 28. Lge.Zooplankton | 18.343 | 3.433 | 19.500 | 1.095 | 0.176 |
| 29. Sm.Zooplankton | 21.700 | 6.200 | 20.667 | 1.729 | 0.300 |
| 30. Phytoplankton | 26.860 | 93.100 | 0.000 | 0.232 | - |
| 31. Detritus | 389.000 | - | - | 0.576 | - |

Table 2b. Diet matrix for the initial Ecopath model.

| Prey \Predator | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 1. Whales | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Table 2b. Diet matrix for the initial Ecopath model (cont.).

| Prey \ Predator | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 | 27 | 28 | 29 | 30 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1. Whales | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 00 |
| 2. Harp Seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3. Hooded Seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 4. Seabirds | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 5. C | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 6. Cod $<=35$ | 0.000 | 0.000 | 0.000 | 0.002 | 0.019 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 7. G.halibut $>40 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 8. G.halibut $<=40 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 9. Aplaice $>35 \mathrm{~cm}$ | 0.0 | 0.00 | 0.000 | 0. | 0. | 0. | 0. | 0. | 0.000 | 0. | 0.000 | 0.000 | 0.000 | 0.000 | 0 |
| 10. Aplaice $<=35 \mathrm{c}$ | 0.000 | 0.000 | 0.000 | 0.00 | 0.00 | 0.0 | 0.0 | 0. | 0.000 | 0. | 0.000 | 0.000 | 0.000 | 0.000 | 0 |
| 11. Flounder | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 12. Skates | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 13. Red | 0.000 | 0.00 | 0.000 | 0.0 | 0.0 | 0. | 0. | 0.000 | 0.000 | 0. | 0.000 | 0.000 | 0.000 | 0.000 | 0 |
| 14 | 0.0 | 0.00 | 0.00 | 0. | 0. | 0. | 0. | 0. | 0. | 0. | 0. | 0.000 | 0. | 0 | 0 |
| 15. S.Dem.Feeder | 0.000 | 0.000 | 0.000 | 0.035 | 0.000 | 0.000 | 0.00 | 0.050 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 16. Capelin | 0.010 | 0.000 | 0.038 | 0.075 | 0.698 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 17. Sand lanc | 0.0 | 0.00 | 0.000 | 0.08 | 0. | 0.00 | 0.0 | 0. | 0.000 | 0. | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 18. Arctic | 0.0 | 0.000 | 0.002 | 0.00 | 0.0 | 0. | 0. | 0.000 | 0.000 | 0. | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 19. L.Pel.Feede | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 20. Pisc. SPF | 0.000 | 0.000 | 0.000 | 0.273 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 21. Plankt. S | 0.000 | 0.000 | 0.000 | 0.18 | 0.083 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 22. Shrimp | 0.000 | 0.000 | 0.000 | 0.012 | 0.008 | 0.005 | 0.000 | 0.049 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 23. Large Crustacea | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.111 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 24. Echinoderms | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.200 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 25. Molluscs | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.198 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 26. Polychaetes | 0.000 | 0.000 | 0.000 | 0.003 | 0.000 | 0.000 | 0.015 | 0.196 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 27. O.Benthic Inver | 0.000 | 0.000 | 0.000 | 0.019 | 0.000 | 0.050 | 0.015 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 28. L.Zooplankton | 0.434 | 0.350 | 0.640 | 0.295 | 0.168 | 0.526 | 0.120 | 0.049 | 0.000 | 0.000 | 0.000 | 0.000 | 0.050 | 0.000 | 0.000 |
| 29. S.Zooplankton | 0.546 | 0.650 | 0.320 | 0.012 | 0.013 | 0.419 | 0.240 | 0.049 | 0.000 | 0.000 | 0.000 | 0.000 | 0.480 | 0.000 | 0.000 |
| 30. Phytoplankton | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.085 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.370 | 1.000 | 0.000 |
| 31. Detritus | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.525 | 0.097 | 1.000 | 1.000 | 1.000 | 1.000 | 0.100 | 0.000 | 0.000 |

## 1. Whales and porpoises (Stenson, Lawson and Bundy)

## Background

The main piscivorous whales in the study area are humpback (Megaptera novaeangliae), fin (Balaenoptera physalus), minke (Balaenoptera acutorostrata), sei (Balaenoptera borealis), sperm (Physeter catodon) and pilot (Globicephala melaena). The main porpoise species is the harbour porpoise (Phocoena phocoena). The only baleen whale in the area that does not feed on fish is the planktivorous blue whale (Balaenoptera musculus) (Gaskin 1982).

## Catch

There are no catches of these species recorded for the study area for the 1985-87 period. Canada ceased commercial whaling in 1972.

## Biomass

Population sizes have been estimated at: humpback - 3,300, backcalculated from a total of 10,000-11,000 in the NW Atlantic in 1992/3 and assuming 45-50\% in Newfoundland waters and a 6.7 \% growth rate (Barlow and Clapman 1997; P. Hammond, Sea Mammal Research Unit, UK. pers. comm.); fin - 1,000 (Hay 1982; Mitchell 1974); minke - 5,000 (based on a range of 74,700 to 145,000 in the northern Atlantic, Anon 1991); sei-1,000 (Mitchell and Chapman 1977); sperm - 1,000 (Braham 1984); pilot -9,000 (Nelson and Lien 1996, projected from 4,000 in 1972 using a $6 \%$ net recruitment rate); harbour porpoise - 20,000 (Gaskin 1992, backcalculated from 1992 using $3 \%$ growth rate and assuming $50 \%$ of the Newfoundland population was in 2J3KLNO, B. Sjare, pers. comm.); blue whale - 200 (Mitchell 1974; the 1974 value was taken as representative of the average population in the 1985-1987 period).

A range of mean body weights appear in the literature, see Table 3. In the absence of definitive mean weights, an average of these values was taken. The Martin (1990) estimate for the sperm whale was used however, because only males migrate as far north as Newfoundland and these are heavier than the females (P. Hammond, pers. comm.).

Table 3 Mean body weights ( t ) of whales from various literature sources.

|  | Miscellaneous Kenney et al (1997) | Martin (1990) | Average | Revised |  |
| :--- | ---: | ---: | ---: | ---: | ---: |
| Humpback | $32^{1}$ | 25 | 36 | 31 | 31 |
| Fin | $40.5^{2}$ | 30 | 45 | 38.5 | 38.5 |
| Minke | $6.3^{2}$ | 4.5 | 6 | 5.6 | 5.6 |
| Sei | $15^{2}$ | 13 | 15 | 14.3 | 14.3 |
| Sperm | $28^{2}$ | 20 | 45 | 31 | 45 |
| Pilot | $1.8^{2}$ | 0.85 | 1.5 | 1.4 | 1.4 |
| Harbour | $0.065^{2}$ | 0.045 | 0.05 | 0.05 | 0.05 |
| Blue | $70^{3}$ | 70 | 90 | 76.7 | 76.7 |

${ }^{1}$ Hay (1985), ${ }^{2}$ Lien (1985), ${ }^{3}$ Lien (1985), Ichihara (1966)
This gives biomass values ( t ) of: humpback - 102,300, fin - 38,500 , minke $-28,000$, sei 14,333 , sperm $-45,000$, pilot $-12,450$, harbour porpoise $-1,067$, and blue whale $-15,333$ and
a total biomass for the area of $251,983 \mathrm{t}$. Assuming a residency time of 125 days for the blue whale and 180 days for all other whales within the study area, the average annual biomass was $124,421 \mathrm{t}, 0.251 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Production:Biomass

Annual production within the study area was calculated by assuming a production to biomass ratio of $10 \%$, giving 17,993 t.

## Consumption:Biomass

Annual consumption by species was calculated assuming a residency time of 180 days ( 125 for blue whale) and percentages of body weight or tonnes consumed per day per individual of: humpback, fin, minke, and sperm - 3\% (Lockyer 1981), sei - 0.9 t (Lien 1985), pilot - $3 \%$ (assuming that its rate is the same as other cetaceans whose mean weight is greater than 0.6 t , Gaskin (1982:85)), harbour porpoise -0.0045 t (Lien 1985) and blue whale, 2.7 t in the Northwest Atlantic (Lien 1985). This gives an annual consumption by species group of: humpback - 552,420 t; fin - 207,900 t; minke - 151,200 t; sei - 162,000 t; sperm-243,000 t; pilot $-67,230 t$; harbour porpoise $-16,200 \mathrm{t}$ and blue whale $67,500 \mathrm{t}$. This gives a mean annual consumption of $1,467,450 \mathrm{t}$ and an annual $\mathrm{Q} / \mathrm{B}$ ratio of $11.79 \mathrm{yr}^{-1}$.

## Diet

There are few quantitative descriptions of diet. Where the literature refers to a prey using the terms "preponderant" or "predominant", or similar terms, it was assumed that this species represents at least $75 \%$ of consumption by weight (Table 4). If other prey species are reported, the remaining consumption was divided equally among them. Based on the literature, the following diets were used in the analysis: humpback - capelin, euphausiids, squid and sand lance (Mitchell 1973); fin - capelin, sand lance, herring and euphausiids (Mitchell 1975); minke - capelin, pre-recruited cod, herring, squid and euphausiids (Horwood 1990, after Mitchell 1974); sei - copepods, euphausiids, capelin and sand lance (Mitchell 1975); sperm - squid, large demersal feeders, small demersal feeders (Gaskin 1982); pilot squid, pre-recruited cod and capelin (Lien 1985); harbour porpoise - capelin, pre-recruited cod and mackerel (Lien 1985); blue whale - large zooplankton species Thysanoessa inermis, Temora longicornis and Meganyctiphanes norvegica (Gaskin 1982; Mitchell 1975).

In order to calculate the overall proportion of each prey type by weight in the whale diet, total consumption by prey type was first calculated, and then the overall proportions, as shown in Table 4. These are shown by the groupings used in the Ecopath model.

Table 4. Diet composition of whales in the study area.

| Ecopath Group | Humpback | Fin | Minke | Sei | Sperm | Pilot | Harbour Seal |  | Mean Diet |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 6 Cod <= 35 cm | 0 | 0 | 0.050 | 0 | 0 | 0.125 | 0.125 | 0 | 0.012 |
| 14 Demersal feeders (misc. large) | 0 | 0 | 0 | 0 | 0.200 | 0 | 0 | 0 | 0.033 |
| 15 Demersal feeders (misc. small) | 0 | 0 | 0 | 0 | 0.200 | 0 | 0 | 0 | 0.033 |
| 16 Capelin | 0.750 | 0.750 | 0.750 | 0.083 |  | 0.125 | 0.750 | 0 | 0.489 |
| 17 Sand lance | 0.083 | 0.083 | 0 | 0.083 | 0 | 0 | 0 | 0 | 0.052 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.083 | 0 | 0.050 | 0 | 0.550 | 0.750 | 0.125 | 0 | 0.163 |
| 21 Plankt. Pelagic feeders (misc. small) | 0 | 0.083 | 0.1 | 0 | 0.050 | 0 | 0 | 0 | 0.030 |
| 28 Zooplankton (large) | 0.083 | 0.083 | 0.050 | 0.083 | 0 | 0 |  | 1.000 | 0.104 |
| 29 Zooplankton (small) | 0 | 0 |  | 0.750 | 0 | 0 | 0 | 0 | 0.083 |
| TOTAL | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |  |

## 2. Harp seals (Stenson and Bundy)

## Background

The northwest Atlantic population of harp seals (Phoca groenlandica) comprises two herds, one that breeds in the Gulf of St. Lawrence (Gulf herd) and on that breeds on the pack ice off southern Labrador (Front herd). Generally, harp seals summer in the Canadian Arctic or Greenland, and winter in Atlantic Canadian waters. Both herds are assumed to spend a portion of the year in the model area ( 2 J 3 KLNO ) although the residency period for Gulf seals is brief.

## Catch

Annual removals (total numbers) from the Front herd were taken as reported in the NAFO Statistical Bulletins (NAFO 1987, 1989, 1990). The number of young of the year removed was obtained directly from the catch statistics. The age structure of older (1 year of age and older) seals in the harvest was estimated using aged sub-samples of the catch for the period 1985-1987 (Sjare et al. 1996). Numbers at age were multiplied by the mean weight at age, (Chabot et al. 1996, estimated from a Gompertz growth curve). The total mean annual catch for $1985-1987$ was estimated as 441.6 t , or $0.0009 \mathrm{t} \cdot \mathrm{km}^{-2}$ for the total area of 2 J 3 KLNO .

## Biomass

Biomass was estimated from the abundance, multiplied by the mean weight at age as described above. The number of seals in each age class was obtained from the population model of Shelton et al. (1996a). Formulation 2 was used, where pup mortality was assumed to be 3 times the mortality rate of the rest of the population. The number of seals in each age group present in the 2 J 3 KLNO area was estimated from the seasonal distribution of harp seals, given in Hammill and Stenson (1998). The mean annual biomass for 1985-1987 was calculated as $91,198.4 \mathrm{t}$, or $0.184 \mathrm{t} \cdot \mathrm{km}^{-2}$ for the total area.

## Production:Biomass

It was assumed that the production:biomass ratio is equivalent to the total mortality rate, Z (Allen 1971). Total mortality was estimated from a weighted mean of the survival rates in Shelton et al. (1996a), giving a P/B of $0.102 \mathrm{yr}^{-1}$, and a total annual production of 9340.2 t .

## Consumption:Biomass

Total annual consumption of each prey species in the study area was estimated from the consumption model of Stenson et al. (1997) and modified by Hammill and Stenson (1998). The average annual consumption was $1,588,111 \mathrm{t}$, producing a consumption:biomass ratio of $17.4 \mathrm{yr}^{-1}$.

## Diet

The harp seal diet is an average diet derived from samples obtained from 1982-1995. Diets were applied seasonally and separate diets are used for seals in nearshore (Lawson and Stenson, 1995; Lawson et al. 1995) and offshore areas (Lawson and Stenson 1997). The diets assumed are summarized in Hammill and Stenson (1998). The relative proportions of the split pools prey groups (Atlantic cod, Greenland halibut and American plaice) are estimated from selectivity data in Lawson and Stenson (1997) and Lawson et al. (1995). The composition of the "other fish" and "other invertebrate" groups in Hammill and Stenson (1998) were determined from the original papers. The diet composition is given in Table 2b.

## 3. Hooded seals (Stenson, Hammill and Bundy)

## Background

Hooded seals (Cystophora cristata) are larger than harp seals but much less abundant within the study area. Whelping takes place on pack ice off northeast Newfoundland and in Davis Strait and the Gulf of St. Lawrence. Little is known about stock relationships between the 3 groups. All pups remain in Arctic waters, having migrated there after whelping, and most remain as juveniles (1-3 years old). Hooded seals migrate north out of the study area in April and return in the fall.

## Catch

The annual number of hooded seals harvested in 1985-87 was obtained from Anon (1998). An average of 891 seals were caught each year but numbers varied greatly among years. In 1986, for example, there was a very low catch of hooded seals. The number of pups harvested was 21 while in 1987 it was 1321 . The number of $1+$ seals in the catch was estimated by dividing the catch by the proportion at age in the population obtained from a population model (Hammill and Stenson 1998). Catch weight was then estimated using the Gompertz growth equation for hooded seals described in Hammill and Stenson (1998). The mean annual weight of hooded seals caught was 19.5 t of pups and 69.1 t of $1+$ seals, making a total catch of 88.7 t of hooded seals, or $0.000179 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Biomass

The number of seals in each age class was estimated using the population model presented in Hammill and Stenson (1998), back calculated to the period 1985-1987 using an annual growth rate of $4.8 \%$, and adjusted for seasonal distribution (Hammill and Stenson 1998). Numbers were then converted to biomass-at-age using the Gompertz growth equation, as above. Hooded seal pups are only resident in the area for about a month. To allow for this, the population number of hooded seal pups was divided by twelve to obtain the mean annual biomass in the area. Pups were assumed to undergo their post-weaning fast when they are in 2J3KLNO and therefore are not included in the consumption estimate below. The total mean annual biomass of hooded seals is $16,810 \mathrm{t}$, or $0.034 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Production:Biomass

It was assumed that the production:biomass ratio is equivalent to the total mortality rate, Z (Allen 1971). Total mortality was estimated from the survival rates in Hammill and Stenson (1998), weighted by the biomass of juveniles and adults. This gives a mean $\mathrm{P} / \mathrm{B}$ of $0.109 \mathrm{yr}^{-1}$, and a total annual production of 1829 t .

## Consumption:Biomass

Total annual consumption was estimated from the consumption model of Hammill and Stenson (1998) and back-calculated from 1990 to the 1985-1987 period (Stenson, unpubl.). The average annual consumption was $220,209 \mathrm{t}$, producing a consumption:biomass ratio of $13.1 \mathrm{yr}^{-1}$.

## Diet

The diet information was taken from Hammill and Stenson (1998), which is based on Ross (1993) and Lawson and Stenson (unpublished data). Ross (1993) reconstructed the diet of hooded seals using 132 prey-containing stomachs (collected mainly in springs of 1982-1991). The Atlantic cod and Greenland halibut components of the diet were split into the two size groups, based on selectivity data in Ross (1993). The main components of the diet are young Greenland halibut, flounders and small pelagic feeders, see Table 2b.

## 4. Seabirds (Montevecchi)

## Background

There are large seasonal fluxes of seabird species and populations in the study area throughout the year. The major species that breed in the region are Leach's storm-petrel (Oceanodroma leucorhoa), northern gannet (Sula bassana), ring-billed gull (Larus delawarensis), herring gull (L. argentatus), great black-backed gull (L. marinus), black-legged kittiwake (Rissa tridactyla), common tern (Sterna hirundo), arctic tern (S. paradisaea), common murre (Uria aalge), thick-billed murre (U. lomvia), razorbill (Alca torda), and Atlantic puffin (Fratercula arctica) (Montevecchi and Tuck 1987, Cairns et al. 1989). Other species that breed in the study area in small numbers are northern fulmar (Fulmarus glacialis), Manx shearwater (Puffinus puffinus), great cormorant (Phalacrocorax carbo), double-crested cormorant ( $P$. auritus), common black-headed gull (Larus ridibundus), Caspian tern (Sterna caspia), and
black guillemot (Cepphus grylle).
The major species that move into the region in large numbers for substantial periods, primarily during summer, but do not breed in the study area, are the transequatorial migrants - greater shearwater (Puffinus gravis), and sooty shearwater (P. griseus). Species that breed in the Arctic and Northeast Atlantic and move into the region in large numbers during winter are northern fulmar, Iceland gull (Larus glaucoides), glaucous gull (L. hyperboreus), black-legged kittiwake, dovekie (Alle alle) and thick-billed murre. In the present treatment, the seaducks and common eider (Somateria dresseri) are included with species that breed in the region. Common eider, oldsquaw (Clangula hyemalis) and scoters (Melanitta spp.) are included with species that breed primarily outside the study area.

## Catch/Anthropogenic Mortality

There are three primary additive sources of anthropogenic mortality on seabirds in the region: 1) by-catch in fishing gear, 2) hunting, and 3) oil pollution (Montevecchi and Tuck 1987). Considerable numbers of seabirds, (mostly alcids, i.e. murres and puffins, but also others, e.g. gannets) are caught as by-catch in fishing gear. The significance of this mortality may be indicated in part by population increases of murres in the Witless Bay Seabird Ecological Reserve in eastern Newfoundland that appear to have occurred since 1992, when fixed fishing gear was removed from coastal waters following the Eastern Canadian Ground-fishery Moratorium. It has been estimated that each winter as many as 500,000 thick-billed (and common) murres are killed by hunters in Newfoundland and Labrador (Montevecchi and Tuck 1987, Elliot et al. 1990). This hunting pressure has likely decreased in the 1990s with the introduction of seasonal and bag limits by the Canadian Wildlife Service. Additionally, tens of thousands of seabirds and seaducks are estimated to be killed each year by illegal discharges of oil by ships, and this mortality appears to be greatest in winter. On the basis of this information, a "guesstimate" of 500,000 seabirds x mean mass of $0.93 \mathrm{~kg}=465 \mathrm{t}, 0.001$ $t \cdot \mathrm{~km}^{-2}$, of seabirds being removed from the study area annually through by-catch, hunting and oiling.

## Biomass

Data on body masses and population estimates were derived from Montevecchi et al. (1984, 1987, 1988a), Nettleship and Chapdelaine (1988), Diamond et al. (1986, 1993), Cairns et al. (1989), Sklepkovych and Montevecchi (1989), Storey and Lien (1985), Stenhouse and Montevecchi (1999) and Montevecchi (unpublished data). For species that breed within NAFO Areas 2J3KLNO, population estimates $=$ estimated breeding pairs x $2+[0.30 \mathrm{x}$ breeding pairs x 2$]$ non-breeders $+[0.80 \mathrm{x}$ breeding pairs (fulmar, Manx shearwater, stormpetrel, gannet, cormorants, murres, puffin)] or [1.0 x breeding pairs (gulls, terns, guillemot) or 2.0 x breeding pairs (eider)].

Population estimates and durations of seabird species are summarised in Table 5. The average body mass and total biomass are given for each species. The overall biomass estimate for the study area is 6101 t , or $0.012 \mathrm{t} \cdot \mathrm{km}^{2}$.

Table 5. Approximate occupation dates, estimated population numbers, average body masses and average annual biomass for the main species of seabird that occur within the study area (NAFO Divisions 2J3KLNO) and A) breed there or B) breed primarily or completely outside of the study area. Note that northern fulmar, common eider, thick-billed murre and dovekie occur in both partitions of the table.
A) Seabird species that breed within NAFO Divisions 2J3KLNO.

| Species | Occupation dates | Population (number) | Individual mass <br> (kg) | $\qquad$ biomass over year (t) |
| :---: | :---: | :---: | :---: | :---: |
| Northern fulmar | Jan-Dec | 181 | 0.80 | 0.14 |
| Manx shearwater | Mar-Nov | 340 | 0.48 | 0.12 |
| Leach's storm-petrel | Apr-Oct | 15,340,636 | 0.05 | 447.44 |
| Northern gannet | Apr-Oct | 48,806 | 3.20 | 91.10 |
| Great cormorant | Mar-Nov | 601 | 2.25 | 1.01 |
| Double-crested cormorant | Mar-Oct | 1,048 | 2.33 | 1.63 |
| Common eider | Jan-Dec | 2,344 | 2.23 | 5.23 |
| Black-headed gull | Jan-Dec | 25 | 0.28 | 0.01 |
| Ring-billed gull | Apr-Oct | 23,062 | 0.50 | 6.73 |
| Herring gull | Jan-Dec | 151,787 | 1.12 | 170.00 |
| Great black-backed gull | Jan-Dec | 12,460 | 1.68 | 20.93 |
| Black-legged kittiwake | Jan-Dec | 293,822 | 0.44 | 129.28 |
| Caspian tern | May-Oct | 108 | 0.61 | 0.03 |
| Common tern | May-Oct | 11,128 | 0.12 | 0.67 |
| Arctic tern | May-Oct | 16,358 | 0.11 | 0.90 |
| Common murre | Jan-Dec | 1,912,857 | 0.99 | 1893.73 |
| Thick-billed murre | Jan-Dec | 40,800 | 0.93 | 37.94 |
| Razorbill | Jan-Dec | 37,305 | 0.69 | 25.74 |
| Black guillemot | Jan-Dec | 54,000 | 0.40 | 21.60 |
| Atlantic puffin | Jan-Dec | 1,032,855 | 0.46 | 475.11 |
| TOTAL | Jan-Dec | 18,970,651 |  | 3329.35 |

B) Seabird species that occur within and breed mostly or completely outside of NAFO

Divisions 2J3KLNO.

| Species | Occupation <br> dates | Population <br> (number) | Individual mass <br> $(\mathrm{kg})$ | Average biomass <br> over year(t) |
| :--- | :---: | ---: | ---: | ---: |
| Wilson's storm-petrel | May-Oct | 50,000 | 0.04 | 1.00 |
| Northern fulmar | Jan-Dec | 300,000 | 0.80 | 240.00 |
| Greater shearwater | May-Oct | $1,500,000$ | 0.89 | 667.50 |
| Sooty shearwater | May-Oct | 300,000 | 0.79 | 11.50 |
| Oldsquaw | Nov-Apr | 15,000 | 0.76 | 5.70 |
| Scoter spp. | Oct-Apr | 40,000 | 1.11 | 25.90 |
| Common eider | Nov-Mar | 50,000 | 2.23 | 46.46 |
| Iceland gull | Oct-Apr | 100,000 | 0.86 | 50.17 |
| Glaucous gull | Oct-Apr | 50,000 | 1.70 | 49.58 |
| Black-legged kittiwake | Oct-Apr | 500,000 | 0.44 | 128.33 |
| Thick-billed murre | Oct-Apr | $1,500,000$ | 0.93 | 813.75 |
| Dovekie | Nov-Mar | $10,000,000$ | 0.15 | 625.00 |
| TOTAL | Jan-Dec | $14,475,000$ |  | 2771.89 |
|  |  |  |  |  |
| GRAND TOTAL |  |  |  |  |

## Production:Biomass

Assuming a production to biomass ratio of 0.25 generates a value of annual production within the study area of 1525 t .

## Consumption:Biomass

Assuming that birds consume $15 \%$ of their body mass per day, the annual consumption for the area is $334,043 \mathrm{t}$, and $\mathrm{Q} / \mathrm{B}$ is $54.75 \mathrm{yr}^{-1}$.

## Diet

Seabirds within the study area feed at a diversity of trophic levels with most prey being small pelagic fishes, cephalopods and crustaceans (Rice 1992; Montevecchi 1993). However, gannets, the largest breeding seabird species in the region, prey on large pelagic species such as mackerel (Scomber scombrus), Atlantic herring (Clupea harengus) and short-finned squid (Illex illecebrosus) that are the same sizes as those captured by commercial fisheries (Montevecchi et al. 1988b). Storm-petrels, the smallest and most abundant breeding species in region, feed primarily on myctophids and crustaceans (amphipods, euphausids) in deep shelf and inshore waters (Montevecchi et al. 1992). Dovekies, the smallest seabird species that occurs in the region but breeds outside of it, is primarily a planktivore (Roby et al. 1981; Montevecchi et al. in prep.). Common murres, gannets, puffins, razorbills, gulls, shearwaters and fulmars feed on capelin that is in the same range of sizes as those exploited by the commercial fishery. Common murres feed their chicks on an almost exclusive diet of female capelin, most of which are gravid (Montevecchi and Myers unpubl. data). Common eiders and scoters prey extensively on blue mussel (Mytilus edulis) (Goudie and Ankney 1989). Dietary proportions for all seabird species combined (Table 2b) are extrapolated and estimated from Goudie and Ankney (1989), Cairns et al. (1990), Montevecchi et al. (1992) and Montevecchi and Myers (1997).

## 5 \& 6. Cod (Lilly, Bundy, Shelton and Dalley)

## Background

The Atlantic cod (Gadus morhua) has provided Atlantic Canada's most important fishery resource for the last 400 years (Templeman 1966; Lear and Parsons 1993). Within the study area two stocks are recognised for assessment purposes, one in the north in 2J3KL (the "northern" cod) and one on southern Grand Bank in 3NO. Unexploited biomass may have been as high as four million tons. During the late 1950s and 1960s, the catches by distant water fleets increased dramatically and both stocks declined to low levels by the mid to late 1970s. There was some recovery in the early 1980s and the 1985-1987 period was one of relative stability in stock size and catches (Stansbury et al. 1995, Shelton et al. 1996b). Both stocks declined to very low levels in the late 1980s and early 1990s. The Atlantic cod has long been regarded as a major component of the Newfoundland-Labrador ecosystem.

For this modelling exercise, the Atlantic cod are split into adults and juveniles, or large and small fish, with the division at 35 cm . This is approximately the size at which cod become more piscivorous. It is also the approximate size of first capture in the commercial fishery, so
that the $>35 \mathrm{~cm}$ group represents the exploitable biomass. However, the split at 35 cm is too low for an effective division into immatures and matures, since the size of $50 \%$ maturity in 1985-1987 was around 41-42 cm for males and $50-51 \mathrm{~cm}$ for females (Shelton et al. 1996b). Only about $62 \%$ of the fish in the $>35 \mathrm{~cm}$ group are mature. For the purposes of this exercise, it is assumed that fish $\leq 35 \mathrm{~cm}$ in length are of ages 2 and less, and fish $>35 \mathrm{~cm}$ are $3+$.

## Catch

## Cod $>35 \mathrm{~cm}$

The average total catch of cod in Divisions 2J3KL and Divisions 3NO for the period 19851987 was calculated from data in Shelton et al. (1996b) and Stansbury et al. (1995) to be $289,031 \mathrm{t}$. To this is added the discarded catch of $9,563 \mathrm{t}$ of cod from the northern cod and shrimp offshore fisheries (Kulka 1997). The total mean annual catch of cod in 1985-1987 is $298,594 \mathrm{t}$, or $0.603 \mathrm{t} \cdot \mathrm{km}^{-2}$.

Cod $\leq 35 \mathrm{~cm}$
There was an estimated mean annual discarded catch of 1.25 t in the Northern cod and shrimp offshore fisheries (Kulka 1997).

## Biomass

Cod $>35 \mathrm{~cm}$
The biomass of $3+$ cod is based on ADAPT estimates of cod biomass at the beginning of the year (1985-1988) reported in Shelton et al. (1996b) and Stansbury et al. (1995). The average biomass of age $3+$ cod in the study area during the period 1985-1987 was calculated as $1,011,907 \mathrm{t}$, or $2.044 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Cod $\leq 35 \mathrm{~cm}$

The biomass of the 0,1 and 2 year olds was estimated from the mean number of 3 year olds on January 1. Averaged over 1986-1988, this was $224,394 \times 10^{3}$ individuals. Assuming an annual instantaneous rate of natural mortality of 0.6 , since young fish are likely to have a higher natural mortality rate than older fish, and assuming catches are inconsequential, this gives $302,901 \times 10^{3} 2$ year olds and $551,921 \times 10^{3} 1$ year olds in mid-year and $865,584 \times 10^{3}$ 0 year olds at the start of the third quarter. Assuming a mean body weight of $0.003,0.025$ and 0.100 kg for cod of ages $0-2$ respectively, this gives an average biomass of $46,685 \mathrm{t}$ of age $0-2$ cod in the study area, or $0.094 \mathrm{t} \cdot \mathrm{km}^{-2}$. This estimate was relatively insensitive to the value used for the annual instantaneous rate of natural mortality, when $\mathrm{M}<1$.

## Production:Biomass

## Cod $>35 \mathrm{~cm}$

$\mathrm{P} / \mathrm{B}$ was estimated from a catch curve analysis of RV groundfish survey data from 2J3KL for the years 1983-1988. The estimates of total mortality (slopes of the regression lines fitted to the downward slope of the catch curve) for different year combinations are shown in Table 6. In each case, the line is fitted to the combined data points for each year combination. The best fit occurs when the 1986 year is omitted, due to the anomalously high catch rates in that year (Table 6, Fig. 3). The value of $0.651 \mathrm{yr}^{-1}$ is taken to represent the $\mathrm{P} / \mathrm{B}$ ratio of cod in 19851987. It is equivalent to an annual production of $658,752 \mathrm{t}$.

Table 6. Results of catch curve analysis of northern cod.

|  | $1983-1988$ | $1985-1987$ | $1984 / 1985 / 1987$ |
| :---: | :---: | :---: | :---: |
| $\mathrm{Z}=$ | 0.648 | 0.646 | 0.651 |
| $\mathrm{R}^{2}$ | 0.894 | 0.881 | 0.925 |

## Cod $\leq 35 \mathrm{~cm}$

The $\mathrm{P} / \mathrm{B}$ ratio for juvenile cod is assumed to be equivalent to the total mortality assumed above in the backcalculation of numbers at age from age 3 . The mean $\mathrm{P} / \mathrm{B}$ is therefore $0.6 \mathrm{yr}^{-1}$, and the annual production is $28,011 \mathrm{t}$.

## Consumption:Biomass

## Cod $>35 \mathrm{~cm}$

Consumption estimates of Atlantic cod in the literature are highly variable. Schulz (1990) compared annual consumption rates estimated for cod from different ICES areas by age. These varied by as much as a factor of 4 . There are few studies in Canadian waters which can be used here. Lilly et al. (1981) estimated consumption by cod, on the basis of an estimated production (mean number at age times somatic growth at age times reproductive growth at age) and a factor equivalent to gross efficiency, to estimate the consumption required to support this production. Mean Q/B estimates for $1980-81$ in 2 J 3 KL were 1.58 to $3.16 \mathrm{yr}^{-1}$ and in 3 NO were 2.2 to $4.4 \mathrm{yr}^{-1}$, depending on whether the gross growth efficiency was taken as 0.2 or 0.1 .

Waiwood et al. (1980) estimated consumption by cod at age in the southern Gulf of St. Lawrence in 1979 from stomach contents and application of the Ursin fish growth theory. The mean $\mathrm{Q} / \mathrm{B}$ estimate for $3+$ age groups was $2.14 \mathrm{yr}^{-1}$.

Published values of $\mathrm{Q} / \mathrm{B}$ in Pauly (1989) range from $1.41 \mathrm{yr}^{-1}$ (north of Norway) to $2.19 \mathrm{yr}^{-1}$ (Georges Bank). There is no clear rationale for using any of the above estimates and all are approximate. Lilly et al. (1981) estimates using a GE of 0.1 may be too high, and in any case would be equivalent to applying a GE of 0.1 or 0.2 to the production estimated above. This would give a $\mathrm{Q} / \mathrm{B}$ of $6.51 \mathrm{yr}^{-1}$ when $\mathrm{GE}=0.1,3.255 \mathrm{yr}^{-1}$ when $\mathrm{GE}=0.2$ or $4.34 \mathrm{yr}^{-1}$ when GE $=0.15$. Again, these estimates are high compared to literature values and the Waiwood et al. (1980) estimate.

A range of $\mathrm{Q} / \mathrm{B}$ values was used. It is assumed that GE is not less than 0.15 , and that $\mathrm{Q} / \mathrm{B}$ is in the range of 2.14 to $4.34 \mathrm{yr}^{-1}$. This is equivalent to an annual consumption between 2,165,481 t and $4,391,676 \mathrm{t}$. A mid-point $\mathrm{Q} / \mathrm{B}$ of $3.24 \mathrm{yr}^{-1}$, equivalent to a consumption of $3,278,579 \mathrm{t}$, was used.

## Cod $\leq 35 \mathrm{~cm}$

Using the data in Waiwood et al. (1980), the $\mathrm{Q} / \mathrm{B}$ for this size group is $8.18 \mathrm{yr}^{-1}$. Assuming a gross growth efficiency of $0.15, \mathrm{Q} / \mathrm{B}$ is $4 \mathrm{yr}^{-1}$. Consumption per unit biomass is likely to be greater in smaller fish. Therefore this range is possible. Total annual consumption would be 186,739 to $381,882 \mathrm{t}$. A mid-point $\mathrm{Q} / \mathrm{B}$ of $6.09 \mathrm{yr}^{-1}$, equivalent to a consumption of $284,311 \mathrm{t}$,
was used.

## Diet

## Cod $>35 \mathrm{~cm}$

Stomach content data are available from stomach collections during Canadian resource assessment bottom-trawl surveys in Divisions 3L, 3N and 3 O during spring (1985-1987), Divisions 2J, 3K and 3L in the autumn (1985-1987), and special surveys in Division 3L in the winter (1985-1986) and summer (1985). All sampling was length-stratified. A portion of the cod in $2 \mathrm{~J}, 3 \mathrm{~K}$ and 3 L are landward of the surveys throughout the year. It is assumed here that this portion is small and may be ignored as a separate entity. In addition, some of the cod which overwinter offshore move into shallow coastal waters in the late spring and early summer and return offshore in the autumn. There was very little sampling inshore in 19851987, but weekly sampling of the commercial fishery was conducted at Bonavista in northern 3L in 1983 and 1984, and the data collected at this location from June to August are assumed to be representative of the summer inshore diet of cod all along the coast. It is further assumed (with little support) that $50 \%$ of the cod in $2 \mathrm{~J}, 3 \mathrm{~K}$ and 3 L participate in the inshoreoffshore migration. For simplicity, an annual diet was estimated for each of three areas (3L alone, 2 J and 3 K combined, and 3 N and 3 O combined) and then the diets from these three areas were combined to represent the diet of cod in 2 J 3 KLNO .

The method by which the various samples were combined is explained in the diet tables. Diets for Division 3L by season and in total are provided in Appendix 2, Table 1a. Diets for Divisions 2 J and 3 K combined, Divisions 3 N and 3 O combined, and then all Divisions combined are provided in Appendix 2, Table 1b. It must be recognized that seasonal coverage is weak, except in Division 3L in 1985, and sampling of coastal waters landward of the area covered by the bottom-trawl surveys is especially weak. In addition, the diets have not been appropriately adjusted for variability in population numbers at length and for variability in digestive rate associated with prey type and temperature.

Additional information on the diet of cod in the study area may be found in Lilly $(1984,1987$, 1991).

## Cod $\leq 35 \mathrm{~cm}$

The diet of small cod in the offshore was determined as described above for large cod. The diet of small cod in the inshore could not be obtained from sampling the commercial fishery, as had the diet of large cod, because small ( $\leq 35 \mathrm{~cm}$ ) cod are not landed by the fishery. However, observations of stomach contents of small cod are available from individuals caught in pelagic traps set for research purposes at two sites in Division 3K and two sites in Division 3L in 1992-1994. Data from all sites and years were combined and divided into two groups by cod length ( $<16 \mathrm{~cm}$ and $\geq 16 \mathrm{~cm}$ ). The frequencies of occurrence of the various prey categories were summed, and the frequency of occurrence of each category was expressed as a percentage of the total. It was then assumed that this number was also appropriate as a percentage by weight. The values for the two size-groups were averaged to yield the inshore diet.

The method by which the various samples were combined is explained in the diet tables.

Diets for Division 3L by season and in total are provided in Appendix 2, Table 2a. Diets for Divisions 2 J and 3 K combined, Divisions 3 N and 3 O combined, and then all Divisions combined are provided in Appendix 2, Table 2b. It must be recognized that the data available for small cod may be less appropriate than that available for large cod. A large but unknown portion of small cod in Divisions $2 \mathrm{~J}, 3 \mathrm{~K}$ and 3 L reside in the inshore and gradually move offshore at ages 1-3. The diet of small cod in inshore waters has received little attention. In addition, the bottom-trawl employed in 1985-1987 was very poor at sampling those small cod that were offshore, and the data may underrepresent the contribution by the smaller individuals within the $<35 \mathrm{~cm}$ range.

## 7 \& 8. Greenland halibut (Morgan, Brodie, Bowering and Lilly)

## Background

Greenland halibut (Reinhardtius hippoglossoides) is a deep water flatfish, usually found at bottom temperatures from -0.5 to $3^{\circ} \mathrm{C}$, in the North Atlantic and the North Pacific. In the northwest Atlantic, Greenland halibut are distributed from Arctic regions to Georges Bank, but are most abundant in deeper waters from northern Labrador to the Grand Bank (Bowering 1983).

The main biomass of Greenland halibut within the study area is in NAFO Divisions 2J3KL. An inshore gillnet fishery began in the 1960s, and as catches declined in various bays, the fishery moved offshore to the deepwater channels between the banks. Catches hovered around $30,000 \mathrm{t}$ during the 1970s and decreased gradually during the 1980s. Since the early 1990s, catches have risen steeply, reflecting a new intense fishery in the deepwater areas in 3LM (Brodie et al. 1997). However, catches have since declined to around $15,000-20,000 \mathrm{t}$. There was little or no catch in Divisions 3NO during the mid-1980s.

Greenland halibut were split into adults and juveniles, or large and small fish. Greenland halibut $>40 \mathrm{~cm}$ in length are equivalent to $6+$ fish (Brodie 1991). Fish $\leq 40 \mathrm{~cm}$ in length, essentially juveniles, are regarded as the 0 to 5 year olds. This is the size which approximately corresponds with the size of first capture (Brodie 1991). The main ontogenic change in diet occurs when Greenland halibut are about 20 cm (Bowering and Lilly 1992), and they mature at lengths $63-98 \mathrm{~cm}$ for females and $51-96 \mathrm{~cm}$ for males (Morgan and Bowering 1995). Thus the 3 criteria for demarcating small or juvenile fish from large or adult fish occur at different sizes and ages. The 40 cm size category was chosen because it splits the biomass into exploitable and non-exploitable biomass (C. Walters, pers. comm.) and it is a useful size to split Greenland halibut in terms of their availability as prey.

## Catch

Estimates of catch were obtained from a compilation of landings (Brodie et al. 1997) and from by-catch data from the shrimp fishery, (D. Kulka, pers. comm, unpublished data). The catch was divided into the two size groups using biomass-at-age data in Bowering and Brodie (1988).

## Greenland halibut $>40 \mathrm{~cm}$

Most (98 \%) of the Greenland halibut catch is greater than 40 cm for the period 1985-87. This gives a total of $17,481 \mathrm{t}$, or $.035 \mathrm{t} \cdot \mathrm{km}^{-2}, 50 \%$ of which was taken in $2 \mathrm{~J}, 30 \%$ in 3 K and $20 \%$ in 3L. Catches from 3 NO were very small.

Greenland halibut $\leq 40 \mathrm{~cm}$
Only $2 \%$ of the catch was 40 cm or less, giving an annual catch of 373 t . However, it is assumed that the average by-catch of 408.5 t of Greenland halibut in the shrimp fishery during 1985 to 1987 was composed of Greenland halibut $\leq 40 \mathrm{~cm}$. The total catch then is 782 t , $0.002 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Biomass

Campelen converted numbers and biomass-at-age survey estimates for 2 J 3 K were used as the basis for the biomass estimate (DFO unpublished data). The ratio of Campelen number-at-age to Engels number-at-age for 2J3K was used to raise the Engels 3L estimates to Campelen values. The biomass in 3NO was assumed to be negligible in 1985-1987. Numbers were converted to biomass using the age-length-weight data in Brodie (1991) and the length-weight relationship in Bowering and Stansbury (1984).

## Greenland halibut $>40 \mathrm{~cm}$

The biomass of Greenland halibut $>40 \mathrm{~cm}$ was estimated as $172,485 \mathrm{t}$ or $0.348 \mathrm{t} \cdot \mathrm{km}^{-2}$ for 2J3KLNO.

## Greenland halibut $\leq 40 \mathrm{~cm}$

The Campelen estimate of biomass of young Greenland halibut was $65,006 \mathrm{t}$. Biomass was also estimated by back-calculating from the number of 6 year olds, assuming a natural mortality of 0.2 on the 3-6 year olds and a natural mortality of 0.6 on the $0-2$ year olds. This gave a biomass estimate of $81,813 \mathrm{t}$. These estimates are similar, although the Campelen estimate is an under-estimate. The higher estimate of $81,813 \mathrm{t}, 0.165 \mathrm{t} \cdot \mathrm{km}^{-2}$ is used here.

## Production:Biomass

## Greenland halibut $>40 \mathrm{~cm}$

Catch curve analysis of numbers-at-age from the RV survey data gave a total mortality and therefore $\mathrm{P} / \mathrm{B}$ estimate of $0.719 \mathrm{yr}^{-1}$. The fit of the estimated data to the observed was quite good, with $\mathrm{R}^{2}=0.897$. Production of Greenland halibut was also estimated using the assumption that production is equivalent to the product of natural mortality and biomass plus the catch. This method, assuming a natural mortality of $0.2 \mathrm{yr}^{-1}$, gave an average annual production of $55,097 \mathrm{t}$, and a $\mathrm{P} / \mathrm{B}$ estimate of $0.296 \mathrm{yr}^{-1}$. The difference between the two estimates is quite large. The first is quite high and overestimates the mortality of the 13-17 year olds. This is because older ages are underestimated in the trawl surveys (their selection curve is domed), and Z estimates from catch curves are inflated. It is suggested that a range of values be used, from 0.296 to $0.719 \mathrm{yr}^{-1}$. This is equivalent to an annual production in the range of 55,097 to $124,053 \mathrm{t}$. A midpoint of $0.508 \mathrm{yr}^{-1}$ was used.

## Greenland halibut $\leq 40 \mathrm{~cm}$

It is assumed that the $\mathrm{P} / \mathrm{B}$ of the young Greenland halibut is equivalent to the weighted natural mortality y with age, used in the back-calculation to estimate biomass. This gives a P/B estimate of $0.247 \mathrm{yr}^{-1}$ and a total annual production of $18,678 \mathrm{t}$.

## Consumption:Biomass

Food consumption by Greenland halibut was examined by Chumakov and Podrazhanskaya (1986) on the shelf and slope areas of the Northwest Atlantic from Davis Strait to eastern Newfoundland. From 76,700 stomachs taken during surveys in 1969-81, the daily food requirements were found to range from $1.2 \%$ of body weight at age 5 to $0.5 \%$ at ages $14-17$ years. These data were used to estimate consumption rates for both size groups of Greenland halibut.

## Greenland halibut $>40 \mathrm{~cm}$

A weighted mean, using relative proportions of biomass-at-age for 6 to 20 year olds, was calculated from the daily food requirements. This gave an annual $\mathrm{Q} / \mathrm{B}$ estimate of $1.478 \mathrm{yr}^{-1}$, assuming that Greenland halibut feed for half the year (Chumakov and Podrazhanskaya 1986), or a total annual consumption of $254,722 \mathrm{t}$.

## Greenland halibut $\leq 40 \mathrm{~cm}$

In order to estimate consumption of the younger, smaller fish, it was assumed that the log linear relationship seen in consumption with age for 5 to 17 year old Greenland halibut also applied to 0 to 5 year old fish. The natural log of consumption was regressed against natural log of age, and the regression estimate used to estimate consumption for the 0 to 5 year olds, $\left(R^{2}=0.99\right)$. $\mathrm{Q} / \mathrm{B}$ estimates ranged from 5.1 for 4 year olds to 15.4 for 1 year olds. A weighted mean, using the biomass-at-age estimated by back-calculation above, gave a mean annual $\mathrm{Q} / \mathrm{B}$ of $3.401 \mathrm{yr}^{-1}$ and an annual consumption of $278,054 \mathrm{t}$.

## Diet

Stomach content data are available from stomach collections during Canadian resource assessment bottom-trawl surveys in Divisions 2J and 3K in the autumn (1985-1987) and seasonal surveys in Division 3L during 1985. The sampling in 2 J and 3 K was conducted to 1000 m . However, the sampling in 3L went only to 366 m , so the information from 3L is representative of the relatively shallow water and is probably less representative of the larger fish, which tend to be in the deeper water. All sampling was length-stratified. It is assumed here that relatively few fish were in 3 NO , and their diet is ignored.

Additional information on the diet of Greenland halibut in the study area may be found in Bowering and Lilly (1992) and Rodríguez-Marín et al. (1995).

## Greenland halibut $>40 \mathrm{~cm}$

The method by which the various samples were combined is explained in Appendix 2, Table 3. The diets have not been appropriately adjusted for variability in population numbers at length and for variability in digestive rate associated with prey type and temperature.

## Greenland halibut $\leq 40 \mathrm{~cm}$

The method by which the various samples were combined is explained in Appendix 2, Table 4.

## 9, 10. American plaice (Morgan, Brodie, Bundy, Lilly and Walsh)

## Background

American plaice (Hippoglossoides platessoides) has a widespread distribution, occurring on both sides of the Atlantic. It is a bottom dwelling flatfish, and in the western Atlantic occurs from the Arctic to USA waters. The study area contains a portion of the $2+3 \mathrm{~K}$ stock and the 3LNO stock. The main fishery on Grand Bank in 3LNO has been exploited since the late 1940s. It was basically a Canadian fishery until the 1960s when other countries also started to fish for American plaice (Pitt 1971). Catches peaked in 1967, then fell until the mid to late 1970s when there was a plateau at around $50,000 \mathrm{t}$. The mid-1980s saw a small peak around $60,000 \mathrm{t}$, but since then catches have declined. Biomass estimates show a similar trend. In 1994 there was no directed fishey and since 1995 there has been a full moratorium on the 3LNO American plaice fishery (Morgan et al. 1997).

The American plaice were split into fish $\leq 35 \mathrm{~cm}$ and fish $>35 \mathrm{~cm}$ in length. This is the length at which there are significant changes in the diet (see below). In addition, the length of 35 cm is approximately the length at maturity of female plaice and the length at first capture. The length of 35 cm corresponds to an age of approximately 7 years, although this varies considerably by area and sex.

## Catch

The average catch for the period 1985-1987 in 3LNO was 57,931 t (Morgan et al. 1997) and in 2J3K was 1611 t (Brodie et al. 1993). Discards from other Newfoundland fisheries increased these totals by 109 t in 2J3K and 1972 t in 3LNO during 1985-1987 (Kulka 1986). In addition, there was a bycatch of 67.9 t from the shrimp fishery (D. Kulka, pers. comm.). This makes a total annual catch of $61,691 \mathrm{t}$.

This total was divided between the two size groups of American plaice using commercial biomass-at-length data in Brodie $(1986,1987)$ for the Canadian fleet and numbers, weight, and length in the commercial catch for the Spanish fleet in 1987 (Brodie 1988). These data are for 3 LNO and it is assumed that they also apply to the smaller fishery in 2 J 3 K .

## American plaice $>35 \mathrm{~cm}$

About $94 \%$ the Canadian catch and $64 \%$ of the Spanish catch of American plaice was greater than 35 cm . It is assumed that all discards are less then 35 cm , and the other fleets' catch was the same as the Spanish catch. This gives a total catch of American plaice $>35 \mathrm{~cm}$ of 49,454 $\mathrm{t}, 0.1 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## American plaice $\leq 35 \mathrm{~cm}$

About $6 \%$ the Canadian catch and $36 \%$ of the Spanish catch of American plaice was 35 cm
or less. Assuming that all discards were 35 cm or less, the total catch was $12,237 \mathrm{t}, 0.025$ $\mathrm{t} \cdot \mathrm{km}^{2}$.

## Biomass

Biomass was estimated from Campelen converted estimates for 1985-1987. At the time of writing, these estimates were available as numbers-at-length for the 3LNO stock only. These were converted to biomass-at-length using a weight-length relationship (Brodie 1985). The biomass in 2 J 3 K was estimated by applying the Engels to Campelen conversion from 3LNO to the numbers at age in 2 J 3 K and then applying the weight-length relationship. The results are shown in Table 7.

Table 7. Mean 1985-1987 American plaice biomass (t) (Campelen converted) in 2J3K and 3LNO.

| Area/ <br> Length group | 2J3K | 3LNO | 2J3KLNO | 2J3KLNO <br> $t \cdot \mathrm{~km}^{-2}$ |
| :---: | :---: | :---: | :---: | :---: |
| $>35 \mathrm{~cm}$ | 57936 | 423269 | 481205 | 0.972 |
| $\leq 35 \mathrm{~cm}$ | 76901 | 311062 | 387963 | 0.784 |
| Total | 134837 | 734331 | 869168 | 1.756 |

## Production:Biomass

## American plaice $>35 \mathrm{~cm}$

It is assumed that $\mathrm{P} / \mathrm{B}=\mathrm{Z}$. Z was estimated from a catch curve analysis of Campelen converted American plaice survey data from 3LNO for the years 1985-1987. The catch curve was fitted for ages 7 to 17 (Fig. 4). The slope of the regression lines fitted to the downward slope of the catch curve is -0.538 , and $\mathrm{R}^{2}=0.988$. The total mortality, and therefore $\mathrm{P} / \mathrm{B}$ estimated from the catch curve, is thus $0.538 \mathrm{yr}^{-1}$.

## American plaice $\leq 35 \mathrm{~cm}$

The total mortality of the small length groups is not known. In the absence of other information, it is assumed that their production is equivalent to biomass multiplied by natural mortality plus the catch. This gives a production of $300,960 \mathrm{t}$ and a $\mathrm{P} / \mathrm{B}$ of $0.625 \mathrm{yr}^{-1}$.

## Consumption:Biomass

## American plaice $>35 \mathrm{~cm}$

Consumption was estimated from daily ration data for American plaice on the tail of the Grand Bank (Zamarro 1992). The diet of American plaice here is comprised mainly of the sand lance (Ammodytus dubius), brittle stars and capelin (Mallotus villosus). Zamarro estimated daily ration using the model of Elliot and Persson (1978). American place were collected from July 1987 to December 1989, thus providing an annual cycle of estimates, at around the time period of interest. Monthly daily rations were taken from Figure 13 in Zamarro (1992). From these data, the mean annual Q/B ratio was estimated as $1.262 \mathrm{yr}^{-1}$

## American plaice $\leq 35 \mathrm{~cm}$

Consumption of juvenile American plaice was estimated from data presented in MacDonald and Waiwood (1987) for American plaice size $14-31 \mathrm{~cm}$ in length. Sampling was conducted in May in Passamaquoddy Bay, New Brunswick, where bottom temperatures were 5.5-6.1 C. The major prey of young American plaice in this area are amphipods and polychaetes. The authors estimated a daily ration of $1.28 \%$ of body weight per day. This translates to a $\mathrm{Q} / \mathrm{B}$ ratio of $4.67 \mathrm{yr}^{-1}$ if feeding is constant throughout the year. Adult American plaice showed a peak in feeding activity during May on the tail of the Grand Bank (Zamarro 1992). The feeding cycle of juvenile American plaice is not known, but if it is assumed to be the same as the adults and assuming a peak daily ration of $1.28 \%, \mathrm{Q} / \mathrm{B}$ is $2.8 \mathrm{yr}^{-1}$. It is suggested that both estimates should be used to described a likely range of $\mathrm{Q} / \mathrm{B}$ values of 2.8 to $4.672 \mathrm{yr}^{-1}$ for the young American plaice. A midpoint of $3.736 \mathrm{yr}^{-1}$ was used.

## Diet

Stomach content data are available from stomach collections made during Canadian resource assessment bottom-trawl surveys in Divisions 2J and 3K in autumn 1977-1978, seasonal surveys in Division 3L during 1985, a spring survey in Divisions 3NO in spring 1984 and juvenile flatfish surveys in 3NO in summer 1985-1986 (Appendix 2, Tables 5a, b, 6a,b).

The sample sizes in 2 J and 3 K were very small. In addition, the autumn 1977 sample in Division 2J had a high proportion of Greenland halibut $<=40 \mathrm{~cm}$. This was probably anomalous, and an arbitrary adjustment has been made to the final 2J3KLNO diet. In addition, sampling of cod stomachs has shown that the quantity of capelin in Divisions 2J and 3K in autumn 1978 was very low (Lilly 1991). For these and other reasons, the data available for 2 J 3 K probably do not well represent the diet in that area in 1985-1987, but it is thought that it is more appropriate to include the data that are available than to ignore the area altogether.

The data for 3 NO in the summers of 1985 and 1986 are from surveys directed toward juvenile flatfish, and do not cover the whole of 3NO.

Additional information on the diet of American plaice in the study area may be found in Pitt (1973) and Zamarro (1992).

## American plaice $>35 \mathrm{~cm}$

The method by which the various samples were combined is explained in the diet tables. Diets for Divisions 2J and 3K and Division 3L (by season and in total) are provided in Appendix 2, Table 5a. Diets for Divisions 3N and 3 O combined, and all then Divisions combined are provided in Appendix 2, Table 5b.

## American plaice $\leq 35 \mathrm{~cm}$

The method by which the various samples were combined is explained in the diet tables. Diets for Divisions 2J and 3K and Division 3L (by season and in total) are provided in Appendix 2, Table 6a. Diets for Divisions 3N and 3O combined, and then all Divisions combined are provided in Appendix 2, Table 6b.

## 11. Flounders (Bowering, Bundy, Walsh and Lilly)

## Background

The Flounders consist of yellowtail flounder (Limanda ferruginea), witch flounder (Glyptocephalus cynoglossus) and winter flounder (Pseudopleuronectes americanus). The flounders were grouped together on the basis of the similarity in their feeding behaviour. All are sedentary demersal flatfish that have small mouths and feed on benthic invertebrates.

Yellowtail flounder dominate the group in terms of biomass. Their distribution ranges from Labrador in the north to Cape Hatteras, North Carolina, in the south. Only one stock, Division 3LNO, lies within the study area. There has been a fishery for yellowtail flounder since the 1960s. Catches peaked in the early 1970s and than again in the mid 1980s (Walsh et al. 1997).

Witch flounder has a more northerly distribution. Two stocks occur within the study area, in Divisions 2J3KL and Division 3NO. Witch flounder is a deep water species. During the model time period 1985-1987 the R.V. survey was mainly in shallow waters, which means that stock sizes will probably be underestimated. There has been a fishery for witch flounder since the 1960s in 2 J 3 KL and since the early 1970s in 3NO. Both had high catches in the mid 1970s, but the 2J3KL stock has since declined, and there has been no directed Canadian fishery since 1994 (Bowering 1997). There was a high peak in catches in the 3NO stock in the mid 1980s, but this stock also declined severely (Bowering and Orr 1997).

Winter flounder is an inshore shallow-water species that undergoes fairly regular inshoreoffshore migrations, moving offshore in the winter and inshore in the summer. Its distribution ranges from the southern Labrador shelf to the Gulf of Maine (Scott and Scott 1988). It is of little commercial importance.

## Catch

The mean annual catch for the period 1985-1987 was calculated for yellowtail flounder from data in Walsh et al. (1997), for witch flounder from data in Bowering (1997) and Bowering and $\operatorname{Orr}(1997)$ and for winter flounder from Anon $(1987,1989,1990)$ (Table 8). Also given are the discarded catches estimated for 1985 (Kulka 1986). The mean flounder catch for 1985-1987 was $39,273 \mathrm{t}$, or $0.079 \mathrm{t} \cdot \mathrm{km}^{-2}$.

Table 8. Mean annual catch (tonnes) for the 3 species of flounder, individually and combined.

|  | Yellowtail <br> Flounder | Witch Flounder | Winter Flounder | Flounders |
| :--- | :---: | :---: | :---: | :---: |
| 2J3KL |  | 3798 | 1280 | 5078 |
| 3LNO | 25151 |  |  |  |
| 3NO |  | 8499 | 10 | 33660 |
| Discards | 516 | 20 |  | 535 |
| Total | 25667 | 12317 | 1290 | 39273 |

## Biomass

Mean biomass of yellowtail and witch flounder was estimated from Campelen converted Engels RV data for 1985-1987. Yellowtail data came from Walsh et al. (1998) and witch data from Bowering (1998).

For yellowtail flounder, RV survey data only sampled 4 to 11 year olds. In order to obtain a biomass estimate for the whole population, numbers were backcalculated from numbers at age 7 to age 0 , and the biomass of age 0 to 6 added to the $7+$ Campelen biomass. A natural mortality of 0.3 (Pitt 1975) was assumed for age groups 3 to 6 and a natural mortality of 0.6 for age groups 0 to 2 . Biomass will be slightly underestimated since there is a commercial catch of 4 to 6 year olds. Backcalculated numbers-at-age were multiplied by mean weights-atage for the 0 to 6 year olds (DFO unpublished data). This backcalculation produced a biomass of yellowtail flounder 0 to 6 years old of $162,304 \mathrm{t}$. The $7+$ Campelen converted biomass was 105,490 t (Table 9).

The mean Campelen converted witch flounder biomass for 1985-1987 was 53,975 t. Again, this figure only includes the older age groups that are selected by the gear. In order to estimate the biomass of young fish, it was assumed that the ratio of young fish to selected fish was the same as for yellowtail flounder. Thus, the biomass estimate for age 0 to 6 fish is $83,045 \mathrm{t}$.

There are no RV survey estimates of winter flounder because it is an inshore species. An approximate biomass was estimated by assuming that the exploitation rate is low, say $10 \%$. This exploitation rate was applied to the highest catch during the time period ( $2,464 \mathrm{t}$ in 1987), giving a biomass estimate of 24,640 t for 1985-1987.

The 1985-1987 biomass estimate for the flounders is $429,454 \mathrm{t}, 0.868 \mathrm{~km}^{-2}$ (Table 9).
Table 9. Mean annual biomass estimates (tonnes) for the 3 species of flounder, individually and combined.

|  | Yellowtail Flounder | Witch Flounder | Winter Flounder | Flounders |
| :--- | :---: | :---: | ---: | ---: |
| Campelen Biomass, age 7+ | 105,490 | 53,975 |  | 159,465 |
| Backcalculated Biomass, age 0-6 | 162,304 | 83,045 |  | 245,349 |
| B from catch assumption |  |  | 24,640 | 24,640 |
| Total | 267,794 | 137,020 | 24,640 | 429,454 |

## Production:Biomass

In order to estimate $\mathrm{P} / \mathrm{B}$, catch curve analysis was first used to estimate total mortality for yellowtail and witch flounders. However, in both cases, the analysis was non-conclusive. Numbers caught at age increase until age 8 or 9 , and then rapidly decrease after age 10. The 2 or 3 age classes that would appear to be fully selected by the gear are not representative of the population and indicate very high mortalities. This method was deemed unsuitable. Instead, it was assumed that production was equivalent to biomass multiplied by natural mortality and the catch added to the product. Natural mortality was assumed to be 0.3 . This gives a total annual production of $106,005 \mathrm{t}$ for yellowtail flounder and $53,423 \mathrm{t}$ for witch flounder. Since
data for winter flounder are poor, its production was not estimated. $\mathrm{P} / \mathrm{B}$ ratios are $0.396 \mathrm{yr}^{-1}$ and $0.390 \mathrm{yr}^{-1}$ for the two species, giving an average of $0.394 \mathrm{yr}^{-1}$ for the flounder group, and a total flounder production of $169,132 \mathrm{t}$.

## Consumption:Biomass

Grosslein et al. (1980) estimated a Q/B of 4.6 for yellowtail flounder on George's Bank, for 1963-1972, weighted over all ages. Alternatively, assuming a gross growth efficiency of 0.15 (Collie 1987), Q/B would equal $2.639 \mathrm{yr}^{-1}$. This same assumption was made for the witch flounder, giving a $\mathrm{Q} / \mathrm{B}$ of $2.599 \mathrm{yr}^{-1}$. No estimate was made for the winter flounder. Since all estimates are only approximations, a range of $\mathrm{Q} / \mathrm{B}$ values, $2.599-4.6 \mathrm{yr}^{-1}$, encompassing these estimates was used. Thus annual consumption from 1985 to 1987 was 734,416 to $1,082,792 \mathrm{t}$. A midpoint of 3.60 was used as the initial input value for $\mathrm{Q} / \mathrm{B}$.

## Diet

The diet of winter flounder (Appendix 2, Table 7a) was compiled from Keats (1990) and Kennedy and Steele (1971). The sampling in both studies was very limited in space. The diet of yellowtail flounder (Appendix 2, Table 7b) was taken from Pitt (1976), who reported on sampling on Grand Bank. The diet of witch flounder (Appendix 2, Table 7b) was taken from Rae (1969), who reported on sampling from the eastern Atlantic.

The diet of the flounder group (Appendix 2, Table 7b) was an average of the diets of the three species, weighted by their biomasses.

## 12. Skates (Bundy, Kulka, Lilly and Methven)

## Background

Skates, of the family Rajidae, are widely distributed through the study area but are most abundant on Grand Bank. There are 8 to 10 species, but $90 \%$ of the RV survey biomass in 3LNO is comprised of the thorny skate (Raja radiata). The smooth skate ( $R$. senta) accounts for $5 \%$ of the biomass and other species, such as winter skate ( $R$. ocellata), spinytail skate ( $R$. spinicauda), and barndoor skate (R. laevis) make up the other 5\%. (Kulka et al. 1996). The thorny skate are fairly long-lived, at least 20 years (Templeman 1984), and are thought to be sedentary (Kulka et al. 1996). They deposit large leathery egg cases which contain a single embryo. A female may lay between 6 and 40 of these per year and egg size is related to parent size. The thorny skate is a bentho-pelagic species and is taken as the representative species for skate group.

## Catch

There was limited interest in skate fishing in the mid-1980s. (This changed in the 1990s when attention was switched to non-traditional species as a consequence of moratoria on traditional species.) The bulk ( $61 \%$ ) of the mean annual catch of $14,846 \mathrm{t}$ for 1985-1987 came from 3 N and $23 \%$ from 3L (NAFO Statistical Bulletins Anon. 1987, 1989, 1990). In addition, Kulka et al. (1996) estimated that another 3000-4000 t of skate were caught as by-catch in other fisheries during the 1980s. This would put the total annual skate catch at around $18,346 \mathrm{t}$, or
$0.037 \mathrm{t} \cdot \mathrm{km}^{-2}$, assuming $3,500 \mathrm{t}$ of by-catch. Kulka et al. (1996) also note that the accuracy of the skate catch statistics is questionable, and that up to $60 \%$ of reported skate catches may have been misreported catches of other species.

## Biomass

Annual biomass estimates of the thorny skate for the period 1985-1987 were obtained from Engels R/V data for Divisions 2J3KLNO for the 1985-1987 period (Kulka et al. 1996; DFO, unpublished data). The average biomass was calculated to be $116,170 \mathrm{t}$. This was raised to Campelen equivalent units as follows. The mean biomass in the last 3 years of the Engels RV survey (1992-1994) was compared to the mean biomass in the second two years of Campelen RV survey 1996-1997. The ratio between the 1992/1994 Engels biomass estimate and the 1996/1997 Campelen biomass estimate was used to raise the 1985-1987 Engels biomass estimate. The raising factor was 2.2 , making the mean annual biomass $255,979 \mathrm{t}$, or 0.517 $\mathrm{t} \cdot \mathrm{km}^{-2}$.

## Production:Biomass

In the absence of information on total mortality of skates in this region and information on catch at age with which to estimate total mortality, it was assumed that production is equivalent to biomass multiplied by natural mortality and the catch added to the product. The natural mortality estimate of 0.214 (Simon and Frank 1996) was used. This gives a total production of $73,126 \mathrm{t}$ and a $\mathrm{P} / \mathrm{B}$ of $0.286 \mathrm{yr}^{-1}$.

## Consumption:Biomass

Consumption rates are not known for the thorny skate in the area. However, studies in two other areas have estimated very similar consumption:biomass values for $R$ radiata. Vinther (1989) estimated a Q/B of 2.865 for $R$. radiata in the North Sea, and Dologov (1997) estimated a $\mathrm{Q} / \mathrm{B}$ of 2.891 for $R$. radiata in the Barents Sea. A mean value of 2.878 was used here, giving a total consumption of $236,669 \mathrm{t}$.

## Diet

The diet of thorny skate is assumed to be representative of the skate group. Thorny skate diet (Appendix 2, Table 8) was modified from data in Templeman (1982).

## 13. Redfish (Power, Bundy, Lilly and Methven)

## Background

Redfish are long-lived, slow growing, semi-pelagic fish that occur in depths from 100-700m. Unlike most other marine fish species, redfish are viviparous. Mating occurs in the fall and females release live young from April to July. They reach a commercial age at approximately 8 to 10 years at a length of 25 cm . There are three redfish stocks ( $2+3 \mathrm{~K}, 3 \mathrm{LN}$ and 30 ) within the study area. The stocks consist of a mixture of Sebastes mentella and S. fasciatus. There was minimal fishing activity in 3O by the Canadian fleet in 1985-1987, although foreign fleets (Russia, Cuba, Japan) did fish in the area.

## Catch

The average annual catch for 1985-1987 for 2J3KLNO was calculated as 86,758 tons (Power 1995, 1997, and Power et al. 1996 for 2J3K, 3LN and 3O respectively). To this is added 1199 t of discarded redfish from other fisheries (Kulka 1986), making a total of $87,957 \mathrm{t}$, equivalent to $0.178 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Biomass

There is no comprehensive Canadian research survey coverage of the whole area to a depth of 1000 m . Since redfish are deepwater species, deeper waters need to be covered in the survey. There are fall RV data for 2 J 3 K (Power 1995) and Campelen converted estimates of RV survey biomass for 3L in each quarter of 1985 and for the first and fourth quarter of 1986 (Power and Maddock Parsons 1998). There are also Russian bottom trawl survey estimates for the second quarter of 1985 to 1987 for 3 L and 3 N (Power and Vaskov 1992). The Canadian RV survey coverage of 30 is insufficient to estimate the biomass of redfish because the survey covered waters only to a depth of 367 m during the time period of interest. All the above sources were used to estimate redfish biomass in 2J3KLNO.

The estimation of biomass for 2 J 3 K and 3L was straightforward. The Campelen converted estimates were used to estimate biomass in 3L in 1985 and 1986. The ratio between the Campelen converted data and Engels data in 3L was used to raise the 2J3K biomass estimate to Campelen converted trawlable biomass estimates. The ratio between the Engels and Russian survey estimates (2.37:1) and between the Russian and Campelen estimates (2.87:1) were used to estimate the 3L 1987 biomass and the 3N biomass for 1985-1987.

Since no survey biomass estimates exist for 30, it was assumed that the exploitation rate in 30 is the same as in 3 LN ;

$$
3 \mathrm{O} \text { Biomass }=(3 \mathrm{LN} \text { Biomass } / 3 \mathrm{LN} \text { catch }) * 3 \mathrm{O} \text { catch }
$$

This may underestimate the biomass since it is likely that the exploitation rate of 30 redfish was less than 3LN in the mid 1980s.

The average annual biomass for the study area obtained using this method was $482,379 \mathrm{t}$, or $0.975 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Production:Biomass

The estimate of $\mathrm{P} / \mathrm{B}$ was first attempted using catch curve analysis of RV survey numbers at age. However, the catch curve analysis did not work well for two reasons: redfish are slow growing and are difficult to age accurately. In addition, RV surveys exhibit considerable variability in estimating relative year class strengths, and selection is dome shaped which means that the mortality of older age groups is overestimated.

Production was calculated as biomass multiplied by natural mortality and the catch added to the product. Natural mortality was assumed to be 0.125 , using the mid-range of values
estimated by Rikhter (1987). This gives an estimated production of $148,255 \mathrm{t}$ and a P/B ratio of $0.33 \mathrm{yr}^{-1}$. This seems high for a slow growing, long lived species and it is likely that the biomass estimate is too low.

## Consumption: Biomass

Dolgov and Drevetnyak (1990) estimated annual food consumption to biomass ratios for Barents Sea deep-water redfish (S. mentella) to vary from 6 for fingerlings down to 1.3 for fish at age 19. This consumption at age data (10+ age groups) was used to estimate consumption of redfish in 2J3KLNO. Relative numbers-at-age in 2J3K (1985-1987) and 3L (1984-1986) (DFO Unpublished data) were multiplied by mean weight-at-age in 2J3K and 3L (19801997, DFO Unpublished data) to obtain a relative biomass-at-age composition. It is assumed that this is representative of the age structure of the population. Relative biomass-at-age was then multiplied by the annual consumption in each year to obtain a relative total consumption per year. This was divided by the relative biomass, summed over all ages, to obtain $\mathrm{Q} / \mathrm{B}$ for ages $10+$ for 2 J 3 K and 3 L . Values ranged from 1.13 to $1.37 \mathrm{yr}^{-1}$ in 2 J 3 K and 1.32 to 1.37 in 3 L . The mean consumption was $1.3 \mathrm{yr}^{-1}$.

However, the average $\mathrm{Q} / \mathrm{B}$ over all ages would be greater than this. For fingerlings to 6 year olds it ranged from 6 to 2 . Thus to allow for the biomass of younger age groups, for which there is no adequate biomass-at-age data, the $\mathrm{Q} / \mathrm{B}$ was raised to an intermediate estimate of 2 $\mathrm{yr}^{-1}$. This gives a mean average annual consumption of $964,758 \mathrm{t}$.

## Diet

The diet of redfish (Appendix 2, Table 9) was based primarily on Lambert (1960) with minor modifications based on Yanulov (1962) and Konchina (1986).

## 14. Large Demersal Feeders (Bundy, Lilly and Methven)

## Background

The large demersal feeders include white hake (Urophycis tenuis), haddock (Melanogrammus aegelfinnus), rocklings, grenadiers (Macrouridae), wolffish (Anarhichas spp.), eelpouts (Zoarcidae), common lumpfish (Cyclopterus lumpus), monkfish (Lophius americanus), and Atlantic halibut (Hippoglossus hippoglossus). Wolffish were distributed throughout 2J3KLNO in the mid 1980s, but their distribution has since contracted (Kulka and Deblois 1996). White hake distribution was limited to the edge of the southern Grand Bank, basically in 3O, and monkfish distribution was limited to the southern Grand Bank (Kulka and Deblois 1996). None of the large demersal feeders has been well studied in 2J3KLNO and most information comes from other sources. Juvenile large demersal feeders are classified as small demersal feeders, see below.

## Catch

Based on the total average catch within the study area of the species listed above for the period 1985-1987 as reported in NAFO Statistical Bulletins (Anon. 1987, 1989, 1990), the average annual catch of large demersal feeders was calculated to be $21,509 \mathrm{t}$. In addition, there was an
average of 94.9 t caught as bycatch in the shrimp fishery during 1985 to 1987 (D. Kulka pers. comm.), and $4,285 \mathrm{t}$ discarded in other fisheries (Kulka 1996). The total catch was $25,889 \mathrm{t}$ or $0.052 \mathrm{t} \cdot \mathrm{km}^{-2}$. Haddock, roundnose grenadier (Coryphaenoides rupestris) and white hake make up $80 \%$ of this total.

## Biomass

Annual biomass estimates of the large demersal feeders for the period 1985-1987 were obtained from R/V data for Divisions 2J3KLNO (DFO, unpublished data). The average biomass was calculated to be $133,907 \mathrm{t}$. This was raised to Campelen equivalent units using the ratio between 1993/1994 Engels biomass estimates for 2J3KLNO and 1996/1997
Campelen biomass estimates. The raising factor was 3.1 , making the mean annual biomass $418,173 \mathrm{t}$, or $0.845 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Production:Biomass

Production was estimated by multiplying the biomass by natural mortality and adding catch. Assuming an instantaneous natural mortality rate of 0.2 , this gave a production of 109,524 , and a $\mathrm{P} / \mathrm{B}$ value of $0.262 \mathrm{yr}^{-1}$.

## Consumption:Biomass

Assuming a production to consumption ratio of 0.15 , consumption was calculated to be $730,161 \mathrm{t}$, and a $\mathrm{Q} / \mathrm{B}$ of $1.747 \mathrm{yr}^{-1}$.

## Diet

The diet of the large demersal feeders is difficult to determine, primarily because the diets of the species within this group are diverse and poorly studied but also because the relative biomasses of the species (or groups) has not yet been determined.

The diet of Atlantic (or striped) wolffish (Anarhichas lupus) has been studied in the offshore by Albikovskaya (1983) and Templeman (1985) and in the inshore by Keats et al. (1986). Diets of spotted wolffish (A. minor) and northern (or broadhead) wolffish (A. denticulatus) have been described by Albikovskaya (1983). An attempt to combine the data from these studies into an average diet for wolffish is presented in Appendix 2, Table 10a.
The diet of the several species of grenadiers (Appendix 2, Table 10b) is compiled from the information in Savvatimsky (1989a) for roughhead grenadier, Gushchin and Podrazhanskaya (1983) for roundnose grenadier and Savvatimsky (1989b) for common grenadier. There is additional information for the above three species in Houston and Haedrich (1986). These data are not used because the samples came from a small area (near Carson Canyon on the southeastern slope of Grand Bank).

The diet of white hake (Appendix 2, Table 10c) is compiled from Petrov (1973).
The diet of haddock (Appendix 2, Table 10c) is compiled from Kohler and Fitzgerald (1969) for the Scotian Shelf and Langton and Bowman (1980) for the southwestern Scotian Shelf and New England, with a modification based on Templeman (1965).

The diet of ocean pout (Appendix 2, Table 10d) was reported by Keats et al. (1987). This diet may be biased toward echinoderms because the fish were collected on sea urchin grounds. Therefore, the echinoderm portion of the diet has been reduced and the difference arbitrarily spread over other groups to reflect the broader prey spectrum seen in studies to the south (Buzulutskaya 1983; references in Scott and Scott 1988). The diet of other eelpouts is a guess based on notes in Scott and Scott (1988).

The diets of lumpfish and monkfish (Appendix 2, Table 10e) are guesses based on notes in Scott and Scott (1988).

The diet of Atlantic halibut (Appendix 2, Table 10e) is compiled from Kohler (1967).
A composite diet for large demersal feeders (Appendix 2, Table 10e) is derived from an average of the eight species or groups discussed above, with weightings based on a guess of relative biomasses.

## 15. Small Demersal Feeders (Bundy and Lilly)

## Background

The small demersal feeders include gunnels (Pholidae), blennies (Stichaeidae), pouts (Lycodes) and wolfeels (Lycenchelys spp.), sculpins (Cottidae), alligatorfish (Agonidae), lumpfish (Cyclopteridae, exluding common lumpfish), and seasnails (Liparidae). None of these species were fished commercially in 1985-1987 and little is known about them. This group also includes juvenile large demersal feeders.

## Catch

There was no targeted catch of small demersal feeders. However, an average of 44.1 t were caught as by-catch and discarded by the shrimp fishery during this period (Kulka pers. comm.).

## Biomass

Annual biomass estimates of the small demersal feeders for the period 1985-1987 were obtained from R/V data for Divisions 2J3KLNO (DFO, unpublished data). The average biomass was calculated to be $10,476 \mathrm{t}$. This was raised to Campelen equivalent units using the ratio between 1993/1994 Engels biomass estimates for 2J3KLNO and 1996/1997 Campelen biomass estimates. The raising factor was 10.7 , making the mean annual biomass $112,327 \mathrm{t}$, or $0.227 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Production:Biomass

In the absence of other data, it was assumed that $\mathrm{P} / \mathrm{B}$ is in the range of 0.2 to $0.4 \mathrm{yr}^{-1}$, giving an annual production of 22,465 to $44,931 \mathrm{t}$. A P/B of $0.3 \mathrm{yr}^{-1}$, equivalent to a production of $33,698 \mathrm{t}$, was used as an initial input value.

## Consumption:Biomass

There are no consumption estimates for the small demersal feeders, so consumption was estimated by assuming a gross growth efficiency of 0.15 , giving a $\mathrm{Q} / \mathrm{B}$ of 1.33 to $2.67 \mathrm{yr}^{-1}$ and an annual consumption of 149,770 to $299,540 \mathrm{t}$. A mean value of $2.0 \mathrm{yr}^{-1}$, equivalent to a consumption of $224,655 \mathrm{t}$, was used as the initial input value.

## Diet

There is very little information on the diet of this group. A guess is made (Appendix 2, Table 11) using general information available in Scott and Scott (1988).

## 16. Capelin (Winters, Carscadden and Lilly)

## Background

The capelin (Mallotus villosus) is perhaps best known for its spawning behaviour. As juveniles they live offshore, but in 2J3KL they migrate to beaches to spawn during late June and early July (Carscadden and Nakashima 1997). In 3NO, they spawn offshore. Capelin is a short-lived, pelagic species that is an important prey for many species in the ecosystem. In the 1980s, capelin was assessed as 3 stocks ( $2+3 \mathrm{~K}, 3 \mathrm{~L}$ and 3 NO ). Since 1993 capelin in $2+3 \mathrm{KL}$ have been treated as one stock. Recruited capelin are considered to be all fish 2 years old and older.

## Catch

The fishery for capelin has a variable history (Carscadden and Nakashima 1997). The fishery was once small and domestic, but in the 1970s it expanded to accommodate, at its peak, a $246,000 \mathrm{t}$ foreign offshore fishery. By the late 1970s, however, the stocks declined and fisheries were either closed or quotas severely reduced. Since the late 1970s there has been a domestic inshore fishery for roe-bearing capelin for the Japanese market. The mean catch for the period from 1985-1987 was $62,362 \mathrm{t}$, or $0.126 \mathrm{t} \cdot \mathrm{km}^{-2}$ in 2J3KLNO (NAFO Statistical Bulletins, Anon, 1987, 1989 and 1990).

## Biomass

The Canadian groundfish RV survey does not give reliable estimates of capelin biomass or abundance (Carscadden and Nakashima 1997). Capelin is a pelagic species and the RV survey is designed to capture demersal fish species. Instead, biomass estimates that came from acoustic surveys of 2J3K, 3L and 3NO were used (Miller 1992, 1993a, and 1993b). The mean average biomass for the population over the whole area in 1985-1987 was $4,100,000 \mathrm{t}$. However, this estimate does not include the 0 and 1 year olds. The biomass of these younger groups was estimated by back-calculating from the mean number of 2 years olds ( 196 billion). Shackell et al. (1994), building on the work of Shelton et al. (1993), analysed survival rates in 3 L capelin by sex and by age. They found that for age 2 capelin, survival rates were not sex specific. Thus a mean of the male and female survival rates for 2 years olds ( $39.7 \%$ ) is used here to represent survival of the 0 and 1 year olds (this assumes that all immature fish, age 2 and less, have the same survival rate). So, assuming that there is no catch of 0 and 1 year olds, there would have been 495 billion 1 year olds at mid-year and 992 billion 0 year olds at the
start of the fourth quarter. Assuming a mean mid-year body weight of 0.001 kg for the 0 year olds and 0.003 kg for the 1 year olds (Carscadden, pers. comm.), the average biomass of capelin younger than age 2 is $2,478,068 \mathrm{t}$. The total average biomass of all capelin is $6,578,068 \mathrm{t}$, or $13.289 \mathrm{t} \cdot \mathrm{km}^{-2}$, in 2 J 3 KLNO . This estimate is relatively insensitive to the survival rate when survival is high. A high range of survival rates, say between 35 and $90 \%$, produce biomass estimates from 5 to 7 million tons. However, if the survival rate decreases below $25 \%$, the biomass estimates increase exponentially.

## Production:Biomass

The production:biomass ratio was assumed to be equivalent to total mortality (Allen 1971). The survival rates in Shackell et al. (1994) for capelin in 3L for 1985-1987 were weighted by sex, age and maturity in the population to give a mean survival rate for the population (age 0 to 6 ) of $31.8 \%$. This is equivalent to a total mortality rate of 1.145 year $^{-1}$.and a mean annual production of 7,532,502 t .

## Consumption:Biomass

There are no consumption rate studies for capelin in Newfoundland waters. However, there are comparative studies from other areas, such as the St. Lawrence Estuary and the Barents Sea. Vesin et al. (1981), for example, found the capelin daily ration of zooplankton in the St. Lawrence Estuary and western Gulf of St. Lawrence was approximately $5 \%$ of the body weight per day in May-September and around $2.5 \%$ in October-November. This is equivalent to an annual $\mathrm{Q} / \mathrm{B}$ of around 13 year $^{-1}$. These estimates are comparable with the studies of Panasenko (1981) in the Barents Sea who used 3 different methods for analysis of diurnal stomach content and estimated daily consumption rates between 1.4 and $13.7 \%$. The average of $6 \%$ body weight per day is equivalent to a $\mathrm{Q} / \mathrm{B}$ of 11 year $^{-1}$. This estimate pertains to spring when the capelin are putting on fat, and may therefore be an overestimate of the annual consumption rate. Ajiad and Pushchaeva (1992) studied Barents Sea capelin in August 1989 and estimated the consumption rate to be between 1.3 and $2.2 \%$. Based on these studies, it is assumed that the daily consumption rate of capelin in 2 J 3 KLNO is $3 \%$ of body weight per day, and that they have an annual feeding period of 7 months (Carscadden pers. comm.), giving an annual $\mathrm{Q} / \mathrm{B}$ of 6.4 year $^{-1}$ and an annual consumption of $42,017,406 \mathrm{t}$. The consumption rate estimate is very sensitive to the assumed \% body weight consumed. For example, $2 \%$ body weight per day would be equivalent to a $Q / B$ of 4.3 year ${ }^{-1}$ and $4 \%$ a day a Q/B of 8.5 year $^{-1}$.

## Diet

The diet of capelin (Appendix 2, Table 12) was compiled from observations on Grand Bank in spring (Kovalyov and Kudrin 1973; Gerasimova 1994) and off Labrador in autumn (Chan and Carscadden 1976).

## 17. Sand lance (Winters, Bundy, and Lilly)

## Background

The sand lance species group is assumed to be all northern sand lance (Ammodytes dubius).

Its geographic range extends from West Greenland to Cape Hatteras in the USA. In Newfoundland and Labrador waters, some small populations occur inshore, but most sand lance in the area occur on the plateau of Grand Bank. They are small, semi-demersal fish and are abundant in areas where the bottom substrate consists of sand and fine gravel. They feed pelagically and are similar to capelin in their role as a forage species (Winters 1983). Main predators on the southern Grand Bank are American plaice and cod.

## Catch

Sand lance are not commercially fished and there is no recorded catch..

## Biomass

Sand lance were very poorly sampled by the RV survey using Engels gear. The mean biomass was estimated as 35 t in 1985-1987 and 31 t in 1992-1994 for 2J3KLNO (DFO, unpublished data). The RV survey using the Campelen gear sampled the sand lance better and the mean biomass in 1996-1997 was 104,091 t (DFO, unpublished data). This is still likely to be an underestimate, for sand lance are considered to be at least as abundant on Grand Bank as capelin, which have a mean biomass of around $300,000 \mathrm{t}$ in 3 NO (Miller 1993b). In order to obtain a more realistic biomass estimate, a 2 step method was used. First it was assumed that the biomass of sand lance has not changed between 1985-1987 and 1996-1997; that is, that the 1996-1997 biomass estimate is representative of the model time period. Second, this biomass estimate was raised by a catchability factor. Scott (1971) estimated a catchability of 0.01 for sand lance on the Scotian Shelf using a Yankee \# 36 otter trawl. It is assumed that the Campelen gear is ten times as efficient as the Yankee \#36 gear at catching sand lance. Thus, the catchability factor is 0.1 . Applying this to the Campelen biomass estimate produces a mean annual biomass estimate of $1,040,912 \mathrm{t}$, or $2.103 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Production:Biomass

Winters (1983) estimated the average instantaneous total mortality rate $(Z)$ of sand lance using catch curve analysis based on catches during research vessel surveys. In 1975-1979, when the biomass of cod was approximately $100,000 \mathrm{t}, \mathrm{Z}$ was $0.62 \mathrm{yr}^{-1} . \mathrm{Z}$ was estimated as $1.15 \mathrm{yr}^{-1}$ from 1968-1974 when cod biomass was approximately $500,000 \mathrm{t}$. In the mid 1980s, the time period of the model, cod biomass on Grand Bank (Divisions 3LNO) was around 500,000 t (Shelton et al. (1996b) and Stansbury et al. (1995). It is thought that cod predation on sand lance is a major contributor to sand lance mortality and variability in abundance (Winters 1983). It is assumed that since cod biomass was high in the mid 1980s, mortality on sand lance was also high and that the Z estimate of $1.15 \mathrm{yr}^{-1}$ applies. Thus, assuming $\mathrm{P} / \mathrm{B}=\mathrm{Z}$ (Allen 1971), $\mathrm{P} / \mathrm{B}=1.15 \mathrm{yr}^{-1}$, and annual production is $1,197,049 \mathrm{t}$.

## Consumption:Biomass

In the absence of information on consumption by sand lance, it was assumed that the gross growth efficiency, the ratio between production and consumption, was 0.15 . Thus, consumption was calculated as $7,980,325 \mathrm{t}$ and $\mathrm{Q} / \mathrm{B}$ was $7.667 \mathrm{yr}^{-1}$.

## Diet

The diet of sand lance (Appendix 2, Table 13) was taken from Scott (1973), who examined fish caught on the Scotian Shelf.

## 18. Arctic cod (Lilly and Bundy)

## Background

The Arctic cod (Boreogadus saida) have a circumpolar distribution. In the northwest Atlantic, they are found from Arctic waters in the north as far south as the southern Gulf of St. Lawrence (Scott and Scott 1988). They are semi-demersal and feed mainly on pelagic invertebrates. The Arctic cod, like capelin, is an important forage species and link in the transfer of energy from zooplankton to other fish, sea mammals and sea birds (Lilly et al. 1994). There is no commercial fishery for Arctic cod and therefore no assessment within the study area. As a consequence, the information on this species is poor.

## Catch

Arctic cod is not commercially fished. There was a small by-catch of $30 t$ in the shrimp fishery during 1985-1987 (D. Kulka, pers. comm.), but this is not regarded here.

## Biomass

Like capelin, the Arctic cod is inadequately sampled by the RV trawl surveys. The mean biomass estimated from these data for 1985 to 1987 in 2J3KLNO was 2, 819 t (Lilly et al. 1994 and DFO, unpublished data). In order to obtain a more realistic estimate of total biomass, capelin biomass data were used to develop a raising factor. The ratio between capelin RV survey biomass and capelin acoustic biomass was used to raise the RV survey biomass estimate of Arctic cod. The assumption is that the RV survey sampled each species equally poorly. This assumption is based on the difference between the biomass estimates obtained from the RV survey using Engels gear in 1992-1994, and using the Campelen gear in 1996-1997. Arctic cod biomass increased from an average of $4,015 \mathrm{t}$ to $61,388 \mathrm{t}$ while capelin biomass increased from 6,344 to $72,141 \mathrm{t}$. It is therefore assumed that the catchability of the two species is similar.

For capelin, the ratio of the acoustic biomass estimate for $2 \mathrm{~J} 3 \mathrm{KLNO}(6,578,068 \mathrm{t})$ to the Campelen RV biomass estimate for $2 \mathrm{~J} 3 \mathrm{KLNO}(13,730 \mathrm{t}$ ) was 479.1. This figure was applied to the mean biomass of Arctic cod in 1985-1987, giving a total biomass estimate of 1,350,713 $\mathrm{t}, 2.729 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Production:Biomass

There is little on which to base a production estimate of Arctic cod. Bradstreet et al. (1986) estimated mortality rates for Arctic cod in the Arctic. They constructed catch curves from numbers of Arctic cod at age found in the stomachs of predators, marine mammals and birds. They estimated a mean mortality of $0.409 \mathrm{yr}^{-1}$ for 1 to 2 year olds and $0.38 \mathrm{yr}^{-1}$ for 2 to 3 year olds. Lilly et al. (1994) found that most Arctic cod in 2J3KLNO were 1 and 2 years olds. On this basis, a mean of the two mortality estimates is taken to represent mortality of the
population. Thus, assuming $\mathrm{P} / \mathrm{B}=\mathrm{Z}$ (Allen 1971), $\mathrm{P} / \mathrm{B}$ is $0.395 \mathrm{yr}^{-1}$ and annual production is 532,856 t.

## Consumption:Biomass

In the absence of information on consumption by Arctic cod, it was assumed that the gross growth efficiency, the ratio between production and consumption, was 0.15 . Thus consumption was calculated as $3,552,376 \mathrm{t}$ and $\mathrm{Q} / \mathrm{B}$ was $2.633 \mathrm{yr}^{-1}$.

## Diet

The diet of Arctic cod (Appendix 2, Table 14) is a guess based on observations of a small number of fish caught off northern (Division 2G) and southern (Division 2J) Labrador in autumn 1978 (Lilly 1980) supplemented with unpublished observations of fish collected in Divisions 2J, 3K and 3L in autumn 1994. The first study revealed that small fish feed primarily on calanoid copepods and larger fish primarily on hyperiid amphipods. The second study revealed the same pattern but also indicated some predation by large Arctic cod on small capelin in Division 3L.

## 19 Large Pelagic Feeders (Bundy and Lilly)

## Background

This group includes the sharks, such as the basking shark (Cetorhinus maximus), the spiny dogfish (Squalus acanthias), tunas such as the bluefin tuna (Thunnus thynnus), pollock (Pollachius virens), silver hake (Merluccius bilinearis), swordfish (Xiphias gladius) and Atlantic salmon (Salmo salar). These are large fish that feed pelagically. None are common in the study area, and pollock and silver hake are found in greater abundance on the Scotian Shelf. The sharks, tunas and salmon are highly migratory (Scott and Scott 1988) and spend only part of the year in the study area. There is very little information for this group and as such, it is poorly defined. For those species that have juveniles in the area, the juveniles are classified as small pelagics (see below).

## Catch

Based on the total average catch within the study area of the species listed above given in NAFO Statistical Bulletins (Anon. 1987, 1989, 1990), and an average of 179 t caught as bycatch in the shrimp fishery (Kulka pers. comm.), the mean catch of the large pelagic feeders in 1985-1987 was calculated to be $2,693 \mathrm{t}$, or $0.006 \mathrm{t} \cdot \mathrm{km}^{-2}$. Seventy-five percent of this catch is Atlantic salmon and silver hake.

## Biomass

There is no biomass information for these species since they are mostly visitors to the area and are present for varying times. For this group, the model was used to estimate biomass by entering a default Ecotrophic efficiency value of 0.95 (Polovina 1984; Christensen and Pauly 1992).

## Production:Biomass

For the most part, these species are fairly long-lived and lightly exploited in the study area. Their $\mathrm{P} / \mathrm{B}$ ratio is thus likely to be low. Assuming the conventional natural morality rate of $0.2 \mathrm{yr}^{-1}$ and assuming $\mathrm{F}=\mathrm{M}$, the $\mathrm{P} / \mathrm{B}$ ratio was taken to be $0.4 \mathrm{yr}^{-1}$.

## Consumption:Biomass

The are very few data on consumption:biomass for the large pelagic feeding species in the study area. Stillwell and Kohler (1985) give a range of 3.4 to $5.8 \mathrm{yr}^{-1}$ for swordfish from Cape Hatteras to the Grand Banks. Pauly (1989) gives estimates of Q/B for several of the large pelagic feeding species in other areas (Table 10). Consumption in colder, northern waters is likely to be lower than these values. The maximum $\mathrm{Q} / \mathrm{B}$ may be around $4 \mathrm{yr}^{-1}$.

Table 10. Q/B estimates from Pauly (1989).

| Species | Location | Q/B $\left(y r^{-1}\right)$ |
| :---: | :---: | :---: |
| Basking Shark | North Sea | 3.70 |
| Spiny Dogfish | New England Coast | 4.77 |
| Pollock | Faeroe and Georges Bank | 4.76 |
| Silver Hake | Georges Bank | 4.26 |
| Swordfish | Cape Hatteras to Grand Banks | $3.4-5.8^{a}$ |

${ }^{\text {a }}$ Stillwell and Kohler (1985)
Assuming a gross efficiency of 0.15 would give a $\mathrm{Q} / \mathrm{B}$ of $2.667 \mathrm{yr}^{-1}$. The likely range of values is then 2.7 to $4 \mathrm{yr}^{-1}$. An intial $\mathrm{Q} / \mathrm{B}$ of $3.333 \mathrm{yr}^{-1}$ was used.

## Diet

The diet of silver hake (Appendix 2, Table 15) is based on Vinogradov (1983) with modifications based on information in Vinogradov (1972) and Swan and Clay (1979) and assumptions regarding prey availability on Grand Bank.

The diet of pollock (Appendix 2, Table 15) is based on Steele (1963) with modifications to reflect assumptions regarding prey availability on Grand Bank.

A diet for large pelagic feeders was calculated as the average of the diets for the above two species. It was assumed that this diet, with its emphasis on pelagic fish and zooplankton, would be appropriate for the combination of other species within the group.

## 20. Piscivorous Small Pelagic Feeders (Winters, Bundy and Lilly)

## Background

This group includes mackerel (Scomber scombrus), piscivorous myctophids and other mesopelagics and the short-finned squid (Illex illecebrosus). Mackerel and short-finned squid are highly migratory and spend only part of the year within the study area. The myctophids and other mesopelagics may remain in the area for the whole year. The group also includes
piscivorous juveniles of the large pelagics.
Mackerel populations in the Northwest Atlantic form a stock complex which overwinters off the New England coast and migrates northwards in May and June after spawning in the Jersey Bight area and the southern Gulf of St. Lawrence (Moores et al. 1975). They sporadically enter Newfoundland waters in late June-early July and begin their southern overwintering migration in the late fall. Only mackerel and short-finned squid are fished commercially.

## Catch

The mean catch of mackerel in NAFO Divisions 2J3KL for 1985-1987 was 7376 tons (Grégoire 1996). The short-finned squid catch statistics were taken from NAFO Statistical Bulletin (Anon. 1987, 1989, 1990), giving a total catch of 7405.7 t , or $0.015 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Biomass

Castonguay and Grégoire (1989) estimated that the mackerel population in the northwest Atlantic was exploited at an average level of about $4 \%$ during the 1985-87 period. If this level of exploitation pertains to the Newfoundland mackerel fishery, it implies an exploitable biomass of 184,411 tons for the period 1985-1987 period. This represents the average annual biomass and takes into account the period of residence in the area (since catch only occurs during periods when mackerel are present).

The biomass of the other species is not known. The biomass of short-finned squid is likely to have been low in 1985-1987 because catches were very small. It is assumed that the combined biomass of short-finned squid and myctophids may range from 10,000 to $30,000 \mathrm{t}$, giving a biomass range of 204,411-244,411, or $0.413-0.494 \mathrm{t} \cdot \mathrm{km}^{-2}$. A midrange value of $224,411 \mathrm{t}, 0.453 \mathrm{t} \cdot \mathrm{km}^{-2}$ was used as an initial input biomass.

## Production:Biomass

A P/B ratio of $0.29 \mathrm{yr}^{-1}$ was calculated for mackerel in the North Sea from MSVPA data (Christensen 1995). This value is similar to the value of $0.24 \mathrm{yr}^{-1}$ produced by assuming that production is equivalent to natural mortality times the biomass plus the catch (or $\mathrm{P} / \mathrm{B}=\mathrm{Z}$ ). These values are taken to represent the $\mathrm{P} / \mathrm{B}$ of the small pelagic feeders, producing a total annual production of between 54,169 and 64,769 t for 1985-1987 when the biomass was $224,411 \mathrm{t}$. A mid-value of 0.265 was used as the initial $\mathrm{P} / \mathrm{B}$ value.

## Consumption:Biomass

There are no consumption figures for the small pelagic feeders in the study area. Pauly (1989) gives a $\mathrm{Q} / \mathrm{B}$ value of $4.4 \mathrm{yr}^{-1}$ for mackerel on Georges Bank. However, assuming a GE of 0.15 would result in a $\mathrm{Q} / \mathrm{B}$ of 1.767 . These two values are quite different. In the absence of further information the default assumption of $\mathrm{GE}=0.15$ is used as an initial input value, but it is noted that $\mathrm{Q} / \mathrm{B}$ may be as high as $4.4 \mathrm{yr}^{-1}$. A Q/B of 1.767 results in a total annual consumption of $396,459 \mathrm{t}$ by the piscivorous small pelagic feeders.

## Diet

The derivation of the diet of the piscivorous small pelagics is given in Appendix 2, Table 16. The diet of mackerel is taken from Moores et al. (1975). The diet of piscivorous myctophids and other mesopleagics ) is a guess, based on diet information in Podrazhanskaya (1993) and Gartner et al. (1997). The diet of short-finned squid was based on sampling in the inshore of eastern Newfoundland (Dawe 1988; Dawe et al. 1997). The time and location of samples of short-finned squid probably yields an underrepresentation of the importance of capelin and crustacean zooplankters, so both groups are adjusted upward and the values for Atlantic cod and sand lance are adjusted downward.

The diet for piscivorous small pelagics was calculated as the weighted average of the diets for mackerel, mesopelagics and squid.

## 21. Planktivorous Small Pelagic Feeders (Winters, Bundy and Lilly)

## Background

This group includes herring (Clupea harengus harengus), planktivorous myctophids and other mesopelagics, and the Atlantic saury (Scomberesox saurus). Also included are Arctic squid (Gonatus sp.) and the planktivorous juveniles of the large pelagics. Herring, myctophids, other mesopelagics and Arctic squid remain in the area for the whole year. The Atlantic saury is highly migratory and spends only part of the year within the study area.

Only herring are commercially fished. Three major stocks are recognized within the study area, viz. White Bay-Notre Dame Bay, Bonavista Bay-Trinity Bay, and Conception BaySouthern Shore (Wheeler and Winters 1984).

## Catch

The average catch of herring in the study area during the 1985-1987 period was $9,000 \mathrm{t}$ (Wheeler et al. 1989). The other catch statistics were taken from NAFO Statistical Bulletins (Anon. 1987, 1989, 1990) giving a total catch of 9322 t , or $0.019 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Biomass

The average biomass of herring during the period 1985-1987 for the White Bay to Trinity Bay area is 212,000 tons (Wheeler et al. 1989). There is no estimate of biomass available for the Conception Bay-Southern Shore stock for this period, but historically this has been a small stock amounting to no more than $10 \%$ of the biomass of the northern stock. Thus, the total herring biomass is approximately 235,000 tons.

The biomass of the other species is not known. The biomass of the Atlantic saury is likely to have been low in these years because catches were very small. However, the biomass of the planktivorous myctophids and other mesopelagics may have been as high or higher than herring. In the absence of further information, a biomass range is used, from 352,500$587,500 \mathrm{t}$, or $0.712-1.187 \mathrm{t} \cdot \mathrm{km}^{-2}$. This allows for up to $60 \%$ of the biomass to be composed of myctophids, mesopelagics and Arctic squid. A mid range value of $470,000 \mathrm{t}, 0.949 \mathrm{t} \cdot \mathrm{km}^{-2}$,
was used as the initial input biomass.

## Production:Biomass

$\mathrm{A} \mathrm{P} / \mathrm{B}$ value of $0.24 \mathrm{yr}^{-1}$ is produced by assuming that production is equivalent to natural mortality times the biomass plus the catch. This is taken to represent the $\mathrm{P} / \mathrm{B}$ of the planktivorous small pelagic feeders, producing a total annual production of $112,800 \mathrm{t}$ for 1985-1987.

## Consumption:Biomass

There are no consumption figures for the small pelagic feeders in the study area. Pauly (1989) gives a $\mathrm{Q} / \mathrm{B}$ value of $4.59 \mathrm{yr}^{-1}$ for herring on Georges Bank. However, assuming a GE of 0.15 would result in a $\mathrm{Q} / \mathrm{B}$ of 1.60 . These two values are quite different. In the absence of further information, the default assumption of $\mathrm{GE}=0.15$ is used as an initial input value, but it is noted that $\mathrm{Q} / \mathrm{B}$ may be as high as $4.59 \mathrm{yr}^{-1}$. A Q/B of 1.6 results in a total annual consumption of $752,000 \mathrm{t}$ by the planktivorous small pelagic feeders.

## Diet

The derivation of the planktivorous small pelagics diet is given in Appendix 2, Table 17. The diet of herring is taken from Messieh et al. (1979), who studied fish from the Gulf of St. Lawrence and adjacent waters. The diet of myctophids and other mesopleagics is a guess, based on diet information in Podrazhanskaya (1993) and Gartner et al. (1997) and the assumption that myctophids and other planktivores comprise most of the biomass.

The diet for small pelagics was calculated as a weighted average of the diets for herring, myctophids and mesopelagics, and Arctic squid.

## 22. Shrimp (Parsons and Bundy)

## Background

Several species of penaeid and caridean shrimp occur within the study area. Northern shrimp (Pandalus borealis) dominates the biomass and is fished commercially. It is taken as the representative species for this group. The fishery during 1985-1987 occurred only in the northern area, that is 2 J 3 K . There was virtually no fishery in 2 J 3 K in 1985 , some in northern 2J in 1986 and a little more throughout 2J with a small bit in 3K in 1987. Northern shrimp off northeastern Newfoundland and southern Labrador generally prefer muddy substratum at depths greater than 200 m where water temperatures are in the range of $2-4^{\circ} \mathrm{C}$. Thus, the potential habitat for the species is extensive throughout the study area. Fishery and survey data suggest that animals occurring in the northern part are highly concentrated during winter and spring, but are more widely dispersed in late summer and early fall.

## Catch

The study period captures the initial stages of a rapid expansion of the fishery. After an increase in catches from the late 1970s to early 1980s, the fishery in Division 2J declined to a minimum in 1985 when only 2 tons were reported. This increased to 1328 t in 1986 and 3263
t in 1987, giving an average annual catch of 1531 tons (DFO, pers. comm). In addition to the directed catch, there was also a by-catch of 210.2 t of other shrimp in the shrimp fishery during 1985 to 1987 (D. Kulka, pers. comm.). This makes the total catch $1,741.2 \mathrm{t}$, or 0.004 $\mathrm{t} \cdot \mathrm{km}^{-2}$.

## Biomass

No biomass estimates for the whole study area are available for this period. Surveys in the 1980s were conducted in depressions of the northeast Newfoundland-Labrador Shelf where the fishery occurred and/or where bottom trawling was possible. Three areas surveyed in 1985 produced an estimate for northern shrimp of around $10,000 \mathrm{t}$ (Parsons and Veitch, 1991). However, this is a very minimal estimate given the unquantified efficiency of the survey gear, the vertical distribution of the shrimp out of range of the gear, and the unknown biomass that exists in untrawlable and other areas. For example, Lilly et al. (1998) described a broader distribution of northern shrimp inferred from both commercial fishery data and the stomach contents of Atlantic cod. They also showed that the distribution of northern shrimp extends south into 3L and the density of the striped pink shrimp ( $P$. montagui) may be greater in 3L than in 2 J 3 K .

Given the uncertainties in the data, it is necessary to assume that the biomass is greater than $10,000 \mathrm{t}$. Recent biomass (1996-97) is estimated at roughly $400,000 \mathrm{t}$ and the standardized catch rate has increased by 3.75 from 1988 to 1997 (Parsons and Veitch, 1998). Assuming the CPUE reflects the general trend in abundance, it is estimated that the mean biomass of northern shrimp from 1985-1987 was about 100,000 t. This extrapolation is for northern shrimp alone, but it is assumed the $100,000 \mathrm{t}, 0.202 \mathrm{t} \cdot \mathrm{km}^{-2}$ is representative of the total shrimp biomass.

## Production:Biomass

Three are no production or mortality estimates from shrimp in the study area. Hopkins and Nilsen (1990) estimated a total mortality of 1.83 from survival curves and length converted catch curves for northern shrimp in northern Norway. Hopkins (1988) estimated a P/B ratio of 1.7, using VBGF parameters and mortality (estimated by the length converted catch curve) amongst other input parameters. However, on the Newfoundland-Labrador shelf (Division 3 K ) northern shrimp grows to a greater size and older age and has a slower growth rate than in northern Norway (Parsons et al. 1986; Hopkins and Nilsen 1990). This indicates that the P/B ratio for the Newfoundland-Labrador Shelf northern shrimp should be less than either of the above estimates. A range of $\mathrm{P} / \mathrm{B}$ values, from 1.2 to 1.7 , is suggested and a value of 1.45 is used as an initial P/B.

## Consumption:Biomass

In the absence of information on the food consumption rate by shrimp, consumption was estimated by assuming a gross growth efficiency of 0.15 . This produces a $\mathrm{Q} / \mathrm{B}$ ratio of 9.667 $\mathrm{yr}^{-1}$ when $\mathrm{P} / \mathrm{B}=1.45 \mathrm{yr}^{-1}$ and a total annual consumption of $966,700 \mathrm{t}$.

## Diet

Feeding occurs in the benthic and the pelagic environments. Previously thought to be
primarily a benthic feeder (see Shumway et al 1985 for a review), northern shrimp is now known to feed extensively in the pelagic zone (e.g. Hopkins et al., 1993; Wienberg, 1981). The benthic diet is comprised mainly of detritus, phytoplankton and benthic invertebrates; the pelagic diet is comprised of detritus, small zooplankton and large zooplankton (DFO, unpublished data). The benthic and pelagic diets were combined assuming that $30 \%$ of the total diet is benthic and $70 \%$ is pelagic. The resultant diet is given in Table 2 b .

## 23. Large Crustaceans (Dawe, Ennis and Bundy)

## Background

The large crustacean group is comprised of the American lobster (Homarus americanus) and the snow crab or queen crab (Chionoecetes opilio), both of which were exploited in the mid1980s, and various non-commercial species such as toad crabs (Hyas spp).

## Catch

The 1985-1987 average catch of American lobster and snow crab was 701 t and 7101 t respectively (NAFO Statistical Bulletins 1986, 1987, 1990). The total large crustacean catch was therefore $7802 \mathrm{t}, 0.016 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Biomass

The biomass of the lobster was first approximated by dividing the catch by the exploitation rate (Ennis et al. 1982, Ennis et al. 1989). The estimate is for the exploited part of the stock and does not include either ovigerous females of commercial size or undersized lobsters, ie., pre-recruits. The biomass estimate was first increased to allow for commercial sized ovigerous females, using size-structure data from commercial sampling, and based on their relative abundance and size composition in research samples. Lobsters smaller than the minimum legal size of 81 mm carapace length (CL) were estimated by back calculating from the number in the $81-90 \mathrm{~mm}$ CL length class, assuming a natural mortality of 0.1 . The process of back-calculation was repeated for the 61-70 $\mathrm{mm}, 51-60 \mathrm{~mm}, 41-50 \mathrm{~mm}$ and $31-40 \mathrm{~mm}$ length groups. Biomass of the $31-40 \mathrm{~mm}$ groups decreased to about $10 \%$ of the $\geq 81 \mathrm{~mm}$ groups. Biomass of the $\leq 31-40 \mathrm{~mm}$ length group was assumed to be $10 \%$ of the $\geq 81 \mathrm{~mm}$ group. This gave the mean annual biomass for 1985-1987 of 2217 t .

Biomass estimates for the snow crab were made using fall bottom-trawl survey estimates for 1995-1997 (Dawe et al. 1997, in prep.) and the relationship between these biomass estimates and commercial CPUE (kg/trap haul) (Taylor and O'Keefe, in prep.) from the trap fishery in each of those years. From this relationship it is possible to estimate the catchability coefficient, q , which in this case represents the proportion of the resource removed by one trap haul ( $\mathrm{q}=$ CPUE/Biomass). Biomass is a conservative estimate of the initial exploitable biomass calculated by summing the catch and the residual fall (minimum trawlable) biomass. It is considered to be conservative because the catchability of the survey trawl is assumed to equal 1. Assuming q to be constant over time, the 1997 estimates of $q$ for each NAFO Division were applied to commercial CPUE data for each of the years 1985-1987 to generate biomass estimates (Biomass=CPUE/q). These biomass estimates were adjusted upward to
account for sublegal-sized males and females. This adjustment factor was based on the estimated maximum contribution of non-commercial snow crabs to the 1995-1997 total biomass estimates ( $55 \%$ in 1995) (Dawe et al. 1997, in prep.). Use of the maximum is justified in that biomass of this component must have been high during 1985-1987, because CPUE began to increase regularly after this period (Taylor and O'Keefe in prep.), indicating increasing recruitment. This too may be conservative if catchability of the survey trawl is lower for smaller crabs. The resultant total biomass of $86,345 \mathrm{t}$ is considered to be a minimum estimate.

The biomass of the other crab species was calculated from the fall bottom-trawl survey estimates for 1995-1997 (Joe Drew, DFO, unpublished data, pers. comm.). It is assumed that since these species were not exploited until the 1990s, these biomass estimates represent a minimum unexploited biomass. The mean minimum trawlable biomass for 1995 to 1997 was $2,901 \mathrm{t}$.

The total large crustacean biomass estimated for 1985-1987 was $91,462 \mathrm{t}$. For the model area this gives an estimate of $0.185 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Production:Biomass

In the absence of mortality estimates for either the lobster or crabs, the production biomass ratio was estimated from catch and biomass data and an assumed natural mortality of 0.1 to 0.3. This produces a mean $\mathrm{P} / \mathrm{B}$ of 0.182 to $0.382 \mathrm{yr}^{-1}$, or total production of 16,668 to $34,961 \mathrm{t}$ per year. The midpoint of this range, $0.282 \mathrm{yr}^{-1}$, was used as the initial input value for $\mathrm{P} / \mathrm{B}$.

## Consumption:Biomass

There were no consumption data available for the snow crab so it was assumed that the consumption rates of lobsters are representative of the crabs. The consumption to biomass ratio for lobsters was calculated from data in Reddin (1973). Reddin studied the consumption of lobsters caught at various parts of the coast around Newfoundland. He maintained them at ambient seawater temperature in the laboratory for one year, and fed them a diet of either $100 \%$ rock crab (Cancer irroratus) or 100\% sea urchin (Strongylocentrotus droebachiensis). The total annual calorific consumption of C. irroratus was $2663 \mathrm{cal} / \mathrm{g}$ lobster and of $S$. droebachiensis was $1425 \mathrm{cal} / \mathrm{g}$ lobster. Reddin estimated the calorific content of C. irroratus as $4211 \mathrm{cal} / \mathrm{g}$ dry wt and of $S$. droebachiensis as $3448 \mathrm{cal} / \mathrm{g}$ dry wt. Converting this to $\mathrm{cal} / \mathrm{g}$ wet wt (assuming 10 g wet $\mathrm{wt}=1 \mathrm{~g}$ dry weight), the weight of prey consumed was estimated as the calorific consumption by lobster divided by the calorific content of prey. This produced an annual consumption rate of $6.32 \mathrm{yr}^{-1}$ when the lobsters were fed $100 \%$ C. irroratus and $4.42 \mathrm{yr}^{-1}$ when the lobsters were fed $100 \%$ S. droebachiensis. Based on the diet composition given below, the weighted mean of these two estimated is $5.85 \mathrm{yr}^{-1}$, giving a total annual consumption of $533,888 \mathrm{t}$.

## Diet

The diet data shown in Table $2 b$ is based on the weighted mean of the lobster and snow crab diets. The lobster diet data from Bonavista Bay in Ennis (1973) were used and adapted slightly.

Diet data from western Newfoundland (Wieczorek and Hooper 1995) and Conception Bay in eastern Newfoundland (Miller and O'Keefe 1981) were used as a basis for the diet of the snow crab. Diet data in Lovrich and Sainte-Marie (1997) were also used for comparative purposes. The data were presented in terms of percentage occurrence whereas what is required here is percentage wt of stomach contents. On the basis of similarity in frequency occurrence in Wieczorek and Hooper (1995) and Miller and O'Keefe (1981), the prey items were allocated a percentage weight value.

## 24-27. The Benthic Invertebrates (Bundy and Gilkinson)

## Background

The benthic invertebrates are treated as four groups: Echinoderms, Molluscs, Polychaetes and Other Benthic Invertebrates (OBI). The other benthic invertebrates include miscellaneous crustaceans, nematodes, and other meiofauna. These groupings reflect the major taxa observed on the Grand Bank (Nesis 1965; Hutcheson et al. 1981). The most recent comprehensive source of information on the benthos is a study carried out in 1980 under the auspices of the Mobil Oil company on the Grand Banks (Hutcheson et al. 1981). Five stations were sampled from May to November 1980. Barrie (1979) and Barrie et al. (1980) conducted a coastal study of marine benthic communities off the Labrador coast in 1977 and 1979. Additional quantitative information on marine benthic communities collected in conjunction with a three year trawling impact experiment on the Grand Bank (Prena et al. 1996; Gordon et al. 1997) is not yet available. The best source of published information is the Mobil study on the Grand Banks. The Labrador coast studies generally corroborate the results of this study. However, it should be stressed that benthic data are poor, and several assumptions are made below in order to obtain model parameter estimates.

Large scale differences in the distribution of benthic species are correlated with water mass type (Nesis 1965). Depth integrates the variables of water column productivity and sediment textural type. It is now clear that infauna-sediment relationships are much more variable than traditionally purported and the distributions of this group cannot be explained on the basis of sediment grain size alone (Snelgrove and Butman 1994). Schneider et al. (1987)
demonstrated that mobility of megafauna determines the range over which they are associated with substrate variability. For the purposes of this model, however, the scale of interest is the southern Labrador shelf, Northeast Newfound Shelf and Grand Bank. Therefore little attention is paid to small scale changes in species distribution. High productivity on Grand Bank is linked to a shallow water depth and mixing of waters from the Labrador current, shelf waters, and southern waters. However, high productivity may be also be due to cold water temperatures which suppress water column microbial metabolism, resulting in more organic matter made available to consumers (Pomeroy and Deibel 1986).

The benthic groups are all common on Grand Bank (Hutcheson et al. 1981). Polychaete worms are numerically the most dominant group, and in particular Exogene hebes, which had a mean numerical density of $220 \mathrm{~m}^{-2}$. Echinoderms and molluscs contribute most to the benthic biomass. The sand dollar, Echinarachnius parma, is a consistent feature of sand assemblages with a mean numerical density of $25 \mathrm{~m}^{-2}$. The bivalve mollusc, Mesodesma
deauratum, occurs at very high densities, up to $22 \mathrm{~kg} \cdot \mathrm{~m}^{-2}$, on the south east portion of the bank. However, this is very localised, and other mollusc species, such as the propeller clam (Cyrtodaria siliqua), which has maximum mean biomass of around $500 \mathrm{~g} \cdot \mathrm{~m}^{-2}\left(=500 \mathrm{t} \cdot \mathrm{km}^{-2}\right)$ on sandy bottoms (K. Gilkinson, unpublished data - this figure was derived from a hydraulic grab that penetrates deeper into the substrate than the van Veen grab-see below), are more common on Grand Bank. Hutcheson et al. (1981) reported that there are no strong seasonal changes in dominant species at the five stations.

The surficial sediments on Grand Bank consist primarily of sands and gravels of various coarseness grades. Sands below 110 m water depth have few associated gravels while those above are interspersed with gravels, originating primarily from Pleistocene glacial deposits. At shallower depths, surficial sediments have been re-worked by a combination of a Holocene marine transgression and modern hydrodynamic processes (Barrie et al. 1984; Barrie and Collins 1989). In the Labrador study (Barrie et al. 1980) the dominant substrate at all sites was fine sand.

## Catch

Polychaetes, echinoderms and OBI were not commercially caught in 1985-1987. The only benthic invertebrate species (as classified in this Ecopath group) that was commercially landed during 1985-1987 was a mollusc, the giant scallop (Placopecten magellanicus), with average annual landings of 121 t , or $0.0003 \mathrm{t} \cdot \mathrm{km}^{-2}$ (Anon 1987, 1989, 1990).

## Biomass

The estimate of mean biomass for each of the four benthic groups was estimated from data from the Mobil study. Hutcheson et al. (1981) sampled five stations on the Grand Banks, including the Hibernia Station, which consisted of 10 stations in the area, and station 24 which was on St. Pierre Bank in NAFO Subdivision 3Ps. The locations of the stations were chosen to reflect the water mass qualities described by Nesis (1965). Station 3 was at the north edge of the bank, station 33 in the middle of the bank in a region of mixed waters, and station 48 was near the edge of the bank on the Southeast Shoal, a region of high productivity. Samples were taken with a $0.1 \mathrm{~m}^{2}$ van Veen grab. Samples were also taken with an epibenthic sled, which, because of operation difficulties, were qualitative. However, they supported the results of the grab samples.

The results demonstrated that benthic communities at station 3, on the northern edge of the Grand Bank, had a strong arctic component and were different from the species at the other stations. It was assumed that the benthic composition at this station was representative of the benthic composition of the Labrador Shelf area, that is, of 2J3K. For 3LNO, roughly the Grand Bank, a weighted average of the mean weights at each station was calculated for each benthic group, based on Figure 3 and Figure 8 in Nesis (1965). Figure 3 gives the distribution of benthos biomass by weight groups ( $<100 \mathrm{~g}, 100-1000 \mathrm{~g},>1000 \mathrm{~g}$ ) and Figure 8 is a map of the distribution of water types in the bottom layers. The stations from the Mobil study were mapped onto these maps and given a representative coverage of the Grand Bank (or 3LNO) based on biomass distribution and water mass type. The area covered by each station was estimated as $55 \%$ for station 3, $44 \%$ for station 33, $2.5 \%$ for Hibernia and $2.5 \%$ for station 48.

The biomass of the molluscs is very large, especially on the Southeast Shoal, but also in 2 J 3 KL . However, only part of this biomass, the soft body tissue, is transferred through the food web. This is confirmed by observations of huge shells beds on the ocean floor (Hutcheson et al. 1981). In order to reduce the biomass and account for soft body tissue only, the ratio of the body weight to whole weight of the locally abundant mollusc Mesodesma deauratum was estimated. The mean ratio between blotted wet weight of tissue weight to whole weight for animals with a shell length between 30 and 35 mm was 0.166 with a standard deviation of $0.023, \mathrm{n}=10$ (Gilkinson, unpubl.).

The estimate of benthic biomass for 3 LNO is $230.6 \mathrm{t} \cdot \mathrm{km}^{-2}$ (Table 11). The mean biomass estimates for 2 J 3 KLNO are the weighted mean of the three areas, and the total mean benthic biomass for 2 J 3 KLNO is $127.7 \mathrm{t} \cdot \mathrm{km}^{-2}$.

Table 11. Estimates of mean benthic biomass in 2J3KLNO.

|  | 2 J or $3 K$ <br> Biomass $\left(t \cdot \mathrm{~km}^{-2}\right)$ | 3LNO <br> Biomass $\left(t \cdot \mathrm{~km}^{-2}\right)$ | Biomass $\left(t \cdot \mathrm{~km}^{-2}\right)$ |
| ---: | :---: | :---: | :---: |
| Echinoderms | 70.6 | 144.8 | 112.3 |
| Molluscs | 16.4 | 62.2 | 42.1 |
| Polychaetes | 8.8 | 11.9 | 10.5 |
| Other Benthic Invertebrates | 2.7 | 11.8 | 7.8 |
| Total Benthos Biomass | 98.5 | 230.6 | 127.7 |

Hutcheson et al. (1981) estimated a total benthic biomass on the Grand Banks of $480.7 \mathrm{t} \cdot \mathrm{km}^{-2}$. The estimate of $230.7 \mathrm{t} \cdot \mathrm{km}^{-2}$ is half this. However, if the mollusc biomass is not corrected for shell weight, as in Hutcheson et al. (1981), then the estimates are similar ( $544.2 \mathrm{t} \cdot \mathrm{km}^{-2}$ ). The Hutcheson et al. estimate of benthic biomass was not used because they omitted certain data from their estimate and it was not clear how they combined the stations into an overall mean. Nesis (1965) estimated a total benthic biomass of $154 \mathrm{t} \cdot \mathrm{km}^{-2}$ for the Grand Bank and the Labrador Shelf. The estimate here is comparable.

The mean biomass estimate in Table 11 may be an underestimate of the total benthic biomass. In particular, the molluscs may be undersampled. The $0.1 \mathrm{~m}^{2}$ van Veen Grab does not deeply penetrate the compacted sands on the northeast Grand Bank and therefore misses some large deep burrowing bivalves (Gilkinson, unpubl.; Prena et al. 1996). However, these are the only currently available published biomass estimates for Grand Bank.

## Production:Biomass

There are some empirical production data for the mollusc group on Grand Bank. For the other benthic groups, all the estimates are taken from literature sources. The resultant estimates are shown in Table 12.

## Echinoderms

Robertson (1979) estimated an annual $\mathrm{P} / \mathrm{B}$ of $0.65 \mathrm{yr}^{-1}$ and Jarre-Teichmann and Guenette (1996) used an estimate of $0.6 \mathrm{yr}^{-1}$ for the southern BC shelf. In the absence of other information, the lower of the two estimates, $0.6 \mathrm{yr}^{-1}$, is used here. This translates into an
annual production of $67.4 \mathrm{t} \cdot \mathrm{km}^{-2} \cdot \mathrm{yr}^{-1}$ of echinoderms.

## Molluscs

Hutcheson et al. (1981) estimated the P/B ratio of the bivalve Mesodesma deauratum from the Southeast Shoal area of Grand Bank as $0.57 \mathrm{yr}^{-1}$ for small individuals and $0.11 \mathrm{yr}^{-1}$ for large ones. Robertson (1979) estimated a P/B of $0.76 \mathrm{yr}^{-1}$ for molluscs and Jarre-Teichmann and Guenette (1996), based on their interpretation of Brey's (1995) database on macrobenthic productivity, estimated a mean $\mathrm{P} / \mathrm{B}$ of $0.7 \mathrm{yr}^{-1}$ for molluscs on the southern shelf of British Columbia. Unfortunately, Hutcheson et al. (1981) do not specify how large a large M. deauratum is, or what proportion of the population is large. To allow for the higher estimates originating in warmer waters (and therefore having a greater productivity), the lower estimate of $0.57 \mathrm{yr}^{-1}$ is used, producing $24 \mathrm{t} \cdot \mathrm{km}^{-2} \cdot \mathrm{yr}^{-1}$ of molluscs.

## Polychaetes

Estimates of polychaete $\mathrm{P} / \mathrm{B}$ ratios range between 2 and $3 \mathrm{yr}^{-1}$ (Mills and Fournier 1979; Collie 1987; Jarre-Teichmann and Guenette 1996). Curtis (1977) estimated a lower value of $1.4 \mathrm{yr}^{-1}$ for polychaetes off West Greenland. To allow for the higher estimates originating in warmer waters (and therefore having a greater productivity), the lower bound of $2 \mathrm{yr}^{-1}$, that is a production of $21 \mathrm{t} \cdot \mathrm{km}^{-2} \cdot \mathrm{yr}^{-1}$, is used.

## Other Benthic Invertebrates (OBI)

Little is know about production of the OBI. The amphipods are assumed to represent this group. P/B estimates for amphipods include 2.5-4.4 $\mathrm{yr}^{-1}$ (on Georges Bank, NE USA) (Collie 1985), $2.5 \mathrm{yr}^{-1}$ (Mills and Fournier 1979) and $2.5 \mathrm{yr}^{-1}$ (Jarre-Teichmann and Guenette 1996). The common estimate of $2.5 \mathrm{yr}^{-1}$ is used, producing $19.521 \mathrm{t} \cdot \mathrm{km}^{-2} \cdot \mathrm{yr}^{-1}$ of other benthic crustaceans.
The total benthic production is $132 \mathrm{t} \cdot \mathrm{km}^{-2} \cdot \mathrm{yr}^{-1}$. The greatest production is by the echinoderms while the other 3 groups have a similar annual production. This compares to a total mean macrobenthic production of between 12 and $533 \mathrm{t} \cdot \mathrm{km}^{-2} \cdot \mathrm{yr}^{-1}$ on Browns Bank, Nova Scotia (Wildish et al. 1989) and $360 \mathrm{t} \cdot \mathrm{km}^{-2} \cdot \mathrm{yr}^{-1}$ on Georges Bank (Sissenwine et al. 1984).

## Consumption:Biomass

The $\mathrm{Q} / \mathrm{B}$ ratio is estimated from the gross efficiency (GE) for all four benthic groups. JarreTeichmann and Guenette (1996) estimated GE for benthic groups on the southern BC Shelf. Using their values, the molluscs, polychaetes and echinoderms, all detritus feeders (see below), have a GE of $9 \%$, or 0.09 . The other benthic invertebrates have a GE of $20 \%$. These GE values equate to the $\mathrm{Q} / \mathrm{B}$ and consumption rates shown in Table 12.

Table 12. Production and consumption estimates for the four benthic groups.

| Benthic Group | $P / B$ <br> $\left(y r^{-1}\right)$ | $P$ <br> $\left(t \cdot k m^{-2} \cdot y r^{-1)}\right.$ | $G E$ | $Q / B$ <br> $\left(y r^{-1}\right)$ | $Q$ <br> $\left(t \cdot k m^{-2} \cdot y r^{-1)}\right.$ |
| ---: | :---: | :---: | :---: | :---: | :---: |
| Echinoderms | 0.6 | 67.4 | 0.09 | 6.7 | 748.5 |
| Molluscs | 0.6 | 24.0 | 0.09 | 6.3 | 266.9 |
| Polychaetes | 2.0 | 21.0 | 0.09 | 22.2 | 233.8 |
| Other Benthic Invertebrates | 2.5 | 19.5 | 0.20 | 12.5 | 97.3 |

The greatest consumption is by echinoderms, followed by molluscs and polychaetes.

## Diet

Much of the diet information on these benthic groups originates in the work of Nesis (1965) and Hutcheson et al. (1981). Here, the diets are described in terms of the groupings used in the Ecopath model. Essentially, they are all described as benthic feeders. It should be noted however, that these data are simplified. It was not possible to review all the available information on congeneric and conspecifics in the literature.

## Echinoderms

The echinoderms are basically detritivores. The sand dollar Echinarchnius parma, the most common echinoderm in the study area, is a surface detritivore (Hutcheson et al. 1981; Mooi and Telford 1982). The sea urchin Stronglyocentrotus pallidus eats animal remains and detritus (Gilkinson et al. 1988) and the brittle star Ophiura robusta is a detritivore (Nesis 1965). The diet of this group then is $100 \%$ detritus.

## Molluscs

Among the bivalves, M. deauratum is a suspension feeder, Macoma calcarea is a deposit feeder, and Liocyma fluctuosa is a suspension or detrital feeder (Hutcheson et al. 1981). Suspension feeders feed on organic detrital matter which is re-suspended in the water immediately above the sediment surface. Deposit feeders can be considered detrital feeders. Thus, the molluscs can be said to be detrital feeders of various forms and the diet of the mollusc group is $100 \%$ detritus.

## Polychaetes

The polychaetes are detritivores and have a diet of $100 \%$ detritus (Nesis 1965; Fauchuld and Jumars 1979).

## Other Benthic Invertebrates (OBI)

Gammarid amphipods, the representative taxon for this group, feed on organic detritus (Nesis 1965; Hutcheson et al. 1981). The Cumacea and Isopoda are also detritivores (Hutcheson et al. 1981).

## 28-29. Zooplankton (Bundy and Deibel)

## Background

Zooplankton are an important link in marine food webs. They transfer organic carbon from phytoplankton to fish, marine mammals and birds higher in the food chain, they are a food source for a broad spectrum of species and they contribute faecal matter and dead zooplankton to the benthic communities. The most recent comprehensive source of information on the zooplankton in the Newfoundland region is a study carried out on the Grand Banks in 1980 under the auspices of the Mobil Oil company (Strong 1981). Strong (1981) sampled the Grand Banks on fifteen oceanographic cruises from March 1980 to February 1981, using a neuston net of $333 \mu \mathrm{~m}$ to sample surface dwelling zooplankton, a bongo sampler equipped
with twinned $333 \mu \mathrm{~m}$ mesh nets, and a ring net with mesh size of $80 \mu \mathrm{~m}$, designed to capture microzooplankton. Over eighty-six species of zooplankton were caught, and arctic and boreal species numerically dominated the samples. The copepod Calanus finmarchius dominated the $333 \mu \mathrm{~m}$ bongo and neuston samples. Two other Calanus species, C. glacialis and C. hyberboreus, characteristic of cold arctic waters, were also found in smaller numbers beyond the 200 m isobath. Other species included tunicates (Oikopleura spp.), chaetognaths (Parasagitta elegans), amphipods (Parathemisto gaudichaudii) and molluscs (Limacina spp.). The cyclopoid copepod Oithona similis dominated the ring net samples, and were most abundant within the 200 m isobath. Few euphausiids were caught and apparently they play a minor role on the Grand Banks (Strong 1981), although this is probably an artifact due to sampling technique. Gelatinous zooplankton, such as the tunicates, are generally underestimated by sampling gear, but are considered to have a very important role in marine ecosystems (Alldredge 1984; Deibel 1988; Knoechel and Steel-Flynn 1989; Madin and Deibel 1998).

The zooplankton are treated here as two groups. The large zooplankton group are greater than 5 mm in length and include euphausiids, chaetognaths, hyperiid amphipods, Cnidarians and Ctenophores (jellyfish), mysids, tunicates $>5 \mathrm{~mm}$ and icthyoplankton. This group feeds herbivorously (some euphausiid species), omnivorously (most euphuasiids, hyperiid amphipods, mysiids and large tunicates) and carnivorously (chaetognaths and jellyfish). The small zooplankton includes zooplankton less than or equal to 5 mm in length. Copepods, mainly Calanus finmarchicus and Oithona similis, are the most numerous small plankton. Other small plankton include tunicates $<5 \mathrm{~mm}$ and meroplankton. C. finmarchicus and $O$. similis are omnivorous.

## Catch

None.

## Biomass

Strong (1981) estimated an mean annual standing crop of zooplankton from the bongo net of $4.38 \mathrm{kcal} \cdot \mathrm{m}^{-2}$, which is equivalent to $7.3 \mathrm{t} \cdot \mathrm{km}^{-2}$, using the conversion factor $0.6: 1$ for zooplankton, estimated by Mills and Fournier (1979). This figure is for zooplankton within the 200 m isobath. Incorporating the microzooplankton from the ring net sample with an approximate calculation, Strong (1981) estimated the total zooplankton biomass to be 35.8 $\mathrm{t} \cdot \mathrm{km}^{-2}$.

This estimate however, does not include a fair representation of the gelatinous zooplankton which are damaged in the net, the jellyfish that are too large, and larval crustaceans (meroplankton) and chaetognaths that are too fast (Strong 1981). In addition, it is only for the Grand Banks.

An alternative method to estimate biomass was used here. Since 1994, a broad-scale survey has been carried out over 2 J 3 KLNO in late summer by DFO using two ships (Anderson and Dalley 1997). The survey captures zooplankton from size 0.4-10 mm and nekton sized 10-200 mm . The surveys use a $333 \mu \mathrm{~m}$ bongo net and an IYGPT trawl (International Young Gadoids Pelagic Trawl). Although the survey is aimed at young-of-the-year fish and older capelin, it
also represents a broad-scale measure of the zooplankton communities in the late summer. Zooplankton biomass, sampled by the bongos, was highest over the Northeast Newfoundland Shelf (Anderson and Dalley 1997). Indeed, Anderson and Dalley (1997) suggest that pelagic productivity, including production by nekton, may be greater over the Northeast Newfoundland Shelf than on Grand Bank. For this reason, and the better sampling technology, zooplankton biomass is estimated from these data in preference to the earlier data of Strong (1981), which only covered Grand Bank out to the 200 m isobath.

## Large Zooplankton

The biomass of the large zooplankton was estimated from the IYGPT trawl data. These data are split into three categories: fish, jellyfish and "other" which are basically euphausiids and amphipods. It is assumed that the jellyfish and "other" categories are representative of the large zooplankton biomass. Unfortunately, there are data only for 1996 and 1997 for these groups and it is assumed that they are representative of the period 1985-1987. The mean biomass of the jellyfish was 1.6 and $0.8 \mathrm{~g} \cdot \mathrm{~m}^{-2}\left(=\mathrm{t} \cdot \mathrm{km}^{-2}\right)$ in 1996 and 1997, and for the "others" was $0.6 \mathrm{~g} \cdot \mathrm{~m}^{-2}$ in both years (data supplied by Denise Davis, DFO).

Koslow et al. (1997) estimate that the catchability of the IYGPT trawl is 0.14 , meaning that the trawl catches $14 \%$ of the animals that it encounters. This catchability factor was used to raise the biomass estimates to 11.14 and $5.79 \mathrm{t} \cdot \mathrm{km}^{-2}$ for jellyfish in 1996 and 1997 and 4.11 and $4.61 \mathrm{t} \cdot \mathrm{km}^{-2}$ for the "others". A mean of the 1996 and 1997 values was taken for both groups and the figures adjusted for seasonality as described above. This gives a total mean biomass estimate of $18.343 \mathrm{t} \cdot \mathrm{km}^{-2}$ for the large zooplankton.

## Small Zooplankton

The biomass of the small zooplankton was estimated from bongo data from 1994-1997. The high contribution of the microzooplankton to the total zooplankton biomass estimated by Strong (1981) is ignored since they are not sampled by the bongo net. Zooplankton samples have a high variance and it is assumed that this variance will allow for the microzooplankton. Thus, the bongo net sample is assumed to represent the small zooplankton.

During the period 1994-1997, small zooplankton have increased inshore, in the north over the Northeast Newfoundland Shelf and in the south over Grand Bank. The mean biomass for each year is $11.8,16.3,17.3$ and $30 \mathrm{t} \cdot \mathrm{km}^{-2}$ (data supplied by Denise Davis, DFO). These figures compare to the $7.3 \mathrm{t} \cdot \mathrm{km}^{-2}$ estimated by Strong (1981) for the Grand Banks.

It is assumed, for want of better information, that the catchability of the bongos is 1 .
However, the biomass figures above are the estimates for late summer, after the spring bloom (Strong 1981). In order to account for seasonality and thus obtain a representative annual biomass estimate, the late summer figures were adjusted for seasonality using the ratio of the monthly displacement volumes of zooplankton from the bongo net samples to the late summer value (Strong 1981, Table 5-2). The resultant estimates are given in Table 13.

Table 13. Small zooplankton biomass estimates for $2 \mathrm{~J} 3 \mathrm{KNLO}\left(\mathrm{t} \cdot \mathrm{km}^{-2}\right)$.

|  | 1994 | 1995 | 1996 | 1997 |
| ---: | :---: | :---: | :---: | :---: |
| Small Zooplankton Biomass | 11.8 | 16.3 | 17.3 | 30.0 |
| Seasonally adjusted Biomass | 16.9 | 23.4 | 24.8 | 42.9 |
|  |  |  |  |  |

The average of the 1994-1996 biomass estimates, $21.7 \mathrm{t} \cdot \mathrm{km}^{-2}$, is used to represent the small zooplankton biomass for the period 1985-1987. The year 1997 is not included since it is clearly much higher than the other years, and there is no reason to assume that there was particularly high small zooplankton biomass in 1985-1987. The biomass estimate of 21.7 $\mathrm{t} \cdot \mathrm{km}^{-2}$ may be compared to a copepod biomass of $25.1 \mathrm{~g} \cdot \mathrm{~m}^{-2}\left(=\mathrm{t} \cdot \mathrm{km}^{-2}\right)$ estimated from production and P/B estimates from the Emerald Bank (Scotian Shelf) in 1979-1980 by McClaren et al. (1989).

## Production:Biomass

## Large Zooplankton

Mauchline (1985) estimated P/B ratios of euphausiid species in the Rockall Trough, in the Atlantic west of Ireland. Three of the species are known to occur in 2J3KLNO; Thysanopoda acutifrons, Meganyctiphanes norvegica and Thysanoessa longicaudata. The P/B ratios are $2.3,1.6$ and $6.4 \mathrm{yr}^{-1}$ respectively. In the absence of other information for the large zooplankton, an average value of $3.433 \mathrm{yr}^{-1}$ is taken as the $\mathrm{P} / \mathrm{B}$ ratio.

## Small Zooplankton

McClaren et al. (1989) estimated annual $\mathrm{P} / \mathrm{B}$ ratios for a range of copepod species on the Scotian Shelf, which is similar to the southern Grand Bank. It is assumed that C. finmarchius is the representative species for this group. It has a P/B ratio of $8.4 \mathrm{yr}^{-1}$. Sakshaug (1997), however, estimated a P/B of 4.0 for copepods in the Barents Sea, a system similar to the northern part of the study area. It should be noted that the cyclopoid copepod Oithona similis, which also belongs to the small zooplankton, has a much higher $\mathrm{P} / \mathrm{B}$ ratio of $23 \mathrm{yr}^{-1}$ on the Scotian Shelf and the $\mathrm{P} / \mathrm{B}$ ratios for tunicates $\leq 5 \mathrm{~mm}$ are not known. A range of $\mathrm{P} / \mathrm{B}$ ratios, with a minimum of $4.0 \mathrm{yr}^{-1}$ and maximum of $8.4 \mathrm{yr}^{-1}$, is used to represent the small zooplankton. A mid-range value of $6.2 \mathrm{yr}^{-1}$ was used as the initial input $\mathrm{P} / \mathrm{B}$ ratio.

## Consumption:Biomass

## Large Zooplankton

There is some information on Q/B values for the large zooplankton. Sameoto (1976) examined the energy budgets of 3 euphausiid species in the Gulf of St. Lawrence. These species also occur in 2J3KLNO. Using his figures, a Q/B of $5 \mathrm{yr}^{-1}$ was estimated for Meganyctiphanes norvegica, $16 \mathrm{yr}^{-1}$ for Thysanoessa inermis and $37 \mathrm{yr}^{-1}$ for Thysanoessa raschii. Heyraud (1979) collected M. norvegica from the waters of Nice, France and kept them under laboratory conditions at $13^{\circ} \mathrm{C}$ and fed them Artemia salina. This produced a $\mathrm{Q} / \mathrm{B}$ estimate of $40.15 \mathrm{yr}^{-1}$. This should be regarded as an upper limit due to the high temperature in which these euphausiids were reared. Jarre-Teichmann (1996) used a GE of 0.16 for krill in the Alaska Gyre. Purcell (1996) estimated a Q/B of $30 \mathrm{yr}^{-1}$ for salps and used a GE of 0.3 for
jellyfish in the Alaska Gyre.
An average $\mathrm{Q} / \mathrm{B}$ of $19.5 \mathrm{yr}^{-1}$ was estimated for the 3 euphausiid species (Sameoto 1976) and is used to represent the $\mathrm{Q} / \mathrm{B}$ of the large zooplankton. This fits well with assuming a GE of 0.16 , which would give a $\mathrm{Q} / \mathrm{B}$ of $21.15 \mathrm{yr}^{-1}$. The value may seem a little high, but it allows for the average $\mathrm{Q} / \mathrm{B}$ of the group to reflect the tunicates which are voracious consumers (Knoechel and Steel-Flynn 1989).

## Small Zooplankton

There is minimal information on the $\mathrm{Q} / \mathrm{B}$ values for copepods and other members of this group. Christensen (1995) assumed a gross efficiency of 0.3. Purcell (1996) estimated a Q/B of $90.4 \mathrm{yr}^{-1}$ for small copepods in the Alaska Gyre, based on copepod ingestion and clearance rates. This value seems rather high. Assuming a GE of 0.3 would result in a $\mathrm{Q} / \mathrm{B}$ of 20.67 $\mathrm{yr}^{-1}$.

## Diet

The two size classes of zooplankton feed on both autotrophic microplankton and heterotrophic microplankton, in addition to other diet components. However, in this model, both types of microplankton are grouped simply as phytoplankton. It is recognised that this ignores the dynamics of the microbial loop (Pomeroy 1979) and the potential contribution from the microbial loop to higher trophic levels. Thus, the diet of the small zooplankton would be mistakenly interpreted as herbivorous when taken at face value.

## Large Zooplankton

Euphausiids eat detritus, phytoplankton, chaetognaths, amphipods and crustaceans (copepods) (Mauchline 1980). Chaetognaths eat copepods (Sullivan 1980) and jellyfish eat copepods (Smayda 1993). Since the relative proportions of these species in the large zooplankton group are unknown, a representative diet for the group might be $38 \%$ small zooplankton, $37 \%$ phytoplankton, 20\% detritus and 5\% cannibalism.

## Small Zooplankton

The diet of the small zooplankton is assumed to be $100 \%$ phytoplankton.

## Acknowledgements

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## 30. Phytoplankton (Bundy and Caverhill)

## Background

Most information on phytoplankton in 2J3KLNO is derived from a study carried out in 1980 under the auspices of the Mobil Oil company on the Grand Banks (Hollibaugh and Booth 1981). Arctic species of phytoplankton are associated with the inshore and offshore branches of the Labrador current, and are widespread over the banks in winter. Phytoplankton require light and nutrients and the main determinant of primary production is water stratification.

When the water column is unstratified, vertical mixing occurs and whilst there may be plentiful nutrients in the water column, phytoplankton may be vertically displaced from the euphotic zone. In a stratified water column, phytoplankton remain in the euphotic zone and have adequate light for photosynthesis and production. With time, however, stability can reduce the supply of nutrients causing a decrease in productivity. The onset of stratification on the Grand Banks occurs around April or May, when there is a spring bloom of diatom species such as Chaetoceros decipiens and Thalassiosira spp. Phytoplankton biomass and productivity decrease in June, due to depletion of nutrients and to predation by copepods (Hollibaugh and Booth 1981). Dinoflagellates and nanoflagellates are the main summer species. At this stage, there may be a significant amount of die off of phytoplankton, which will sink to the bottom and provide energy to benthic communities and/or be incorporated into the microbial food web (Turner and Roff 1993). The most significant feature of phytoplankton biomass and productivity is the narrow band of increased biomass and productivity along the shelf break stations (Hollibaugh and Booth 1981). Phytoplankton have rapid growth rates and show large spatial and temporal variability in biomass distribution (Hollibaugh and Booth 1981).

Biomass and production are the only two parameters required for the Ecopath model. There is no catch. There is no consumption and no diet because they are autotrophs.

## Biomass

Phytoplankton biomass is measured as chlorophyll "a" biomass. On the Grand Bank three sources of information on chlorophyll "a" biomass were used: the Mobil study noted above; oceanographic research cruises carried out by the Bedford Institute of Oceanography (BIO) (Irwin et al. 1986, 1988; C. Caverhill unpubl.); and remote sensing, CZCS data (Feldman et al. 1995). On the Labrador Shelf area, two sources of information were used: the BIO research cruises (Irwin et al. 1989, 1990; C. Caverhill unpubl.) and the CZCS data.

The Mobil data came from a monthly series of research cruises where samples were taken from March 1980 to February 1981 across the Grand Bank. Other than the March 1980 sample of 12, all other monthly samples ranged in size from 40-55 observations. The BIO research cruises on the Grand Bank were carried out in April 1984 ( $\mathrm{n}=4$ ), May 1988 ( $\mathrm{n}=1$ ), May 1994 ( $n=2$ ), May 1996 ( $n=2$ ), May 1997 ( $n=3$ ), July 1995 ( $n=3$ ) and September 1985 $(\mathrm{n}=9)$. These were integrated over 50 m in order to compare them with the Mobil data and up to 100 m . The CZCS estimates of chlorophyll "a" are estimated from 8 years of data (19781986). The CZCS chlorophyll measure is taken as surface chlorophyll. This is then integrated over depth by using a stored chloropyll depth profile, which is based on monthly empirical ship-based studies for the region (for further explanation of the method, see Platt et al. 1991, Longhurst et al. 1995 and Sathyendranath et al. 1995).

There are large differences among the monthly values of these data sets (Table 14). This highlights both the variability of phytoplankton abundance and the differences in the methods used. The monthly CZCS chloropyll "a" data do not correspond at all with either the Mobil or BIO ship-based data. In the CZCS data there is no spring bloom but an early winter bloom. This would indicate that the CZCS data are not representative of the actual conditions, since a spring bloom is known to occur. For this reason the CZCS data are not used to estimate
chlorophyll biomass for the Grand Bank.
The BIO ship-based chloropyll estimates are higher than the Mobil estimates in any given month. However, since these are representative of the period (including only the data from 1984-1988 would increase the May estimate), and since they are of a similar magnitude to the Labrador Shelf data below, these data were used to estimate biomass. These data cover only the months of April, May, August and September. In order to estimate a mean annual chlorophyll "a" biomass, it was assumed that the only significant difference between the 19841988 period represented by the BIO ship-based data and the 1980 period represented by the Mobil data was in the magnitude of the spring bloom. Thus the BIO ship-based values were used for April, May, August and September, and the Mobil values were used for the other months. Chlorophyll "a" biomass was then converted to phytoplankton biomass using a monthly Chl:C ratio (Hollibaugh and Booth 1981). This produced a mean average biomass of $24.6 \mathrm{~g} \cdot \mathrm{~m}^{-2}$, ie., $24.6 \mathrm{t} \cdot \mathrm{km}^{-2}$.

Table 14. Comparison of chlorophyll "a" biomass estimates from different sources for Grand Bank.

| G. Bank | Mobil Study | BIO Ship-based Data |  | CZCS |
| :--- | :---: | :---: | :---: | :---: |
| MONTH | $\mathrm{mg} \mathrm{CHL} \cdot \mathrm{m}^{\wedge} 2<=50 \mathrm{~m}$ | $\mathrm{mg} \mathrm{CHL} \cdot \mathrm{m}^{\wedge} 2<=50 \mathrm{~m}$ | $\mathrm{mg} \mathrm{CHL} \cdot \mathrm{m}^{\wedge} 2$ | $\mathrm{mg} \mathrm{CHL} \cdot \mathrm{m}^{\wedge} 2$ |
|  |  |  | $<=100 \mathrm{~m}$ | $<10 \mathrm{~m}$ |
| March | 23.80 |  | 440.54 | 440.54 |
| April | 9.09 | 195.10 | 331.90 | 49.1 |
| May | 130.13 |  |  | 63.9 |
| June | 27.52 | 13.80 | 18.70 | 47.8 |
| July | 17.60 |  |  | 51.9 |
| August | 16.65 | 31.10 | 31.40 | 63.9 |
| September | 11.36 |  |  | 60.9 |
| October | 11.36 |  |  | 87.0 |
| November | 30.50 |  |  | 137.1 |
| December | 30.50 |  |  | 88.1 |
| January | 11.79 |  |  | 109.7 |
| February | 11.79 |  |  | 62.7 |

The CZCS data for the Labrador shelf show both a spring bloom and a fall bloom (Table 15) and thus is indicative of expected seasonal trends in phytoplankton biomass (Longhurst 1995). There are BIO ship-based data for May 1988 ( $\mathrm{n}=7$ ), May 1994 ( $\mathrm{n}=2$ ), May 1996 ( $\mathrm{n}=4$ ), May 1997 (n=3), June 1984 (n=9), June 1997 (n=3), July 1984 (n=3), July 1985 (n=4), July 1995 $(\mathrm{n}=5)$ and November $1995(\mathrm{n}=3)$. These data confirm both the trend and the magnitude shown in CZCS data. Both data sets were used to estimate the mean annual biomass. For the BIO ship-based data, the same procedure was followed as described for the Grand Bank. For the CZCS data, mean annual chlorophyll biomass was estimated as the mean of the monthly biomass. As above, chlorophyll "a" biomass was converted to phytoplankton biomass using a monthly Chl:C ratio (Hollibaugh and Booth 1981). The BIO ship-based data produced a mean annual phytoplankton biomass estimate of $28.4 \mathrm{t} \cdot \mathrm{km}^{-2}$. The CZCS estimate was $29.8 \mathrm{t} \cdot \mathrm{km}^{-2}$.

Table 15. Comparison of chlorophyll "a" biomass estimates from different sources for the Labrador Shelf.

| Labrador Shelf | BIO Ship-based Data <br> $\mathrm{mg} \mathrm{CHL} \cdot \mathrm{m}^{-2}<=100 \mathrm{~m}$ | CZCS <br> MON CHL $\cdot \mathrm{m}^{-2}<=100 \mathrm{~m}$ |
| :--- | :---: | :---: |
| March |  | 50.5 |
| April |  | 51.3 |
| May | 313.0 | 105.2 |
| June | 172.3 | 100.5 |
| July | 130.2 | 88.0 |
| August |  | 64.5 |
| September |  | 60.1 |
| October |  | 105.6 |
| November | 78.6 | 71.5 |
| December |  | 26.9 |
| January |  | 33.6 |
| February |  | 24.2 |

The mean biomass for the entire area was calculated as a mean of the two Labrador estimates and a weighted mean of this estimate and the estimate for the Grand Bank. This produced a total mean biomass estimate of $26.395 \mathrm{t} \cdot \mathrm{km}^{-2}$. The mean annual biomass estimate from the Mobil study for the Grand Bank was $11.593 \mathrm{t} \cdot \mathrm{km}^{-2}$. This is considerably lower than the estimate made here. Given the noted variability in phytoplankton biomass and distribution, it would be wise to use these two figures to represent a range of likely values for mean phytoplankton biomass in 2J3KLNO.

## Primary Production

Primary productivity is estimated from the same data sources as described above. The CZCS data were not used for the Grand Bank, for the reasons given. The Mobil study estimated a primary productivity of $155.6 \mathrm{gC} \cdot \mathrm{m}^{-2} \cdot \mathrm{yr}^{-1}$ for 1980-1981. Prasad and Haedrich (1993), using satellite data, estimated a primary production of $194 \mathrm{gC} \cdot \mathrm{m}^{-2} \cdot \mathrm{yr}^{-1}$ for the same time period for the Grand Bank. There was only one ship-based estimate, in April 1984, of primary production. This was extrapolated to a yearly value on the basis of the Mobil data (that is, each month was raised by the difference between the corresponding month in the Mobil data and the Mobil April estimate). This produced an annual primary productivity of $482.4 \mathrm{gC} \cdot \mathrm{m}^{-2} \cdot \mathrm{yr}^{-1}$. However, this sample came from only one area, east of the Southeast Shoal region of Grand Bank, on the shelf break. Since productivity and biomass levels are known to be high on the shelf break (Hollibaugh and Booth 1981), this value is unlikely to be representative of mean productivity on Grand Bank.

For the Labrador shelf, the CZCS-based estimate of primary productivity was $429 \mathrm{gC} \cdot \mathrm{m}^{-2} \cdot \mathrm{yr}^{-1}$ (Caverhill, unpubl.). The estimate is derived from the estimate of surface chlorophyll "a" biomass on the Labrador Shelf and the chlorophyll depth profiles (see above). A model is
used to calculate surface light, then another model to compute how much light there is at each depth increment in the water column. Productivity by the chlorophyll at each depth and light level is estimated using stored ship-based photosynthesis parameters that are compiled by domain and season. Productivity at depth is then integrated over depth to the euphotic depth to give total productivity (see Platt et al. 1991; Longhurst et al. 1995; Sathyendranath et al. 1995).

Primary productivity data from BIO ship-based information were available for May 1988. This was raised to an annual figure using the Mobil data and the method described above for Grand Bank. The estimated annual productivity was $307.6 \mathrm{gC} \cdot \mathrm{m}^{-2} \cdot \mathrm{yr}^{-1}$.

The productivity estimates for Grand Bank range from 155.6 to $194 \mathrm{gC} \cdot \mathrm{m}^{-2} \cdot \mathrm{yr}^{-1}$, and for the Labrador shelf from 307.6 to $426 \mathrm{gC} \cdot \mathrm{m}^{-2} \cdot \mathrm{yr}^{-1}$. The Mobil study and the Prasad and Haedrich (1993) estimates are lower than the ship-based or CZCS estimates, but are based (in the case of the Mobil data) on monthly samples across the Grand Bank over one year. The BIO shipbased samples are based on data for one month only. The CZCS data for the Labrador Shelf include data for all months over several years. The high BIO ship-based estimate for Grand Bank from the shelf break should not be used, for the reasons given above. However, all the other estimates are valid estimates of productivity. The range of estimates is given in Table 16 below.

Table 16. Primary productivity and $\mathrm{P} / \mathrm{B}$ estimates for Grand Bank and the Labrador Shelf.

|  | Grand Bank |  | Labrador Shelf |  |
| :--- | ---: | ---: | ---: | ---: |
|  | Low | High | Low | High |
| $\mathrm{PP}\left(\mathrm{gC} \cdot \mathrm{cm}^{-2}\right)$ | 155.6 | 194.0 | 307.6 | 429.0 |
| $\mathrm{PP}\left(\mathrm{t} \cdot \mathrm{km}^{-2}\right)$ | 1556.3 | 1940.0 | 3076.0 | 4290.0 |
| Biomass $\left(\mathrm{t} \cdot \mathrm{km}^{-2}\right)$ | 24.6 | 24.6 | 29.1 | 29.1 |
| P/B $\left(\mathrm{yr}^{-1}\right)$ | 63.3 | 78.9 | 105.7 | 147.3 |

The $\mathrm{P} / \mathrm{B}$ ratios are estimated using the biomasses derived above. The range is wide, from 63 to $147 \mathrm{yr}^{-1}$, and the mean weighted $\mathrm{P} / \mathrm{B}$ is 93.1 . The Mobil biomass estimate for the Grand Bank ( $11.6 \mathrm{yr}^{-1}$ ) gives a higher $\mathrm{P} / \mathrm{B}$ of 134.2 to $167.2 \mathrm{yr}^{-1}$. There is considerable uncertainty concerning the $\mathrm{P} / \mathrm{B}$ ratio, and an initial value of $93.1 \mathrm{yr}^{-1}$ is used, although it may be higher than this.

## 31. Detritus

The detritus biomass was estimated using an empirical relationship derived by Pauly et al. (1993). It relates detritus biomass to primary productivity and euphotic depth.

$$
\begin{equation*}
\log _{10} \mathrm{D}=-2.41+0.954 \log _{10} \mathrm{PP}+0.863 \log _{10} \mathrm{E} \tag{4}
\end{equation*}
$$

where,
$\mathrm{D}=$ detritus standing stock $\left(\mathrm{gC} \cdot \mathrm{m}^{-2}\right.$ (grams of carbon per square metre) $), \mathrm{PP}=$ primary productivity $\left(\mathrm{gCm}^{-2}\right.$ year ${ }^{-1}$ ), $\mathrm{E}=$ euphotic depth (m).

The fit of the regression equation to the data is not very good, but as suggested by Pauly et al. (1993:13), it "might be considered sufficient in cases where no other information is available".

Primary production was described above. The annual range is $155.6 \mathrm{gC} \cdot \mathrm{m}^{-2}$ to $483.4 \mathrm{gC} \cdot \mathrm{m}^{-2}$.
The euphotic depth is calculated from the Beer-Bouger Law where,

$$
\begin{equation*}
\ln \mathrm{I}(1)-\ln \mathrm{I}(2)=\mathrm{k}(\mathrm{D}(2)-\mathrm{D}(1)) \tag{5}
\end{equation*}
$$

and,
$\mathrm{I}(1)=100 \%$ irradiance (at the surface), $\mathrm{I}(2)=1 \%$ irradiance (at the euphotic depth), $\mathrm{D}(1)=$ depth at surface $(0 \mathrm{~m}), \mathrm{D}(2)=$ euphotic depth, $\mathrm{k}=$ light attenuation co-efficient.

Seasonal values of the light attenuation co-efficient for 1980 on Grand Bank were taken from Hollibaugh and Booth (1981, Table 4_2). The average value was substituted into equation (5) and the mean annual euphotic depth on the Grand Bank estimated as 54.7 m .

This euphotic depth and the two extremes of the PP estimates were substituted into equation (4) giving a range of detritus biomass estimates from $14.9 \mathrm{gC} \cdot \mathrm{m}^{-2}$ to $38.9 \mathrm{gC} \cdot \mathrm{m}^{-2}$, or $149 \mathrm{t} \cdot \mathrm{km}^{-2}$ to $389 \mathrm{t} \cdot \mathrm{km}^{-2}$, using a conversion factor of 10 g wet weight $=1 \mathrm{gC}$ (Christensen and Pauly 1992:20). This range is wide and applies only to Grand Bank. Given the approximate nature of the estimation method, the range should be regarded as a gross approximation. For the purposes of the Ecopath model, a midprange of $26.9 \mathrm{t} \cdot \mathrm{km}^{-2}$ is used as the detritus biomass estimate.

## Results

## The Unbalanced Model

Table 17 shows the main model results and Table 18 gives the breakdown of mortality for each group. The model is not balanced; there are negative flows to the detritus and some ecotrophic efficiencies (EE) are greater than 1. EE is the proportion of production that is passed up the food web, that is, consumed or harvested. If EE is greater than 1, then the consumption of group $i$ is greater than its production. Two components can contribute to high EE's: high consumption rate and/or biomass of the predator, and low production rate and/or biomass of the prey. The EE of the following groups was greater than $1: \operatorname{cod} \leq 35 \mathrm{~cm}$, Greenland halibut $\leq 40 \mathrm{~cm}$, flounders, redfish, small demersal feeders, sand lance, piscivorous small pelagic feeders, planktivorous small pelagic feeders, shrimp, large crustacea, large zooplankton and small zooplankton. The total biomass of harvested fish species is 27.39 $\mathrm{t} \cdot \mathrm{km}^{-2}$.

The total mortality for each group is equivalent to the $\mathrm{P} / \mathrm{B}$ ratio. The fishing mortality is calculated as the catch divided by the biomass. Predation mortality, M2, is calculated as the consumption of each group (total predator catch) divided by biomass. The "other mortality", which is attributed to disease, starvation and old age, is estimated by the model (and is
equivalent to $1-\mathrm{EE}$ ) and balances the mortalities so that $\mathrm{Z}=\mathrm{F}+\mathrm{M} 2+\mathrm{MO}$. In Table 18 , many of the MOs are negative. This is another indication that the model is not balanced, and that M2 must be reduced.

There are several key predators in this system. The mortality accounted for by the top vertebrate predators is shown in Figure 5 for all prey (a) and for vertebrate prey (b). Cod $>35$ cm , harp seals and skates account for almost $50 \%$ of total mortality imposed by vertebrate predators on all prey. Harp seals, cod $>35 \mathrm{~cm}$ and whales account for over $50 \%$ of total mortality imposed by vertebrate predators on vertebrate prey. From the fish point of view then, these are the main predators. Skates, hooded seals and Greenland halibut $>40 \mathrm{~cm}$ are also important predators. The impact of these predators on the groups whose EE is greater than 1 is shown in Table 19.

Harp seals, hooded seals, cod $>35 \mathrm{~cm}$, whales and skates account for over $80 \%$ of the total predation mortality on cod $\leq 35 \mathrm{~cm}$, Greenland halibut $<40 \mathrm{~cm}$, flounders and redfish. For the small demersal feeders, American plaice $\leq 35 \mathrm{~cm}$ and large demersal feeders are important predators in addition to the predators noted above. Capelin, cod $>35 \mathrm{~cm}$ and American plaice are the most important predators on sand lance. Whales impose a significant predation mortality on the piscivorous small pelagic feeders, while harp seals are also an important predator. The greatest predation on shrimp is by $\operatorname{cod}>35 \mathrm{~cm}$, followed by large demersal feeders and harp seals. Predation on large crustaceans is largely by skates, cod $>35 \mathrm{~cm}$ and cannibalism. The principal predators of large zooplankton and small zooplankton are capelin and large zooplankton.

Ecopath is a top-down model, so model estimates are determined by the biomass and consumption rates of the top predators. Either these are too high for the predators highlighted in Table 19 or the $\mathrm{P} / \mathrm{B}$ rates of the prey groups are too low. Both possibilities may be contributing to the imbalances.

Table 17. Results of the first (unbalanced) Ecopath model of Newfoundland-Labrador, NAFO divisions 2J3KLNO, 1985-1987.

| Group Name | Biomass $\left(t \cdot k m^{-2}\right)$ | $\begin{gathered} P / B \\ \left(y r^{-1}\right) \end{gathered}$ | $\begin{aligned} & \text { Q/B } \\ & \left(y r^{-1}\right) \end{aligned}$ | EE | GE | Harvest ( $t \cdot \mathrm{~km}^{-2}$ ) | Flow to Detritus ( $t \cdot \mathrm{~km}^{-2}$ ) | Trophic Level |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1. Whales | 0.25 | 0.10 | 11.79 | 0 | 0.008 | 0 | 0.62 | 4.24 |
| 2. Harp Seals | 0.18 | 0.10 | 17.41 | 0.05 | 0.006 | 0.001 | 0.66 | 4.36 |
| 3. Hooded Seals | 0.03 | 0.11 | 13.10 | 0 | 0.008 | 0 | 0.09 | 4.68 |
| 4. Seabirds | 0.01 | 0.25 | 54.75 | 0.33 | 0.005 | 0.001 | 0.13 | 4.20 |
| 5. $\mathrm{Cod}>35 \mathrm{~cm}$ | 2.04 | 0.65 | 3.24 | 0.52 | 0.201 | 0.603 | 1.97 | 4.16 |
| 6. $\mathrm{Cod}<=35 \mathrm{~cm}$ | 0.09 | 0.60 | 6.09 | 7.74 | 0.099 | 0 | -0.27 | 3.87 |
| 7. G. halibut $>40 \mathrm{~cm}$ | 0.35 | 0.51 | 1.48 | 0.36 | 0.344 | 0.035 | 0.22 | 4.53 |
| 8. G. halibut $<=40 \mathrm{~cm}$ | 0.16 | 0.25 | 3.40 | 9.64 | 0.073 | 0.002 | -0.24 | 4.25 |
| 9. Aplaice $>35 \mathrm{~cm}$ | 0.97 | 0.54 | 1.26 | 0.19 | 0.426 | 0.100 | 0.67 | 3.66 |
| 10. Aplaice $<=35 \mathrm{~cm}$ | 0.78 | 0.63 | 3.74 | 0.53 | 0.167 | 0.025 | 0.82 | 3.70 |
| 11. Flounders | 0.87 | 0.39 | 3.60 | 1.15 | 0.109 | 0.079 | 0.57 | 3.09 |
| 12. Skates | 0.52 | 0.29 | 2.88 | 0.26 | 0.099 | 0.037 | 0.41 | 4.15 |
| 13. Redfish | 0.98 | 0.33 | 2.00 | 1.79 | 0.165 | 0.178 | 0.14 | 3.66 |
| 14. L.Dem.Feeders | 0.85 | 0.26 | 1.75 | 0.68 | 0.150 | 0.052 | 0.37 | 3.44 |
| 15. S.Dem.Feeders | 0.23 | 0.30 | 2.00 | 15.14 | 0.150 | 0 | -0.87 | 3.11 |
| 16. Capelin | 13.29 | 1.15 | 6.39 | 0.74 | 0.179 | 0.126 | 20.96 | 3.27 |
| 17. Sand lance | 2.10 | 1.15 | 7.67 | 1.20 | 0.150 | 0 | 2.75 | 3.20 |
| 18. Arctic cod | 2.73 | 0.40 | 2.63 | 0.91 | 0.150 | 0 | 1.53 | 3.41 |
| 19. L.Pel.Feeders | 0.03 | 0.40 | 3.33 | 0.95 | 0.120 | 0.006 | 0.02 | 4.24 |
| 20. Pisc. SPF | 0.45 | 0.27 | 1.77 | 6.81 | 0.150 | 0.015 | -0.54 | 4.14 |
| 21. Plankt. SPF | 0.95 | 0.24 | 1.60 | 4.38 | 0.150 | 0.019 | -0.47 | 3.30 |
| 22. Shrimp | 0.20 | 1.45 | 9.67 | 2.76 | 0.150 | 0.004 | -0.12 | 2.46 |
| 23. Large Crustacea | 0.19 | 0.28 | 5.85 | 20.87 | 0.048 | 0.016 | -0.82 | 3.13 |
| 24. Echinoderms | 112.30 | 0.60 | 6.67 | 0.02 | 0.090 | 0 | 215.67 | 2.00 |
| 25. Molluscs | 42.10 | 0.57 | 6.33 | 0.04 | 0.090 | 0 | 76.48 | 2.00 |
| 26. Polychaetes | 10.50 | 2.00 | 22.22 | 0.12 | 0.090 | 0 | 65.23 | 2.00 |
| 27. O.Benthic Inver | 7.80 | 2.50 | 12.50 | 0.12 | 0.200 | 0 | 36.76 | 2.00 |
| 28. L.Zooplankton | 18.34 | 3.43 | 19.50 | 1.10 | 0.176 | 0 | 65.58 | 2.56 |
| 29. S.Zooplankton | 21.70 | 6.20 | 20.67 | 1.73 | 0.300 | 0 | -8.45 | 2.00 |
| 30. Phytoplankton | 26.86 | 93.10 | 0 | 0.23 - |  | 0 | 1919.68 | 1.00 |
| 31. Detritus | 389.00 | - |  | 0.58 - |  | 0 - |  | 1.00 |

Table 18. Distribution of different types of mortality on the Ecopath groups $\left(\mathrm{yr}^{-1}\right) . \mathrm{Z}=$ total mortality, $\mathrm{F}=$ fishing mortality, $\mathrm{MO}=$ other mortality and $\mathrm{M} 2=$ predation mortality.

| Group Name | $Z$ | $F$ | $M 0$ | $M 2$ |
| :--- | ---: | ---: | ---: | ---: |
| 1. Whales | 0.10 | 0 | 0.10 | 0 |
| 2. Harp Seals | 0.10 | 0.01 | 0.10 | 0 |
| 3. Hooded Seals | 0.11 | 0 | 0.11 | 0 |
| 4. Seabirds | 0.25 | 0.08 | 0.17 | 0 |
| 5. Cod $>35 \mathrm{~cm}$ | 0.65 | 0.30 | 0.32 | 0.04 |
| 6. Cod $<=35 \mathrm{~cm}$ | 0.60 | 0 | -4.05 | 4.65 |
| 7. G.halibut $>40 \mathrm{~cm}$ | 0.51 | 0.10 | 0.32 | 0.08 |
| 8. G.halibut <=40 cm | 0.25 | 0.01 | -2.13 | 2.37 |
| 9. Aplaice $>35 \mathrm{~cm}$ | 0.54 | 0.10 | 0.44 | 0 |
| 10. Aplaice <=35 cm | 0.63 | 0.03 | 0.29 | 0.30 |
| 11. Flounders | 0.39 | 0.09 | -0.06 | 0.36 |
| 12. Skates | 0.29 | 0.07 | 0.21 | 0 |
| 13. Redfish | 0.33 | 0.18 | -0.26 | 0.41 |
| 14. L.Dem.Feeders | 0.26 | 0.06 | 0.08 | 0.12 |
| 15. S.Dem.Feeders | 0.30 | 0 | -4.24 | 4.54 |
| 16. Capelin | 1.15 | 0.01 | 0.30 | 0.84 |
| 17. Sand lance | 1.15 | 0 | -0.23 | 1.38 |
| 18. Arctic cod | 0.40 | 0 | 0.03 | 0.36 |
| 19. L.Pel.Feeders | 0.40 | 0.21 | 0.02 | 0.17 |
| 20. Pisc. SPF | 0.27 | 0.03 | -1.54 | 1.77 |
| 21. Plankt. SPF | 0.24 | 0.02 | -0.81 | 1.03 |
| 22. Shrimp | 1.45 | 0.02 | -2.55 | 3.98 |
| 23. Large Crustacea | 0.28 | 0.09 | -5.60 | 5.80 |
| 24. Echinoderms | 0.60 | 0 | 0.59 | 0.01 |
| 25. Molluscs | 0.57 | 0 | 0.55 | 0.02 |
| 26. Polychaetes | 2.00 | 0 | 1.77 | 0.23 |
| 27. O.Benthic Inver | 2.50 | 0 | 2.21 | 0.29 |
| 28. L.Zooplankton | 3.43 | 0 | -0.32 | 3.76 |
| 29. S.Zooplankton | 6.20 | 0 | -4.52 | 10.72 |
| 30. Phytoplankton | 93.10 | 0 | 71.47 | 21.63 |

Table 19. Predator mortality on individual prey groups $\left(\mathrm{yr}^{-1}\right)$. Predators in bold account for $80 \%+$ of the mortality on each prey.

| PREDATORS | PREY |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} 6 \\ \operatorname{Cod}<= \\ 35 \mathrm{~cm} \end{gathered}$ | $\begin{aligned} & 8 \\ & \text { G.halibut } \\ & <=40 \mathrm{~cm} \end{aligned}$ | 11 Flounders | 13 <br> Redfish | $\begin{gathered} 15 \\ \text { S.Dem. } \\ \text { Feeders } \end{gathered}$ | 17 <br> Sand <br> lance | $\begin{gathered} 20 \\ P i s c \\ S P F \end{gathered}$ | 21 Plankt. SPF | $\underset{p}{22}$ | $\begin{aligned} & 23 \\ & \text { L.C. } \end{aligned}$ | $\begin{gathered} 28 \\ L Z P \end{gathered}$ | $\begin{gathered} 29 \\ S Z P \end{gathered}$ |
| 1. Whales | 0.38 |  |  |  | 0.43 | 0.07 | 1.07 | 0.09 |  |  | 0.02 | 0.01 |
| 2. Harp Seals | 1.16 | 0.87 | 0.18 | 0.02 | 0.66 | 0.14 | 0.36 | 0.13 | 0.57 |  | 0.01 |  |
| 3. Hooded Seals |  | 0.78 | 0.11 | 0.02 | 0.24 |  | 0.03 | 0.07 |  |  |  |  |
| 4. Seabirds | 0.03 |  |  |  | 0.05 | 0.02 | 0.02 | 0.02 | 0.02 |  | 0 | 0 |
| 5. $\mathrm{Cod}>35 \mathrm{~cm}$ | 1.27 | 0.24 | 0.01 | 0.06 | 1.17 | 0.33 | 0.01 | 0.03 | 1.21 | 1.72 | 0.02 |  |
| 6. $\mathrm{Cod}<=35 \mathrm{~cm}$ | 0.04 | 0.01 |  |  | 0.06 | 0.01 | 0.01 | 0.01 | 0.22 | 0.07 | 0.01 |  |
| 7. G.halibut $>40 \mathrm{~cm}$ | 0.42 | 0.40 | 0 | 0.14 | 0.19 |  | 0.01 | 0.01 | 0.03 |  | 0 |  |
| 8. G. halibut $<=40 \mathrm{~cm}$ | 0.05 | 0.02 |  |  | 0.02 |  | 0.01 | 0.02 | 0.06 |  | 0 |  |
| 9. Aplaice $>35 \mathrm{~cm}$ |  | 0.01 |  |  | 0.05 | 0.10 |  |  | 0.02 | 0.30 | 0 |  |
| 10. Aplaice <=35 cm | 0.12 | 0.04 | 0.03 | 0 | 0.30 | 0.13 |  |  | 0.19 | 0.33 | 0.01 |  |
| 11. Flounders |  |  |  |  | 0.10 | 0.04 |  |  | 0.11 |  | 0 |  |
| 12. Skates | 0.79 | 0.01 | 0.01 | 0.13 | 0.73 | 0.10 | 0.19 | 0.01 | 0.16 | 2.50 | 0 | 0 |
| 13. Redfish | 0.04 |  |  | 0.01 | 0.01 | 0 |  | 0.50 | 0.34 |  | 0.06 | 0.01 |
| 14. L.Dem.Feeders | 0.17 |  | 0.02 | 0.02 | 0.27 | 0.02 | 0 | 0.05 | 0.66 | 0.21 | 0.01 | 0 |
| 15. S.Dem.Feeders |  |  |  |  | 0.02 | 0 | 0 | 0 | 0.04 | 0.02 | 0 | 0 |
| 16. Capelin |  |  |  |  |  | 0.40 |  |  |  |  | 2.01 | 2.14 |
| 17. Sand lance |  |  |  |  |  |  |  |  |  |  | 0.31 | 0.48 |
| 18. Arctic cod |  |  |  |  |  |  |  |  |  |  | 0.25 | 0.11 |
| 19. L.Pel.Feeders | 0 |  |  | 0 | 0.01 | 0 | 0.06 | 0.02 | 0.01 |  | 0 | 0 |
| 20. Pisc. SPF | 0.16 |  |  |  |  | 0 |  | 0.07 | 0.03 |  | 0.01 | 0 |
| 21. Plankt. SPF |  |  |  |  |  |  |  |  | 0.04 |  | 0.04 | 0.03 |
| 22. Shrimp |  |  |  |  |  |  |  |  |  |  | 0.01 | 0.02 |
| 23. Large Crustacea |  |  |  |  | 0.24 |  |  |  | 0.26 | 0.65 | 0 | 0 |
| 28. L.Zooplankton |  |  |  |  |  |  |  |  |  |  | 0.98 | 7.91 |
| TOTAL | 4.63 | 2.38 | 0.36 | 0.40 | 4.55 | 1.36 | 1.77 | 1.03 | 3.97 | 5.8 | 3.75 | 10.71 |

The quality of the data used for each group in the model is variable. For some, there are empirical estimates available from samples taken from within the model area and time frame (e.g., $\operatorname{cod}>35 \mathrm{~cm}$, harp seals). For others, it was necessary to use empirical data from other areas and/or time frames (e.g., flounders, skate) or less specific information (e.g., whales, redfish, large pelagic feeders).

The groups with the most reliable data are the predator groups, specifically harp seals, cod $>$ 35 cm and Greenland halibut $>40 \mathrm{~cm}$. Of the prey groups, $\operatorname{cod}<35 \mathrm{~cm}$ are probably the best studied, but our knowledge about the biomass, production and consumption of any of these groups is poor. For example, for the young cod and Greenland halibut, there is uncertainty concerning numbers-at-age, actual mortality with age or size and mean weight at age, all of which are fundamental to the calculations made to estimate biomass.

This means that greater confidence is placed on the consumption demands by the predators than on the biomass of prey estimated to meet these demands. That is, the balance of the Ecopath model should be constrained to meet the consumption requirements of the groups for which we have the most certainty about the parameter estimates. It is a robust finding that many large ecosystem energetic models do not balance the consumption of the top predators with the production of prey (e.g., Mills and Fournier 1979; Jones 1984; Welch et al. 1992).

## Other problems with the initial model

In addition to the basic imbalance of the model, there were some inconsistencies in the gross efficiencies, GE. GE is the ratio of production to consumption and for most groups should have values between 0.1 and 0.3 (Christensen and Pauly 1992). The marine mammals and the seabirds are exceptions. Once these animals reach maturity their growth efficiency is very low (Gaskin 1982). This is seen in Table 17.

Cod $\leq 35 \mathrm{~cm}$, Greenland halibut $\leq 40 \mathrm{~cm}$, and large crustaceans also have low gross efficiencies. This could be due to either low production or high consumption. In order to increase the large crustacean gross efficiency of 0.048 , the production estimate was increased to the maximum of the range ( $0.382 \mathrm{yr}^{-1}$ ) and the consumption estimate was reduced to the minimum of the range ( $4.42 \mathrm{yr}^{-1}$ ). This produced a GE estimate of 0.08 , which is similar to estimates made by Jarre-Teichmann and Guenette (1996). For cod $\leq 35 \mathrm{~cm}$ and Greenland halibut $\leq 40 \mathrm{~cm}$, the GEs were altered in the process of reducing EE to less than zero, see below.

The gross efficiencies of Greenland halibut $\leq 40 \mathrm{~cm}$ and American plaice $>35 \mathrm{~cm}$ were too high. In the case of the Greenland halibut, the $\mathrm{P} / \mathrm{B}$ ratio was uncertain, with a described range of 0.296 to $0.719 \mathrm{yr}^{-1}$. A mid-point was used above. However, since this results in a GE that is too large, this indicates that the lower end of the range is more likely. For this reason the $\mathrm{P} / \mathrm{B}$ of $0.296 \mathrm{yr}^{-1}$ (estimated by assuming $\mathrm{P}=\mathrm{M}^{*} \mathrm{~B}+\mathrm{C}$ ) was used, giving a GE of 0.200.

For American plaice, the $\mathrm{P} / \mathrm{B}$ estimate was derived from catch curve analysis and considered better determined than the $\mathrm{Q} / \mathrm{B}$ estimate. The $\mathrm{Q} / \mathrm{B}$ estimate was estimated from Zammaro's (1992) work on the Tail of the Grand Bank. The value of $1.262 \mathrm{yr}^{-1}$ was raised to $2.0 \mathrm{yr}^{-1}$, an increase of almost $60 \%$, giving a GE of $0.269 \mathrm{yr}^{-1}$.

## Addressing imbalances in the model

The first step to balance the model is to reduce the EE's to less than 1 for all the groups with $\mathrm{EE}>1$. It was noted above that the data for these groups is not as certain as the data for the top predators (cod $>35 \mathrm{~cm}$ and harp seal). Thus, the rationale used is to alter the input parameters (biomass or production) of the prey groups to produce EE's of less than 1.

## $\mathbf{C o d} \leq \mathbf{3 5} \mathrm{cm}$

The biomass of small cod was estimated by back-calculating from numbers at age 3 . The mean number at age 3 (on Jan 1) was $224,394 * 10^{3}$, and the natural mortality was assumed as 0.6 for all age groups. The biomass was raised here by altering the assumptions made about natural mortality and numbers at age 3 . Natural mortality was raised to 0.65 on age $2,1.2$ on age 1 and 3.0 on age 0 . In addition, the number at age 3 was raised by a factor of 2 . It was assumed that the sequential population analysis (Shelton et al. 1996b) underestimates the number of these fish since it used a natural mortality of 0.2 for all ages.

The results of the back-calculation are shown in Table 20. A biomass of 168,112 t is equivalent to $0.340 \mathrm{t} \cdot \mathrm{km}^{-2}$. Since mortality at age was increased, the mean M, and therefore $\mathrm{P} / \mathrm{B}$, was also increased to $1.597 \mathrm{yr}^{-1}$. When this biomass and $\mathrm{P} / \mathrm{B}$ are re-entered in the model, EE is 0.923 and the GE is 0.201 .

Table 20. Back-calculation of $\operatorname{cod} \leq 35 \mathrm{~cm}$ biomass.

| AGE | M yr |  |  |  |  |
| :---: | ---: | ---: | ---: | ---: | ---: |
| 3 |  | Nos (OOO's) <br> (mid year) | Mn wt (kg) | Biomass (t) | W'ted mean M |
|  | 0.65 | 605,789 |  |  |  |
| 1 | 1.20 | $2,011,831$ | 0.100 | 60,580 | 0.216 |
| 0 | 3.00 | $19,082,980$ | 0.025 | 50,283 | 0.359 |
| SUM | $21,700,112$ |  | 168,112 | 1.022 |  |

${ }^{1}$ Numbers on Jan 1st, raised by 2

## Greenland halibut $\leq 40 \mathrm{~cm}$

The biomass of small Greenland halibut was estimated by back-calculating from numbers-atage 6. The mean numbers at age 6 (from the fall RV Survey) was $105,135^{*} 10^{3}$, and the natural mortality was assumed to be 0.2 on ages 3-6 and 0.6 on ages $0-2$. As above, the biomass was raised by altering the assumptions made about natural mortality and numbers-atage. Natural mortality was raised to 0.4 on ages $3-6$ and to 1.6 on ages 0 to 2 . In addition, the number at age 6 was raised by a factor of 1.4. For Greenland halibut, the numbers at age were derived from an RV survey. It is assumed that the catchability of the survey was not 1 , and that it may have been around 0.7 .

The results of the back-calculation are shown in Table 21. A biomass of $247,165 \mathrm{t}$ is equivalent to $0.499 \mathrm{t} \cdot \mathrm{km}^{-2}$. Since mortality-at-age was increased, the mean $M$, and therefore $\mathrm{P} / \mathrm{B}$, was also increased to $0.872 \mathrm{yr}^{-1}$. $\mathrm{Q} / \mathrm{B}$ also changed, since it is estimated using mean weight-at-age, to $4.476 \mathrm{yr}^{-1}$. When the new biomass, $\mathrm{P} / \mathrm{B}$, and $\mathrm{Q} / \mathrm{B}$ are re-entered in the
model, EE is 0.931 and the GE is 0.195 .
Table 21. Back-calculation of Greenland halibut $\leq 40 \mathrm{~cm}$ biomass.

| AGE | M Nos (000's) |  | Mn wt (kg) | Biomass (t) | W'ted Z |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 6 | 0.4 | 116,192 |  |  |  |
| 5 | 0.4 | 173,339 | 0.386 | 66,909 | 0.108 |
| 4 | 0.4 | 258,591 | 0.191 | 49,505 | 0.080 |
| 3 | 0.4 | 385,772 | 0.087 | 33,494 | 0.054 |
| 2 | 1.6 | 1,910,741 | 0.028 | 54,432 | 0.352 |
| 1 | 1.6 | 9,463,964 | 0.004 | 40,118 | 0.260 |
| 0 | 1.6 | 31,421465 | 8.618E-05 | 2,708 | 0.018 |
| SUM |  | 43,730,064 |  | 247,165 | 0.872 |

## Flounders

The yellowtail flounder biomass was estimated as the Campelen converted biomass for 7+ fish and the backcalculated biomass at age for ages 0 to 6 . The ratio between young and old fish was then used to estimate biomass of witch flounder ages 0 to 6 . Biomass was re-estimated for the younger age groups, using an M of 0.9 for ages 0 to 2, and assuming a catchability of 0.7 for age 6 fish in the RV survey. This produced a total flounder biomass of $551,362 \mathrm{t}$, $1.114 \mathrm{t} \cdot \mathrm{km}^{-2}$. The $\mathrm{P} / \mathrm{B}$ ratio is decreased slightly to $0.372 \mathrm{yr}^{-1}\left(\right.$ from $\left.0.394 \mathrm{yr}^{-1}\right)$, and the EE is 0.949 .

Table 22. Re-estimated flounder biomass

|  | Yellowtail Flounder | Witch | Winter | Flounders |
| :--- | ---: | ---: | ---: | ---: |
| Campelen Biomass 7+ | 105,490 | 53,975 |  | 159,465 |
| Backcalculated Biomass (0-6) | 242,949 | 124,308 |  | 367,257 |
| B from catch assumption |  | 24,640 | 24,640 |  |
| Total | 348,439 | 178,283 | 24,640 | 551,362 |

## Redfish

The EE of the redfish was too high. However, the main predator of redfish is Greenland halibut $<40 \mathrm{~cm}$, and its diet is considered more reliable than the parameters derived for the redfish, which are poorly known. EE was set equal to 0.95 and the biomass was estimated by the Ecopath model as $1.876 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Small demersal feeders

This is a mixed group of small demersal feeding fish, including the juveniles of the large demersal feeders. It is poorly defined and consumption is estimated by assuming a GE of 0.15 . EE was very high. In order to reduce EE, production was first increased to the maximum of the range, from 0.3 to 0.4 year $^{-1}$ (this has the effect of increasing GE to 2). There was no rational for increasing the biomass, so the EE was made equal to 0.95 , and the biomass estimated by the model. The required biomass was $5.06 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Sand lance

The biomass of this group was a gross approximation. EE was set to 0.95 and biomass estimated from the model. The required biomass was $2.692 \mathrm{t} \cdot \mathrm{km}^{-2}$.

## Piscivorous Small Pelagic Feeders

The $\mathrm{P} / \mathrm{B}$ of this group was raised to $0.6 \mathrm{yr}^{-1}$ from $0.265 \mathrm{yr}^{-1}$ since the large pelagic feeders had a P/B of 0.4 and smaller fish would be expected to have a higher $\mathrm{P} / \mathrm{B}$ ratio. The ecotrophic efficiency was set to 0.95 . The model estimated a biomass of $1.509 \mathrm{t} \cdot \mathrm{km}^{-2}$ and a GE of 0.34 . The higher estimate for GE agrees with the gross conversion efficiency estimate for squid in O'Dor and Dawe (1998).

## Planktivorous Small Pelagic Feeders

The P/B of this group was raised to $0.5 \mathrm{yr}^{-1}$ from $0.265 \mathrm{yr}^{-1}$ since the large pelagic feeders had a P/B of 0.4 and smaller fish would be expected to have a higher $\mathrm{P} / \mathrm{B}$ ratio. The ecotrophic efficiency was set to 0.95 . The model estimated a biomass of $3.568 \mathrm{t} \cdot \mathrm{km}^{-2}$ and a $\mathrm{Q} / \mathrm{B}$ of 3.33 $\mathrm{yr}^{-1}$.

## Shrimp

The EE of shrimp was set equal to 0.95 and a biomass of $1.459 \mathrm{t} \cdot \mathrm{km}^{-2}$ estimated by the model.

## Large Crustaceans

The large crustaceans are a mixed group comprising crabs and lobsters. Using the method employed above, that is, allowing biomass to be estimated by the model, resulted in an extremely large estimate due to cannibalism within the group. Large crustaceans made up 11 $\%$ of the diet. Only a small part of this is likely to be real cannibalism. Most will be one species of crab eating another species of crab. There are two ways to resolve this problem. One is to create two large crustacean groups, one predator and one prey. However, there is insufficient information to do this. The other short cut approach is to reduce cannibalism. Cannibalism was reduced to $1 \%$ of the diet and the difference distributed over the echinoderms, molluscs, polychaetes and other benthic invertebrates by their relative weights.

When the ecotrophic efficiency was set to 0.95 , the model estimated a biomass of $3.548 \mathrm{t} \cdot \mathrm{km}^{-2}$. This is still more than an order of magnitude greater than the original estimate.

## Large Zooplankton

EE was set equal to 0.95 and a biomass of $22.46 \mathrm{t} \cdot \mathrm{km}^{-2}$ estimated by the model.

## Small Zooplankton

Production of the small zooplankton was increased from 6.2 to $8.4 \mathrm{yr}^{-1}$, the maximum of the range. Since GE was set equal to 0.15 , this resulted in an increase in consumption biomass ratio from 20.67 to 28 year ${ }^{-1}$. EE was set equal to 0.95 and a biomass of $33.67 \mathrm{t} \cdot \mathrm{km}^{-2}$ estimated by the model.

## Changes to other groups

## Arctic Cod

Arctic cod was not among the original set of groups that were out of balance. However, predation on this group increased due to the increased biomass of $\operatorname{cod} \leq 35 \mathrm{~cm}$ and Greenland halibut $\leq 40 \mathrm{~cm}$. In order to redress this imbalance, the biomass was increased by $10 \%$. In addition, the $\mathrm{P} / \mathrm{B}$ was re-estimated as the weighted average mortality of the 1 to 2 years olds and the 2 to 3 years olds, assuming the ratio was $3: 1$, to better reflect the preponderance of 1 and 2 year old Arctic cod in 2J3KLNO (Lilly et al. 1994). This produced a biomass of $1,485,785 \mathrm{t}, 3.0 \mathrm{t} \cdot \mathrm{km}^{-2}$, a P/B of $0.402 \mathrm{yr}^{-1}$ and an EE of 0.967 .

## Capelin

Predation on capelin also increased as a result of the increased biomass estimates of cod $\leq 35$ cm , Greenland halibut $\leq 40 \mathrm{~cm}$ and piscivorous small pelagic feeders. EE was set to 0.95 and the model estimated a biomass of $13.61 \mathrm{t} \cdot \mathrm{km}^{-2}$. In addition, capelin consumption was reduced from a $\mathrm{Q} / \mathrm{B}$ of 6.4 to 4.3 year $^{-1}$. The original estimate was based on the fish eating $3 \%$ body weight per day. However, this produced a Q/B larger than the small pelagics and this was considered to be unlikely. The effect of reducing $\mathrm{Q} / \mathrm{B}$ is to reduce predation on their zooplankton prey.

## Model 1 - The first balanced model

The input parameters and selected results of the balanced model are shown in Table 23. The balanced model, under the assumptions described above, produced a total finfish biomass of $38.8 \mathrm{t} \cdot \mathrm{km}^{-2}, 41 \%$ higher than the original estimate. Since biomass estimates of the lower trophic levels were raised to meet the consumption requirements of the top predators, this result was predictable. It indicates that there may be a greater fish biomass in 2J3KLNO than has been previously estimated. This result is typical of large scale ecosystem energetics models.

The changes made to the consumption and production estimates were discussed above and all were within the original range of estimates described for the model. However, many of the biomasses in Table 23 were estimated by the model, that is, they are the biomasses required to balance the model. These biomass estimates are discussed below to determine whether they are feasible. The percentage differences in biomass estimated by the unbalanced and balanced model are shown in Figure 6.

Table 23. Input parameters and selected results of the balanced Newfoundland-Labrador model, NAFO divisions 2J3KLNO, 1985-1987.

| Group Name | Biomass ( $t \cdot \mathrm{~km}^{-2}$ ) | $\begin{gathered} P / B \\ \left(y r^{-1}\right) \end{gathered}$ | $\begin{aligned} & Q / B \\ & \left(y r^{-1}\right) \end{aligned}$ | EE | GE | Harvest $\left(t \cdot k m^{-2}\right)$ | Flow to Detritus $\left(t \cdot \mathrm{~km}^{-2}\right)$ | Trophic Level |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1. Whales | 0.25 | 0.10 | 11.79 | 0 | 0.008 | 0 | 0.62 | 4.24 |
| 2. Harp Seals | 0.18 | 0.10 | 17.41 | 0.05 | 0.006 | 0.001 | 0.66 | 4.36 |
| 3. Hooded Seals | 0.03 | 0.11 | 13.10 | 0 | 0.008 | 0 | 0.09 | 4.68 |
| 4. Seabirds | 0.01 | 0.25 | 54.75 | 0.33 | 0.005 | 0.001 | 0.13 | 4.20 |
| 5. Cod $>35 \mathrm{~cm}$ | 2.04 | 0.65 | 3.24 | 0.52 | 0.201 | 0.603 | 1.97 | 4.16 |
| 6. $\mathrm{Cod}<=35 \mathrm{~cm}$ | 0.34 | 1.60 | 6.09 | 0.92 | 0.262 | 0 | 0.46 | 3.87 |
| 7. G. halibut $>40 \mathrm{~cm}$ | 0.35 | 0.30 | 1.48 | 0.62 | 0.200 | 0.035 | 0.14 | 4.53 |
| 8. G. halibut $<=40 \mathrm{~cm}$ | 0.50 | 0.87 | 4.48 | 0.93 | 0.195 | 0.002 | 0.48 | 4.25 |
| 9. Aplaice $>35 \mathrm{~cm}$ | 0.97 | 0.54 | 2.00 | 0.19 | 0.269 | 0.100 | 0.71 | 3.65 |
| 10. Aplaice $<=35 \mathrm{~cm}$ | 0.78 | 0.63 | 3.74 | 0.53 | 0.167 | 0.025 | 0.82 | 3.70 |
| 11. Flounders | 1.11 | 0.37 | 3.60 | 0.95 | 0.103 | 0.079 | 0.82 | 3.09 |
| 12. Skates | 0.52 | 0.29 | 2.88 | 0.26 | 0.099 | 0.037 | 0.41 | 4.11 |
| 13. Redfish | 1.88 | 0.33 | 2.00 | 0.95 | 0.165 | 0.178 | 0.78 | 3.66 |
| 14. L.Dem.Feeders | 0.85 | 0.26 | 1.75 | 0.68 | 0.150 | 0.052 | 0.37 | 3.44 |
| 15. S.Dem.Feeders | 5.06 | 0.40 | 2.00 | 0.95 | 0.200 | - | 2.13 | 3.11 |
| 16. Capelin | 13.61 | 1.15 | 4.30 | 0.95 | 0.266 | 0.126 | 12.49 | 3.27 |
| 17. Sand lance | 2.69 | 1.15 | 7.67 | 0.95 | 0.150 | 0 | 4.28 | 3.20 |
| 18. Arctic cod | 3.00 | 0.40 | 2.63 | 0.97 | 0.153 | 0 | 1.62 | 3.41 |
| 19. L.Pel.Feeders | 0.03 | 0.40 | 3.33 | 0.95 | 0.120 | 0.006 | 0.02 | 4.24 |
| 20. Pisc. SPF | 1.51 | 0.6 | 1.77 | 0.95 | 0.150 | 0.015 | 0.58 | 4.14 |
| 21. Plankt. SPF | 3.57 | 0.5 | 3.33 | 0.95 | 0.340 | 0.019 | 2.47 | 3.30 |
| 22. Shrimp | 1.46 | 1.45 | 9.67 | 0.95 | 0.150 | 0.004 | 2.93 | 2.46 |
| 23. Large Crustacea | 3.55 | 0.38 | 4.42 | 0.95 | 0.086 | 0.016 | 3.20 | 3.02 |
| 24. Echinoderms | 112.30 | 0.60 | 6.67 | 0.10 | 0.090 | 0 | 210.60 | 2.00 |
| 25. Molluscs | 42.10 | 0.57 | 6.33 | 0.20 | 0.090 | 0 | 72.53 | 2.00 |
| 26. Polychaetes | 10.50 | 2.00 | 22.22 | 0.38 | 0.090 | 0 | 59.80 | 2.00 |
| 27. O.Benthic Inver | 7.80 | 2.50 | 12.50 | 0.42 | 0.200 | 0 | 30.76 | 2.00 |
| 28. L.Zooplankton | 22.46 | 3.43 | 19.50 | 0.95 | 0.176 | 0 | 91.45 | 2.56 |
| 29. S.Zooplankton | 33.67 | 8.40 | 20.67 | 0.95 | 0.406 | 0 | 153.31 | 2.00 |
| 30. Phytoplankton | 26.86 | 93.10 | 0 | 0.34 - |  | 0 | 1641.55 | 1.00 |
| 31. Detritus | 389.00 - |  |  | 0.61 - |  | 0 |  | 1.00 |

## Biomass

## Cod $\leq \mathbf{3 5} \mathbf{c m}$, Greenland halibut $\leq \mathbf{4 0} \mathbf{c m}$, Flounders

The biomasses of these groups were re-estimated using the back-calculation method. Tables 20 and 21 detailed the assumptions necessary to generate the required 200-300 \% biomass increases in $\operatorname{cod} \leq 35 \mathrm{~cm}$ and Greenland halibut $\leq 40 \mathrm{~cm}$ (Figure 6). The biomass generated by the back-calculation method is sensitive to natural mortality at age when it is greater than 1. Since such high mortalities are necessary on the youngest age groups of fish, these biomasses should be regarded as rough approximations only.

Predation on cod $\leq 35 \mathrm{~cm}$ is shown in Table 24. Cannibalism by larger cod and predation by harp seals are the largest sources of mortality, accounting for $46 \%$ of the total morality. Skates and piscivorous small pelagic feeders are also important predators. The pattern of predation is similar to the unbalanced model (Table 19), with the exception that piscivorous small pelagic feeders have gained importance as predators. Their biomass has increased by an order of magnitude to accommodate the consumption requirements of whales and harp seals. It is possible that this biomass is over-estimated (see below) and this would affect the biomass estimated for $\operatorname{cod} \leq 35 \mathrm{~cm}$. However, since the main predators, $\operatorname{cod}>35 \mathrm{~cm}$ and harp seals, are relatively well parameterised in this model, it is reasonable to assume that the biomass estimated for $\operatorname{cod} \leq 35 \mathrm{~cm}$ is of the correct magnitude.

Table 24. Composition of predation mortality on $\operatorname{cod} \leq 35 \mathrm{~cm}$.

| Predator | Mortality percentage <br> $\left(y r^{-1}\right)$ |  |
| :--- | ---: | ---: |
| Cod $>35 \mathrm{~cm}$ | 0.35 | 24.0 |
| Harp Seals | 0.32 | 21.9 |
| Skates | 0.22 | 15.1 |
| Pisc. SPF | 0.15 | 10.3 |
| G.halibut $>40 \mathrm{~cm}$ | 0.11 | 7.5 |
| Whales | 0.10 | 6.8 |
| Others | 0.20 | 14.4 |
| Total | 1.46 | 100.0 |

The main predators of Greenland halibut $\leq 40 \mathrm{~cm}$ are hooded seals and harp seals (Table 25). Cannibalism and predation by cod $>35 \mathrm{~cm}$ are also important sources of mortality. This pattern is the same as that seen in the unbalanced model. Again, since these predators are relatively well parameterised is it reasonable to assume that the estimated biomass for Greenland halibut $\leq 40 \mathrm{~cm}$ is approximately correct.

Harp seals and hooded seals are also the main predators of the flounder group (Table 26). Fishing mortality also contributes to this group, making the total mortality caused by these 3 predators $85.7 \%$. Again, this pattern of mortality is similar to that in the unbalanced model (Table 19). All three predators are well parameterised and thus the assumption is that the estimated biomass is approximately correct.

Table 25. Composition of predation mortality on Greenland halibut $\leq 40 \mathrm{~cm}$.

| Predator | Mortality percentage <br> $\left(y r^{\prime}\right)$ |  |
| :--- | ---: | ---: |
| Harp Seals | 0.29 | 36.3 |
| Hooded Seals | 0.26 | 32.5 |
| G.halibut $>40 \mathrm{~cm}$ | 0.13 | 16.3 |
| Cod $>35 \mathrm{~cm}$ | 0.08 | 10.0 |
| Others | 0.04 | 5.0 |
| Total | 0.80 | 100.0 |

Table 26. Composition of predation mortality on flounders.

| Predator | Mortality <br> $\left(y r^{-1}\right)$ | percentage |
| :--- | ---: | ---: |
| Harp Seals | 0.14 | 40.0 |
| Hooded Seals | 0.09 | 25.7 |
| Fishing | 0.07 | 20.0 |
| Others | 0.05 | 14.3 |
| Total | 0.35 | 100.0 |

## Redfish

The redfish biomass estimated by the model was almost twice the original estimate. The original biomass estimate was based on a range of surveys, both Canadian and Russian. Divisions 2 J 3 K and 3 L were reasonably surveyed, but 3 NO was not. The original estimate may be too low for the following reasons: the conversion between Campelen, Engels and Russian survey data may be inaccurate and/or the biomass estimate does not allow for catchability, i.e., it is the trawlable biomass. Edwards (1968) estimated a catchability of 0.27 for redfish (using Yankee 36). If a catchability of 0.5 is assumed here, assuming that the Campelen gear is twice as efficient as the Yankee 36, then the redfish biomass required by the model is obtained.

The main predators of redfish are shown in Table 27. Over $60 \%$ of this mortality is imposed by the fishery, Greenland halibut $>40 \mathrm{~cm}$, and $\operatorname{cod}>35 \mathrm{~cm}$, groups for which there is reasonably good parameter definition. Thus the biomass estimated to meet consumption demands is likely to be a reasonable approximation.

Table 27. Composition of predation mortality on redfish.

| Predator | Mortality <br> $\left(y r^{-1}\right)$ | percentage |
| :--- | ---: | ---: |
| Fishing | 0.09 | 30.0 |
| G.halibut $>40 \mathrm{~cm}$ | 0.07 | 23.3 |
| Skates | 0.07 | 23.3 |
| Cod $>35 \mathrm{~cm}$ | 0.03 | 10 |
| Others | 0.04 | 13.3 |
| Total | 0.30 | 100.0 |

## Small demersal feeders

The original biomass of this group was estimated as the sum of the trawlable biomass for all species in the group. This was then raised to a Campelen equivalent, using the biomass ratio between Engels 1993/94 and Campelen 1996/1997 (see above). However, the biomass required to balance the Ecopath model $(2,504,700 \mathrm{t})$ is twenty-two times the original estimate. This large biomass is difficult to recreate with existing data. Catchability of these species is low, but it would have to be as low as 0.045 for the Campelen gear or 0.004 for the Engels gear in order to raise the trawlable biomass to the value of the required biomass estimated by the Ecopath model.

The large biomass of small demersal feeders is driven by the consumption demands of the large crustaceans, which account for over $41 \%$ of the total mortality (Table 28). This high mortality is a result of the large biomass estimated for the large crustaceans. This estimate is uncertain, see below. If the large crustacean diet is altered so that small demersal feeders are no longer eaten, the biomass estimate is reduced to $1,424,610 \mathrm{t}$, a biomass 1 million tonnes lower than the estimate above. Small demersal feeders constitute only $5 \%$ of the diet of large crustaceans, and thus the sensitivity of the small demersal feeders to the biomass of the large crustacea is very high. This is discussed further below. Of the total of 13 groups that prey on the small demersal feeders, there is also high uncertainty over the biomass (and thus consumption requirements) of skates and whales.

Table 28. Composition of predation mortality on small demersal feeders.

| Predator | Mortality <br> $\left(y r^{-1}\right)$ | percentage |
| :--- | ---: | ---: |
| Large Crustacea | 0.15 | 41.7 |
| Cod $>35 \mathrm{~cm}$ | 0.05 | 13.9 |
| Harp Seals | 0.03 | 8.3 |
| Skates | 0.03 | 8.3 |
| Whales | 0.02 | 5.6 |
| S.Dem.Feeders | 0.02 | 5.6 |
| Others | 0.06 | 16.7 |
| Total | 0.36 | 100.0 |

## Capelin

The biomass of capelin required by the Ecopath model was $6,738,930 \mathrm{t}$ while the original estimate was $6,578,068 \mathrm{t}$. This represents a minor increase, which is feasible, given the range of estimates described earlier.

Capelin is described as a forage fish, and Table 29 amply demonstrates this point. There are 15 groups which prey on capelin, plus fishing, which is not an important source of mortality. The main predators are cod $>35 \mathrm{~cm}$, piscivorous small pelagic feeders, Greenland halibut $\leq$ 40 cm and whales.

Table 29. Composition of predation mortality on capelin.

| Predator | Mortality <br> $\left(y r^{-1}\right)$ | percentage |
| :--- | ---: | ---: |
| Cod $>35 \mathrm{~cm}$ | 0.29 | 26.9 |
| Pisc. SPF | 0.14 | 13.0 |
| G.halibut $<=40 \mathrm{~cm}$ | 0.14 | 13.0 |
| Whales | 0.11 | 10.2 |
| Harp Seals | 0.08 | 7.4 |
| Cod $<=35 \mathrm{~cm}$ | 0.07 | 6.5 |
| Aplaice $<=35 \mathrm{~cm}$ | 0.07 | 6.5 |
| Aplaice $>35 \mathrm{~cm}$ | 0.04 | 3.7 |
| Capelin | 0.04 | 3.7 |
| Seabirds | 0.03 | 2.8 |
| Arctic cod | 0.02 | 1.9 |
| Fishing | 0.01 | 0.9 |
| G.halibut $>40 \mathrm{~cm}$ | 0.01 | 0.9 |
| Flounders | 0.01 | 0.9 |
| Skates | 0.01 | 0.9 |
| S.Dem.Feeders | 0.01 | 0.9 |
| Total | 1.08 | 100.0 |

## Sand lance

The sand lance biomass estimate was based on sand lance trawlable biomass in the 1990s, which was then raised by a catchability of 0.1 , based on Edwards (1968). The biomass required to balance the model $(1,332,540 \mathrm{t})$ is 1.3 times that biomass. Given the uncertainty in the original estimate, this is feasible simply by assuming that the catchability was 0.08 rather than 0.1.

A total of 14 groups prey on sand lance. $\operatorname{Cod}>35 \mathrm{~cm}$ and capelin are the main predators, while both size groups of American plaice play a role, in addition to harp seals, skates and whales (Table 30). The biomass estimated for sand lance is likely to be a good approximation.

Table 30. Composition of predation mortality on sand lance.

| Predator | Mortality <br> $\left(y r^{-1}\right)$ | percentage |
| :--- | ---: | ---: |
| Cod $>35 \mathrm{~cm}$ | 0.26 | 23.9 |
| Capelin | 0.22 | 20.2 |
| Aplaice $>35 \mathrm{~cm}$ | 0.12 | 11.0 |
| Harp Seals | 0.11 | 10.1 |
| Aplaice $<=35 \mathrm{~cm}$ | 0.10 | 9.2 |
| Skates | 0.08 | 7.3 |
| Whales | 0.06 | 5.5 |
| Others | 0.14 | 12.8 |
| Total | 1.09 | 100.0 |

## Arctic Cod

To estimate the biomass of Arctic cod, it was assumed that the catchability of Arctic cod was the same as the catchability of capelin. The biomass estimated in the balanced model is $10 \%$ greater than the original estimate. Given the approximate nature of the biomass estimation, this is an acceptable margin of error.

Harp seals are the main predators of Arctic cod, accounting for over $60 \%$ of the total mortality (Table 31). There is no fishing mortality on this group.

Table 31. Composition of predation mortality on Arctic cod.

| Predator | Mortality <br> $\left(y r^{-1}\right)$ | percentage |
| :--- | ---: | ---: |
| Harp Seals | 0.23 | 60.5 |
| Cod $>35 \mathrm{~cm}$ | 0.05 | 13.2 |
| G.halibut $<=40 \mathrm{~cm}$ | 0.04 | 10.5 |
| Cod $<=35 \mathrm{~cm}$ | 0.02 | 5.3 |
| S.Dem.Feeders | 0.02 | 5.3 |
| Seabirds | 0.01 | 2.6 |
| Arctic cod | 0.01 | 2.6 |
| Total | 0.38 | 100.0 |

## Piscivorous small pelagic feeders

The biomass of this group increased by over $200 \%$, to $746,955 \mathrm{t}$, in the balanced model. This substantial increase is largely driven by the predatory requirements of whales, which account for over $50 \%$ of the predation mortality (Table 32). (Much of this consumption by whales is on squid.) Harp seals and skates are also important predators. As described above, the biomass of short-finned squid was thought to be low during the years of the study, so this high biomass estimate is suspect. Fishing accounts for a mortality of $0.01 \mathrm{yr}^{-1}$.

Table 32. Composition of predation mortality on piscivorous small pelagic feeders.

| Predator | Mortality <br> $\left(\right.$ yr $\left.^{-1}\right)$ | percentage |
| :--- | ---: | ---: |
| Whales | 0.32 | 56.1 |
| Harp Seals | 0.11 | 19.3 |
| Skates | 0.06 | 10.5 |
| L.Pel.Feeders | 0.02 | 3.5 |
| Others | 0.06 | 10.5 |
| Total | 0.57 | 100.0 |

## Planktivorous small pelagic feeders

The biomass of the planktivorous small pelagic feeders increased by $275 \%$. The greatest predators on this group are redfish and piscivorous small pelagic feeders (Table 33). The biomass of both these groups increased in the balanced model. The planktivorous small pelagic feeders include myctophids and other mesopelagics and Gonatus squid. The biomass of these groups is essentially unknown. The initial estimate above was a "guesstimate", and
the current estimate may be valid.
Table 33. Composition of predation mortality on planktivorous small pelagic feeders.

| Predator | Mortality <br> $\left(y r^{-1}\right)$ | percentage |
| :--- | ---: | ---: |
| Redfish | 0.26 | 55.3 |
| Pisc. SPF | 0.06 | 12.8 |
| Harp Seals | 0.03 | 6.4 |
| Whales | 0.02 | 4.3 |
| Hooded Seals | 0.02 | 4.3 |
| G.halibut <=40 cm | 0.02 | 4.3 |
| Others | 0.06 | 12.8 |
| Total | 0.47 | 100.0 |

## Shrimp

The estimated shrimp biomass of $722,205 \mathrm{t}$ for 2 J 3 KLNO is more than $600 \%$ greater than the original estimate, which applied to 2 J 3 K only. It is also greater than the 1996-1997 biomass estimate of $400,000 \mathrm{t}$ for 2 J 3 K (Parsons and Veitch 1998). The main predators of this group are large crustaceans, cod $>35 \mathrm{~cm}$, small demersal feeders and cod $\leq 35 \mathrm{~cm}$, Table 34 .

Table 34. Composition of predation mortality on shrimp.

| Predator | Mortality <br> $\left(y r^{-1}\right)$ | percentage |
| :--- | ---: | ---: |
| Large Crustacea | 0.53 | 39.0 |
| Cod > 35cm | 0.17 | 12.5 |
| S.Dem.Feeders | 0.14 | 10.3 |
| Cod <= 35 cm | 0.11 | 8.1 |
| Redfish | 0.09 | 6.6 |
| L.Dem.Feeders | 0.09 | 6.6 |
| Harp Seals | 0.08 | 5.9 |
| Plankt. SPF | 0.04 | 2.9 |
| Others | 0.11 | 8.1 |
| Total | 1.36 | 100.0 |

## Large Crustacea

The large crustacean biomass underwent an enormous increase of over $1800 \%$ to $1,756,260 \mathrm{t}$ in the balancing of the model. This increase is largely driven by skates and cod $>35 \mathrm{~cm}$ (Table 35). Cannibalism was reduced to $1 \%$ of the diet, but it still accounts for over $11 \%$ of the total mortality.

The large crustacea are a key group at the lower trophic levels in this system. They directly affect the biomass estimates of the small demersal feeders and shrimp. The biomass estimates for these 2 groups are large, because the initial large crustacea biomass estimate is large. Large crustacea feed on shrimp and small demersal feeders, small demersal feeders feed on shrimp and large crustaceans, and shrimp feed mainly on zooplankton and detritus. A trophic
triangle exists among these groups and the estimate of large crustacea biomass impacts on the other 2 groups.

Table 35. Composition of predation mortality on large crustacea.

| Predator | Mortality <br> $\left(y r^{-1}\right)$ | percentage |
| :--- | ---: | ---: |
| Skates | 0.13 | 37.1 |
| Cod $>35 \mathrm{~cm}$ | 0.09 | 25.7 |
| Large Crustacea | 0.04 | 11.4 |
| S.Dem.Feeders | 0.03 | 8.6 |
| Aplaice $>35 \mathrm{~cm}$ | 0.02 | 5.7 |
| Aplaice $<=35 \mathrm{~cm}$ | 0.02 | 5.7 |
| Cod $<=35 \mathrm{~cm}$ | 0.01 | 2.9 |
| L.Dem.Feeders | 0.01 | 2.9 |
| Total | 0.35 | 100.0 |

## Large Zooplankton and Small Zooplankton

The biomass of the large zooplankton is increased by $22 \%$ from the original estimate and the biomass of the large zooplankton is increased by $55 \%$. Both of these estimates are on the high side. Capelin and cannibalism are the main sources of mortality for large zooplankton, accounting for $65 \%$ of the mortality. Almost all ( $78 \%$ ) of the predation mortality on the small zooplankton, is by the large zooplankton.

## Model 2: Further Improvements to the Model

The balanced model described above is not completely consistent with what is known or surmised about the ecosystem. There are two areas which are of particular concern; the impact of skates and an apparent trophic triangle of large crustacea, small demersal feeders and shrimp.

## The impact of skates

The mortality analysis placed skates as the fifth most important natural predator of vertebrates. They have a high impact on cod $\leq 35 \mathrm{~cm}$, redfish, small demersal feeders, sand lance, piscivorous small pelagic feeders and large crustaceans. They have not previously been considered as major predators in the system, and the results here may be in error. In order to examine this possibility, the input parameters and the sensitivity of the model output were reexamined.

The biomass estimate was derived from the Engels trawlable biomass estimate, raised by a factor of 2.2, to a Campelen equivalent. However, this raising factor may be too high. The mean size of skates has declined since the 1980s (Kulka et al. 1996) and the Campelen gear catches more small fish than the Engels gear. Thus, in the 1990s, when the Engels:Campelen ratio was estimated (see above), the catch consisted of a greater proportion of small fish, which the Campelen gear caught efficiently, while the Engels gear did not. This would artificially raise the ratio between the two gears, and would not represent the situation in the
mid-1980s. So, instead of raising the skate biomass to a Campelen equivalent, it is assumed that the Engels biomass estimate is a better indicator of skate biomass in the mid-1980s. In order to allow for some error, a $10 \%$ error term was used. Thus, the biomass range was 104,553-127,787 t. The high end of the range was used below. This reduction in the biomass of the skates produced biomass reductions in the groups noted above, ranging from $5 \%$ for sand lance, $10 \%$ for $\operatorname{cod} \leq 23 \mathrm{~cm}$, to $20 \%$ for the large crustacea.

The skate $\mathrm{P} / \mathrm{B}$ ratio and $\mathrm{Q} / \mathrm{B}$ ratio were left unchanged since there was no rationale for altering them. Sensitivity to a change in the $\mathrm{Q} / \mathrm{B}$ ratio was the same as sensitivity to a change in skate biomass.

The skate diet was adapted from the aggregate diet for thorny skate (see Appendix 2, Table 8 and Templeman 1982). Piscivory in this diet was reduced by $20 \%$, and the difference attributed to the invertebrate groups. However, predation on the large crustaceans in particular was very high. In order to reduce this, the $20 \%$ originally due to piscivory and distributed to invertebrate groups, was redistributed to the polychaetes and amphipods (large zooplankton) only. Both groups are prevalent in the diet of smaller skates (Templeman 1982).

## The large crustacea - small demersal feeders - shrimp trophic triangle

The model estimates of biomass for the large crustacea, small demersal feeders and shrimp were very high. These groups form a trophic triangle: large crustacea and small demersal feeders prey on each other, where predation by large crustacea is much stronger than predation by small demersal feeders; shrimp are prey to both species. This mutual predation drives the biomass of both groups. The consequent high consumption of shrimp drives the biomass of shrimp upwards.

The diet of the large crustacea was re-examined. In the original diet (Table 2b), small demersal feeders composed $5 \%$ of the crab diet. It was noted that the diets from which this was derived (Wieczorek and Hooper 1995; Miller and O'Keefe 1981) were expressed in percentage frequency of occurrence, whereas here percentage weight is used. The diet was reestimated using a different rationale. The fish that were present in the diet have been reclassified as detritus, on the assumption that these fish were scavenged, see Appendix 2, Table 18. In addition, the $\mathrm{P} / \mathrm{B}$ was raised from 0.38 to $0.45 \mathrm{yr}^{-1}$.

In addition, the consumption to biomass ratio of $\operatorname{cod}>35 \mathrm{~cm}$ was reduced from the middle of the input range to the lower end of the input range, that is, from 3.24 to $2.2 \mathrm{yr}^{-1}$.

The resultant changes in biomass for large crustacea, small demersal feeders and shrimp are shown in Table 36. The biomass estimates are reduced by more than $50 \%$ for large crustacea and small demersal feeders.

The difference between the biomass estimates from Models 1 and 2 and the unbalanced model are shown in Figure 7. In addition to the large reductions in biomass estimates for the large crustacea, small demersal feeders and shrimp, the estimated biomass of several other species were also reduced, making Model 2 a better fit to the original parameters. These species include $\operatorname{cod} \leq 35 \mathrm{~cm}$, Greenland halibut $\leq 40 \mathrm{~cm}$, flounders, redfish, sand lance, Arctic cod,
piscivorous small pelagic feeders, planktivorous small pelagic feeders, large zooplankton and small zooplankton.

Table 36. Comparison of biomass before and after changes to the large crustacea and skate groups.

|  | Before changes <br> $t \cdot \mathrm{~km}^{-2}$ | After changes <br> $t \cdot \mathrm{~km}^{-2}$ |
| :--- | :---: | :---: |
| Large crustacea | 3.548 | 1.732 |
| Small demersal feeders | 5.060 | 2.380 |
| Shrimp | 1.459 | 0.816 |

Differences also resulted in the predator spectrum of several of the prey groups, Table 37 and Tables 24 to 35 . The predatory impact of $\operatorname{cod}>35 \mathrm{~cm}$, skates, large crustacea and small demersal feeders was reduced on all prey groups. For the small demersal feeders, large crustacea and shrimp, predation by skates and large crustacea was reduced while the role of other predators such as harp seals became more important. In addition, for $\operatorname{cod} \leq 35 \mathrm{~cm}$, harp seals became the main predator. This is due to the reduction in the consumption of cod $>35$ cm . Large demersal feeders have a predatory impact on skates, whereas when the skate biomass was larger, this predation was negligible.

For many groups, the total mortality remained the same in Models 1 and 2. However, there were a few exceptions. Predation mortality on the large crustacea increased overall. This is due to their decreased biomass and increased predation by all predators except small demersal feeders and skates. The predation mortality on capelin and the other invertebrate groups (echinoderms, molluscs, polychaetes and other benthic invertebrates) decreased. Predation mortality on capelin decreased due to decreased predation by $\operatorname{cod}>35 \mathrm{~cm}, \operatorname{cod} \leq 35 \mathrm{~cm}$, Greenland halibut $>40 \mathrm{~cm}$ and skates. The predation mortality on the other invertebrate groups decreased due to decreased predation by small demersal feeders and large crustacea

## Mortality

The distribution of predation mortality in Model 2 is shown in Figure 8. The top chart (Figure 8a) gives the distribution of predation mortality amongst the vertebrate predators on all the prey groups. Compared to the unbalanced model, the same key predators are present here, that is, cod $>35 \mathrm{~cm}$, harp seals, skates and capelin, but they appear in a different order. Capelin now accounts for the greatest amount of predation mortality, followed by harp seals and cod > 35 cm . Other groups such as sand lance, large crustacea, and small demersal feeders have become more important predators, whilst the level of predation imposed by the skates has decreased.

Compared to Figure 5a, a greater number of groups contribute to the predation mortality. For example, in Figure 5a, the top $50 \%$ is comprised of only 3 groups, whereas in Figure 8a it is composed of 5 groups (excluding fishing). The top $75 \%$ is composed of 6 groups in Figure 5a and 12 groups in Figure 8a. Thus the balanced model is less dominated by a small number of predators.

The distribution of predators on vertebrate prey or fish prey is shown in Figure 8b. Harp seals are now the top predators in the system, followed by fishing. After fishing, cod $>35 \mathrm{~cm}$, whales, hooded seals and Greenland halibut $>40 \mathrm{~cm}$ are the greatest predators of fish.

Table 37. Predator mortality on individual prey groups $\left(\mathrm{yr}^{-1}\right)$ : Model 2.

| PREDATORS | PREY |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{gathered} 6 \\ \operatorname{Cod} \\ <=35 \mathrm{~cm} \\ \hline \end{gathered}$ | $\begin{gathered} 8 \\ \text { G.halibut } \\ <=40 \mathrm{~cm} \end{gathered}$ | 11 <br> Flounders | 13 <br> Redfish | $\begin{gathered} 15 \\ \text { S.Dem. } \\ \text { Feeders } \end{gathered}$ | 17 <br> Sand <br> lance | $\begin{gathered} 20 \\ \text { Pisc. } \\ \text { SPF } \end{gathered}$ | 21 Plankt. SPF | $\begin{gathered} 22 \\ \text { Shrimp } \end{gathered}$ | $\begin{aligned} & 23 \\ & L C \end{aligned}$ | $\begin{gathered} 28 \\ L Z P \end{gathered}$ | $\stackrel{29}{S Z P}$ |
| 1. Whales | 0.13 |  |  |  | 0.04 | 0.07 | 0.34 | 0.03 |  |  | 0.02 | 0.01 |
| 2. Harp Seals | 0.40 | 0.31 | 0.15 | 0.01 | 0.06 | 0.13 | 0.12 | 0.04 | 0.14 |  | 0 |  |
| 3. Hooded Seals |  | 0.27 | 0.09 | 0.01 | 0.02 |  | 0.01 | 0.02 |  |  |  |  |
| 4. Seabirds | 0.01 |  |  |  | 0 | 0.01 | 0.01 | 0.01 | 0 |  | 0 | 0 |
| 5. Cod $>35 \mathrm{~cm}$ | 0.29 | 0.06 | 0 | 0.03 | 0.08 | 0.20 | 0 | 0.01 | 0.2 | 0.12 | 0.01 |  |
| 6. $\mathrm{Cod}<=35 \mathrm{~cm}$ | 0.04 | 0.01 |  |  | 0.02 | 0.02 | 0.01 | 0.01 | 0.16 | 0.02 | 0.02 |  |
| 7. G.halibut $>40 \mathrm{~cm}$ | 0.14 | 0.14 | 0 | 0.08 | 0.02 |  | 0 | 0 | 0.01 |  | 0 |  |
| 8. G.halibut $<=40 \mathrm{~cm}$ | 0.07 | 0.02 |  |  | 0.01 |  | 0.01 | 0.02 | 0.06 |  | 0 |  |
| 9. Aplaice $>35 \mathrm{~cm}$ |  | 0 |  |  | 0.01 | 0.14 |  |  | 0.01 | 0.05 | 0 |  |
| 10. Aplaice $<=35 \mathrm{~cm}$ | 0.04 | 0.01 | 0.02 | 0 | 0.03 | 0.12 |  |  | 0.05 | 0.04 | 0.01 |  |
| 11. Flounders |  |  |  |  | 0.01 | 0.04 |  |  | 0.03 |  | 0.01 |  |
| 12. Skates | 0.13 | 0 | 0 | 0.04 | 0.03 | 0.05 | 0.03 | 0 | 0.01 | 0.09 | 0 | 0 |
| 13. Redfish | 0.02 |  |  | 0.01 | 0 | 0.01 |  | 0.25 | 0.14 |  | 0.08 | 0.02 |
| 14. L.Dem.Feeders | 0.06 |  | 0.01 | 0.01 | 0.03 | 0.02 | 0 | 0.02 | 0.16 | 0.02 | 0.01 | 0 |
| 15. S.Dem.Feeders |  |  |  |  | 0.02 | 0.02 | 0.01 | 0 | 0.12 | 0.03 | 0.01 | 0.01 |
| 16. Capelin |  |  |  |  |  | 0.25 |  |  |  |  | 1.22 | 1.02 |
| 17. Sand lance |  |  |  |  |  |  |  |  |  |  | 0.30 | 0.38 |
| 18. Arctic cod |  |  |  |  |  |  |  |  |  |  | 0.23 | 0.08 |
| 19. L.Pel.Feeders | 0 |  |  | 0 | 0 | 0 | 0.02 | 0.01 | 0 |  | 0 | 0 |
| 20. Pisc. SPF | 0.17 |  |  |  |  | 0.01 |  | 0.07 | 0.02 |  | 0.02 | 0 |
| 21. Plankt. SPF |  |  |  |  |  |  |  |  | 0.06 |  | 0.27 | 0.15 |
| 22. Shrimp |  |  |  |  |  |  |  |  |  |  | 0.05 | 0.06 |
| 23. Large Crustacea |  |  |  |  |  |  |  |  | 0.19 | 0.04 | 0.01 | 0 |
| 28. L.Zooplankton |  |  |  |  |  |  |  |  |  |  | 0.98 | 6.26 |
| TOTAL | 1.50 | 0.82 | 0.27 | 0.19 | 0.38 | 1.09 | 0.56 | 0.49 | 1.36 | 0.41 | 3.25 | 7.99 |

## Fishing versus predation

In terms of overall predation, the fishery is the fourth top predator (excluding invertebrate predators). For predation on vertebrate prey only, the fishery is the second top predator; harp seals take a greater biomass of fish from the system. Figure 9 illustrates the distribution of fishing mortality across the prey groups and compares this to the predation mortality for each group. The two different mortalities are distributed to the left and the right of the plot. The fishery focuses on the large fish groups and the predators on the smaller fish groups. Hence, fishing removes more spawning biomass but less pre-recruits and immature fish than the natural predators.

The fishery is the only predator on harp seals ${ }^{2}$ (not shown) and American plaice $>35 \mathrm{~cm}$. Fishing greatly exceeds predation mortality on $\operatorname{cod}>35 \mathrm{~cm}$, and exceeds predation mortality on skates, large pelagic feeders and Greenland halibut $>40 \mathrm{~cm}$. For American plaice $\leq 35 \mathrm{~cm}$, flounders, redfish, large demersal feeders, capelin and piscivorous small pelagic feeders, predation mortality is greater than fishing mortality. There is some exploitation of planktivorous small pelagic feeders, shrimp and large crustacea, but the fishing mortality is low due to the large biomass estimated for these groups. Note that no predators of seabirds were included in the model, and catch included various types of anthropogenic mortality, such as hunting or oil discharges at sea. For these reasons, they are not included in this comparison.

These results indicate that the fishery has a large impact on several species. It includes the key commercial species such as cod $>35 \mathrm{~cm}$, Greenland halibut $>40 \mathrm{~cm}$ and American plaice $>$ 35 cm , but also other species such as skates, large pelagic feeders and redfish. However, there are also many species for which the harp seals and cod $>35 \mathrm{~cm}$ impose higher mortality. These include cod $\leq 35 \mathrm{~cm}$, Greenland halibut $\leq 40 \mathrm{~cm}$, American plaice $\leq 35 \mathrm{~cm}$, flounders, small demersal feeders, capelin, sand lance, Arctic cod, small pelagic feeders, shrimps and large crustacea (Figure 9). Clearly, fishing, harp seals and cod $>35 \mathrm{~cm}$ are major predators in this system. However, for most of the species that are fished, the fishery takes the lion's share.

## Discussion

The results described above indicate a system for which our information is uncertain. As a consequence, there are several, if not many, possible versions of the model. Three are described here: an unbalanced model; a model that is balanced by allowing the model to estimate the biomass of some prey groups (Model 1); and a model in which further modifications are made to the skates and large crustacea in Model 1 (Model 2).

The uncertainty over the input parameters for some of the groups at the lower trophic levels has been discussed above. However, there are also several areas of uncertainty for groups at higher trophic levels. These uncertainties are transmitted down the food web, since all production and losses must be balanced for each group. A high consumption at the top of the food web requires high production at all lower levels. The results of a sensitivity analysis indicate that the model is sensitive to some of the groups.

## Sensitivity Analysis

Sensitivity analyses of Models 1 and 2, where the basic input parameters B, P/B, Q/B and EE were varied between $-50 \%$ and $+50 \%$, indicate that for many groups, sensitivity is relatively low (see Appendix 3, Tables 1 and 2). A $50 \%$ change in the input parameters of some groups effected a response of less than $15 \%$ in the estimated parameters of other groups. These groups include cod $\leq 35 \mathrm{~cm}$, Greenland halibut $>40 \mathrm{~cm}$, Greenland halibut $\leq 40 \mathrm{~cm}$, American plaice $>35 \mathrm{~cm}$, American plaice $\leq 35 \mathrm{~cm}$, redfish, large demersal feeders, Arctic

[^1]cod, piscivorous small pelagic feeders, and planktivorous small pelagic feeders. The model is insensitive to several groups, such as the seabirds, large pelagic feeders, shrimp, echinoderms, molluscs, polychaetes and other benthic invertebrates. The linkages between groups are shown graphically in the foodweb in Figure 10.

The model is most sensitive to the groups at the top of the food web and at the bottom of the food web. For example, a $50 \%$ decrease in the biomass of the whales resulted in a $35 \%$ decrease in the estimated biomass of the large demersal feeders and piscivorous small pelagic feeders. Arctic cod show a similar sensitivity to the harp seal biomass. The estimated biomass of the flounders and Greenland halibut $\leq 40 \mathrm{~cm}$ decreased by $20 \%$, and $\operatorname{cod} \leq 35 \mathrm{~cm}$ and large pelagic feeders decreased by $15 \%$, in response to a decrease of $50 \%$ in the biomass of harp seals. Several groups are sensitive to the biomass of cod $>35 \mathrm{~cm}$ : American plaice $\leq$ 35 cm is the most sensitive, then large crustacea, small demersal feeders, capelin, sand lance and molluscs, shrimp and cod $\leq 35 \mathrm{~cm}$.

The greatest sensitivities, however, are to the large crustacea and the large zooplankton. A 50 $\%$ reduction in the $\mathrm{P} / \mathrm{B}$ ratio of the large crustacea results in: a $63 \%$ increase in the estimated biomass of small demersal feeders and shrimp; a $100 \%$ increase in the estimated EE of the echinoderms and molluscs; and a $75 \%$ increase in the estimated EE of polychaetes (Figure 11). Thus the change effected is greater than the change to the input parameter. Sensitivity to the large crustacea was reduced in Model 2 as a result of the reduction of the large crustacea biomass. Since the small demersal feeders were no longer included as live fish in the large crustacea diet, they are no longer sensitive to the large crustacea. The sensitivity of the other groups above was reduced to between 50 and $70 \%$ increases in their estimated parameters.

Sensitivity to the large zooplankton was considerably greater. The biomass of the small zooplankton and the EE of the phytoplankton increased by $200 \%$ when the $\mathrm{P} / \mathrm{B}$ of large zooplankton was reduced by $50 \%$. This was the same for Models 1 and 2 .

In general, there was less sensitivity in Model 2 than Model 1. This is due to the decreased biomass and $\mathrm{Q} / \mathrm{B}$ ratios of some of the predator groups. A few prey groups whose biomass was reduced, such as the shrimp, were more sensitive. Unfortunately, the models were most sensitive to one of the groups that we know least about, that is, the large crustacea. The input parameters for this group were changed in the balancing of the models. However, these parameters are still highly uncertain and impact on several groups. Fortunately, the model is not sensitive to the groups on which they impact, that is, shrimp, echinoderms, molluscs, polychaetes and other benthic invertebrates. However, if the assumptions made about the small demersal feeders in the diet of the large crustacea are erroneous, this ceases to be true. Shrimp and large crustacea are both sensitive to the small demersal feeders input parameters. The models are even more sensitive to the large zooplankton. However this only affects groups to which the model is not sensitive, that is, the small zooplankton and the phytoplankton. Essentially, neither model simulates these lowest trophic levels well; they are highly aggregated and serve as producers for the higher trophic levels. However, if for example there was greater certainty about the biomass of the large zooplankton, then this biomass could act as a limiting factor for the productivity of higher trophic levels.

## Uncertainties in the data

## Whales

The input data for the whales were broad estimates. Biomass was estimated from numbers, generally rounded to the nearest 1,000 , and multiplied by mean weight. However, it was noted that the consumption of squid by whales requires a high biomass estimate of the piscivorous small pelagic feeders. The biomass of short-finned squid was not considered to be this high.

The main species of whale that consume piscivorous small pelagics are sperm whales and pilot whales (potheads). The biomass and diet compositions of both species are poorly known, yet they have a large influence on the model estimates. Their consumption requirements of squid drive the biomass of the piscivorous small pelagic feeders estimated to meet this consumption demand. This, in turn, increases the predation mortality on $\operatorname{cod} \leq 35 \mathrm{~cm}$ and therefore their biomass. Pilot whales in Newfoundland are described as following the squid to inshore waters where squid is their main food (Sergeant 1962). In years when there are few squid, there are few pilot whales. Squid have not been abundant in Newfoundland waters since the late 1970s and early 1980s (Black et al. 1987; Montevecchi 1993). J. Lien (pers. comm.) has not observed any pilot whales in Newfoundland waters in the past 3 years and it is thought that they have moved to the Gulf of St. Lawrence. The numbers of sperm whale are not well known.

Thus, it is conceivable that the whale biomass is lower than estimated, that the piscivorous small pelagic feeders component of the whale diet composition is lower than described, and that the biomass of the piscivorous small pelagic feeders is lower than estimated above.

## Harp Seal

There are a number of areas of uncertainty surrounding the parameter estimates of the harp seals (Shelton et al. 1997; Warren et al. 1997). Uncertainties include residency times, population estimates, diet data and energy requirements (G. Stenson, pers. comm). The biomass of the harp seals is estimated from the model of Shelton et al. (1996a). Warren et al. (1997) analysed uncertainty in population size with respect to pup production from survey estimates. Their estimated $95 \%$ confidence range allows for a $10 \%$ variation in the population estimate of $91,184 \mathrm{t}$. If the uncertainty over residency time, the proportion of seals in the Front herd, the Gulf herd and those which remain in the Arctic are included, the range of uncertainty in the estimate may extend to $20 \%$. Shelton et al. (1997) found that estimates of cod consumption by harp seals had CVs as high as $35 \%$.

There is also uncertainty over the diet composition. The composite diet is a combined inshore and offshore diet. It is assumed that $45 \%$ of the harp seals are inshore and $55 \%$ are offshore, and that they are randomly distributed across the area. It is assumed that the inshore diet is representative of seals eating within 100 km of the shoreline (G. Stenson, pers. comm.). However, these inshore:offshore ratios may be incorrect. Additionally, the data for the offshore diet for the 1980s are poor, and it is assumed that the diet is the same as during the 1990s. Even so, the offshore diet data for the 1990s are poorer than the inshore diet data and it
is feasible that the greater part of the diet occurs offshore (G. Stenson, pers. comm.). If this were the case, consumption of $\operatorname{cod} \leq 35 \mathrm{~cm}$ and Arctic cod would decrease and consumption of capelin and sand lance would increase.

There are similar uncertainties for hooded seals.

## American plaice

The estimated biomass of American plaice is surprisingly high when compared to the biomass of $\operatorname{cod}(869,168 \mathrm{t}$ and $1,044,656 \mathrm{t}$ respectively). The American plaice biomass was estimated from Campelen converted data, using numbers at length and a length-weight relationship, then extrapolating from 3 LNO to find the biomass for 2 J 3 K . The 3LNO biomass estimate is in agreement with the DFO working group estimate (W. Brodie, pers. comm). In contrast, the biomass of cod was taken from SPAs. It is possible that biomass estimates from Campelen trawl surveys exceed those from SPAs for both American plaice and cod.

In all models, the EE of American plaice of both size classes is low. This is caused by low mortality on both adults and juveniles. The predation mortality on both American plaice $>35$ cm and American plaice $\leq 35 \mathrm{~cm}$ is low compared to other groups in the system. This may be due to the lack of definition of American plaice in the stomach samples used to estimate diets. Flatfish are difficult to identify and thus a large proportion of flatfish is "unidentified". In the construction of the diets, some of these unidentified flatfish may have been American plaice that were mistakenly assigned to large demersal feeders (G. Lilly, pers. comm.).

There is also an anomaly in the feeding habits of American plaice. Fish are more important in the diet of the small size group than in that of the large size group. Fish as a percentage of the diet is $48.7 \%$ for the small American plaice and $47.8 \%$ for the large size. Moreover, the smaller size group has a greater diversity of fish in its diet $(\operatorname{cod} \leq 35 \mathrm{~cm}$, Greenland halibut $\leq$ 40 cm , American plaice $\leq 35 \mathrm{~cm}$, flounders, redfish, small demersal feeders, capelin, sand lance and Arctic cod) than the large size group (Greenland halibut $\leq 40 \mathrm{~cm}$, American plaice $\leq$ 35 cm , small demersal feeders, capelin, sand lance and Arctic cod).

Small demersal feeders, large crustacea, large zooplankton
In the balancing of Models 1 and 2 , the parameters of the small demersal feeders, large crustacea and large zooplankton were changed from the original estimates to accommodate the apparent consumption demands of higher trophic levels. Again, this highlights how little we know about the lower trophic levels. Unfortunately, the model is sensitive to all these groups.

Existing information on the small demersal feeders and the large crustacea is scant. The biomass estimated by the model is much larger than the biomass estimated from trawl survey and other data. In order to increase the credibility of these models, it is important to get better data for these groups.

In the case of the large zooplankton, only the small zooplankton and phytoplankton are directly sensitive to the input parameters. However, as discussed further below, the large zooplankton form the base of the forage production. If there were better estimates of large
zooplankton, small zooplankton and phytoplankton biomass for the mid 1980s, these data could constrain the allowable productivity at higher trophic levels.

## Additional comments on diet

Weaknesses in diet data have been noted in many sections of this paper. Additional comments are warranted regarding small juvenile fish in the diets of predators and the degree to which an exercise such as this can provide insightful analysis of the contribution of various predators to the mortality of a commercially important top predator such as Atlantic cod.

There are problems with how small (say $<6-7 \mathrm{~cm}$ ) fish are included within the model, both as predators and as prey. For most fish species, their role as predators while they are small ( $<6-7$ cm ) has not been included explicitly in the present model. Their role as prey has not been treated equally for all species. For example, the model results show considerable predation by capelin on sand lance and capelin (cannibalism). This seems at first somewhat surprising, because capelin is considered to be planktivorous. The information regarding predation on sand lance and capelin comes from careful observations of stomach contents of capelin at sea in the spring (Gerasimova 1994). If there were similar observations of the diet of other species at the time when juveniles of potential prey are in the water column or just settling out of the water column, then more predation on small ( $<6-7 \mathrm{~cm}$ ) individuals of many species might be recognized. For example, studies in the Barents Sea have pointed to the importance of predation by herring on small capelin. Such predation may occur unrecognized in Newfoundland waters because there have been no careful studies of herring stomach contents during the period between the hatching of capelin and their flushing out of inshore waters.

At the time of writing of this paper, there is considerable interest in the contribution of various predators, most notably harp seals, to the mortality of Atlantic cod. The present model provides estimates of the contributions by various predators or groups of predators, but it must be clear that the diet information for all of these predators is inadequate for obtaining accurate estimates of their consumption of minor prey. For example, the information on predation by whales comes mainly from reports of cod in the diet of pilot whales, and yet the actual proportion of cod in the diet of pilot whales is not known. In addition, pilot whales may not have been very abundant in Newfoundland waters in 1985-1987 because there were very few short-finned squid (see section on Whales above). On the other hand, minke whales are reported to prey on Atlantic cod in the Barents Sea, but there is no recent information on the diet of minke whales in Newfoundland waters. After the analyses reported in this paper were completed, the estimate of the quantity of Atlantic cod consumed by harp seals was considerably reduced (Lilly et al. 1999), illustrating the uncertainty involved in these calculations. Predation on Atlantic cod by skates remains very uncertain because diet data for skates are sparse and perhaps unrepresentative. The importance of cannibalism in Atlantic cod needs to be calculated more carefully. It has been suggested that Arctic cod might eat small Atlantic cod. Although this has not yet been observed, it must be noted that the stomach contents of Arctic cod have not been examined during the late summer and early autumn when pelagic juvenile Atlantic cod would be in the water column. The information on predation by short-finned squid on Atlantic cod came from very few sites and there is no quantitative information on the biomass of squid entering Newfoundland waters each year. For each of the various predators discussed above, Atlantic cod is a minor component of the diet, and an
extensive and intensive stomach sampling program would be required to obtain accurate estimates of the number/quantity of Atlantic cod consumed.

There is also considerable uncertainty about how the model might take into account predation on eggs and larvae. For example, eggs and larvae of Atlantic cod could be preyed upon by planktivorous fish (capelin, herring, Arctic cod) and various groups of invertebrates, including jellyfish, euphausiids, shrimp and crab larvae. Such predation is not included in the models described in this paper. The consequence of excluding such predation is not known but likely to be small as long as one is interested in biomass flows. However, such links between species and species groups might be very important if the model outputs are to be used as input to dynamic models.

## Network Flow Analysis

The biomass flows between Ecopath groups in Model 2 are shown in Figure 10. The trophic levels are given in Table 23, and indicated on the $y$-axis of the figure. Hooded seals are at the top of the food web, reflecting the high proportion of fish such as Greenland halibut $\leq 40 \mathrm{~cm}$ and cod $\leq 35 \mathrm{~cm}$ in the diet. There are several other groups (including Greenland halibut, harp seals, cod and skates) with a trophic level higher than four, reflecting the high number of top predators in this system.

This high level of predation is supported by a large pelagic forage base, which includes capelin, sand lance, Arctic cod, and planktivorous small pelagic feeders. Large zooplankton and small zooplankton form the basis of this forage network. The main production by large zooplankton and small zooplankton is channelled through capelin. Capelin are thus the critical converter of energy produced at lower trophic levels to a form available to the higher trophic levels. The important role of capelin in the food web is shown in Figure 12. To fulfil this role, capelin require almost $11 \%$ of the primary production in the system.

As shown in Figure 10, there are many linkages in the model. The pelagic side of the food web dominates the system in terms of energy flow. For example, around $80 \%$ of the energy required by harp seals and $\operatorname{cod}>35 \mathrm{~cm}$ is derived from the pelagic side of the foodweb. Harp seals require $5 \%$ of primary production and $\operatorname{cod}>35 \mathrm{~cm}$ require $4.6 \%$ of primary production. However, little empirical data are available for any of the pelagic species that link the zooplankton to the top predators of the system.

Benthic productivity is even less well understood. As modelled here, less than $1 \%$ of the energy consumed by echinoderms and molluscs is transferred to the next trophic level. This is an extremely low transfer efficiency. It is very likely that most of the interactions at this level are lost in the aggregation process and are thus not well modelled. The benthic invertebrate groups are at a trophic level of 2, and their consumers are all at trophic level 3 or higher. A more realistic model would include consumers that spanned these two trophic levels.

## Concluding remarks

The Ecopath model presented here provides a summary of our current knowledge of the biomass, consumption, production, food web and trophic flows in the ecosystem exploited by the Newfoundland fishery in divisions 2J3KLNO. The study confirms some of our knowledge, such as the important forage role played by capelin and the dominance of predators such as harp seals and cod. It suggests that harp seals and cod $>35 \mathrm{~cm}$ are the two major predators of $\operatorname{cod} \leq 35 \mathrm{~cm}$.

The model also highlights many of the uncertainties in our knowledge. Unfortunately, these uncertainties extend over all trophic levels and many of the constituent groups of the models, including those that are regularly surveyed and assessed. It was necessary to make several assumptions to balance the model. For some model estimates, such as the biomass of the small demersal feeders, there is little known to either substantiate or discredit their magnitude. The sensitivity analysis indicated that the model was not very sensitive to the biomass of the small demersal feeders. However, if the biomass of this group was reduced to the original estimate, the model would not be balanced. Thus the estimated biomass is critical to the model, as parameterised.

The biomasses of several other groups, such as juveniles of commercial species, piscivorous small pelagic feeders, planktivorous small pelagic feeders, shrimp, and large crustacea, were also raised during the balancing process. Thus, the model indicates a greater biomass at lower trophic levels than has been estimated to date. This result has been found in other ecosystem modelling studies. For example, Welch et al. (1992) modelled the Lancaster Sound with a bioenergetics model. Their model predicted a greater biomass of Arctic cod than estimated by acoustic and trawl surveys. Earlier work by Jones (1984) focussed on primary and secondary producers in the North Sea and Georges Bank and showed that with more than one to two pathways from prey to predator, there was insufficient production to sustain the predator.

There are two major implications of this study. The first is that we know little about productivity at lower trophic levels. Since there is continued debate over whether different species and trophic levels are food limited or predator limited, knowledge of the productivity of lower trophic levels is clearly important. Secondly, this study demonstrates the utility of studying top predators to estimate the system energy flows required to sustain those top predators. We cannot begin to understand the ecosystem effects of fishing until we understand the nature of interactions between different species or groups, at all levels of the food web.

## Recommendations

There are two main areas where better data are required: diet and biomass estimation.

## Uncertainty over diet data

For many of the groups in the model, empirical diet data from the model area were sparse or non-existent. Where data did not exist information was taken from the literature. However, even for groups with good diet data, numerous shortcomings were noted. The following recommendations are made for new diet studies:

Sampling should include the full size spectrum of each predator and extend over the predator's full geographic distribution with sufficient temporal coverage to ensure that all major seasonal changes in diet are represented. Stomach content examinations should be conducted at a taxonomic level sufficient to accommodate at least the finest level of aggregation desired in the models and should include information on the size, number and weight of the prey. Additional information that might be helpful includes state of digestion of the prey. During calculation of the average diet composition, samples should be weighted by the predator population numbers at length. Variation in digestion rate associated with factors such as temperature, type and size of prey, and degree of stomach fullness should be taken into account.

## Biomass estimation

Biomass estimates serve to scale the model but the biomass estimates are poor for many groups. This is especially so at lower trophic levels, but also occurs at higher levels such as the whales, whose biomass estimates are crude approximations derived from the products of estimated numbers and mean weights. The models would be more robust if more confidence could be placed in the biomass estimates.

As a final word, we wish to note that the process of constructing models such as these is essentially open-ended. The data available for inputs are constantly being added to and revised. The models presented here should be considered a first step. We invite others to critique the model structure, the input data and the assumptions, so that the models can be improved during subsequent exercises. The data used in the current models are available from the authors upon request.

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Fig. 1a. Map of the shelf off Labrador and Newfoundland, indicating the major shelf areas and banks.


Fig. 1b. Map of the shelf off Labrador and Newfoundland, indicating NAFO divisions and subdivisions.


Fig. 1c. Map of the study area from southern Labrador to the southern Grand Bank (divisions 2J3KLNO).


Fig. 2. Stock structure of major commercial fish species on the Labrador Shelf, the Northeast Newfoundland Shelf and Grand Bank. Dotted lines north of 2J indicate that the stock boundary extends to the northern end of Subarea 2. The dotted line between 3K and 3L in the $2+3 \mathrm{KL}$ capelin stock is a reminder that during 1985-1987 the stock was assessed as two stocks; $2+3 \mathrm{~K}$ and 3 L .


Fig. 3. Catch curve analysis of cod survey data (numbers per tow), ages 3-13, for years 1984, 1985, and 1987.


Fig. 4. Catch curve analysis of American plaice survey data, ages 7-17, 1985-1987.


Fig. 5. Composition of vertebrate predators on (A) all prey and (B) fish prey.


Fig. 6. Percentage difference between the biomass in the unbalanced model and the biomass in the balanced Model 1.


Fig. 7. Percentage difference between the biomass in the unbalanced model and the biomass in the balanced Models 1 and 2.


Fig. 8. Distribution of predation mortality among vertebrate predators on all prey (A) and all predators on vertebrate prey (B). The percentage figures include fishing.


Fig. 9. Comparison of fishing mortality and predation mortality on individual prey groups.


Figure 10 Flow diagram of the Newfoundland-Labrador Ecopath model depicting the
linkages between Ecopath groups.


Fig. 11. Sensitivity of the shrimp, echinoderms, molluscs, polychaetes and other benthic invertebrates to the large crustacea input parameters PB and EE.


Figure 12 Flow diagram of the Newfoundland-Labrador Ecopath model highlighting the role of Capelin.

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Appendix 2. Table 1a. Diet of large ( $>35 \mathrm{~cm}$ ) cod in Division 3L by season and in total.

| Common Name | 3L winter |  |  | 3L spring |  |  |  | 3L summer |  |  |  | 3L autumn |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | offshore | inshore |  |  |  |  |  |  |  |
|  | 1985 | 1986 | Diet 1 |  |  |  |  | 1985 | 1986 | 1987 | Diet 2 | 1985 | 1983 | 1984 | Diet 3 |  | 1985 | 1986 | 1987 | Diet 4 |
| 1 Whales | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 2 Harp seals | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 3 Hooded seals | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 4 Seabirds | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| $5 \mathrm{Cod}>35 \mathrm{~cm}$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.0617 | 0 | 0.0309 | 0.0477 | 0.0179 | 0.0048 | 0.0235 | 0.0195 | 0.0003 | 0 | 0.0001 | 0.0325 | 0.0393 | 0.0139 | 0.0286 | 0.018566 |
| 7 Greenland halibut > 40 cm | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 8 Greenland halibut $<=40 \mathrm{~cm}$ | 0.0001 | 0 | 7E-05 | 0.0034 | 0.0057 | 0.0039 | 0.0043 | 0.0109 | 0 | 7E-05 | 4E-05 | 0.0214 | 0.0019 | 0.0053 | 0.0095 | 0.003883 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 10 American plaice $<=35 \mathrm{~cm}$ | 0.0624 | 0.0136 | 0.038 | 0.0376 | 0.0233 | 0.0156 | 0.0255 | 0.0163 | 0 | 0.001 | 0.0005 | 0.0712 | 0.0114 | 0.0171 | 0.0332 | 0.021137 |
| 11 Flounders | 0.0089 | 0.0004 | 0.0047 | 0 | 0.001 | 0 | 0.0003 | 0 | 1E-05 | 0 | 5E-06 | 0 | 0 | 0 | 0 | 0.000998 |
| 12 Skates | 0 | 0 | 0 | 0.0002 | 0 | 0 | 5E-05 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0.0000 |
| 13 Redfish | 0.0163 | 0 | 0.0081 | 0.0002 | 0.0006 | 0 | 0.0003 | 0.0005 | 0 | 0 | 0 | 0.0039 | 0.0142 | 0.0046 | 0.0076 | 0.003239 |
| 14 Demersal feeders (misc. large) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 15 Demersal feeders (misc. small) | 0.1247 | 0.1112 | 0.1179 | 0.0615 | 0.0587 | 0.0316 | 0.0506 | 0.04 | 0.0007 | 0.0002 | 0.0005 | 0.1019 | 0.0707 | 0.0553 | 0.0759 | 0.053043 |
| 16 Capelin | 0.4655 | 0.5043 | 0.4849 | 0.5976 | 0.5432 | 0.5161 | 0.5523 | 0.4832 | 0.9779 | 0.9398 | 0.9588 | 0.3965 | 0.3071 | 0.2414 | 0.315 | 0.606404 |
| 17 Sand lance | 0.0347 | 0.0059 | 0.0203 | 0.1225 | 0.0984 | 0.2603 | 0.1604 | 0.2593 | 0.0067 | 0.0337 | 0.0202 | 0.1147 | 0.1674 | 0.3334 | 0.2052 | 0.109162 |
| 18 Arctic cod | 0.001 | 0.0159 | 0.0084 | 0.0014 | 0.0004 | 0.0001 | 0.0007 | 0.0021 | 0 | 0 | 0 | 0.003 | 0.0031 | 0.0065 | 0.0042 | 0.002869 |
| 19 Pelagic feeders (misc. large) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.0038 | 0.0008 | 0.0023 | 1E-04 | 8E-05 | 0.0005 | 0.0002 | 0.001 | 0.0001 | 0.0017 | 0.0009 | 0.0015 | 0.0003 | 0.0009 | 0.0009 | 0.001057 |
| 21 Plankt. Pelagic feeders (misc. small) | 0.0113 | 0.0023 | 0.0068 | 0.0003 | 0.0002 | 0.0014 | 0.0006 | 0.0029 | 0.0003 | 0.0052 | 0.0028 | 0.0046 | 0.0008 | 0.0028 | 0.0027 | 0.003171 |
| 22 Shrimp | 0.0695 | 0.0939 | 0.0817 | 0.0206 | 0.0179 | 0.0084 | 0.0156 | 0.0146 | 0.0007 | 0.0035 | 0.0021 | 0.0398 | 0.06 | 0.0391 | 0.0463 | 0.030831 |
| 23 Large Crustaceans (crab, lobster) | 0.0496 | 0.1063 | 0.0779 | 0.0591 | 0.1181 | 0.0755 | 0.0842 | 0.0563 | 0.0044 | 0.0069 | 0.0056 | 0.1019 | 0.144 | 0.137 | 0.1276 | 0.065277 |
| 24 Echinoderms | 0.0046 | 0.008 | 0.0063 | 0.0022 | 0.0048 | 0.0021 | 0.0031 | 0.0016 | 0.0059 | 0.0017 | 0.0038 | 0.0062 | 0.0094 | 0.0058 | 0.0071 | 0.0046 |
| 25 Molluscs (benthic) | 0.0131 | 0.052 | 0.0325 | 0.03 | 0.054 | 0.0305 | 0.0382 | 0.0542 | 0.0004 | 0.002 | 0.0012 | 0.0318 | 0.0928 | 0.0488 | 0.0578 | 0.031466 |
| 26 Polychaetes | 0.009 | 0.0272 | 0.0181 | 0.0032 | 0.0073 | 0.003 | 0.0045 | 0.0066 | 0.0002 | 0.0006 | 0.0004 | 0.0125 | 0.0127 | 0.0118 | 0.0124 | 0.007777 |
| 27 Benthic invertebrates (other) | 0.0485 | 0.0233 | 0.0359 | 0.0105 | 0.0137 | 0.0148 | 0.013 | 0.024 | 0.0021 | 0.0014 | 0.0017 | 0.02 | 0.0323 | 0.031 | 0.0278 | 0.018252 |
| 28 Zooplankton (large) | 0.0152 | 0.0349 | 0.025 | 0.0021 | 0.0347 | 0.0313 | 0.0227 | 0.007 | 0.0003 | 0.0022 | 0.0012 | 0.0397 | 0.0325 | 0.0455 | 0.0392 | 0.0185 |
| 29 Zooplankton (small) | 0.0001 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0 | 0.0000 | 0 | 0.0000 | 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 30 Phytoplankton | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 31 Detritus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| No. of stomachs | 1289 | 589 |  | 1843 | 2898 | 2413 |  | 1057 | 246 | 1019 |  | 1529 | 1025 | 1034 |  |  |

Diet 1. Division 3L winter diet. Average of 1985 and 1986. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small).
Diet 2. Division 3L spring diet. Average of 1985, 1986 and 1987. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small) Diet 3. Division 3L summer inshore diet. Average of 1983 and 1984. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small) Diet 4. Division 3L autumn diet. Average of 1985, 1986 and 1987. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small)
Diet 5. Division 3L total diet. The winter, spring, summer offshore, summer inshore and autumn diets are given weightings of $0.2,0.2,0.1,0.3$, and 0.2 .
This reflects the following weightings for population size (first number) and feeding rate (second number): ( $1^{*} 1$ ), ( $1^{*} 1$ ), ( $0.5^{*} 1$ ), ( $0.5^{*} 3$ ), ( $1^{* 1}$ ).
The summer inshore diet is given a weighting 3 times the offshore diets because of higher stomach fullness and possibly higher temperature.
NB For all diets, the two small pelagic groups were originally one. Here they are split on the assumption that $75 \%$ are planktivorous feeders and $25 \%$ are piscivorous feeders.

Appendix 2. Table 1b. Diet of large ( $>35 \mathrm{~cm}$ ) cod in divisions $2 \mathrm{~J} 3 \mathrm{~K}, 3 \mathrm{NO}$ and 2 J 3 KLNO combined.

| Common Name | summer inshore Diet 3 | 2J3K autumn |  |  |  | $\begin{gathered} \hline 2 \mathrm{~J}+3 \mathrm{~K} \\ \text { total } \\ \text { Diet } 7 \\ \hline \end{gathered}$ | 3NO spring |  |  |  | $\begin{gathered} \hline 2 \mathrm{~J} 3 \mathrm{~K} \\ \text { total } \\ (\text { Diet } 7) \\ \hline \end{gathered}$ |  | $\begin{gathered} \hline 3 \mathrm{NO} \\ \text { total } \\ \text { (Diet 8) } \end{gathered}$ | $\begin{gathered} \hline \text { 2J3KLNO } \\ \text { total } \\ \text { Diet } 9 \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1985 | 1986 | 1987 | Diet 6 |  | 1985 | 1986 | 1987 | Diet 8 |  |  |  |  |
| 1 Whales | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 Harp seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 3 Hooded seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 4 Seabirds | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $5 \mathrm{Cod}>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.0001 | 0.0195 | 0.0236 | 0.0232 | 0.0221 | 0.0138 | 0.0200 | 0.0422 | 0.0380 | 0.0334 | 0.0138 | 0.0186 | 0.0334 | 0.0177 |
| 7 Greenland halibut > 40 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 8 Greenland halibut <= 40 cm | 0.0000 | 0.0209 | 0.0085 | 0.0083 | 0.0126 | 0.0078 | 0.0000 | 0.0015 | 0.0000 | 0.0005 | 0.0078 | 0.0039 | 0.0005 | 0.0057 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 10 American plaice $<=35 \mathrm{~cm}$ | 0.0005 | 0.0015 | 0.0051 | 0.0244 | 0.0103 | 0.0066 | 0.0401 | 0.0649 | 0.0599 | 0.0550 | 0.0066 | 0.0211 | 0.0550 | 0.0171 |
| 11 Flounders | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0048 | 0.0134 | 0.0000 | 0.0061 | 0.0000 | 0.0010 | 0.0061 | 0.0011 |
| 12 Skates | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0046 | 0.0000 | 0.0015 | 0.0000 | 0.0000 | 0.0015 | 0.0002 |
| 13 Redfish | 0.0000 | 0.0510 | 0.0020 | 0.0051 | 0.0194 | 0.0120 | 0.0060 | 0.0083 | 0.0012 | 0.0052 | 0.0120 | 0.0032 | 0.0052 | 0.0086 |
| 14 Demersal feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 15 Demersal feeders (misc. small) | 0.0005 | 0.0300 | 0.0743 | 0.0532 | 0.0525 | 0.0327 | 0.0784 | 0.0213 | 0.0268 | 0.0422 | 0.0327 | 0.0530 | 0.0422 | 0.0399 |
| 16 Capelin | 0.9588 | 0.5229 | 0.6460 | 0.5582 | 0.5757 | 0.7213 | 0.0795 | 0.0260 | 0.0882 | 0.0646 | 0.7213 | 0.6064 | 0.0646 | 0.6026 |
| 17 Sand lance | 0.0202 | 0.0003 | 0.0000 | 0.0002 | 0.0002 | 0.0078 | 0.5166 | 0.6606 | 0.4068 | 0.5280 | 0.0078 | 0.1092 | 0.5280 | 0.1048 |
| 18 Arctic cod | 0.0000 | 0.0850 | 0.0567 | 0.0329 | 0.0582 | 0.0361 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0361 | 0.0029 | 0.0000 | 0.0218 |
| 19 Pelagic feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.0009 | 0.0028 | 0.0023 | 0.0018 | 0.0023 | 0.0018 | 0.0000 | 0.0002 | 0.0000 | 0.0001 | 0.0018 | 0.0011 | 0.0001 | 0.0013 |
| 21 Plankt. Pelagic feeders (misc. small) | 0.0028 | 0.0084 | 0.0069 | 0.0053 | 0.0069 | 0.0053 | 0.0000 | 0.0005 | 0.0001 | 0.0002 | 0.0053 | 0.0032 | 0.0002 | 0.0040 |
| 22 Shrimp | 0.0021 | 0.0752 | 0.0705 | 0.0818 | 0.0759 | 0.0478 | 0.0013 | 0.0010 | 0.0043 | 0.0022 | 0.0478 | 0.0308 | 0.0022 | 0.0370 |
| 23 Large Crustaceans (crab, lobster) | 0.0056 | 0.0295 | 0.0499 | 0.0563 | 0.0452 | 0.0302 | 0.0963 | 0.0661 | 0.1154 | 0.0926 | 0.0302 | 0.0653 | 0.0926 | 0.0485 |
| 24 Echinoderms | 0.0038 | 0.0022 | 0.0020 | 0.0024 | 0.0022 | 0.0028 | 0.0094 | 0.0066 | 0.0148 | 0.0103 | 0.0028 | 0.0046 | 0.0103 | 0.0043 |
| 25 Molluscs (benthic) | 0.0012 | 0.0022 | 0.0010 | 0.0016 | 0.0016 | 0.0014 | 0.1105 | 0.0629 | 0.1249 | 0.0994 | 0.0014 | 0.0315 | 0.0994 | 0.0229 |
| 26 Polychaetes | 0.0004 | 0.0041 | 0.0022 | 0.0063 | 0.0042 | 0.0028 | 0.0330 | 0.0084 | 0.0067 | 0.0160 | 0.0028 | 0.0078 | 0.0160 | 0.0059 |
| 27 Benthic invertebrates (other) | 0.0017 | 0.0109 | 0.0076 | 0.0122 | 0.0102 | 0.0070 | 0.0039 | 0.0024 | 0.0096 | 0.0053 | 0.0070 | 0.0183 | 0.0053 | 0.0100 |
| 28 Zooplankton (large) | 0.0012 | 0.1337 | 0.0411 | 0.1268 | 0.1005 | 0.0628 | 0.0002 | 0.0092 | 0.1032 | 0.0375 | 0.0628 | 0.0185 | 0.0375 | 0.0466 |
| 29 Zooplankton (small) | 0.0000 | 0.0000 | 0.0000 | 0.0001 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 30 Phytoplankton | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 31 Detritus | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |

## No. of stomachs

 $\begin{array}{lll}2,024 & 1,768 & 1,924\end{array}$ 925 1,091 1,572
## Diet 3. From Table 1a

Diet 6. Division 2J3K autumn diet. Average of 1985, 1986 and 1987. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small). Diet 7. Division 2 J 3 K total diet. The summer inshore and autumn diets are given weightings of 0.38 and 0.62 .
This reflects the following weightings for population size (first number) and feeding rate (second number) for the winter, spring, summer offshore,
summer inshore and autumn: ( $1^{*}$ ), ( $\left.1^{1} 1\right),\left(0.5^{* 1}\right),\left(0.5^{*} 3\right),\left(1^{*} 1\right)$.
It is assumed for simplicity that these cod do not feed in the winter (Lilly 1987) and that the autumn diet is also appropriate for spring and summer offshore.
Diet 8. Division 3NO spring diet. It is assumed that this diet is appropriate for all seasons.Average of 1985, 1986 and 1987. Demersal feeders (large)
are considered to be juveniles and classified as demersal feeders (small)
Diet 9 . Average of $2 \mathrm{~J} 3 \mathrm{~K}, 3 \mathrm{~L}$, and 3 NO diets, weighted by population sizes. The weightings are $0.58,0.29$, and 0.13 . This is based on 0.87 of the average biomass being in 2 J 3 KL , and the assumption that 0.33 of the 2 J 3 KL biomass is in 3 L .
NB For all diets, the two small pelagic groups were originally one. Here they are split on the assumption that $75 \%$ of small pelagics in the diet are planktivorous feeders and $25 \%$ are piscivorous feeders.

Appendix 2. Table 2a. Diet of small ( $<=35 \mathrm{~cm}$ ) cod in Division 3L by season and in total.

| Common Name | 3L winter |  |  | 3L spring |  |  |  | 3L summer |  | 3L autumn |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  |  | $\frac{\text { offshore }}{1985}$ | $\begin{array}{r} \hline \text { inshore } \\ \hline \text { Diet } 3 \\ \hline \end{array}$ |  |  |  |  |  |
|  | 1985 | 1986 | Diet 1 |  |  | 1985 | 1986 | 1987 | Diet 2 | 1985 | 1986 | 1987 | Diet 4 |  |
| 1 Whales | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 Harp seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 3 Hooded seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 4 Seabirds | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $5 \mathrm{Cod}>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.0035 | 0.0000 | 0.0017 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0142 | 0.0000 | 0.0000 | 0.0362 | 0.0121 | 0.0088 |
| 7 Greenland halibut > 40 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 8 Greenland halibut <= 40 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0094 | 0.0000 | 0.0031 | 0.0004 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 10 American plaice <=35 cm | 0.0009 | 0.0003 | 0.0006 | 0.0002 | 0.0000 | 0.0036 | 0.0013 | 0.0005 | 0.0000 | 0.0071 | 0.0016 | 0.0068 | 0.0052 | 0.0009 |
| 11 Flounders | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0001 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 12 Skates | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 13 Redfish | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0050 | 0.0017 | 0.0002 |
| 14 Demersal feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 15 Demersal feeders (misc. small) | 0.0191 | 0.0509 | 0.0350 | 0.0101 | 0.0022 | 0.0007 | 0.0043 | 0.0083 | 0.0115 | 0.0536 | 0.0340 | 0.0264 | 0.0380 | 0.0165 |
| 16 Capelin | 0.5730 | 0.4442 | 0.5086 | 0.6848 | 0.6573 | 0.4742 | 0.6054 | 0.6978 | 0.3778 | 0.4806 | 0.4468 | 0.4828 | 0.4701 | 0.4741 |
| 17 Sand lance | 0.0539 | 0.0166 | 0.0352 | 0.0213 | 0.0193 | 0.0227 | 0.0211 | 0.0399 | 0.0107 | 0.0190 | 0.0332 | 0.1467 | 0.0663 | 0.0257 |
| 18 Arctic cod | 0.0022 | 0.0000 | 0.0011 | 0.0104 | 0.0000 | 0.0000 | 0.0035 | 0.0000 | 0.0000 | 0.0000 | 0.0167 | 0.0000 | 0.0056 | 0.0013 |
| 19 Pelagic feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0001 | 0.0000 | 0.0000 | 0.0000 | 0.0102 | 0.0023 | 0.0000 | 0.0000 | 0.0008 | 0.0052 |
| 21 Plankt. Pelagic feeders (misc. sma | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0004 | 0.0000 | 0.0001 | 0.0000 | 0.0305 | 0.0069 | 0.0000 | 0.0000 | 0.0023 | 0.0156 |
| 22 Shrimp | 0.1639 | 0.2151 | 0.1895 | 0.0802 | 0.0559 | 0.0654 | 0.0671 | 0.0647 | 0.0024 | 0.1504 | 0.1376 | 0.1110 | 0.1330 | 0.0580 |
| 23 Large Crustaceans (crab, lobster) | 0.0100 | 0.0219 | 0.0160 | 0.0684 | 0.0786 | 0.0885 | 0.0785 | 0.0125 | 0.0040 | 0.0104 | 0.0058 | 0.0081 | 0.0081 | 0.0164 |
| 24 Echinoderms | 0.0004 | 0.0000 | 0.0002 | 0.0004 | 0.0000 | 0.0008 | 0.0004 | 0.0000 | 0.0003 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0002 |
| 25 Molluscs (benthic) | 0.0104 | 0.0091 | 0.0098 | 0.0137 | 0.0078 | 0.0147 | 0.0121 | 0.0225 | 0.0006 | 0.0149 | 0.0015 | 0.0118 | 0.0094 | 0.0070 |
| 26 Polychaetes | 0.0401 | 0.0441 | 0.0421 | 0.0221 | 0.0239 | 0.0118 | 0.0193 | 0.0158 | 0.0065 | 0.0471 | 0.0511 | 0.0380 | 0.0454 | 0.0186 |
| 27 Benthic invertebrates (other) | 0.0585 | 0.0917 | 0.0751 | 0.0548 | 0.0512 | 0.1142 | 0.0734 | 0.0885 | 0.2589 | 0.0921 | 0.0802 | 0.0869 | 0.0864 | 0.1699 |
| 28 Zooplankton (large) | 0.0632 | 0.1059 | 0.0845 | 0.0324 | 0.1033 | 0.2032 | 0.1129 | 0.0495 | 0.2727 | 0.1156 | 0.1823 | 0.0401 | 0.1127 | 0.1813 |
| 29 Zooplankton (small) | 0.0010 | 0.0002 | 0.0006 | 0.0013 | 0.0001 | 0.0003 | 0.0006 | 0.0000 | 0.0000 | 0.0001 | 0.0000 | 0.0001 | 0.0001 | 0.0002 |
| 30 Phytoplankton | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 31 Detritus | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |

Diet 1. Division 3L winter diet. Average of 1985 and 1986. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small).
Diet 2. Division 3L spring diet. Average of 1985, 1986 and 1987. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small). Diet 3. Division 3L summer inshore diet. Derived from frequency of occurrence data in 1992-1994. See text for details. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small).
Diet 4. Division 3L autumn diet. Average of 1985, 1986 and 1987. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small). Diet 5 . Division 3L total diet. The winter, spring, summer offshore, summer inshore and autumn diets are given weightings of $0.125,0.125,0.125,0.5$ and 0.125 .
This reflects the following weightings for population size (first number) and feeding rate (second number) for the offshore in winter, spring, summer and autumn and the inshore for winter, spring, summer and autumn: ( $0.5^{*} 1$ ), ( $0.5^{*} 1$ ), ( $\left.0.5^{*} 1\right)$, ( $0.5^{*} 1$ ), ( $\left.0.5^{*} 0\right)$, ( $\left.0.5^{*} 1\right)$, ( $0.5^{*} 2$ ), ( $0.5^{*} 1$ ).
It is assumed that the inshore cod do not feed in winter and that the summer inshore diet is applicable for spring inshore and autumn inshore.
NB For all diets, the two small pelagic groups were originally one. Here they are split on the assumption that $75 \%$ are planktivorous feeders and $25 \%$ are piscivorous feeders.

Appendix 2. Table 2b. Diet of small ( $<=35 \mathrm{~cm}$ ) cod in divisions $2 \mathrm{~J} 3 \mathrm{~K}, 3 \mathrm{NO}$ and 2 J 3 KLNO combined.

| Common Name | Diet 3 | 2J3K autumn |  |  |  | $\begin{gathered} \hline 2 \mathrm{~J} 3 \mathrm{~K} \\ \text { total } \\ \text { Diet } 7 \\ \hline \end{gathered}$ | 3NO spring |  |  |  | $\begin{gathered} \hline 2 \mathrm{~J} 3 \mathrm{~K} \\ \text { total } \\ \text { (Diet 7) } \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { 3L } \\ \text { total } \\ \text { (Diet 5) } \\ \hline \end{gathered}$ | $\begin{gathered} \hline 3 \mathrm{NO} \\ \text { total } \\ \text { (Diet 8) } \\ \hline \end{gathered}$ | 2J3KLNO total Diet 9 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 1985 | 1986 | 1987 | Diet 6 |  | 1985 | 1986 | 1987 | Diet 8 |  |  |  |  |
| 1 Whales | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 Harp seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 3 Hooded seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 4 Seabirds | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $5 \mathrm{Cod}>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.0142 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0054 | 0.0000 | 0.0181 | 0.0000 | 0.0060 | 0.0054 | 0.0088 | 0.0060 | 0.0065 |
| 7 Greenland halibut > 40 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 8 Greenland halibut <= 40 cm | 0.0000 | 0.0129 | 0.0041 | 0.0000 | 0.0057 | 0.0035 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0035 | 0.0004 | 0.0000 | 0.0023 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 10 American plaice $<=35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0009 | 0.0000 | 0.0003 |
| 11 Flounders | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 12 Skates | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 13 Redfish | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0002 | 0.0000 | 0.0001 |
| 14 Demersal feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 15 Demersal feeders (misc. small) | 0.0115 | 0.0108 | 0.0294 | 0.0790 | 0.0397 | 0.0290 | 0.0032 | 0.0121 | 0.0219 | 0.0124 | 0.0290 | 0.0165 | 0.0124 | 0.0239 |
| 16 Capelin | 0.3778 | 0.5753 | 0.4126 | 0.4383 | 0.4754 | 0.4383 | 0.1819 | 0.1580 | 0.1292 | 0.1563 | 0.4383 | 0.4741 | 0.1563 | 0.4297 |
| 17 Sand lance | 0.0107 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0041 | 0.4524 | 0.2629 | 0.1585 | 0.2913 | 0.0041 | 0.0257 | 0.2913 | 0.0309 |
| 18 Arctic cod | 0.0000 | 0.1240 | 0.0764 | 0.0475 | 0.0826 | 0.0512 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0512 | 0.0013 | 0.0000 | 0.0322 |
| 19 Pelagic feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.0102 | 0.0011 | 0.0010 | 0.0017 | 0.0012 | 0.0046 | 0.0000 | 0.0035 | 0.0006 | 0.0014 | 0.0046 | 0.0052 | 0.0014 | 0.0046 |
| 21 Plankt. Pelagic feeders (misc. small) | 0.0305 | 0.0032 | 0.0029 | 0.0051 | 0.0037 | 0.0139 | 0.0000 | 0.0106 | 0.0019 | 0.0042 | 0.0139 | 0.0156 | 0.0042 | 0.0137 |
| 22 Shrimp | 0.0024 | 0.0751 | 0.2028 | 0.1770 | 0.1516 | 0.0949 | 0.0036 | 0.0398 | 0.0373 | 0.0269 | 0.0949 | 0.0580 | 0.0269 | 0.0787 |
| 23 Large Crustaceans (crab, lobster) | 0.0040 | 0.0011 | 0.0056 | 0.0177 | 0.0081 | 0.0065 | 0.1404 | 0.2746 | 0.1465 | 0.1871 | 0.0065 | 0.0164 | 0.1871 | 0.0222 |
| 24 Echinoderms | 0.0003 | 0.0000 | 0.0003 | 0.0000 | 0.0001 | 0.0002 | 0.0000 | 0.0017 | 0.0000 | 0.0006 | 0.0002 | 0.0002 | 0.0006 | 0.0002 |
| 25 Molluscs (benthic) | 0.0006 | 0.0018 | 0.0011 | 0.0005 | 0.0012 | 0.0009 | 0.1380 | 0.0049 | 0.0113 | 0.0514 | 0.0009 | 0.0070 | 0.0514 | 0.0064 |
| 26 Polychaetes | 0.0065 | 0.0094 | 0.0323 | 0.0212 | 0.0209 | 0.0155 | 0.0222 | 0.0239 | 0.0254 | 0.0238 | 0.0155 | 0.0186 | 0.0238 | 0.0170 |
| 27 Benthic invertebrates (other) | 0.2589 | 0.0238 | 0.0737 | 0.0717 | 0.0564 | 0.1333 | 0.0342 | 0.0546 | 0.0558 | 0.0482 | 0.1333 | 0.1699 | 0.0482 | 0.1387 |
| 28 Zooplankton (large) | 0.2727 | 0.1616 | 0.1573 | 0.1398 | 0.1529 | 0.1984 | 0.0242 | 0.1354 | 0.4116 | 0.1904 | 0.1984 | 0.1813 | 0.1904 | 0.1926 |
| 29 Zooplankton (small) | 0.0000 | 0.0002 | 0.0006 | 0.0005 | 0.0004 | 0.0003 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0003 | 0.0002 | 0.0000 | 0.0002 |
| 30 Phytoplankton | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 31 Detritus | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| No. of stomachs |  | 411 | 353 | 238 |  |  | 284 | 163 | 320 |  |  |  |  |  |

Diet 3. From Table 2a.
Diet 6. Division 2J3K autumn diet. Average of 1985, 1986 and 1987. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small). Diet 7. Division 2J3K total diet. The offshore autumn and inshore diets are given weightings of 0.62 and 0.38 .
This reflects the following weightings for population size (first number) and feeding rate (second number) for the offshore in winter, spring, summer and autumn
and the inshore for winter, spring, summer and autumn: ( $0.5^{*} 0$ ), ( $0.5^{*} 1$ ), ( $\left.0.25^{*} 1\right)$, ( $0.5^{*} 1$ ), ( $\left.0.5^{*} 0\right),\left(0.5^{*} 1\right),\left(0.75^{*} 2\right),\left(0.5^{*} 1\right)$.
It is assumed that there is no feeding in winter, that the summer inshore diet is applicable for spring inshore and autumn inshore.
and that the offshore autumn diet is also appropriate for offshore spring and offshore summer.
Diet 8. Division 3NO spring diet. It is assumed that this diet is appropriate for all seasons. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small).
Diet 9. Average of 2J3K, 3L, and 3NO diets, weighted by population sizes. The weightings are $0.62,0.31$, and 0.07 . This is based on 0.93 of the average biomass being in 2 J 3 KL , and the assumption that 0.33 of the 2 J 3 KL biomass is in 3 L .
NB For all diets, the two small pelagic groups were originally one. Here they are split on the assumption that $75 \%$ are planktivorous feeders and $25 \%$ are piscivorous feeders.

Appendix 2. Table 3. Diet of large ( $>40 \mathrm{~cm}$ ) Greenland halibut in Divisions 2J3K, 3L, and 2J3KL together.

| Common Name | 2J3K autumn |  |  |  |  |  | 3L 1985 |  |  |  |  | 2J3KL total Diet 3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1985 | 1986 | 1987 | Diet 1 | Diet 1.1 | Diet 1.2 | winter | spring | summer | autumn |  |  |
| 1 Whales | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 Harp seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 3 Hooded seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 4 Seabirds | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 5 Cod > 35 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.0695 | 0.1166 | 0.0646 | 0.0836 | 0.0695 | 0.081 | 0.0548 | 0.0432 | 0.0000 | 0.0215 | 0.0299 | 0.0764 |
| 7 Greenland halibut > 40 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 8 Greenland halibut <= 40 cm | 0.1176 | 0.1111 | 0.2910 | 0.1732 | 0.1176 | 0.137 | 0.0485 | 0.0000 | 0.0521 | 0.0011 | 0.0254 | 0.1269 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 10 American plaice <= 35 cm | 0.0139 | 0.0000 | 0.0000 | 0.0046 | 0.0000 | 0 | 0.0000 | 0.0000 | 0.0424 | 0.0000 | 0.0106 | 0.0010 |
| 11 Flounders | 0.0023 | 0.0000 | 0.0048 | 0.0024 | 0.0023 | 0.003 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0024 |
| 12 Skates | 0.0037 | 0.0000 | 0.0008 | 0.0015 | 0.0008 | 0.001 | 0.0000 | 0.0000 | 0.0044 | 0.0000 | 0.0011 | 0.0010 |
| 13 Redfish | 0.2146 | 0.3718 | 0.2380 | 0.2748 | 0.2380 | 0.277 | 0.1412 | 0.0116 | 0.0064 | 0.0312 | 0.0476 | 0.2565 |
| 14 Demersal feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 15 Demersal feeders (misc. small) | 0.1001 | 0.0632 | 0.0743 | 0.0792 | 0.0743 | 0.087 | 0.2576 | 0.0118 | 0.0376 | 0.0104 | 0.0793 | 0.0859 |
| 16 Capelin | 0.4100 | 0.2976 | 0.2589 | 0.3222 | 0.2976 | 0.347 | 0.3956 | 0.8592 | 0.8314 | 0.9092 | 0.7489 | 0.3828 |
| 17 Sand lance | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 18 Arctic cod | 0.0333 | 0.0030 | 0.0246 | 0.0203 | 0.0246 | 0.029 | 0.0066 | 0.0239 | 0.0002 | 0.0000 | 0.0077 | 0.0267 |
| 19 Pelagic feeders (misc. large) | 0.0055 | 0.0004 | 0.0018 | 0.0025 | 0.0018 | 0.002 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0019 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.0035 | 0.0044 | 0.0063 | 0.0047 | 0.0044 | 0.0051 | 0.0209 | 0.0096 | 0.0020 | 0.0060 | 0.0096 | 0.0055 |
| 21 Plankt. Pelagic feeders (misc. small) | 0.0105 | 0.0131 | 0.0190 | 0.0142 | 0.0131 | 0.0153 | 0.0627 | 0.0287 | 0.0061 | 0.0180 | 0.0289 | 0.0165 |
| 22 Shrimp | 0.0113 | 0.0124 | 0.0119 | 0.0119 | 0.0119 | 0.014 | 0.0103 | 0.0090 | 0.0097 | 0.0016 | 0.0077 | 0.0133 |
| 23 Large Crustaceans (crab, lobster) | 0.0000 | 0.0020 | 0.0000 | 0.0007 | 0.0000 | 0 | 0.0000 | 0.0004 | 0.0000 | 0.0000 | 0.0001 | 0.0000 |
| 24 Echinoderms | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 25 Molluscs (benthic) | 0.0000 | 0.0012 | 0.0000 | 0.0004 | 0.0000 | 0.000 | 0.0000 | 0.0000 | 0.0006 | 0.0000 | 0.0002 | 0.0000 |
| 26 Polychaetes | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 27 Benthic invertebrates (other) | 0.0004 | 0.0004 | 0.0025 | 0.0011 | 0.0004 | 0.001 | 0.0015 | 0.0024 | 0.0044 | 0.0007 | 0.0022 | 0.0007 |
| 28 Zooplankton (large) | 0.0037 | 0.0029 | 0.0014 | 0.0027 | 0.0029 | 0.003 | 0.0003 | 0.0001 | 0.0027 | 0.0003 | 0.0009 | 0.0031 |
| 29 Zooplankton (small) | 0.0000 | 0.0000 | 0.0001 | 0.0000 | 0.0000 | 0.000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 30 Phytoplankton | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 31 Detritus | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |

Diet 1. Division 2J3K autumn diet. Average of 1985, 1986 and 1987. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small). Diet 1.1 Division 2J3K autumn diet expressed as the median of 1985, 1986 and 1987.
Diet 1.2 Diet 1.1 normalised
Diet 2. Division 3L diet. Average of the four seasons. Sampling went only to 366 m . Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small).
Diet 3. Division 2J3KL diet. The 2J3K autumn diet is assumed to be appropriate for all seasons. The 2J3K autumn diet (Diet 1.2) and the 3L
annual diet (Diet 2) are averaged with weightings of 0.91 and 0.09 to reflect the relative biomasses in 2 J 3 K and 3 L as determined during
autumn research surveys (Fig. 14 of Bowering and Brodie, 1995).
NB For all diets, the two small pelagic groups were originally one. Here they are split on the assumption that $75 \%$ are planktivorous feeders and $25 \%$ are piscivorous feeders.

Appendix 2. Table 4. Diet of small ( $<=40 \mathrm{~cm}$ ) Greenland halibut in Divisions 2J3K, 3L, and 2J3KL together.

| Common Name | 2J3K autumn |  |  |  | 3L 1985 |  |  |  |  | 2J3KL total Diet 3 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1985 | 1986 | 1987 | Diet 1 | winter | spring | summer | autumn |  |  |
| 1 Whales | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 Harp seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 3 Hooded seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 4 Seabirds | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $5 \mathrm{Cod}>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.0085 | 0.0149 | 0.0000 | 0.0078 | 0.0000 | 0.0733 | 0.0000 | 0.0000 | 0.0183 | 0.0087 |
| 7 Greenland halibut > 40 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 8 Greenland halibut <= 40 cm | 0.0041 | 0.0127 | 0.0013 | 0.0060 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0055 |
| 9 American plaice > 35 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 10 American plaice <= 35 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 11 Flounders | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 12 Skates | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 13 Redfish | 0.0001 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 14 Demersal feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 15 Demersal feeders (misc. small) | 0.0005 | 0.0155 | 0.0087 | 0.0082 | 0.0000 | 0.0026 | 0.0000 | 0.0000 | 0.0007 | 0.0076 |
| 16 Capelin | 0.8354 | 0.8279 | 0.8048 | 0.8227 | 0.9451 | 0.9030 | 0.9749 | 0.9604 | 0.9458 | 0.8338 |
| 17 Sand lance | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0089 | 0.0000 | 0.0000 | 0.0000 | 0.0022 | 0.0002 |
| 18 Arctic cod | 0.0916 | 0.0303 | 0.0404 | 0.0541 | 0.0078 | 0.0124 | 0.0035 | 0.0000 | 0.0059 | 0.0498 |
| 19 Pelagic feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.0077 | 0.0114 | 0.0125 | 0.0105 | 0.0056 | 0.0000 | 0.0000 | 0.0079 | 0.0034 | 0.0099 |
| 21 Plankt. Pelagic feeders (misc. sma | 0.0231 | 0.0341 | 0.0376 | 0.0316 | 0.0168 | 0.0000 | 0.0000 | 0.0238 | 0.0101 | 0.0297 |
| 22 Shrimp | 0.0182 | 0.0308 | 0.0223 | 0.0238 | 0.0050 | 0.0070 | 0.0093 | 0.0042 | 0.0064 | 0.0222 |
| 23 Large Crustaceans (crab, lobster) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 24 Echinoderms | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0032 | 0.0000 | 0.0000 | 0.0000 | 0.0008 | 0.0001 |
| 25 Molluscs (benthic) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 26 Polychaetes | 0.0000 | 0.0001 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0001 | 0.0000 | 0.0000 | 0.0000 |
| 27 Benthic invertebrates (other) | 0.0030 | 0.0060 | 0.0017 | 0.0036 | 0.0007 | 0.0004 | 0.0066 | 0.0030 | 0.0027 | 0.0035 |
| 28 Zooplankton (large) | 0.0079 | 0.0164 | 0.0707 | 0.0317 | 0.0068 | 0.0014 | 0.0057 | 0.0008 | 0.0037 | 0.0291 |
| 29 Zooplankton (small) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 30 Phytoplankton | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 31 Detritus | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |

Diet 1. Division 2J3K autumn diet. Average of 1985, 1986 and 1987. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small).
Diet 2. Division 3L diet. Average of the four seasons. Sampling went only to 366 m .
Diet 3. Division 2J3KL diet. The 2J3K autumn diet is assumed to be appropriate for all seasons. The 2J3K autumn diet
(Diet 1) and the 3L annual diet (Diet 2) are averaged with weightings of 0.91 and 0.09 to reflect the relative biomasses in 2 J 3 K and 3L as determined during autumn research surveys (Fig. 14 of Bowering and Brodie, 1995).
NB For all diets, the two small pelagic groups were originally one. Here they are split on the assumption that $75 \%$ of small pelagics in the diet are planktivorous feeders and $25 \%$ are piscivorous feeders.

Appendix 2. Table 5a. Diet of large ( $>35 \mathrm{~cm}$ ) American plaice in Divisions 2J3K and 3L.

| Common Name | 2J3K |  |  |  |  | 3L |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 J autumn |  |  | $\frac{3 \mathrm{~K} \text { aut. }}{\text { pl3k78au }}$ | total <br> Diet 2 | 3L 1985 |  |  |  | total Diet 3 |
|  | 1977 | 1978 | Diet 1 |  |  | winter | spring | summer | autumn |  |
| 1 Whales | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 Harp seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 3 Hooded seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 4 Seabirds | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $5 \mathrm{Cod}>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.0000 | 0.0024 | 0.0012 | 0.0000 | 0.0006 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 7 Greenland halibut > 40 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 8 Greenland halibut <= 40 cm | 0.1558 | 0.0065 | 0.0812 | 0.0000 | 0.0406 | 0.0000 | 0.0000 | 0.0011 | 0.0007 | 0.0005 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 10 American plaice <= 35 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0063 | 0.0017 |
| 11 Flounders | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0006 | 0.0002 |
| 12 Skates | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 13 Redfish | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 14 Demersal feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 15 Demersal feeders (misc. small) | 0.0377 | 0.0094 | 0.0235 | 0.0000 | 0.0118 | 0.0352 | 0.0034 | 0.0024 | 0.0053 | 0.0090 |
| 16 Capelin | 0.6729 | 0.0000 | 0.3364 | 0.0000 | 0.1682 | 0.4596 | 0.4141 | 0.5446 | 0.4969 | 0.4872 |
| 17 Sand lance | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0849 | 0.1430 | 0.1367 | 0.0126 | 0.0959 |
| 18 Arctic cod | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0042 | 0.0011 |
| 19 Pelagic feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 20 Pisc. Pelagic feeders (misc. small | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0004 | 0.0000 | 0.0001 |
| 21 Plankt. Pelagic feeders (misc. smi | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0011 | 0.0000 | 0.0004 |
| 22 Shrimp | 0.0046 | 0.0000 | 0.0023 | 0.0000 | 0.0012 | 0.0104 | 0.0007 | 0.0023 | 0.0077 | 0.0048 |
| 23 Large Crustaceans (crab, lobster) | 0.0000 | 0.0007 | 0.0003 | 0.0000 | 0.0002 | 0.0160 | 0.1838 | 0.0124 | 0.0018 | 0.0496 |
| 24 Echinoderms | 0.0888 | 0.5045 | 0.2966 | 0.6654 | 0.4810 | 0.3118 | 0.1509 | 0.1995 | 0.3175 | 0.2393 |
| 25 Molluscs (benthic) | 0.0000 | 0.1441 | 0.0720 | 0.0000 | 0.0360 | 0.0056 | 0.0424 | 0.0614 | 0.0437 | 0.0428 |
| 26 Polychaetes | 0.0206 | 0.2688 | 0.1447 | 0.0059 | 0.0753 | 0.0107 | 0.0028 | 0.0125 | 0.0105 | 0.0094 |
| 27 Benthic invertebrates (other) | 0.0196 | 0.0636 | 0.0416 | 0.3287 | 0.1852 | 0.0483 | 0.0431 | 0.0187 | 0.0562 | 0.0395 |
| 28 Zooplankton (large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0177 | 0.0158 | 0.0069 | 0.0359 | 0.0186 |
| 29 Zooplankton (small) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 30 Phytoplankton | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 31 Detritus | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| No. of stomachs | 60 | 77 |  | 42 |  | 1030 | 2001 | 1589 | 1765 |  |

Diet 1. Division 2 J autumn diet. Average of 1977 and 1978. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small). Diet 2. Division 2J3K autumn diet. Average of 2J diet (Diet 1) and 3K diet. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small).
Diet 3. Division 3L diet. Average of 4 seasons, weighted by $0.17,0.23,0.33,0.27$ (winter to autumn). Each weighting is the average of weightings calculated for large and small fish, where the weightings are ratios of the square roots of seasonal average total stomach fullness indices.
Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small).
The fullness indices were calculated as described in Lilly (1991) and Bowering and Lilly (1992).
NB For all diets, the two small pelagic groups were originally one. Here they are split on the assumption that $75 \%$ are planktivorous feeders and $25 \%$ are piscivorous feeders

Appendix 2. Table 5b. Diet of large ( $>35 \mathrm{~cm}$ ) American plaice in Divisions 3NO and 2J3KLNO together.

| Common Name | 3NO |  |  |  |  |  |  | $\begin{gathered} 3 \mathrm{NO} \\ \text { total } \\ \text { (Diet 5) } \end{gathered}$ | $\begin{gathered} \text { 2J3KLNO } \\ \text { total } \end{gathered}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3 NO spr | 3NO summer |  |  | total Diet 5 |  |  |  |  |  |
|  | 1984 | 1985 | 1986 | Diet 4 |  |  |  |  | Diet 6 | Diet 7 |
| 1 Whales | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 Harp seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 3 Hooded seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 4 Seabirds | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $5 \mathrm{Cod}>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0006 | 0.0003 | 0.0002 | 0.0006 | 0.0000 | 0.0002 | 0.0001 | 0.0001 |
| 7 Greenland halibut > 40 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 8 Greenland halibut <= 40 cm | 0.0000 | 0.0020 | 0.0000 | 0.0010 | 0.0005 | 0.0406 | 0.0005 | 0.0005 | 0.0053 | 0.0010 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 10 American plaice <= 35 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0017 | 0.0000 | 0.0009 | 0.0009 |
| 11 Flounders | 0.0000 | 0.0000 | 0.0016 | 0.0008 | 0.0004 | 0.0000 | 0.0002 | 0.0004 | 0.0002 | 0.0002 |
| 12 Skates | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 13 Redfish | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 14 Demersal feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 15 Demersal feeders (misc. small) | 0.0005 | 0.0008 | 0.0342 | 0.0175 | 0.0090 | 0.0118 | 0.0090 | 0.0090 | 0.0093 | 0.0094 |
| 16 Capelin | 0.0924 | 0.0259 | 0.0000 | 0.0129 | 0.0527 | 0.1682 | 0.4872 | 0.0527 | 0.2960 | 0.2973 |
| 17 Sand lance | 0.3155 | 0.3640 | 0.3389 | 0.3514 | 0.3334 | 0.0000 | 0.0959 | 0.3334 | 0.1680 | 0.1687 |
| 18 Arctic cod | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0011 | 0.0000 | 0.0006 | 0.0006 |
| 19 Pelagic feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 20 Pisc. Pelagic feeders (misc. small | 0.0000 | 0.0000 | 0.0010 | 0.0005 | 0.0003 | 0.0000 | 0.0001 | 0.0003 | 0.0002 | 0.0002 |
| 21 Plankt. Pelagic feeders (misc. sma | 0.0000 | 0.0000 | 0.0031 | 0.0016 | 0.0008 | 0.0000 | 0.0004 | 0.0008 | 0.0005 | 0.0005 |
| 22 Shrimp | 0.0000 | 0.0010 | 0.0012 | 0.0011 | 0.0005 | 0.0012 | 0.0048 | 0.0005 | 0.0028 | 0.0029 |
| 23 Large Crustaceans (crab, lobster) | 0.0946 | 0.0112 | 0.0061 | 0.0086 | 0.0516 | 0.0002 | 0.0496 | 0.0516 | 0.0444 | 0.0446 |
| 24 Echinoderms | 0.3076 | 0.3102 | 0.3506 | 0.3304 | 0.3190 | 0.4810 | 0.2393 | 0.3190 | 0.2964 | 0.2976 |
| 25 Molluscs (benthic) | 0.1185 | 0.2276 | 0.1441 | 0.1858 | 0.1522 | 0.0360 | 0.0428 | 0.1522 | 0.0805 | 0.0808 |
| 26 Polychaetes | 0.0075 | 0.0023 | 0.0104 | 0.0063 | 0.0069 | 0.0753 | 0.0094 | 0.0069 | 0.0164 | 0.0165 |
| 27 Benthic invertebrates (other) | 0.0238 | 0.0430 | 0.0786 | 0.0608 | 0.0423 | 0.1852 | 0.0395 | 0.0423 | 0.0579 | 0.0582 |
| 28 Zooplankton (large) | 0.0397 | 0.0121 | 0.0297 | 0.0209 | 0.0303 | 0.0000 | 0.0186 | 0.0303 | 0.0205 | 0.0206 |
| 29 Zooplankton (small) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 30 Phytoplankton | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 31 Detritus | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| No. of stomachs | 1034 | 637 | 847 |  |  |  |  |  |  |  |

Diet 4. Division 3NO summer diet. Average of 1985 and 1986. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small).
Diet 5. Division 3NO average diet. Average of spring diet and summer diet (Diet 4).
Diet 6. Division 2J3KLNO diet. Average of the 2J3K, 3L and 3NO diets, weighted by $0.12,0.53$ and 0.35 , which are the relative biomasses in these areas.
Diet 7. Arbitrary adjustment to Diet 6 . The value of 0.0053 for Greenland halibut $<=40 \mathrm{~cm}$ seems high. It derives from a very high value
( 0.1558 ) in the 2 J autumn 1997 sample. The sample size was very small. A high value for Greenland halibut was also recorded for cod
stomachs sampled at the same time (Lilly 1984). This value was clearly anomalous compared with other years (Lilly 1984, 1991). For the
American plaice diet, the value of Greenland halibut $<=40 \mathrm{~cm}$ is reduced to 0.001 , and the values for all other boxes are increased
proportionally to bring the total to 1 .
NB For all diets, the two small pelagic groups were originally one. Here they are split on the assumption that $75 \%$ are planktivorous feeders and $25 \%$ are piscivorous feeders.

Appendix 2. Table 6a. Diet of small $(<=35 \mathrm{~cm})$ American plaice in Divisions 2J3K and 3L.

| Common Name | 2J3K |  |  |  |  | 3L |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 2 J autumn |  |  | $\frac{3 \text { K aut. }}{1978}$ | total <br> Diet 2 | 3L 1985 |  |  |  | total Diet 3 |
|  | 1977 | 1978 | Diet 1 |  |  | winter | spring | summer | autumn |  |
| 1 Whales | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 Harp seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 3 Hooded seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 4 Seabirds | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $5 \mathrm{Cod}>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 7 Greenland halibut > 40 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 8 Greenland halibut <= 40 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0144 | 0.0039 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 10 American plaice <= 35 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0003 | 0.0006 | 0.0221 | 0.0062 |
| 11 Flounders | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 12 Skates | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 13 Redfish | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0061 | 0.0016 |
| 14 Demersal feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 15 Demersal feeders (misc. small) | 0.0151 | 0.0315 | 0.0233 | 0.0000 | 0.0117 | 0.0154 | 0.0096 | 0.0041 | 0.0066 | 0.0079 |
| 16 Capelin | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.6925 | 0.4528 | 0.6885 | 0.6034 | 0.6120 |
| 17 Sand lance | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0402 | 0.1465 | 0.0448 | 0.0078 | 0.0574 |
| 18 Arctic cod | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0056 | 0.0000 | 0.0236 | 0.0077 |
| 19 Pelagic feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 21 Plankt. Pelagic feeders (misc. sma | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 22 Shrimp | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0718 | 0.0094 | 0.0135 | 0.0222 | 0.0248 |
| 23 Large Crustaceans (crab, lobster) | 0.0152 | 0.0000 | 0.0076 | 0.0000 | 0.0038 | 0.0083 | 0.0620 | 0.0018 | 0.0012 | 0.0166 |
| 24 Echinoderms | 0.2879 | 0.1171 | 0.2025 | 0.4909 | 0.3467 | 0.0289 | 0.0626 | 0.0701 | 0.0816 | 0.0644 |
| 25 Molluscs (benthic) | 0.0152 | 0.0000 | 0.0076 | 0.0000 | 0.0038 | 0.0041 | 0.0190 | 0.0228 | 0.0086 | 0.0149 |
| 26 Polychaetes | 0.3030 | 0.1577 | 0.2303 | 0.3818 | 0.3061 | 0.0897 | 0.0674 | 0.0655 | 0.0483 | 0.0654 |
| 27 Benthic invertebrates (other) | 0.3636 | 0.6937 | 0.5287 | 0.1273 | 0.3280 | 0.0369 | 0.1529 | 0.0734 | 0.0773 | 0.0865 |
| 28 Zooplankton (large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0122 | 0.0118 | 0.0150 | 0.0768 | 0.0305 |
| 29 Zooplankton (small) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0002 | 0.0000 | 0.0000 | 0.0001 |
| 30 Phytoplankton | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 31 Detritus | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |


| No. of stomachs | 81 | 49 | 48 | 1003 | 1670 | 1249 | 1737 |
| :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |

Diet 1. Division 2J autumn diet. Average of 1977 and 1978.
Diet 2. Division 2 J 3 K autumn diet. Average of 2 J diet (Diet 1 ) and 3 K diet.
Diet 3. Division 3L diet. Average of 4 seasons, weighted by $0.17,0.23,0.33,0.27$ (winter to autumn). Each weighting is the average of weightings calculated for large and small fish, where the weightings are ratios of the square roots of seasonal average total stomach fullness indices.
Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small).
The fullness indices were calculated as described in Lilly (1991) and Bowering and Lilly (1992).
NB For all diets, the two small pelagic groups were originally one. Here they are split on the assumption that $75 \%$ are planktivorous feeders and $25 \%$ are piscivorous feeders.

Appendix 2. Table 6b. Diet of small ( $<=35 \mathrm{~cm}$ ) American plaice in Divisions 3NO and 2J3KLNO together.

| Common Name | 3NO |  |  |  |  | 2J3K total (Diet 2) |  |  | $\begin{gathered} \begin{array}{c} 2 \mathrm{~J} 3 \mathrm{KLNO} \\ \text { total } \end{array} \\ \hline \text { Diet } 6 \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 3NO spr | 3NO summer |  |  | total Diet 5 |  |  |  |  |
|  | 1984 | 1985 | 1986 | Diet 4 |  |  |  |  |  |
| 1 Whales | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 2 Harp seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 3 Hooded seals | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 4 Seabirds | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $5 \mathrm{Cod}>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0538 | 0.0269 | 0.0134 | 0.0000 | 0.0000 | 0.0134 | 0.0043 |
| 7 Greenland halibut > 40 cm | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 8 Greenland halibut $<=40 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0039 | 0.0000 | 0.0019 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 10 American plaice $<=35 \mathrm{~cm}$ | 0.0000 | 0.2119 | 0.0112 | 0.1115 | 0.0558 | 0.0000 | 0.0062 | 0.0558 | 0.0209 |
| 11 Flounders | 0.0000 | 0.0000 | 0.0980 | 0.0490 | 0.0245 | 0.0000 | 0.0000 | 0.0245 | 0.0079 |
| 12 Skates | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 13 Redfish | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0016 | 0.0000 | 0.0008 |
| 14 Demersal feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 15 Demersal feeders (misc. small) | 0.0167 | 0.0786 | 0.0947 | 0.0866 | 0.0517 | 0.0117 | 0.0079 | 0.0517 | 0.0227 |
| 16 Capelin | 0.2175 | 0.0515 | 0.0000 | 0.0258 | 0.1216 | 0.0000 | 0.6120 | 0.1216 | 0.3334 |
| 17 Sand lance | 0.1312 | 0.3728 | 0.1478 | 0.2603 | 0.1958 | 0.0000 | 0.0574 | 0.1958 | 0.0905 |
| 18 Arctic cod | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0077 | 0.0000 | 0.0037 |
| 19 Pelagic feeders (misc. large) | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.0001 | 0.0000 | 0.0000 | 0.0000 | 0.0001 | 0.0000 | 0.0000 | 0.0001 | 0.0000 |
| 21 Plankt. Pelagic feeders (misc. smal | 0.0004 | 0.0000 | 0.0000 | 0.0000 | 0.0002 | 0.0000 | 0.0000 | 0.0002 | 0.0001 |
| 22 Shrimp | 0.0024 | 0.0005 | 0.0041 | 0.0023 | 0.0024 | 0.0000 | 0.0248 | 0.0024 | 0.0127 |
| 23 Large Crustaceans (crab, lobster) | 0.0705 | 0.0005 | 0.0166 | 0.0086 | 0.0395 | 0.0038 | 0.0166 | 0.0395 | 0.0214 |
| 24 Echinoderms | 0.0591 | 0.0005 | 0.0330 | 0.0167 | 0.0379 | 0.3467 | 0.0644 | 0.0379 | 0.1118 |
| 25 Molluscs (benthic) | 0.0195 | 0.1554 | 0.0940 | 0.1247 | 0.0721 | 0.0038 | 0.0149 | 0.0721 | 0.0311 |
| 26 Polychaetes | 0.0636 | 0.0192 | 0.0847 | 0.0519 | 0.0577 | 0.3061 | 0.0654 | 0.0577 | 0.1106 |
| 27 Benthic invertebrates (other) | 0.0431 | 0.1001 | 0.1547 | 0.1274 | 0.0852 | 0.3280 | 0.0865 | 0.0852 | 0.1339 |
| 28 Zooplankton (large) | 0.3758 | 0.0091 | 0.2075 | 0.1083 | 0.2420 | 0.0000 | 0.0305 | 0.2420 | 0.0923 |
| 29 Zooplankton (small) | 0.0002 | 0.0000 | 0.0000 | 0.0000 | 0.0001 | 0.0000 | 0.0001 | 0.0001 | 0.0001 |
| 30 Phytoplankton | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| 31 Detritus | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 | 0.0000 |
| No. of stomachs | 506 | 325 | 617 |  |  |  |  |  |  |

Diet 4. Division 3 NO summer diet. Average of 1985 and 1986. Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small).
Diet 5. Division 3NO average diet. Average of spring diet and summer diet (Diet 4). Demersal feeders (large) are considered to be juveniles and classified as demersal feeders (small).
Diet 6. Division 2 J 3 KLNO diet. Average of the $2 \mathrm{~J} 3 \mathrm{~K}, 3 \mathrm{~L}$ and 3 NO diets, weighted by $0.20,0.48$ and 0.32 , which are the relative
biomasses in these areas.
NB For all diets, the two small pelagic groups were originally one. Here they are split on the assumption that $75 \%$ are planktivorous feeders and $25 \%$ are piscivorous feeders.

Appendix 2. Table 7a. Diet of flounders (winter flounder).

|  | Winter flounder |  |  |  |  |  |
| :--- | :---: | ---: | ---: | ---: | ---: | ---: |
|  | Diet 1 | Diet 1.1 | Diet 1.2 | Diet 2 | Diet 2.1 | Diet 3 |
| 1 Whales | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 Harp seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 Hooded seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 4 Seabirds | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 5 Cod > 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 6 Cod <= 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 7 Greenland halibut > 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 8 Greenland halibut <= 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 9 American plaice > 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 10 American plaice <= 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 11 Flounders | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 12 Skates | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 13 Redfish | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 14 Demersal feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 15 Demersal feeders (misc. small) | 0.000 | 0.000 | 0.000 | 0.077 | 0.118 | 0.071 |
| 16 Capelin | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 17 Sand lance | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 18 Arctic cod | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 19 Pelagic feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 21 Plankt. Pelagic feeders (misc. small) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 22 Shrimp | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 23 Large Crustaceans (crab, lobster) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 24 Echinoderms | 0.097 | 0.244 | 0.244 | 0.005 | 0.008 | 0.102 |
| 25 Molluscs (benthic) | 0.085 | 0.215 | 0.215 | 0.068 | 0.104 | 0.149 |
| 26 Polychaetes | 0.000 | 0.000 | 0.000 | 0.143 | 0.220 | 0.132 |
| 27 Benthic invertebrates (other) | 0.214 | 0.539 | 0.541 | 0.358 | 0.550 | 0.546 |
| 28 Zooplankton (large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 29 Zooplankton (small) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 30 Phytoplankton | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 31 Detritus | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| Total |  |  |  |  |  |  |

Diet 1 is from Table 1 of Keats (1990). Data (\% weight) are from fish collected in June-August 1983 at 5-15 m in Conception Bay, primarily from rocky bottom.
Diet 1 includes benthic fish eggs (0.0697) which make up a portion of the benthic invertebrate (other) catagory.
Diet 1 does not add up to 1.0 because unidentified material including fleshy algae (Total=0.602) was not included.
Diet $1.1=(\text { Diet } 1)^{*}((0.3958+0.602) / 0.3958)$. i.e. the unidentified material (including fleshy algae) is distributed over all the other prey catagories.
Diet 1.2 is Diet 1.1 standardized to 1.0
Diet 2 is from Appendix Table 1 of Kennedy and Steele (1971). Data (\% volume) are from fish collected at all seasons in 1962-1963 at 1-2 m in a small inlet in Conception Bay, mainly from muddy sand.
The benthic invertebrates (other) is dominated ( 0.294 of 0.358 ) by eggs of fish (mainly capelin). In addition, 0.153 of the stomach contents was fish remains. It is not clear if this is natural food or offal.

As a guess, it is assumed that half of this was natural food, which is assigned to small demersal feeders. Diet 2.1 is Diet 2 standardized to 1 .
Diet 3 is an average of Diets 1.2 and 2.1, with arbitrary weightings of 0.4 and 0.6. The weightings reflect the small sample size and limited seasonal coverage of Diet 1.

Appendix 2. Table 7b. Diet of flounders (yellowtail, witch and all three species combined).

|  | Winter |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Yellowtail |  | Witch | flounder | Flounders |  |
|  | Diet 4 | Diet 4.1 | Diet 4.2 | Diet 5 | (Diet 3) | Diet 6 |
| 1 Whales | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 Harp seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 Hooded seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 4 Seabirds | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 5 Cod > 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 6 Cod <= 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 7 Greenland halibut > 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 8 Greenland halibut <= 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 9 American plaice > 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 10 American plaice <= 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 11 Flounders | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 12 Skates | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 13 Redfish | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 14 Demersal feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 15 Demersal feeders (misc. small) | 0.000 | 0.000 | 0.000 | 0.009 | 0.071 | 0.007 |
| 16 Capelin | 0.037 | 0.038 | 0.039 | 0.000 | 0.000 | 0.025 |
| 17 Sand lance | 0.038 | 0.040 | 0.040 | 0.000 | 0.000 | 0.025 |
| 18 Arctic cod | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 19 Pelagic feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 21 Plankt. Pelagic feeders (misc. small | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 22 Shrimp | 0.000 | 0.000 | 0.000 | 0.021 | 0.000 | 0.007 |
| 23 Large Crustaceans (crab, lobster) | 0.000 | 0.000 | 0.000 | 0.001 | 0.000 | 0.000 |
| 24 Echinoderms | 0.069 | 0.072 | 0.073 | 0.006 | 0.102 | 0.053 |
| 25 Molluscs (benthic) | 0.028 | 0.029 | 0.030 | 0.024 | 0.149 | 0.035 |
| 26 Polychaetes | 0.380 | 0.395 | 0.404 | 0.660 | 0.132 | 0.470 |
| 27 Benthic invertebrates (other) | 0.348 | 0.362 | 0.370 | 0.279 | 0.546 | 0.351 |
| 28 Zooplankton (large) | 0.040 | 0.042 | 0.043 | 0.001 | 0.000 | 0.027 |
| 29 Zooplankton (small) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 30 Phytoplankton | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 31 Detritus | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| Total |  |  |  |  |  |  |
|  | 0.940 | 0.978 | 1.000 | 1.000 |  | 1.000 |

Diet 4 is from Table 2 of Pitt (1976). These \% weight data are from $10-49 \mathrm{~cm}$ yellowtail collected in 3 L and 3 N in 1968-1973. Data may have been collected throughout the year as in Pitt (1973).
Diet 4.1 is (Diet 4$)^{*}((0.94+0.038) / 0.94)$. i.e. unidentified material and plants, stones etc. were distributed proportionally over all prey catagories.
Diet 4.2 is Diet 4.1 standardized to 1.0
Diet 5 is from Table 4 of Rae (1969). Since the author says that sampling was proportional to the size composition of the catch, the prey volumes were summed across predator size groups prior to calculation of proportions. The crustacea group included some decapods. The proportion of these was estimated from the frequency of occurrence data in Table $5\left[(16.6 / 76.7)^{*} 0.199=0.043\right]$. Of this, it was assumed that half (i.e. 0.021 ) was shrimp and 0.001 was large crustacea. The remainer ( 0.177 ) was placed in the other benthic invertebrates.
Diet 6 is the average for the three species, weighted by their respective biomasses ( $0.624: 0.319: 0.057$ ).

Appendix 2. Table 8. Diet of thorny skate.

|  | Diet 1 | Diet 1.1 | Diet 1.2 | Diet 1.3 | Diet 1.4 |
| :--- | ---: | ---: | ---: | ---: | ---: |
| 1 Whales | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 Harp seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 Hooded seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 4 Seabirds | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 5 Cod > 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 6 Cod <= 35 cm | 0.041 | 0.041 | 0.049 | 0.062 | 0.050 |
| 7 Greenland halibut > 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 8 Greenland halibut <= 40 cm | 0.001 | 0.001 | 0.001 | 0.002 | 0.001 |
| 9 American plaice > 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 10 American plaice <= 35 cm | 0.001 | 0.001 | 0.001 | 0.002 | 0.001 |
| 11 Flounders | 0.006 | 0.006 | 0.007 | 0.009 | 0.007 |
| 12 Skates | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 13 Redfish | 0.176 | 0.179 | 0.216 | 0.108 | 0.086 |
| 14 Demersal feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 15 Demersal feeders (misc. small) | 0.196 | 0.200 | 0.240 | 0.139 | 0.111 |
| 16 Capelin | 0.021 | 0.021 | 0.026 | 0.100 | 0.080 |
| 17 Sand lance | 0.118 | 0.120 | 0.144 | 0.182 | 0.145 |
| 18 Arctic cod | 0.001 | 0.001 | 0.001 | 0.002 | 0.001 |
| 19 Pelagic feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.047 | 0.048 | 0.058 | 0.072 | 0.058 |
| 21 Plankt. Pelagic feeders (misc. small | 0.006 | 0.006 | 0.007 | 0.009 | 0.007 |
| 22 Shrimp | 0.010 | 0.010 | 0.012 | 0.015 | 0.022 |
| 23 Large Crustaceans (crab, lobster) | 0.143 | 0.143 | 0.172 | 0.216 | 0.311 |
| 24 Echinoderms | 0.002 | 0.002 | 0.002 | 0.003 | 0.004 |
| 25 Molluscs (benthic) | 0.007 | 0.007 | 0.008 | 0.011 | 0.015 |
| 26 Polychaetes | 0.037 | 0.037 | 0.045 | 0.056 | 0.080 |
| 27 Benthic invertebrates (other) | 0.006 | 0.006 | 0.007 | 0.009 | 0.013 |
| 28 Zooplankton (large) | 0.002 | 0.002 | 0.002 | 0.003 | 0.004 |
| 29 Zooplankton (small) | 0.000 | 0.000 | 0.000 | 0.001 | 0.001 |
| 30 Phytoplankton | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 31 Detritus | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| Total of above |  |  |  |  |  |
| Total of identified fish | 0.821 | 0.831 | 1.0000 | 0.9996 | 0.9996 |
|  | 0.614 |  | 0.751 | 0.685 | 0.548 |

DIET 1 is from Table 2 of Templeman (1982) (the final column, combining all depths (17-740 m) and all sizes of skates (21-102 cm). The stomachs were collected in 1947-1967 from Greenland to Georges Bank but mainly from 3KLNOP. The prey identified as fish offal are not included.
The value of 0.196 for Demersal Feeders (misc. small) comprises the assumed juveniles of the demersal feeders (large) (0.176) and the true demersal feeders (small) (0.020)
DIET 1.1 has the fish categories in DIET 1 adjusted to account for the unidentified fish in the diet (0.011). That is, Diet $1.1=(\operatorname{Diet} 1) *(0.614+0.011) / 0.614$, but only for the fish portion of the diet.
DIET $1.2=($ Diet 1.1)/ 0.832 , to adjust diet to 1.0 .
Diet 1.3 has arbitrary adjustments to Diet 1.2. It is thought that the relative abundance of small redfish and haddock was much lower in 1985-1987 than during the period when the stomachs were collected, so the diet proportions for redfish and demersal feeders large are reduced by half. The diet proportion for capelin seems very low, and is raised to 0.1. Zooplankton (small) is assigned 0.001.All other prey categories are adjusted upward proportionally ((Diet 1.2)*0.14/0.544) so that the categories sum to 1 . Diet 1.4-Diets 1-1.3 above are based on the diet data for larger thorny skates which eat more fish than smaller thorny skates which feed mainly on cephalopods, polychaetes and amphipods (Scott and Scott 1988). Other skate species are also less pisciverous than the thorny skate. To allow for this, the proportion of fish in the diet was reduced by $20 \%$ and distributed over the inverebrates groups by their relative contribution to the diet.

## Appendix 2. Table 9. Diet of redfish.

|  | Diet 1 | Diet 1.1 | Diet 1.2 | Diet 2 |
| :---: | :---: | :---: | :---: | :---: |
| 1 Whales |  |  | 0.000 | 0.000 |
| 2 Harp seals |  |  | 0.000 | 0.000 |
| 3 Hooded seals |  |  | 0.000 | 0.000 |
| 4 Seabirds |  |  | 0.000 | 0.000 |
| $5 \mathrm{Cod}>35 \mathrm{~cm}$ |  |  | 0.000 | 0.000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ |  | 0.007 | 0.007 | 0.002 |
| 7 Greenland halibut > 40 cm |  |  | 0.000 | 0.000 |
| 8 Greenland halibut <= 40 cm |  |  | 0.000 | 0.000 |
| 9 American plaice > 35 cm |  |  | 0.000 | 0.000 |
| 10 American plaice <= 35 cm |  |  | 0.000 | 0.000 |
| 11 Flounders |  |  | 0.000 | 0.000 |
| 12 Skates |  |  | 0.000 | 0.000 |
| 13 Redfish |  | 0.007 | 0.007 | 0.007 |
| 14 Demersal feeders (misc. large) |  |  | 0.000 | 0.000 |
| 15 Demersal feeders (misc. small) |  |  | 0.000 | 0.001 |
| 16 Capelin |  | 0.007 | 0.007 | 0.007 |
| 17 Sand lance |  |  | 0.000 | 0.004 |
| 18 Arctic cod |  |  | 0.000 | 0.000 |
| 19 Pelagic feeders (misc. large) |  |  | 0.000 | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) |  |  | 0.000 | 0.000 |
| 21 Plankt. Pelagic feeders (misc. smal | 0.012 | 0.233 | 0.245 | 0.245 |
| 22 Shrimp | 0.035 |  | 0.035 | 0.035 |
| 23 Large Crustaceans (crab, lobster) |  |  | 0.000 | 0.000 |
| 24 Echinoderms |  |  | 0.000 | 0.000 |
| 25 Molluscs (benthic) |  |  | 0.000 | 0.000 |
| 26 Polychaetes |  |  | 0.000 | 0.000 |
| 27 Benthic invertebrates (other) |  |  | 0.000 | 0.000 |
| 28 Zooplankton (large) | 0.538 |  | 0.538 | 0.538 |
| 29 Zooplankton (small) | 0.161 |  | 0.161 | 0.161 |
| 30 Phytoplankton |  |  | 0.000 | 0.000 |
| 31 Detritus |  |  | 0.000 | 0.000 |
| Total | 0.746 | 0.254 | 1.000 | 1.000 |

Diet 1 is from Tables $1 \& 2$ of Lambert (1960). The diet can only be constructed for the entire area sampled which is the waters all around Newfoundland including the Gulf of St. Lawrence. The invertebrate diet was copied directly from Table 2 into Diet 1. It does not sum to 1 because the fish are not included.
Diet 1.1 is the fish diet only. Nine species of fish make up the diet (Table 1). Each fish species is ranked by abundance from 1 to 3 . One is very rare, 2 is rare and 3 is common. It is assumed that the relative contribution to stomach contents is 1:3:9. There are a total of 35 abundance rankings distributed among the nine species. Therefore each ranking is $=0.254 / 35=0.0073$. Capelin, cod and redfish are eached ranked with one (i.e. each is 0.007 ). The other 6 species are mesopelagic and make up 0.232 . Diet 1.2 is the sum of Diet 1 and Diet 1.1 .

Diet 2 is a modification of Diet 1.2 based partly on guess and partly on Yanulov (1962) and Konchina (1986).
The proportion for cod $<=35 \mathrm{~cm}$ is reduced from 0.007 to 0.002 , and the balance of 0.005 is distributed among sand lance ( 0.004 ) and demersal feeders (small) ( 0.001 ), which are the juveniles of the demersal feeders (large).

Appendix 2. Table 10a. Diet of wolffish (a component of the large demersal feeders).

|  | Atlantic wolffish |  |  |  |  |  |  | Spotted | Northern |  | Average |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Diet 1 | Diet 1.1 | Diet 2 | Diet 2.1 | Diet 2.2 | Diet 3 | Diet 4 | Diet 5 | Diet 6 | Diet 7 | Diet 8 |
| 1 Whales | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 Harp seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 Hooded seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 4 Seabirds | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| $5 \mathrm{Cod}>35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.032 | 0.032 | 0.000 | 0.000 | 0.000 | 0.000 | 0.011 | 0.000 | 0.000 | 0.000 | 0.005 |
| 7 Greenland halibut > 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 8 Greenland halibut <= 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 10 American plaice $<=35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.019 | 0.021 | 0.021 | 0.000 | 0.007 | 0.000 | 0.000 | 0.000 | 0.003 |
| 11 Flounders | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 12 Skates | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.007 | 0.002 | 0.000 | 0.000 | 0.000 | 0.001 |
| 13 Redfish | 0.000 | 0.000 | 0.120 | 0.132 | 0.132 | 0.163 | 0.098 | 0.000 | 0.000 | 0.000 | 0.049 |
| 14 Demersal feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 15 Demersal feeders (misc. small) | 0.005 | 0.005 | 0.000 | 0.000 | 0.000 | 0.000 | 0.002 | 0.000 | 0.000 | 0.000 | 0.001 |
| 16 Capelin | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 17 Sand lance | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 18 Arctic cod | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 19 Pelagic feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 21 Plankt. Pelagic feeders (misc. sma | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 22 Shrimp | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.025 | 0.008 | 0.004 | 0.028 | 0.028 | 0.012 |
| 23 Large Crustaceans (crab, lobster) | 0.022 | 0.022 | 0.121 | 0.121 | 0.121 | 0.000 | 0.048 | 0.204 | 0.000 | 0.000 | 0.075 |
| 24 Echinoderms | 0.764 | 0.772 | 0.262 | 0.262 | 0.261 | 0.528 | 0.520 | 0.495 | 0.757 | 0.657 | 0.548 |
| 25 Molluscs (benthic) | 0.118 | 0.119 | 0.356 | 0.356 | 0.355 | 0.234 | 0.236 | 0.011 | 0.056 | 0.056 | 0.135 |
| 26 Polychaetes | 0.000 | 0.000 | 0.001 | 0.001 | 0.001 | 0.005 | 0.002 | 0.008 | 0.019 | 0.019 | 0.008 |
| 27 Benthic invertebrates (other) | 0.048 | 0.049 | 0.111 | 0.111 | 0.111 | 0.038 | 0.066 | 0.279 | 0.140 | 0.140 | 0.138 |
| 28 Zooplankton (large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.100 | 0.025 |
| 29 Zooplankton (small) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 30 Phytoplankton | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 31 Detritus | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| Total | 0.989 |  | 0.990 | 1.004 | 1.000 |  |  |  |  |  |  |
| Total of identified fish |  |  | 0.139 |  |  |  |  |  |  |  |  |

Diet 1 is from Table 1 of Keats et al. (1986). Data (\% weight) are from fish collected in 1983-1984 at 5-15 m, primarily from hard bottom dominated by sea urchins.
The value for cod is derived from one large cod in one wolffish. Demersal fish eggs are included in the benthic invertebrate (other) category.
Diet 1.1 is Diet 1 adjusted to a total of 1.0
Diet 2 is from Table 2 of Templeman (1984). Data (\% volume) are from fish collected in 1946-1966 from West Greenland to the Scotian Shelf.
Diet 2.1 is the unidentified fish (0.014) apportioned over the identified fish.
Diet 2.2 is Diet 2.1 standardized to 1.0.
Diet 3 is from Table 2 of Albikovskaya (1983). Data (fullness indices) were collected in 2J3KL in 1981
The two sources of data for each Division were averaged, and then the data from the 3 Divisions were averaged and converted to proportions.
Diet 4 is the average of the diets from the three studies of Atlantic wolfish.
Diets 5 and 6 are from Albikovskaya (1983), calculated as described for Diet 3.
Diet 7 is Diet 6 modified to give 0.1 to large zooplankton, since northern wolffish feed on ctenophores and medusae
Diet 8 is the average for the three species, weighted arbitrarily by $0.5,0.25$ and 0.25 .

Appendix 2. Table 10b. Diet of grenadier (a component of the large demersal feeders).

|  | Roughhead |  | Roundnose |  |  | Common |  | Average Diet 12 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Diet 9 | Diet 9.1 | Diet 10 D | iet 10.10 | iet 10.2 | Diet 11 | Diet 11.1 |  |
| 1 Whales | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 2 Harp seals | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 3 Hooded seals | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 4 Seabirds | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| $5 \mathrm{Cod}>35 \mathrm{~cm}$ | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 7 Greenland halibut > 40 cm | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 8 Greenland halibut < $=40 \mathrm{~cm}$ | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 10 American plaice $<=35 \mathrm{~cm}$ | 0.1 | 0.001 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 11 Flounders | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 12 Skates | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 13 Redfish | 0.5 | 0.004 | 0.024 | 0.030 | 0.036 | 0.0 | 0.000 | 0.016 |
| 14 Demersal feeders (misc. large) | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 15 Demersal feeders (misc. small) | 0.2 | 0.002 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.001 |
| 16 Capelin | 9.6 | 0.081 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.032 |
| 17 Sand lance | 4.7 | 0.040 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.016 |
| 18 Arctic cod | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 19 Pelagic feeders (misc. large) | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 21 Plankt. Pelagic feeders (misc. smal | 0.7 | 0.006 | 0.244 | 0.300 | 0.150 | 0.0 | 0.000 | 0.062 |
| 22 Shrimp | 8.9 | 0.075 | 0.387 | 0.387 | 0.477 | 13.3 | 0.154 | 0.252 |
| 23 Large Crustaceans (crab, lobster) | 1.3 | 0.011 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.004 |
| 24 Echinoderms | 30.6 | 0.259 | 0.000 | 0.000 | 0.000 | 7.1 | 0.082 | 0.120 |
| 25 Molluscs (benthic) | 11.7 | 0.099 | 0.000 | 0.000 | 0.000 | 9.1 | 0.105 | 0.061 |
| 26 Polychaetes | 19.0 | 0.161 | 0.005 | 0.005 | 0.006 | 25.7 | 0.297 | 0.126 |
| 27 Benthic invertebrates (other) | 22.4 | 0.190 | 0.057 | 0.057 | 0.070 | 25.8 | 0.298 | 0.164 |
| 28 Zooplankton (large) | 7.2 | 0.061 | 0.168 | 0.168 | 0.207 | 5.0 | 0.058 | 0.119 |
| 29 Zooplankton (small) | 1.3 | 0.011 | 0.042 | 0.042 | 0.052 | 0.5 | 0.006 | 0.026 |
| 30 Phytoplankton | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| 31 Detritus | 0.0 | 0.000 | 0.000 | 0.000 | 0.000 | 0.0 | 0.000 | 0.000 |
| Total | 118.2 | 1.000 | 0.927 | 0.989 | 0.999 | 86.5 | 1.000 | 1.000 |

Diet 9 is from Table 7 of Savvatimsky (1989a). Data (\% occurrence) are from 11-90 cm fish collected in 1969-1983 in the areas of Baffin Island, Labrador and Newfoundland. The value of 0.2 for the demersal feeders (small) is due to the juvenile large demersal feeders.
Diet 9.1 The frequencies were summed and the frequency for each prey category was expressed as a proportion of the total. Diet 10 is from Table 3 of Gushchin and Podrazhanskaya (1983). Data (\% weight) are provided for various areas.
The data here are averages of the values given for southern Labrador, northern Newfoundland and northeastern slope of Grand Bank. It is assumed that the Decapoda are all shrimp.
Diet 10.1 is Diet 1 with the fish components adjusted for the unidentified fish (0.062).
Diet 10.2 It is also assumed that the value for "pelagic feeders small" is much too high, since it is based almost entirely on an extremely high value for myctophids in a sample with a small sample size. The value for this group is arbitrarily reduced to 0.15 , and all other groups are adjusted to bring the total to 1.0
Diet 11 is from Table 5 of Savvatimsky (1989b). Data (\% occurrence) are from fish collected in 1969 and 1979 in 3LNOP. Diet 11.1 The frequencies were summed and the frequency for each prey category was expressed as a proportion of the total. Diet 12 is an average of the diets of the three species, with arbitrary weightings of $0.4,0.4$ and 0.2 , based somewhat on relative individual size.

Appendix 2. Table 10c. Diet of white hake and haddock (components of the large demersal feeders).

|  | white hake |  | haddock |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Diet 13 | Diet 13.1 | Diet 14 | Diet 14.1 | Diet 15 | Diet 15.1 | Diet 16 | Diet 17 |
| 1 Whales | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 Harp seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 Hooded seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 4 Seabirds | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 5 Cod > 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.062 | 0.090 | 0.003 | 0.004 | 0.003 | 0.004 | 0.004 | 0.004 |
| 7 Greenland halibut $>40 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 8 Greenland halibut <= 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 10 American plaice <= 35 cm | 0.062 | 0.090 | 0.020 | 0.024 | 0.000 | 0.000 | 0.012 | 0.012 |
| 11 Flounders | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 12 Skates | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 13 Redfish | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 14 Demersal feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 15 Demersal feeders (misc. small) | 0.186 | 0.270 | 0.005 | 0.006 | 0.014 | 0.017 | 0.012 | 0.012 |
| 16 Capelin | 0.062 | 0.090 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.090 |
| 17 Sand lance | 0.062 | 0.090 | 0.083 | 0.098 | 0.000 | 0.000 | 0.049 | 0.049 |
| 18 Arctic cod | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 19 Pelagic feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.011 | 0.016 | 0.002 | 0.002 | 0.000 | 0.000 | 0.001 | 0.001 |
| 21 Plankt. Pelagic feeders (misc. sma | 0.073 | 0.106 | 0.039 | 0.046 | 0.129 | 0.158 | 0.102 | 0.012 |
| 22 Shrimp | 0.027 | 0.039 | 0.078 | 0.092 | 0.015 | 0.018 | 0.055 | 0.055 |
| 23 Large Crustaceans (crab, lobster) | 0.010 | 0.015 | 0.031 | 0.037 | 0.004 | 0.005 | 0.021 | 0.021 |
| 24 Echinoderms | 0.008 | 0.012 | 0.210 | 0.248 | 0.299 | 0.367 | 0.308 | 0.308 |
| 25 Molluscs (benthic) | 0.007 | 0.010 | 0.058 | 0.068 | 0.031 | 0.038 | 0.053 | 0.053 |
| 26 Polychaetes | 0.005 | 0.007 | 0.168 | 0.198 | 0.176 | 0.216 | 0.207 | 0.207 |
| 27 Benthic invertebrates (other) | 0.000 | 0.000 | 0.116 | 0.137 | 0.116 | 0.143 | 0.140 | 0.140 |
| 28 Zooplankton (large) | 0.114 | 0.165 | 0.034 | 0.040 | 0.027 | 0.033 | 0.037 | 0.037 |
| 29 Zooplankton (small) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 30 Phytoplankton | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 31 Detritus | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| Total | 0.689 | 1.000 | 0.847 | 1.000 | 0.814 | 1.000 | 1.000 | 1.000 |

Diet 13 is from Table 2 of Petrov (1973). Data (\% occurrence) are from fish collected in 1969-1972 along the southwest slope of Grand Bank in Division 30. The frequency of occurrence values for the three months in the table were averaged. The fish component was assigned equally to the 8 species or groups indicated in the text, viz. cod, haddock, flatfish (assumed to be American plaice), sand lance, capelin, myctophids, grenadiers, and gobies.
The demersal feeders (small) are comprised of 0.124 juvenile demersal feeders (large) and 0.062 demersal feeders (small).
Diet 13.1 The frequencies were summed and the frequency for each prey category was expressed as a proportion of the total. Diet 14 is from Table 2 of Kohler and Fitzgerald (1969). Data (percentage volume) are from fish collected in Jan.-Apr. 1959-1962 on the Scotian Shelf (Division 4W). The values for each size group and depth range were averaged without weighting. Unidentified material was assigned as seemed appropriate.
The demersal feeders (small) are comprised of 0.002 juvenile demersal feeders (large) and 0.003 demersal feeders (small). Diet 14.1 The frequencies were summed and the frequency for each prey category was expressed as a proportion of the total. Diet 15 is from Table 1 of Langton and Bowman (1980). Data (percentage weight) are from fish collected in 1969-1972 from the southwestern Scotian Shelf to south of Georges Bank.
Diet 15.1 The proportions were adjusted to a total of 1.0.
Diet 16 is the average of Diet 14.1 and Diet 15.1.
Diet 17 is the final haddock diet. It is the same as Diet 16 , except that 0.09 is removed from small pelagic feeders to capelin to reflect the difference in prey availablity between herring in the Langton and Bowman (1980) study and capelin on the southern Grand Bank (Templeman 1965).

Appendix 2. Table 10d. Diet of zoarcids (a component of the large demersal feeders).

|  | ocean pout |  |  | eelpouts | zoarcids |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Diet 18 | Diet 18.1 | Diet 18.2 | Diet 19 | Diet 20 |
| 1 Whales | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 Harp seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 Hooded seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 4 Seabirds | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| $5 \mathrm{Cod}>35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 7 Greenland halibut > 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 8 Greenland halibut <= 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 10 American plaice $<=35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 11 Flounders | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 12 Skates | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 13 Redfish | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 14 Demersal feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 15 Demersal feeders (misc. small) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 16 Capelin | 0.104 | 0.107 | 0.107 | 0.000 | 0.021 |
| 17 Sand lance | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 18 Arctic cod | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 19 Pelagic feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 21 Plankt. Pelagic feeders (misc. small | 0.092 | 0.095 | 0.095 | 0.000 | 0.019 |
| 22 Shrimp | 0.000 | 0.000 | 0.010 | 0.010 | 0.010 |
| 23 Large Crustaceans (crab, lobster) | 0.002 | 0.002 | 0.032 | 0.003 | 0.009 |
| 24 Echinoderms | 0.688 | 0.709 | 0.509 | 0.220 | 0.278 |
| 25 Molluscs (benthic) | 0.003 | 0.003 | 0.063 | 0.160 | 0.141 |
| 26 Polychaetes | 0.000 | 0.000 | 0.050 | 0.160 | 0.138 |
| 27 Benthic invertebrates (other) | 0.081 | 0.084 | 0.134 | 0.440 | 0.379 |
| 28 Zooplankton (large) | 0.000 | 0.000 | 0.000 | 0.006 | 0.005 |
| 29 Zooplankton (small) | 0.000 | 0.000 | 0.000 | 0.001 | 0.001 |
| 30 Phytoplankton | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 31 Detritus | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| Total | 0.970 |  |  | 1.000 | 1.000 |

Diet 18 is from Table 1 of Keats et al. (1987). Data (\% weight) are from fish collected in 1983-1984 at $5-15 \mathrm{~m}$ from hard bottom dominated by sea urchins. The benthic invertebrate (other) category is entirely eggs of fish (lumpfish and capelin).
Diet 18.1 is Diet 18 adjusted to a total of 1.0
Diet 18.2 is Diet 18.1 with arbitrary changes to spread the echinoderm value over other groups.
An amount of 0.2 is deducted from echinoderms and assigned to shrimp ( 0.01 ), large crustaceans ( 0.03 ), molluscs ( 0.06 ), polychaetes ( 0.05 ) and other benthic invertebrates ( 0.05 ). Diet 19 is fictional based on information in Scott and Scott (1988).
Diet 20 is an average of Diet 18.2 and Diet 19, weighted by 0.2 and 0.8 to reflect a guess at relative biomasses.

Appendix 2. Table 10e. Diet of lumpfish, monkfish and Atlantic halibut, and composite diet of the large demersal feeders.

|  | wolffish | grenadier | white <br> hake | haddock | eelpouts | lumpfish | monkfish | Atlantic halibut | average |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | (Diet 8) | (Diet 12) | (Diet 13.1) | (Diet 17) | (Diet 20) | Diet 21 | Diet 22 | Diet 23 | Diet 24 |
| 1 Whales | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 Harp seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 Hooded seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 4 Seabirds | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 5 Cod > 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.005 | 0.000 | 0.090 | 0.004 | 0.000 | 0.000 | 0.030 | 0.039 | 0.011 |
| 7 Greenland halibut > 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 8 Greenland halibut <= 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.005 | 0.000 | 0.000 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 10 American plaice $<=35 \mathrm{~cm}$ | 0.003 | 0.000 | 0.090 | 0.012 | 0.000 | 0.000 | 0.020 | 0.136 | 0.017 |
| 11 Flounders | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.040 | 0.136 | 0.011 |
| 12 Skates | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.010 | 0.000 | 0.001 |
| 13 Redfish | 0.049 | 0.016 | 0.000 | 0.000 | 0.000 | 0.000 | 0.040 | 0.039 | 0.016 |
| 14 Demersal feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 15 Demersal feeders (misc. small) | 0.001 | 0.001 | 0.270 | 0.012 | 0.000 | 0.000 | 0.200 | 0.195 | 0.042 |
| 16 Capelin | 0.000 | 0.032 | 0.090 | 0.090 | 0.021 | 0.020 | 0.150 | 0.000 | 0.039 |
| 17 Sand lance | 0.000 | 0.016 | 0.090 | 0.049 | 0.000 | 0.000 | 0.210 | 0.000 | 0.027 |
| 18 Arctic cod | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 19 Pelagic feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.000 | 0.000 | 0.016 | 0.001 | 0.000 | 0.000 | 0.000 | 0.000 | 0.001 |
| 21 Plankt. Pelagic feeders (misc. sma | 0.000 | 0.062 | 0.106 | 0.012 | 0.019 | 0.020 | 0.100 | 0.000 | 0.034 |
| 22 Shrimp | 0.012 | 0.252 | 0.039 | 0.055 | 0.010 | 0.010 | 0.020 | 0.324 | 0.090 |
| 23 Large Crustaceans (crab, lobster) | 0.075 | 0.004 | 0.015 | 0.021 | 0.009 | 0.000 | 0.050 | 0.094 | 0.026 |
| 24 Echinoderms | 0.548 | 0.120 | 0.012 | 0.308 | 0.278 | 0.010 | 0.020 | 0.008 | 0.199 |
| 25 Molluscs (benthic) | 0.135 | 0.061 | 0.010 | 0.053 | 0.141 | 0.000 | 0.020 | 0.017 | 0.067 |
| 26 Polychaetes | 0.008 | 0.126 | 0.007 | 0.207 | 0.138 | 0.010 | 0.020 | 0.013 | 0.080 |
| 27 Benthic invertebrates (other) | 0.138 | 0.164 | 0.000 | 0.140 | 0.379 | 0.010 | 0.060 | 0.000 | 0.141 |
| 28 Zooplankton (large) | 0.025 | 0.119 | 0.165 | 0.037 | 0.005 | 0.870 | 0.005 | 0.000 | 0.185 |
| 29 Zooplankton (small) | 0.000 | 0.026 | 0.000 | 0.000 | 0.001 | 0.050 | 0.000 | 0.000 | 0.014 |
| 30 Phytoplankton | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 31 Detritus | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Diets 21 and 22 are guesses based on information in Scott and Scott (1988).
The demersal feeders (small) in diet 22 are comprised of 0.06 juvenile demersal feeders (large) and 0.14 demersal feeders (small).
Diet 23 is based on frequency of occurrence data in Kohler (1967). Allocation to major groups is based on observations of stomachs of halibut taken by research otter trawls on the Scotian Shelf (Table 5). Subdivision of the fish component ( 0.544 ) is according to the occurrence of fish in stomachs of halibut taken by longliners in the Grand Bank area (Table 8). It is curious that there is no mention of pelagic fish.
Diet 24 is the diet for the large demersal feeders. It is an average of diets $8,12,13.1,17,20,21,22$, and 23 , with weightings of $0.16,0.22,0.06,0.12,0.16,0.16,0.06$ and 0.06 .

Appendix 2. Table 11. Diet of small demersal feeders.

|  | Diet 1 |
| :--- | ---: |
| 1 Whales | 0.000 |
| 2 Harp seals | 0.000 |
| 3 Hooded seals | 0.000 |
| 4 Seabirds | 0.000 |
| 5 Cod $>35 \mathrm{~cm}$ | 0.000 |
| 6 Cod <= 35 cm | 0.000 |
| 7 Greenland halibut $>40 \mathrm{~cm}$ | 0.000 |
| 8 Greenland halibut $<=40 \mathrm{~cm}$ | 0.000 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.000 |
| 10 American plaice <= 35 cm | 0.000 |
| 11 Flounders | 0.000 |
| 12 Skates | 0.000 |
| 13 Redfish | 0.000 |
| 14 Demersal feeders (misc. large) | 0.000 |
| 15 Demersal feeders (misc. small) | 0.010 |
| 16 Capelin | 0.020 |
| 17 Sand lance | 0.010 |
| 18 Arctic cod | 0.005 |
| 19 Pelagic feeders (misc. large) | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.002 |
| 21 Plankt. Pelagic feeders (misc. smal | 0.003 |
| 22 Shrimp | 0.020 |
| 23 Large Crustaceans (crab, lobster) | 0.010 |
| 24 Echinoderms | 0.100 |
| 25 Molluscs (benthic) | 0.100 |
| 26 Polychaetes | 0.200 |
| 27 Benthic invertebrates (other) | 0.420 |
| 28 Zooplankton (large) | 0.050 |
| 29 Zooplankton (small) | 0.050 |
| 30 Phytoplankton | 0.000 |
| 31 Detritus | 0.000 |

Diet 1 is a guess based on information in Scott and Scott (1988).

Appendix 2. Table 12. Diet of capelin.

|  | spring |  |  | autumn | average |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | Diet 1 | Diet 2 | Diet 3 | Diet 4 | Diet 5 |
| 1 Whales | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 Harp seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 Hooded seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 4 Seabirds | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 5 Cod > 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 7 Greenland halibut > 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 8 Greenland halibut <= 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 10 American plaice $<=35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 11 Flounders | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 12 Skates | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 13 Redfish | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 14 Demersal feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 15 Demersal feeders (misc. small) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 16 Capelin | 0.028 | 0.013 | 0.021 | 0.000 | 0.010 |
| 17 Sand lance | 0.028 | 0.013 | 0.021 | 0.000 | 0.010 |
| 18 Arctic cod | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 19 Pelagic feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 21 Plankt. Pelagic feeders (misc. smal | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 22 Shrimp | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 23 Large Crustaceans (crab, lobster) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 24 Echinoderms | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 25 Molluscs (benthic) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 26 Polychaetes | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 27 Benthic invertebrates (other) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 28 Zooplankton (large) | 0.445 | 0.599 | 0.522 | 0.345 | 0.434 |
| 29 Zooplankton (small) | 0.499 | 0.375 | 0.437 | 0.655 | 0.546 |
| 30 Phytoplankton | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 31 Detritus | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Diet 1 is from Table 3 of Gerasimova (1994). Data (\% occurrence) are from fish collected in April-May 1987-1990 on Grand Bank. The frequency for each prey type was averaged across the 4 years and 2 maturity stages and expressed as a proportion of the sum of the frequencies for all prey categories. The "fish juvenile" portion was divided evenly between capelin and sand lance. Diet 2 is from Table 4 of Kovalyov and Kudrin (1973). Data (\% occurrence) are from fish collected in March-June 1972 on Grand Bank. The frequency for each prey type was expressed as a proportion of the sum of the frequencies for all prey categories. The "fish larvae" portion was divided evenly between capelin and sand lance.
Diet 3 is an average spring Grand Bank diet (average of Diet 1 and Diet 2).
Diet 4 is from Chan and Carscadden (1976), who examined capelin caught mainly off southern Labrador in August-November 1973. Percentage weights for large zooplankton and small zooplankton were taken from Table 4. The data for the 4 months were averaged, with weightings by monthly stomach fullness (Table 1).
Diet 5 is the average of the spring Grand Bank diet and the autumn Labrador diet (average of Diet 3 and Diet 4)

Appendix 2. Table 13. Diet of sand lance.

|  | Diet 1 |
| :--- | ---: |
| 1 Whales | 0.000 |
| 2 Harp seals | 0.000 |
| 3 Hooded seals | 0.000 |
| 4 Seabirds | 0.000 |
| 5 Cod > 35 cm | 0.000 |
| 6 Cod <= 35 cm | 0.000 |
| 7 Greenland halibut > 40 cm | 0.000 |
| 8 Greenland halibut <= 40 cm | 0.000 |
| 9 American plaice > 35 cm | 0.000 |
| 10 American plaice <= 35 cm | 0.000 |
| 11 Flounders | 0.000 |
| 12 Skates | 0.000 |
| 13 Redfish | 0.000 |
| 14 Demersal feeders (misc. large) | 0.000 |
| 15 Demersal feeders (misc. small) | 0.000 |
| 16 Capelin | 0.000 |
| 17 Sand lance | 0.000 |
| 18 Arctic cod | 0.000 |
| 19 Pelagic feeders (misc. large) | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.000 |
| 21 Plankt. Pelagic feeders (misc. smal | 0.000 |
| 22 Shrimp | 0.000 |
| 23 Large Crustaceans (crab, lobster) | 0.000 |
| 24 Echinoderms | 0.000 |
| 25 Molluscs (benthic) | 0.000 |
| 26 Polychaetes | 0.000 |
| 27 Benthic invertebrates (other) | 0.000 |
| 28 Zooplankton (large) | 0.350 |
| 29 Zooplankton (small) | 0.650 |
| 30 Phytoplankton | 0.000 |
| 31 Detritus | 0.000 |
|  |  |
|  |  |

Diet information is from Table 1 of Scott (1973).
Data (\% volume) are from fish collected in May 1967
on the Scotian Shelf.

Appendix 2. Table 14. Diet of Arctic cod.

|  | Diet 1 |
| :--- | :---: |
| 1 Whales | 0.000 |
| 2 Harp seals | 0.000 |
| 3 Hooded seals | 0.000 |
| 4 Seabirds | 0.000 |
| 5 Cod $>35 \mathrm{~cm}$ | 0.000 |
| 6 Cod <= 35 cm | 0.000 |
| 7 Greenland halibut $>40 \mathrm{~cm}$ | 0.000 |
| 8 Greenland halibut <= 40 cm | 0.000 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.000 |
| 10 American plaice <= 35 cm | 0.000 |
| 11 Flounders | 0.000 |
| 12 Skates | 0.000 |
| 13 Redfish | 0.000 |
| 14 Demersal feeders (misc. large) | 0.000 |
| 15 Demersal feeders (misc. small) | 0.000 |
| 16 Capelin | 0.038 |
| 17 Sand lance | 0.000 |
| 18 Arctic cod | 0.002 |
| 19 Pelagic feeders (misc. large) | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.000 |
| 21 Plankt. Pelagic feeders (misc. small) | 0.000 |
| 22 Shrimp | 0.000 |
| 23 Large Crustaceans (crab, lobster) | 0.000 |
| 24 Echinoderms | 0.000 |
| 25 Molluscs (benthic) | 0.000 |
| 26 Polychaetes | 0.000 |
| 27 Benthic invertebrates (other) | 0.000 |
| 28 Zooplankton (large) | 0.640 |
| 29 Zooplankton (small) | 0.320 |
| 30 Phytoplankton | 0.000 |
| 31 Detritus | 0.000 |

Diet 1 is a guess based on a small number of observations from northern and southern Labrador in autumn 1978 (Lilly 1980) and additional unreported observations from Divisions 2J, 3K and 3L in autumn 1994.

## Appendix 2. Table 15. Diet of large pelagics.

|  | silver hake |  |  | pollock |  |
| :--- | :--- | :--- | :--- | :--- | ---: |
|  | Diet 1 | Diet 1.1 | Diet 2 | Diet 2.1 | average |
| 1 Whales | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 Harp seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 Hooded seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 4 Seabirds | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 5 Cod > 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 6 Cod <= 35 cm | 0.000 | 0.000 | 0.003 | 0.003 | 0.002 |
| 7 Greenland halibut > 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 8 Greenland halibut <= 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 9 American plaice > 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 10 American plaice <= 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 11 Flounders | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 12 Skates | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 13 Redfish | 0.000 | 0.000 | 0.003 | 0.003 | 0.002 |
| 14 Demersal feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 15 Demersal feeders (misc. small) | 0.070 | 0.070 | 0.000 | 0.000 | 0.035 |
| 16 Capelin | 0.000 | 0.050 | 0.000 | 0.100 | 0.075 |
| 17 Sand lance | 0.000 | 0.050 | 0.322 | 0.122 | 0.086 |
| 18 Arctic cod | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 19 Pelagic feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.390 | 0.340 | 0.156 | 0.206 | 0.273 |
| 21 Plankt. Pelagic feeders (misc. sma | 0.220 | 0.170 | 0.155 | 0.205 | 0.188 |
| 22 Shrimp | 0.020 | 0.020 | 0.003 | 0.003 | 0.012 |
| 23 Large Crustaceans (crab, lobster) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 24 Echinoderms | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 25 Molluscs (benthic) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 26 Polychaetes | 0.005 | 0.005 | 0.000 | 0.000 | 0.003 |
| 27 Benthic invertebrates (other) | 0.003 | 0.003 | 0.035 | 0.035 | 0.019 |
| 28 Zooplankton (large) | 0.284 | 0.274 | 0.316 | 0.316 | 0.295 |
| 29 Zooplankton (small) | 0.000 | 0.018 | 0.000 | 0.007 | 0.013 |
| 30 Phytoplankton | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 31 Detritus | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| Total | 0.992 | 1.000 | 0.993 | 1.000 | 1.000 |
|  |  |  |  |  |  |

Diet 1 is from Table 7 of Vinogradov (1983). Data (\% weight) are from fish collected in 1968-1974 on Georges Bank.
The demersal feeders (small) are comprised of 0.036 juvenile demersal feeders (large) and 0.034 demersal feeders (small).
Diet 1.1 is a guess. It is Diet 1 modified to take account of diet information in Vinogradov (1972) and Swan and Clay (1979), and the prey suite potentially available on Grand Bank.
A value of 0.05 is subtracted from each of "pisc. pelagic feeders" and "plankt. pelagic feeders" and given to capelin and sand lance. In addition, a value of 0.01 is deducted from
"zooplankton (large)" and given to "zooplankton (small)". The 0.008 missing from Diet 1 is also given to "zooplankton (small)".
Diet 2 is from Table 13 of Steele (1963). Data (\% volume) are from fish collected on the Scotian Shelf in July 1959 and April-May 1961. The diets of the 3 size-groups in the 1959 sample were averaged without weighting as were the diets of the 4 size-groups in the 1961 sample.The diets from the 2 years were then averaged.
Diet 2.1 is a guess. It is Diet 2 modified to take into account the prey suite potentially available on Grand Bank. A value of 0.2 is subtracted from "sand lance" and given to capelin (0.1),
"pisc. pelagic feeders (small)" (0.05) and "plankt. pelagic feeders (small)" (0.05) . A value of 0.1 is subtracted from "pelagic feeders (large)" and given to "demersal feeders large". The 0.007 missing from Diet 2 is given to "zooplankton (small)".

Diet 3 is the diet for large pelagic feeders. It is an average of Diet 1.1 and Diet 2.1

Appendix 2. Table 16. Diet of piscivorous small pelagics.

|  | mackerel <br> Diet 1 | mesopelagics <br> Diet 2 | short-finned squid |  | average |
| :--- | ---: | ---: | ---: | ---: | ---: |
|  | Diet 3 | Diet 3.1 | Diet 4 |  |  |
| 1 Whales | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 2 Harp seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 3 Hooded seals | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 4 Seabirds | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 5 Cod > 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 6 Cod <= 35 cm | 0.000 | 0.000 | 0.429 | 0.215 | 0.019 |
| 7 Greenland halibut > 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 8 Greenland halibut <= 40 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 9 American plaice > 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 10 American plaice <= 35 cm | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 11 Flounders | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 12 Skates | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 13 Redfish | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 14 Demersal feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 15 Demersal feeders (misc. small) | 0.000 | 0.000 | 0.001 | 0.001 | 0.000 |
| 16 Capelin | 0.810 | 0.000 | 0.189 | 0.359 | 0.698 |
| 17 Sand lance | 0.000 | 0.000 | 0.252 | 0.126 | 0.011 |
| 18 Arctic cod | 0.000 | 0.000 | 0.003 | 0.003 | 0.000 |
| 19 Pelagic feeders (misc. large) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 21 Plankt. Pelagic feeders (misc. smi | 0.000 | 0.810 | 0.116 | 0.116 | 0.083 |
| 22 Shrimp | 0.000 | 0.090 | 0.000 | 0.000 | 0.008 |
| 23 Large Crustaceans (crab, lobster) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 24 Echinoderms | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 25 Molluscs (benthic) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 26 Polychaetes | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 27 Benthic invertebrates (other) | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 28 Zooplankton (large) | 0.180 | 0.050 | 0.010 | 0.180 | 0.168 |
| 29 Zooplankton (small) | 0.010 | 0.050 | 0.000 | 0.000 | 0.013 |
| 30 Phytoplankton | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |
| 31 Detritus | 0.000 | 0.000 | 0.000 | 0.000 | 0.000 |

Diet 1 is from Table 9 of Moores et al. (1979). Data (\% occurrence) are from fish collected from commercial catches in Newfoundland waters in June-October (presumably in 1970-1973). The sample size was small. Diet 2 is a guess based on Podrazhanskaya (1993) and Gartner et al. (1997)
Diet 3 is based on sampling of short-finned squid in inshore waters of eastern Newfoundland in 1985 and 1987. Proportions of fish, crustacea (assumed to be large zooplankters) and squid in the diet were calculated as an average of the values for Holyrood in August and September, 1985 (Table 2 of Dawe (1988)). The fish component was apportioned according to the percentage of identified otoliths in samples from Holyrood and St. Chad's in 1985 and Leading Tickles in 1987 (Table 3 in Dawe et al. (1997)).
Diet 3.1 is an arbitrary adjustment based on assumption that the time and location of sampling might cause underestimation of the importance of crustacean zooplankters and capelin. The values for Atlantic cod and sand lance are arbitrarily reduced by $50 \%$, and the difference ( 0.340 ) is arbitrarily assigned equally to large zooplankton and capelin.
Diet 4 is an average diet for piscivorous small pelagics. It is the average of Diets 1,2 and 3.1 , weighted by $0.822,0.089$ and 0.089 . These are the proportional biomasses, where the biomass of mackerel is 184411 t and the biomass of the mesopleagics is $20,000 \mathrm{t}$ and the biomass of squid is $20,000 \mathrm{t}$.

Appendix 2. Table 17. Diet of planktivorous small pelagics.

|  | herring Diet 1 | mesopelagics <br> + Arctic squid <br> Diet 2 | average Diet 3 |
| :---: | :---: | :---: | :---: |
| 1 Whales | 0.000 | 0.000 | 0.000 |
| 2 Harp seals | 0.000 | 0.000 | 0.000 |
| 3 Hooded seals | 0.000 | 0.000 | 0.000 |
| 4 Seabirds | 0.000 | 0.000 | 0.000 |
| $5 \mathrm{Cod}>35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 |
| $6 \mathrm{Cod}<=35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 |
| 7 Greenland halibut > 40 cm | 0.000 | 0.000 | 0.000 |
| 8 Greenland halibut <= 40 cm | 0.000 | 0.000 | 0.000 |
| 9 American plaice $>35 \mathrm{~cm}$ | 0.000 | 0.000 | 0.000 |
| 10 American plaice <= 35 cm | 0.000 | 0.000 | 0.000 |
| 11 Flounders | 0.000 | 0.000 | 0.000 |
| 12 Skates | 0.000 | 0.000 | 0.000 |
| 13 Redfish | 0.000 | 0.000 | 0.000 |
| 14 Demersal feeders (misc. large) | 0.000 | 0.000 | 0.000 |
| 15 Demersal feeders (misc. small) | 0.000 | 0.000 | 0.000 |
| 16 Capelin | 0.000 | 0.000 | 0.000 |
| 17 Sand lance | 0.000 | 0.000 | 0.000 |
| 18 Arctic cod | 0.000 | 0.000 | 0.000 |
| 19 Pelagic feeders (misc. large) | 0.000 | 0.000 | 0.000 |
| 20 Pisc. Pelagic feeders (misc. small) | 0.000 | 0.000 | 0.000 |
| 21 Plankt. Pelagic feeders (misc. small) | 0.000 | 0.000 | 0.000 |
| 22 Shrimp | 0.000 | 0.010 | 0.005 |
| 23 Large Crustaceans (crab, lobster) | 0.000 | 0.000 | 0.000 |
| 24 Echinoderms | 0.000 | 0.000 | 0.000 |
| 25 Molluscs (benthic) | 0.000 | 0.000 | 0.000 |
| 26 Polychaetes | 0.000 | 0.000 | 0.000 |
| 27 Benthic invertebrates (other) | 0.100 | 0.000 | 0.050 |
| 28 Zooplankton (large) | 0.513 | 0.540 | 0.527 |
| 29 Zooplankton (small) | 0.387 | 0.450 | 0.419 |
| 30 Phytoplankton | 0.000 | 0.000 | 0.000 |
| 31 Detritus | 0.000 | 0.000 | 0.000 |

Diet 1 is from Table 4 of Messieh et al. (1979). Data (significance index) are from fish collected from commerc catches in the Gulf of St. Lawrence and adjacent areas in 1970-1973. The values in Table 4 were summed ov all areas and months, and the sum for each prey category was expressed as a proportion of the sum for all pre categories combined.
Diet 2 is a guess based on Podrazhanskaya (1993) and Gartner et al. (1997) and the assumption that myctophids and other planktivores comprise most of the biomass.
It is assumed here that the diet of Arctic squid is similar to the mesopelagic diet.
Diet 3 is an average diet for planktivorous small pelagics. It is the average of Diets 1 and 2 , equally weighted.

Appendix 2. Table 18. Revised diet of the Large Crustacea.

|  |  |  |
| :--- | ---: | ---: |
|  | Original Diet | Revised Diet |
| 1. Whales | 0.000 | 0.000 |
| 2. Harp Seals | 0.000 | 0.000 |
| 3. Hooded Seals | 0.000 | 0.000 |
| 4. Seabirds | 0.000 | 0.000 |
| 5. Cod $>35 \mathrm{~cm}$ | 0.000 | 0.000 |
| 6. Cod $<=35 \mathrm{~cm}$ | 0.000 | 0.000 |
| 7. G.halibut $>40 \mathrm{~cm}$ | 0.000 | 0.000 |
| 8. G.halibut<=40cm | 0.000 | 0.000 |
| 9. Aplaice>35cm | 0.000 | 0.000 |
| 10. Aplaice<=35cm | 0.000 | 0.000 |
| 11. Flounders | 0.000 | 0.000 |
| 12. Skates | 0.000 | 0.000 |
| 13. Redfish | 0.000 | 0.000 |
| 14. L.Dem.Feeders | 0.000 | 0.000 |
| 15. S.Dem.Feeders | 0.050 | 0.000 |
| 16. Capelin | 0.000 | 0.000 |
| 17. Sand lance | 0.000 | 0.000 |
| 18. Arctic cod | 0.000 | 0.000 |
| 19. L.Pel.Feeders | 0.000 | 0.000 |
| 20. Pisc. SPF | 0.000 | 0.000 |
| 21. Plankt. SPF | 0.000 | 0.000 |
| 22. Shrimp | 0.049 | 0.020 |
| 23. Large Crustacea | 0.111 | 0.010 |
| 24. Echinoderms | 0.200 | 0.300 |
| 25. Molluscs | 0.198 | 0.120 |
| 26. Polychaetes | 0.196 | 0.300 |
| 27. O.Benthic Inver | 0.001 | 0.120 |
| 28. L.Zooplankton | 0.049 | 0.020 |
| 29. S.Zooplankton | 0.049 | 0.010 |
| 30. Phytoplankton | 0.000 | 0.000 |
| 31. Detritus | 0.097 | 0.100 |
|  |  |  |

Appendix 3 Table 1. Results of the Sensitivity Analysis - Model 1.

|  |  | \% change in input parameter |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Input parameters |  | Estimated | -50 | -40 | -30 | -20 | -10 | 0 | 10 | 20 | 30 | 40 | 50 |
| 1B | 1QB | 6 B | -0.07 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.07 |
| 1B | 1QB | 14EE | -0.33 | -0.26 | -0.2 | -0.13 | -0.07 | 0 | 0.07 | 0.13 | 0.2 | 0.26 | 0.33 |
| 1B | 1QB | 16B | -0.09 | -0.07 | -0.05 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.05 | 0.07 | 0.09 |
| 1B | 1QB | 17B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 1B | 1QB | 20B | -0.28 | -0.22 | -0.17 | -0.11 | -0.06 | 0 | 0.06 | 0.11 | 0.17 | 0.22 | 0.28 |
| 1B | 1QB | 21B | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 1B | 1QB | 28B | -0.07 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.07 |
| 1B | 1QB | 29B | -0.07 | -0.06 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.06 | 0.07 |
| 1B | 1QB | 30EE | -0.07 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.07 |
| 2B | 2PB | 2EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 2B | 2QB | 6 B | -0.13 | -0.1 | -0.08 | -0.05 | -0.03 | 0 | 0.03 | 0.05 | 0.08 | 0.1 | 0.13 |
| 2B | 2QB | 7EE | -0.1 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.1 |
| 2B | 2QB | 8B | -0.18 | -0.14 | -0.11 | -0.07 | -0.04 | 0 | 0.04 | 0.07 | 0.11 | 0.14 | 0.18 |
| 2B | 2QB | 10EE | -0.06 | -0.05 | -0.04 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.04 | 0.05 | 0.06 |
| 2B | 2QB | 11 B | -0.2 | -0.16 | -0.12 | -0.08 | -0.04 | 0 | 0.04 | 0.08 | 0.12 | 0.16 | 0.2 |
| 2B | 2QB | 15 B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 2B | 2QB | 16B | -0.09 | -0.07 | -0.05 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.05 | 0.07 | 0.09 |
| 2B | 2QB | 17B | -0.08 | -0.07 | -0.05 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.05 | 0.07 | 0.08 |
| 2B | 2QB | 18B | -0.33 | -0.27 | -0.2 | -0.13 | -0.07 | 0 | 0.07 | 0.13 | 0.2 | 0.27 | 0.33 |
| 2B | 2QB | 19B | -0.15 | -0.12 | -0.09 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.09 | 0.12 | 0.15 |
| 2B | 2QB | 20B | -0.1 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.1 |
| 2B | 2QB | 21B | -0.07 | -0.06 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.06 | 0.07 |
| 2B | 2QB | 22B | -0.06 | -0.05 | -0.04 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.04 | 0.05 | 0.06 |
| 2B | 2QB | 26EE | -0.07 | -0.06 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.06 | 0.07 |
| 2B | 2QB | 27EE | -0.08 | -0.06 | -0.05 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.05 | 0.06 | 0.08 |
| 2B | 2QB | 28B | -0.11 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.11 |
| 2B | 2QB | 29B | -0.1 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.1 |
| 2B | 2QB | 30EE | -0.1 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.1 |
| 3B | 3QB | 7EE | -0.13 | -0.1 | -0.08 | -0.05 | -0.03 | 0 | 0.03 | 0.05 | 0.08 | 0.1 | 0.13 |
| 3B | 3QB | 8B | -0.16 | -0.13 | -0.1 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.1 | 0.13 | 0.16 |
| 3B | 3QB | 11 B | -0.12 | -0.1 | -0.07 | -0.05 | -0.02 | 0 | 0.02 | 0.05 | 0.07 | 0.1 | 0.12 |
| 3B | 3QB | 21B | -0.04 | -0.03 | -0.02 | -0.01 | -0.01 | 0 | 0.01 | 0.01 | 0.02 | 0.03 | 0.04 |
| 3B | 3QB | 26EE | -0.04 | -0.03 | -0.02 | -0.01 | -0.01 | 0 | 0.01 | 0.01 | 0.02 | 0.03 | 0.04 |
| 3B | 3QB | 27EE | -0.03 | -0.03 | -0.02 | -0.01 | -0.01 | 0 | 0.01 | 0.01 | 0.02 | 0.03 | 0.03 |
| 4B | 4QB | QEE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 5B | 5PB | 5EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 5B | 5QB | 6 B | -0.13 | -0.1 | -0.08 | -0.05 | -0.03 | 0 | 0.03 | 0.05 | 0.08 | 0.1 | 0.13 |
| 5B | 5QB | 8 B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 5B | 5QB | 10EE | -0.22 | -0.17 | -0.13 | -0.09 | -0.04 | 0 | 0.04 | 0.09 | 0.13 | 0.17 | 0.22 |
| 5B | 5QB | 13B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 5B | 5QB | 15 B | -0.15 | -0.12 | -0.09 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.09 | 0.12 | 0.15 |
| 5B | 5QB | 16B | -0.16 | -0.13 | -0.1 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.1 | 0.13 | 0.16 |
| 5B | 5QB | 17B | -0.16 | -0.13 | -0.1 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.1 | 0.13 | 0.16 |
| 5B | 5QB | 18 B | -0.08 | -0.07 | -0.05 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.05 | 0.07 | 0.08 |
| 5B | 5QB | 21B | -0.04 | -0.03 | -0.02 | -0.01 | -0.01 | 0 | 0.01 | 0.01 | 0.02 | 0.03 | 0.04 |


| 5B | 5QB | 22B | -0.15 | -0.12 | -0.09 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.09 | 0.12 | 0.15 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5B | 5QB | 23B | -0.16 | -0.13 | -0.1 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.1 | 0.13 | 0.16 |
| 5B | 5QB | 24EE | -0.12 | -0.1 | -0.07 | -0.05 | -0.02 | 0 | 0.02 | 0.05 | 0.07 | 0.1 | 0.12 |
| 5B | 5QB | 25EE | -0.15 | -0.12 | -0.09 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.09 | 0.12 | 0.15 |
| 5B | 5QB | 26EE | -0.11 | -0.09 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.09 | 0.11 |
| 5B | 5QB | 27EE | -0.1 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.1 |
| 5B | 5QB | 28B | -0.13 | -0.1 | -0.08 | -0.05 | -0.03 | 0 | 0.03 | 0.05 | 0.08 | 0.1 | 0.13 |
| 5B | 5QB | 29B | -0.13 | -0.1 | -0.08 | -0.05 | -0.03 | 0 | 0.03 | 0.05 | 0.08 | 0.1 | 0.13 |
| 5B | 5QB | 30EE | -0.13 | -0.1 | -0.08 | -0.05 | -0.03 | 0 | 0.03 | 0.05 | 0.08 | 0.1 | 0.13 |
| 6 PB | 6 EE | 6 B | 1.06 | 0.7 | 0.45 | 0.26 | 0.11 | 0 | -0.09 | -0.17 | -0.24 | -0.29 | -0.34 |
| 6 PB | 6 EE | 16 B | 0.07 | 0.05 | 0.03 | 0.02 | 0.01 | 0 | -0.01 | -0.01 | -0.0 | -0.02 | -0.02 |
| 6 PB | 6 EE | 18 B | 0.06 | 0.04 | 0.03 | 0.02 | 0.01 | 0 | -0.01 | -0.01 | -0.01 | -0.02 | -0.02 |
| 6 PB | 6 EE | 22 B | 0.11 | 0.07 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 6 PB | 6 EE | 27 EE | 0.07 | 0.05 | 0.03 | 0.02 | 0.01 | 0 | -0.01 | -0.01 | -0.02 | -0.02 | -0.02 |
| 6 PB | 6 EE | 28 B | 0.06 | 0.04 | 0.03 | 0.02 | 0.01 | 0 | -0.01 | -0.01 | -0.01 | -0.02 | -0.02 |
| 6 PB | 6 EE | 29 B | 0.06 | 0.04 | 0.03 | 0.02 | 0.01 | 0 | -0.01 | -0.01 | -0.01 | -0.02 | -0.02 |
| 6 PB | 6 EE | 30 EE | 0.06 | 0.04 | 0.03 | 0.02 | 0.01 | 0 | -0.01 | -0.01 | -0.01 | -0.02 | -0.02 |
| 6 QB | 6 EE | 16 B | -0.03 | -0.03 | -0.02 | -0.01 | -0.01 | 0 | 0.01 | 0.01 | 0.02 | 0.03 | 0.03 |
| 6 QB | 6 EE | 22 B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 6 QB | 6 EE | 27 EE | -0.03 | -0.03 | -0.02 | -0.01 | -0.01 | 0 | 0.01 | 0.01 | 0.02 | 0.03 | 0.03 |
| 7B |  | 6 B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 7B | 7PB | 7EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 7B | 7QB | 8 B | -0.08 | -0.06 | -0.05 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.05 | 0.06 | 0.08 |
| 7B | 7QB | 13B | -0.12 | -0.09 | -0.07 | -0.05 | -0.02 | 0 | 0.02 | 0.05 | 0.07 | 0.09 | 0.12 |
| 7B | 7QB | 19B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.0 | 0.04 | 0.05 |
| 7B | 7QB | 21B | -0.07 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.07 |
| 8PB | 8 EE | 8 B | 1.06 | 0.7 | 0.45 | 0.26 | 0.11 | 0 | -0.09 | -0.17 | -0.24 | -0.29 | -0.34 |
| 8 PB | 8 EE | 16B | 0.15 | 0.1 | 0.06 | 0.04 | 0.02 | 0 | -0.01 | -0.02 | -0.03 | -0.04 | -0.05 |
| 8 PB | 8 EE | 18 B | 0.1 | 0.07 | 0.04 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 8 PB | 8 EE | 28B | 0.09 | 0.06 | 0.04 | 0.02 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 8 PB | 8 EE | 29B | 0.1 | 0.06 | 0.04 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 8 PB | 8 EE | 30EE | 0.1 | 0.06 | 0.04 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 8QB |  | 16B | -0.07 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.07 |
| 8QB |  | 18 B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 8QB |  | 28B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.0 | 0.04 | 0.05 |
| 8QB |  | 29B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 8QB |  | 30EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 9B | 9PB | 9EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 9B | 9QB | 17B | -0.06 | -0.05 | -0.04 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.04 | 0.05 | 0.06 |
| 9B | 9QB | 23B | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 9B | 9QB | 24EE | -0.07 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.07 |
| 9B | 9QB | 25EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 10B |  | 10EE | 0.76 | 0.51 | 0.33 | 0.19 | -0.08 | 0 | -0.07 | -0.13 | -0.18 | -0.22 | -0.25 |
| 10QB |  | 10EE | -0.12 | -0.09 | -0.07 | -0.05 | -0.02 | 0 | 0.02 | 0.05 | 0.07 | 0.09 | 0.12 |
| 10B | 10QB | 16B | -0.04 | -0.03 | -0.02 | -0.01 | -0.01 | 0 | 0.01 | 0.01 | 0.02 | 0.03 | 0.04 |
| 10B | 10QB | 17B | -0.06 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.06 |
| 10B | 10QB | 24EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 10B | 10QB | 25EE | -0.04 | -0.03 | -0.02 | -0.01 | -0.01 | 0 | 0.01 | 0.01 | 0.02 | 0.03 | 0.04 |
| 10B | 10QB | 26EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 10B | 10QB | 27EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 10PB |  | 10EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |


| 11PB | 11 EE | 11 B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 11PB | 11 EE | 26EE | 0.24 | 0.16 | 0.1 | 0.06 | 0.03 | 0 | -0.02 | -0.04 | -0.06 | -0.07 | -0.08 |
| 11PB | 11EE | 27EE | 0.18 | 0.12 | 0.08 | 0.04 | 0.02 | 0 | -0.02 | -0.03 | -0.04 | -0.05 | -0.06 |
| 11QB |  | 26EE | -0.12 | -0.09 | -0.07 | -0.05 | -0.02 | 0 | 0.02 | 0.05 | 0.07 | 0.09 | 0.12 |
| 11QB |  | 27EE | -0.09 | -0.17 | -0.05 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.05 | 0.17 | 0.09 |
| 12 B |  | 6B | -0.08 | -0.07 | -0.05 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.05 | 0.07 | 0.08 |
| 12B | 12PB | 12EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 12B | 12QB | 13B | -0.11 | -0.09 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.09 | 0.11 |
| 12B | 12QB | 15 B | -0.14 | -0.11 | -0.09 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.09 | 0.11 | 0.14 |
| 12B | 12QB | 17B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 12B | 12QB | 20B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 12B | 12QB | 21B | -0.08 | -0.06 | -0.05 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.05 | 0.06 | 0.08 |
| 12B | 12QB | 22B | -0.12 | -0.1 | -0.07 | -0.05 | -0.02 | 0 | 0.02 | 0.05 | 0.07 | 0.1 | 0.12 |
| 12B | 12QB | 23B | -0.22 | -0.18 | -0.13 | -0.09 | -0.04 | 0 | 0.04 | 0.09 | 0.13 | 0.18 | 0.22 |
| 12B | 12QB | 24EE | -0.13 | -0.1 | -0.08 | -0.05 | -0.03 | - | 0.03 | 0.05 | 0.08 | 0.1 | 0.13 |
| 12B | 12QB | 25EE | -0.13 | -0.1 | -0.08 | -0.05 | -0.03 | 0 | 0.03 | 0.05 | 0.08 | 0.1 | 0.13 |
| 12B | 12QB | 26EE | -0.09 | -0.07 | -0.05 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.05 | 0.07 | 0.09 |
| 12B | 12QB | 27EE | -0.11 | -0.09 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.09 | 0.11 |
| 12B | 12QB | 28EE | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 12B | 12QB | 29EE | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 12B | 12QB | 30EE | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 13PB | 13EE | 13B | 1.1 | 0.72 | 0.46 | 0.26 | 0.12 | 0 | -0.09 | -0.17 | -0.24 | -0.3 | -0.34 |
| 13PB | 13EE | 21B | 0.6 | 0.39 | 0.25 | 0.14 | 0.06 | 0 | -0.05 | -0.09 | -0.13 | -0.16 | -0.19 |
| 13PB | 13EE | 22B | 0.09 | 0.06 | 0.04 | 0.02 | 0.01 | 0 | -0.01 | -0.01 | -0.02 | -0.02 | -0.03 |
| 13PB | 13EE | 28B | 0.12 | 0.08 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.03 | -0.04 |
| 13PB | 13EE | 29B | 0.11 | 0.07 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 13PB | 13EE | 30EE | 0.11 | 0.07 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 13QB |  | 21B | -0.28 | -0.22 | -0.17 | -0.11 | -0.06 | 0 | 0.06 | 0.11 | 0.17 | 0.22 | 0.28 |
| 13QB |  | 22B | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 13QB |  | 28B | -0.06 | -0.05 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.05 | 0.06 |
| 13QB |  | 29B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 13QB |  | 30EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 14B | 14PB | 14 EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 14B | 14QB | 10EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 14B | 14QB | 22 B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 14B | 14QB | 24 EE | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 14B | 14QB | 27 EE | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 15 PB | 15 EE | 15 B | 1.22 | 0.78 | 0.49 | 0.28 | 0.12 | 0 | -0.1 | -0.18 | -0.25 | -0.31 | -0.35 |
| 15 PB | 15 EE | 22B | 0.17 | 0.11 | 0.07 | 0.04 | 0.02 | 0 | -0.01 | -0.02 | -0.03 | -0.04 | -0.05 |
| 15 PB | 15 EE | 23B | 0.11 | 0.08 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.03 | -0.04 |
| 15 PB | 15 EE | 24EE | 0.26 | 0.17 | 0.1 | 0.06 | 0.03 | 0 | -0.02 | -0.04 | -0.05 | -0.06 | -0.07 |
| 15 PB | 15 EE | 25EE | 0.33 | 0.21 | 0.13 | 0.08 | 0.03 | 0 | -0.03 | -0.05 | -0.07 | -0.08 | -0.1 |
| 15 PB | 15 EE | 26EE | 0.36 | 0.23 | 0.15 | 0.08 | 0.04 | 0 | -0.03 | -0.05 | -0.07 | -0.09 | -0.1 |
| 15 PB | 15 EE | 27EE | 0.64 | 0.41 | 0.26 | 0.15 | 0.07 | 0 | -0.05 | -0.1 | -0.13 | -0.16 | -0.19 |
| 15QB |  | 15 B | -0.05 | -0.04 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.04 | 0.05 |
| 15QB |  | 22B | -0.07 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.07 |
| 15QB |  | 23B | -0.05 | -0.04 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.04 | 0.05 |
| 15QB |  | 24EE | -0.11 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.11 |
| 15QB |  | 25EE | -0.14 | -0.11 | -0.09 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.09 | 0.11 | 0.14 |
| 15QB |  | 26EE | -0.15 | -0.13 | -0.09 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.09 | 0.13 | 0.15 |
| 15QB |  | 27EE | -0.28 | -0.22 | -0.17 | -0.11 | -0.06 | 0 | 0.0 | 0.1 | 0.17 | 0.22 | 0.28 |


| 16PB | 16EE | 16B | 1.09 | 0.71 | 0.45 | 0.26 | 0.12 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 16PB | 16EE | 17B | 0.22 | 0.14 | 0.09 | 0.05 | 0.02 | 0 | -0.02 | -0.03 | -0.05 | -0.06 | -0.07 |
| 16PB | 16EE | 28B | 0.57 | 0.37 | 0.24 | 0.14 | 0.06 | 0 | -0.05 | -0.09 | -0.12 | -0.15 | -0.18 |
| 16PB | 16EE | 29B | 0.58 | 0.38 | 0.24 | 0.14 | 0.06 | 0 | -0.05 | -0.09 | -0.13 | -0.16 | -0.18 |
| 16PB | 16EE | 30EE | 0.58 | 0.38 | 0.24 | 0.14 | 0.06 | 0 | -0.05 | -0.09 | -0.13 | -0.16 | -0.18 |
| 16QB |  | 17B | -0.1 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.1 |
| 16QB |  | 28B | -0.27 | -0.21 | -0.16 | -0.11 | -0.05 | 0 | 0.05 | 0.11 | 0.16 | 0.21 | 0.27 |
| 16QB |  | 29B | -0.27 | -0.21 | -0.16 | -0.11 | -0.05 | 0 | 0.05 | 0.11 | 0.16 | 0.21 | 0.27 |
| 16QB |  | 30EE | -0.27 | -0.21 | -0.16 | -0.11 | -0.05 | 0 | 0.05 | 0.11 | 0.16 | 0.21 | 0.27 |
| 17PB | 17EE | 17B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 17PB | 17EE | 28B | 0.14 | 0.09 | 0.06 | 0.04 | 0.02 | 0 | -0.01 | -0.02 | -0.03 | -0.04 | -0.05 |
| 17PB | 17EE | 29B | 0.16 | 0.11 | 0.07 | 0.04 | 0.02 | 0 | -0.01 | -0.03 | -0.04 | -0.05 | -0.05 |
| 17PB | 17EE | 30EE | 0.16 | 0.11 | 0.07 | 0.04 | 0.02 | 0 | -0.01 | -0.03 | -0.04 | -0.04 | -0.05 |
| 17QB |  | 28B | -0.07 | -0.06 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.06 | 0.07 |
| 17QB |  | 29B | -0.08 | -0.06 | -0.05 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.05 | 0.06 | 0.08 |
| 17QB |  | 30EE | -0.08 | -0.06 | -0.05 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.05 | 0.06 | 0.08 |
| 18PB | 18EE | 18 B | 1.03 | 0.68 | 0.44 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.34 |
| 18B | 18EE | 28B | 0.11 | 0.08 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.03 | -0.04 |
| 18B | 18EE | 29B | 0.1 | 0.07 | 0.04 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 18B | 18EE | 30EE | 0.1 | 0.07 | 0.04 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 18QB |  |  | -0.06 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.06 |
| 18QB |  |  | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 18QB |  |  | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 19PB | 19EE | 19B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 20PB | 20EE | 6 B | 0.1 | 0.07 | 0.04 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 20PB | 20EE | 16B | 0.14 | 0.09 | 0.06 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.04 | -0.05 |
| 20PB | 20EE | 20B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 20PB | 20EE | 21B | 0.13 | 0.09 | 0.06 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.04 | -0.04 |
| 20PB | 20EE | 28B | 0.1 | 0.07 | 0.04 | 0.02 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 20PB | 20EE | 29B | 0.1 | 0.07 | 0.04 | 0.02 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 20PB | 20EE | 30EE | 0.1 | 0.07 | 0.04 | 0.02 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 20QB |  | 6 B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 20QB |  | 16B | -0.07 | -0.06 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.06 | 0.07 |
| 20QB |  | 21B | -0.07 | -0.06 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.06 | 0.07 |
| 20QB |  | 28B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 20QB |  | 29B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 20QB |  | 30EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 21PB | 21EE | 21B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 21PB | 21EE | 27EE | 0.07 | 0.05 | 0.03 | 0.02 | 0.01 | 0 | -0.01 | -0.01 | -0.02 | -0.02 | -0.02 |
| 21PB | 21EE | 28B | 0.12 | 0.08 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.04 | -0.04 |
| 21PB | 21EE | 29B | 0.11 | 0.08 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.03 | -0.04 |
| 21PB | 21EE | 30EE | 0.12 | 0.08 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.03 | -0.04 |
| 21QB |  | 27EE | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 21QB |  | 28B | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 21QB |  | 29B | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 21QB |  | 30EE | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 22PB | 22EE | 22B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 23PB |  | 15 B | 0.63 | 0.39 | 0.24 | 0.13 | 0.06 | 0 | -0.05 | -0.08 | -0.11 | -0.14 | -0.16 |
| 23PB |  | 22B | 0.62 | 0.38 | 0.24 | 0.13 | 0.06 | 0 | -0.05 | -0.08 | -0.11 | -0.14 | -0.16 |
| 23PB |  | 23B | 1.45 | 0.9 | 0.55 | 0.31 | 0.13 | 0 | -0.11 | -0.19 | -0.26 | -0.32 | -0.37 |
| 23PB |  | 24EE | 1 | 0.62 | 0.38 | 0.21 | 0.09 | 0 | -0.07 | -0.13 | -0.18 | -0.22 | -0.26 |


| 23PB |  | 25EE | 1.07 | 0.67 | 0.41 | 0.23 | 0.1 | 0 | -0.08 | -0.14 | -0.19 | -0.24 | -0.27 |
| :---: | :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 23PB |  | 26EE | 0.75 | 0.46 | 0.28 | 0.16 | 0.07 | 0 | -0.05 | -0.1 | -0.13 | -0.16 | -0.19 |
| 23PB |  | 27EE | 0.46 | 0.28 | 0.17 | 0.1 | 0.04 | 0 | -0.03 | -0.06 | -0.08 | -0.1 | -0.12 |
| 23QB |  | 15 B | -0.23 | -0.19 | -0.15 | -0.1 | -0.05 | 0 | 0.05 | 0.1 | 0.15 | 0.19 | 0.23 |
| 23QB |  | 22B | -0.23 | -0.19 | -0.14 | -0.1 | -0.05 | 0 | 0.05 | 0.1 | 0.14 | 0.19 | 0.23 |
| 23QB |  | 23B | -0.08 | -0.07 | -0.05 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.05 | 0.07 | 0.08 |
| 23QB |  | 24EE | -0.37 | -0.3 | -0.23 | -0.16 | -0.08 | 0 | 0.08 | 0.15 | 0.22 | 0.3 | 0.38 |
| 23QB |  | 25EE | -0.4 | -0.33 | -0.25 | -0.17 | -0.09 | 0 | 0.09 | 0.15 | 0.23 | 0.31 | 0.4 |
| 23QB |  | 26EE | -0.28 | -0.23 | -0.17 | -0.12 | -0.06 | 0 | 0.06 | 0.12 | 0.14 | 0.19 | 0.25 |
| 23QB |  | 27EE | -0.17 | -0.14 | -0.11 | -0.07 | -0.04 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.06 |
| 24 B | 24 PB | 24 Ee | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 25 B | 25 PB | 25 Ee | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 26 B | 26 PB | 26 Ee | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 27 B | 27 PB | 27 Ee | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 28PB | 28EE | 28B | 2.49 | 1.33 | 0.75 | 0.4 | 0.17 | 0 | -0.12 | -0.22 | -0.3 | -0.36 | -0.42 |
| 28PB | 28EE | 29B | 1.95 | 1.04 | 0.59 | 0.31 | 0.13 | 0 | -0.1 | -0.17 | -0.23 | -0.28 | -0.33 |
| 28PB | 28EE | 30EE | 2.05 | 1.09 | 0.62 | 0.33 | 0.14 | 0 | -0.1 | -0.18 | -0.25 | -0.3 | -0.34 |
| 28QB |  | 28B | -0.18 | -0.15 | -0.11 | -0.08 | -0.04 | 0 | 0.04 | 0.09 | 0.15 | 0.21 | 0.27 |
| 28QB |  | 29B | -0.46 | -0.38 | -0.3 | -0.21 | -0.11 | 0 | 0.12 | 0.24 | 0.38 | 0.54 | 0.71 |
| 28QB |  | 30EE | -0.48 | -0.4 | -0.31 | -0.22 | -0.11 | 0 | 0.12 | 0.26 | 0.4 | 0.57 | 0.75 |
| 29PB | 29EE | 29B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 29PB | 29EE | 30EE | 0.81 | 0.54 | 0.35 | 0.2 | 0.09 | 0 | -0.07 | -0.13 | -0.19 | -0.23 | -0.27 |
| 29QB |  | 30EE | -0.4 | -0.32 | -0.24 | -0.16 | -0.08 | 0 | 0.08 | 0.16 | 0.24 | 0.32 | 0.4 |
| 30B |  | 30EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |

Appendix 3 Table 2. Results of the Sensitivity Analysis - Model 2.

|  |  |  | \% change in input parameter |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Input parameters |  | Estimated parameter | -50 | -40 | -30 | -20 | -10 | 0 | 10 | 20 | 30 | 40 | 50 |
| 1B | 1QB | 6B | -0.08 | -0.06 | -0.05 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.05 | 0.06 | 0.08 |
| 1B | 1QB | 14EE | -0.33 | -0.26 | -0.2 | -0.13 | -0.07 | 0 | 0.07 | 0.13 | 0.2 | 0.26 | 0.33 |
| 1B | 1QB | 15B | -0.06 | -0.05 | -0.04 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.04 | 0.05 | 0.06 |
| 1B | 1QB | 16B | -0.09 | -0.07 | -0.05 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.05 | 0.07 | 0.09 |
| 1B | 1QB | 17B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 1B | 1QB | 20B | -0.28 | -0.22 | -0.17 | -0.11 | -0.06 | 0 | 0.06 | 0.11 | 0.17 | 0.22 | 0.28 |
| 1B | 1QB | 21B | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 1B | 1QB | 28B | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 1B | 1QB | 29B | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 1B | 1QB | 30EE | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 2B | 2PB | 2EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 2B | 2QB | 6B | -0.16 | -0.13 | -0.09 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.09 | 0.13 | 0.16 |
| 2B | 2QB | 7EE | -0.1 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.1 |
| 2B | 2QB | 8B | -0.18 | -0.14 | -0.11 | -0.07 | -0.04 | 0 | 0.04 | 0.07 | 0.11 | 0.14 | 0.18 |
| 2B | 2QB | 10EE | -0.06 | -0.05 | -0.04 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.04 | 0.05 | 0.06 |
| 2B | 2QB | 11B | -0.2 | -0.16 | -0.12 | -0.08 | -0.04 | 0 | 0.04 | 0.08 | 0.12 | 0.16 | 0.2 |
| 2B | 2QB | 15B | -0.11 | -0.09 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.09 | 0.11 |
| 2B | 2QB | 16 EE | -0.1 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.1 |
| 2B | 2QB | 17B | -0.08 | -0.06 | -0.05 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.05 | 0.06 | 0.08 |
| 2B | 2QB | 18B | -0.36 | -0.29 | -0.22 | -0.14 | -0.07 | 0 | 0.07 | 0.14 | 0.22 | 0.29 | 0.36 |
| 2B | 2QB | 19B | -0.15 | -0.12 | -0.09 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.09 | 0.12 | 0.15 |
| 2B | 2QB | 20B | -0.1 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.1 |
| 2B | 2QB | 21B | -0.08 | -0.06 | -0.05 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.05 | 0.06 | 0.08 |
| 2B | 2QB | 22 B | -0.1 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.1 |
| 2B | 2QB | 25 EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 2B | 2QB | 26 EE | -0.09 | -0.07 | -0.05 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.05 | 0.07 | 0.09 |
| 2B | 2QB | 27 EE | -0.1 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.1 |
| 2B | 2QB | 28B | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 2B | 2QB | 29B | -0.06 | -0.05 | -0.04 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.04 | 0.05 | 0.06 |
| 2B | 2QB | 30EE | -0.06 | -0.05 | -0.04 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.04 | 0.05 | 0.06 |
| 3B | 3QB | 7EE | -0.13 | -0.1 | -0.08 | -0.05 | -0.03 | 0 | 0.03 | 0.05 | 0.08 | 0.1 | 0.13 |
| 3B | 3QB | 8 B | -0.16 | -0.13 | -0.1 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.1 | 0.13 | 0.16 |
| 3B | 3QB | 11B | -0.12 | -0.1 | -0.07 | -0.05 | -0.02 | 0 | 0.02 | 0.05 | 0.07 | 0.1 | 0.12 |
| 3B | 3QB | 15 B | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 3B | 3QB | 21 B | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 3B | 3QB | 26 EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 3B | 3QB | 27 EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | , | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 4B | 4PB | 4EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 5B | 5PB | 5EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 5B | 5PB | 6 B | -0.1 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.1 |
| 5B | 5QB | 8 B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 5B | 5QB | 10EE | -0.17 | -0.14 | -0.1 | -0.07 | -0.03 | 0 | 0.03 | 0.07 | 0.1 | 0.14 | 0.17 |
| 5B | 5QB | 13B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 5B | 5QB | 15 B | -0.11 | -0.09 | -0.07 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.07 | 0.09 | 0.11 |


| 5B | 5QB | 16EE | -0.12 | -0.09 | -0.07 | -0.05 | -0.02 | 0 | 0.02 | 0.05 | 0.07 | 0.09 | 0.12 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5B | 5QB | 17B | -0.1 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.1 |
| 5B | 5QB | 18 B | -0.06 | -0.05 | -0.04 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.04 | 0.05 | 0.06 |
| 5B | 5QB | 22B | -0.12 | -0.1 | -0.07 | -0.05 | -0.02 | 0 | 0.02 | 0.05 | 0.07 | 0.1 | 0.12 |
| 5B | 5QB | 23B | -0.18 | -0.14 | -0.11 | -0.07 | -0.04 | 0 | 0.04 | 0.07 | 0.11 | 0.14 | 0.18 |
| 5B | 5QB | 24EE | -0.11 | -0.09 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.09 | 0.11 |
| 5B | 5QB | 25EE | -0.13 | -0.1 | -0.08 | -0.05 | -0.03 | 0 | 0.03 | 0.05 | 0.08 | 0.1 | 0.13 |
| 5B | 5QB | 26EE | -0.09 | -0.07 | -0.05 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.05 | 0.07 | 0.09 |
| 6PB | 6EE | 6B | 1.06 | 0.7 | 0.45 | 0.26 | 0.11 | 0 | -0.09 | -0.17 | -0.24 | -0.29 | -0.34 |
| 6PB | 6EE | 16 EE | 0.06 | 0.04 | 0.03 | 0.02 | 0.01 | 0 | -0.01 | -0.01 | -0.01 | -0.02 | -0.02 |
| 6 PB | 6EE | 22 B | 0.14 | 0.09 | 0.06 | 0.03 | 0.02 | 0 | -0.01 | -0.02 | -0.03 | -0.04 | -0.04 |
| 6PB | 6EE | 23 B | 0.06 | 0.04 | 0.03 | 0.02 | 0.01 | 0 | -0.01 | -0.01 | -0.01 | -0.02 | -0.02 |
| 6PB | 6EE | 27 EE | 0.07 | 0.05 | 0.03 | 0.02 | 0.01 | 0 | -0.01 | -0.01 | -0.02 | -0.02 | -0.02 |
| 6 QB |  | 22 B | -0.07 | -0.05 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.05 | 0.07 |
| 6 QB |  | 27 EE | -0.03 | -0.03 | -0.02 | -0.01 | -0.01 | 0 | 0.01 | 0.01 | 0.02 | 0.03 | 0.03 |
| 7B |  | 6B | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 7B | 7PB | 7EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 7B | 7QB | 8 B | -0.08 | -0.06 | -0.05 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.05 | 0.06 | 0.08 |
| 7B | 7QB | 13B | -0.14 | -0.11 | -0.08 | -0.05 | -0.03 | 0 | 0.03 | 0.05 | 0.08 | 0.11 | 0.14 |
| 7B | 7QB | 19B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 7B | 7QB | 21B | -0.08 | -0.06 | -0.05 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.05 | 0.06 | 0.08 |
| 7B | 7QB | 22B | -0.03 | -0.03 | -0.02 | -0.01 | -0.01 | 0 | 0.01 | 0.01 | 0.02 | 0.03 | 0.03 |
| 8 PB | 8 EE | 8B | 1.06 | 0.7 | 0.45 | 0.26 | 0.11 | 0 | -0.09 | -0.17 | -0.24 | -0.29 | -0.34 |
| 8 PB | 8 EE | 16 EE | 0.15 | 0.1 | 0.06 | 0.04 | 0.02 | 0 | -0.01 | -0.02 | -0.03 | -0.04 | -0.05 |
| 8 PB | 8 EE | 18 B | 0.11 | 0.07 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 8QB |  | 16 EE | -0.07 | -0.05 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.05 | 0.07 |
| 8QB |  | 18 B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 9B | 9PB | 9EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 9B | 9QB | 17B | -0.07 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.07 |
| 9B | 9QB | 23B | -0.07 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.07 |
| 9B | 9QB | 24EE | -0.11 | -0.09 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.09 | 0.11 |
| 9B | 9QB | 25EE | -0.07 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.07 |
| 10B |  | 10EE | 0.72 | 0.48 | 0.31 | 0.18 | -0.08 | 0 | -0.07 | -0.12 | -0.17 | -0.21 | -0.24 |
| 10QB |  | 10EE | -0.14 | -0.11 | -0.08 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.08 | 0.11 | 0.14 |
| 10B | 10QB | 15B | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 10B | 10QB | 16 EE | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 10B | 10QB | 17B | -0.06 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.06 |
| 10B | 10QB | 23 B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 10B | 10QB | 24EE | -0.07 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.07 |
| 10B | 10QB | 25EE | -0.06 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.06 |
| 10B | 10QB | 26EE | -0.06 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.06 |
| 10B | 10QB | 27EE | -0.06 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.06 |
| 10PB |  | 10EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 11PB |  | 11B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 11PB |  | 25 EE | 0.08 | 0.05 | 0.03 | 0.02 | 0.01 | 0 | -0.01 | -0.01 | -0.02 | -0.02 | -0.03 |
| 11PB |  | 26EE | 0.32 | 0.21 | 0.14 | 0.08 | 0.04 | 0 | -0.03 | -0.05 | -0.07 | -0.09 | -0.11 |
| 11PB |  | 27EE | 0.24 | 0.16 | 0.1 | 0.06 | 0.03 | 0 | -0.02 | -0.04 | -0.06 | -0.07 | -0.08 |
| 11 QB |  | 25 EE | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 11QB |  | 26EE | -0.16 | -0.13 | -0.1 | -0.06 | -0.03 | 0 | 0.03 | 0.06 | 0.1 | 0.13 | 0.16 |
| 11QB |  | 27EE | -0.12 | -0.1 | -0.07 | -0.05 | -0.02 | 0 | 0.02 | 0.05 | 0.07 | 0.1 | 0.12 |
| 12 B |  | 6B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |


| 12B | 12PB | 12EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
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| 12B | 12QB | 13B | -0.07 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.07 |
| 12B | 12QB | 15 B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 12B | 12QB | 21B | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 12B | 12QB | 22B | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 12B | 12QB | 23B | -0.13 | -0.1 | -0.08 | -0.05 | -0.03 | 0 | 0.03 | 0.05 | 0.08 | 0.1 | 0.13 |
| 12B | 12QB | 24EE | -0.08 | -0.06 | -0.05 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.05 | 0.06 | 0.08 |
| 12B | 12QB | 25EE | -0.07 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.07 |
| 12B | 12QB | 26EE | -0.07 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.07 |
| 12B | 12QB | 27 EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 13PB | 13EE | 13B | 1.1 | 0.72 | 0.46 | 0.26 | 0.12 | 0 | -0.09 | -0.17 | -0.24 | -0.3 | -0.34 |
| 13PB | 13EE | 21B | 0.6 | 0.39 | 0.25 | 0.14 | 0.06 | 0 | -0.05 | -0.09 | -0.13 | -0.16 | -0.19 |
| 13PB | 13EE | 22B | 0.14 | 0.09 | 0.06 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.04 | -0.04 |
| 13PB | 13EE | 28B | 0.11 | 0.07 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 13PB | 13EE | 29B | 0.1 | 0.07 | 0.04 | 0.02 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 13PB | 13EE | 30EE | 0.1 | 0.07 | 0.04 | 0.02 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 13QB |  | 21B | -0.28 | -0.22 | -0.17 | -0.11 | -0.06 | 0 | 0.06 | 0.11 | 0.17 | 0.22 | 0.28 |
| 13QB |  | 22B | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 13QB |  | 28B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 13QB |  | 29B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 13QB |  | 30EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 14B | 14QB | 10EE | -0.06 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.06 |
| 14B | 14PB | 14 EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 14B | 14QB | 15 B | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 14B | 14QB | 22 B | -0.07 | -0.05 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.0 | 0.05 | 0.07 |
| 14B | 14QB | 24 EE | -0.06 | -0.05 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.05 | 0.06 |
| 14B | 14QB | 25 EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 14B | 14QB | 26 EE | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 14B | 14QB | 27 EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 15 PB | 15 EE | 15 B | 1.12 | 0.73 | 0.46 | 0.27 | 0.12 | 0 | -0.1 | -0.17 | -0.24 | -0.3 | -0.35 |
| 15 PB | 15 EE | 22B | 0.11 | 0.07 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 15 PB | 15 EE | 23B | 0.08 | 0.05 | 0.03 | 0.02 | 0.01 | 0 | -0.01 | -0.01 | -0.02 | -0.02 | -0.03 |
| 15 PB | 15 EE | 24EE | 0.17 | 0.11 | 0.07 | 0.04 | 0.02 | 0 | -0.01 | -0.03 | -0.04 | -0.05 | -0.05 |
| 15 PB | 15 EE | 25EE | 0.3 | 0.2 | 0.13 | 0.07 | 0.03 | 0 | -0.03 | -0.05 | -0.07 | -0.08 | -0.09 |
| 15 PB | 15 EE | 26EE | 0.21 | 0.14 | 0.09 | 0.05 | 0.02 | 0 | -0.02 | -0.03 | -0.05 | -0.06 | -0.07 |
| 15 PB | 15 EE | 27EE | 0.39 | 0.26 | 0.16 | 0.09 | 0.04 | 0 | -0.03 | -0.06 | -0.08 | -0.1 | -0.12 |
| 15QB |  | 22B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 15QB |  | 23B | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 15QB |  | 24EE | -0.08 | -0.06 | -0.05 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.05 | 0.06 | 0.08 |
| 15QB |  | 25EE | -0.14 | -0.11 | -0.08 | -0.05 | -0.03 | 0 | 0.03 | 0.05 | 0.08 | 0.11 | 0.14 |
| 15QB |  | 26EE | -0.1 | -0.08 | -0.06 | -0.04 | -0.02 | 0 | 0.02 | 0.04 | 0.06 | 0.08 | 0.1 |
| 15QB |  | 27EE | -0.18 | -0.14 | -0.11 | -0.07 | -0.04 | 0 | 0.04 | 0.07 | 0.11 | 0.14 | 0.18 |
| 16 B |  | 16 EE | 0.96 | 0.64 | 0.41 | 0.24 | 0.11 | 0 | -0.09 | -0.16 | -0.22 | -0.27 | -0.32 |
| 16 B | 16QB | 17B | -0.11 | -0.09 | -0.07 | -0.05 | -0.02 | 0 | 0.02 | 0.05 | 0.07 | 0.09 | 0.11 |
| 16 B | 16QB | 28B | -0.28 | -0.23 | -0.17 | -0.11 | -0.06 | 0 | 0.06 | 0.11 | 0.17 | 0.23 | 0.28 |
| 16 B | 16QB | 29B | -0.29 | -0.23 | -0.17 | -0.12 | -0.06 | 0 | 0.06 | 0.12 | 0.17 | 0.23 | 0.29 |
| 16 B | 16QB | 30EE | -0.29 | -0.23 | -0.17 | -0.12 | -0.06 | 0 | 0.06 | 0.12 | 0.17 | 0.23 | 0.29 |
| 16 PB |  | 16 EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 17PB | 17EE | 17B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 17PB | 17EE | 28B | 0.14 | 0.09 | 0.06 | 0.04 | 0.02 | 0 | -0.01 | -0.02 | -0.03 | -0.04 | -0.05 |
| 17PB | 17EE | 29B | 0.16 | 0.11 | 0.07 | 0.04 | 0.02 | 0 | -0.01 | -0.03 | -0.04 | -0.05 | -0.05 |


| 17PB | 17EE | 30EE | 0.16 | 0.11 | 0.07 | 0.04 | 0.02 | 0 | -0.01 | -0.03 | -0.04 | -0.04 | -0.05 |
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| 17QB |  | 28B | -0.07 | -0.06 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.06 | 0.07 |
| 17QB |  | 29B | -0.08 | -0.06 | -0.05 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.05 | 0.06 | 0.08 |
| 17QB |  | 30EE | -0.08 | -0.06 | -0.05 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.05 | 0.06 | 0.08 |
| 18PB | 18 EE | 18 B | 1.03 | 0.68 | 0.44 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 18B | 18 EE | 28B | 0.1 | 0.07 | 0.04 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 18B | 18 EE | 29B | 0.09 | 0.06 | 0.04 | 0.02 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 18B | 18 EE | 30EE | 0.09 | 0.06 | 0.04 | 0.02 | 0.01 | 0 | -0.01 | -0.02 | -0.02 | -0.03 | -0.03 |
| 18QB |  | 28B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 18QB |  | 29B | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 18QB |  | 30EE | -0.05 | -0.04 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.04 | 0.05 |
| 19PB | 19EE | 19B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 20PB | 20EE | 6 B | 0.12 | 0.08 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.03 | -0.04 |
| 20PB | 20EE | 16 EE | 0.14 | 0.09 | 0.06 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.04 | -0.05 |
| 20PB | 20EE | 20B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 20PB | 20EE | 21B | 0.13 | 0.09 | 0.06 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.04 | -0.04 |
| 20 QB |  | 6 B | -0.06 | -0.05 | -0.03 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.03 | 0.05 | 0.06 |
| 20QB |  | 16 EE | -0.07 | -0.06 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.06 | 0.07 |
| 20QB |  | 21B | -0.07 | -0.06 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.06 | 0.07 |
| 21PB | 21EE | 21B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 21PB | 21EE | 27EE | 0.09 | 0.06 | 0.04 | 0.02 | 0.01 | 0 | -0.01 | -0.01 | -0.02 | -0.03 | -0.03 |
| 21PB | 21EE | 28B | 0.12 | 0.08 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.04 | -0.04 |
| 21PB | 21EE | 29B | 0.11 | 0.08 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.03 | -0.04 |
| 21PB | 21EE | 30EE | 0.11 | 0.08 | 0.05 | 0.03 | 0.01 | 0 | -0.01 | -0.02 | -0.03 | -0.03 | -0.04 |
| 21QB |  | 27EE | -0.04 | -0.03 | -0.02 | -0.02 | -0.01 | 0 | 0.01 | 0.02 | 0.02 | 0.03 | 0.04 |
| 21QB |  | 28B | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 21QB |  | 29B | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 21QB |  | 30EE | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 22PB | 22EE | 22B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 23PB |  | 22B | 0.51 | 0.32 | 0.2 | 0.11 | 0.05 | 0 | -0.04 | -0.07 | -0.1 | -0.12 | -0.14 |
| 23PB | 23 EE | 23B | 1.26 | 0.81 | 0.5 | 0.29 | 0.13 | 0 | -0.1 | -0.18 | -0.25 | -0.31 | -0.36 |
| 23PB | 23 EE | 24EE | 0.69 | 0.44 | 0.27 | 0.16 | 0.07 | 0 | -0.05 | -0.1 | -0.14 | -0.17 | -0.2 |
| 23PB | 23 EE | 25EE | 0.58 | 0.37 | 0.23 | 0.13 | 0.06 | 0 | -0.05 | -0.08 | -0.12 | -0.14 | -0.16 |
| 23PB | 23 EE | 26EE | 0.5 | 0.32 | 0.2 | 0.11 | 0.05 | 0 | -0.04 | -0.07 | -0.1 | -0.12 | -0.14 |
| 23PB | 23 EE | 27EE | 0.2 | 0.13 | 0.08 | 0.05 | 0.02 | 0 | -0.02 | -0.03 | -0.04 | -0.05 | -0.06 |
| 23QB |  | 22B | -0.07 | -0.06 | -0.04 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.04 | 0.06 | 0.07 |
| 23QB |  | 23B | -0.06 | -0.05 | -0.04 | -0.03 | -0.01 | 0 | 0.01 | 0.03 | 0.04 | 0.05 | 0.06 |
| 23QB |  | 24EE | -0.29 | -0.23 | -0.18 | -0.12 | -0.06 | 0 | 0.06 | 0.12 | 0.18 | 0.23 | 0.29 |
| 23QB |  | 25EE | -0.24 | -0.2 | -0.15 | -0.1 | -0.05 | 0 | 0.05 | 0.1 | 0.15 | 0.2 | 0.24 |
| 23QB |  | 26EE | -0.21 | -0.17 | -0.13 | -0.09 | -0.04 | 0 | 0.04 | 0.09 | 0.13 | 0.17 | 0.21 |
| 23QB |  | 27EE | -0.08 | -0.07 | -0.05 | -0.03 | -0.02 | 0 | 0.02 | 0.03 | 0.05 | 0.07 | 0.08 |
| 24 B | 24 PB | 24 EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 25 B | 25 PB | 25 EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 26 B | 26 PB | 26 EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 27 B | 27 PB | 27 EE | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| 28PB | 28EE | 28B | 2.49 | 1.33 | 0.75 | 0.4 | 0.17 | 0 | -0.12 | -0.22 | -0.3 | -0.36 | -0.42 |
| 28PB | 28EE | 29B | 1.95 | 1.04 | 0.59 | 0.31 | 0.13 | 0 | -0.1 | -0.17 | -0.23 | -0.28 | -0.33 |
| 28PB | 28EE | 30EE | 2.05 | 1.09 | 0.62 | 0.33 | 0.14 | 0 | -0.1 | -0.18 | -0.25 | -0.3 | -0.34 |
| 28QB |  | 28B | -0.18 | -0.15 | -0.11 | -0.08 | -0.04 | 0 | 0.04 | 0.09 | 0.15 | 0.21 | 0.27 |
| 28QB |  | 29B | -0.46 | -0.38 | -0.3 | -0.21 | -0.11 | 0 | 0.12 | 0.24 | 0.38 | 0.54 | 0.71 |
| 28QB |  | 30EE | -0.48 | -0.4 | -0.31 | -0.22 | -0.11 | 0 | 0.12 | 0.26 | 0.4 | 0.57 | 0.75 |


| 29PB | 29EE | 29B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |
| :--- | :--- | :---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 29PB | 29EE | 30EE | 0.81 | 0.54 | 0.35 | 0.2 | 0.09 | 0 | -0.07 | -0.13 | -0.19 | -0.23 | -0.27 |
| 29QB |  | 30EE | -0.4 | -0.32 | -0.24 | -0.16 | -0.08 | 0 | 0.08 | 0.16 | 0.24 | 0.32 | 0.4 |
| 30PB | 30EE | 30B | 1 | 0.67 | 0.43 | 0.25 | 0.11 | 0 | -0.09 | -0.17 | -0.23 | -0.29 | -0.33 |


[^0]:    ${ }^{1}$ For further information, see http:<br>www.ecopath.org

[^1]:    ${ }^{2}$ Note that marine mammals are also caught by fishing gear as by-catch, for example, seals are caught in lumpfish gear and whales in gillnets and traps (B. Montevecchi, pers. comm.).

