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Reference points and assessment update for American Plaice (Hippoglossoides platessoides) in NAFO SA2 + Div. 3K and Subdiv. 3Ps

Points de référence et mise à jour de l'évaluation de la plie canadienne (Hippoglossoides platessoides) de la sous-zone 2 de l'OPANO, division 3K et de la sous-division 3Ps

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#### Abstract

Survey indices from research vessel (RV) and landings information were updated for Subarea (SA) 2 + Div. 3K and Subdiv. 3Ps American Plaice (Hippoglossoides platessoides). Limit reference points (LRP) were determined for these two stocks and stock status updated relative to these points.

In SA2 + Div. 3K the stock declined to very low levels by the early 1990s. Total mortality due to all causes, including fishing, has been decreasing on more recent cohorts but, stock biomass remains at only $10 \%$ of the values of the mid-1980s. An empirical biological LRP was determined from examining stock recruit data from the research vessel survey. Generally recruitment has been impaired when the survey spawning stock biomass (SSB) index is below $70,000 \mathrm{t}$ and therefore this was chosen as the LRP. The 2009 estimate of survey spawning stock biomass (SSB; the most recent year for which age data are available) indicates that the stock is at $24 \%$ of the LRP. It was not possible to determine an upper reference point or a removals ( F -based) reference point for this stock.

In Subdiv. 3Ps a Bayesian Surplus Production Model was applied to landings data from 1960 to 2010 and survey data from 1980 to 2010. Consistent with the Fisheries and Oceans Canada (DFO) Precautionary Approach (PA) policy $\mathrm{B}_{\text {lim }}$ is $40 \% \mathrm{~B}_{\text {msy }}$, the upper stock reference is $80 \%$ $\mathrm{B}_{\text {msy }}$ and $\mathrm{F}_{\text {lim }}$ is $\mathrm{F}_{\text {msy }}$. Stock status relative to these reference points was estimated from the model. Stock size estimated from the surplus production model decreased fairly steadily from the late 1960s to reach a low in 1993 of less than $10 \%$ of $B_{\text {msy }}$. Stock size has been increasing slowly since 1993; however, current biomass is $50 \%$ of $\mathrm{B}_{\text {lim }}$ and therefore the stock is in the critical zone. The probability of being below $\mathrm{B}_{\text {lim }}$ is high (0.94). Current fishing mortality is estimated to be $64 \%$ of $\mathrm{F}_{\text {lim }}$. The probability of being above $\mathrm{F}_{\text {lim }}$ is 0.2 .


## RÉSUMÉ

Les indices des relevés provenant du navire de recherche (NR) et les renseignements sur les débarquements ont été mis à jour pour la plie canadienne (Hippoglossoides platessoides) de la sous-zone 2, division 3K et de la sous-division 3Ps. On a déterminé les points de référence limites (PRL) pour ces deux stocks et on a mis à jour l'état de ces stocks en fonction de ces points.

Dans la sous-zone 2, division 3 K , le stock a diminué à des niveaux très bas vers le début des années 90. La mortalité totale, attribuable à l'ensemble des causes, y compris la pêche, a diminué pour les cohortes plus récentes, mais la biomasse du stock demeure à seulement $10 \%$ des valeurs du milieu des années 80 . Un PRL empirique biologique a été déterminé à partir de l'examen des données sur les recrues du stock tirées du relevé du navire de recherche. En règle générale, le recrutement s'était détérioré quand l'indice du relevé de la biomasse du stock reproducteur (BSR) est inférieur à 70000 t et, par conséquent, ceci a été choisi comme le PRL. L'estimation de 2009 de la biomasse du stock reproducteur (BSR; l'année la plus récente pour laquelle les données sont disponibles) indique que le stock est à $24 \%$ du PRL. Il n'a pas été possible de déterminer un point de référence supérieur ni un point de référence d'exploitation (fondé sur le taux de mortalité par la pêche) pour ce stock.

Un modèle bayésien de production excédentaire a été appliqué aux données de débarquements entre 1960 et 2010 et aux données des relevés effectués entre 1980 et 2010. Conformément à la politique de l'approche de précaution (AP) de Pêches et Océans Canada (MPO), le $\mathrm{B}_{\mathrm{lim}}$ est établi à $40 \%$ de la $\mathrm{B}_{\mathrm{rms}}$, le point de référence supérieur du stock est établi à $80 \%$ de la $B_{r m s}$ et le $F_{\text {lim }}$ correspond au $F_{\text {rms }}$. L'état du stock par rapport à ces points de référence a été estimé à partir du modèle. La taille du stock estimée à partir du modèle de production excédentaire a connu un déclin assez constant à partir de la fin des années 60 pour atteindre son point le plus bas en 1993 avec moins de $10 \%$ de la $\mathrm{B}_{\mathrm{rms}}$. La taille du stock augmente lentement depuis 1993, toutefois, la biomasse actuelle est à $50 \%$ de la valeur du $\mathrm{B}_{\mathrm{lim}}$ et, par conséquent, la valeur du stock est dans la zone critique. La probabilité que la valeur du stock soit inférieure à la valeur du $\mathrm{B}_{\text {lim }}$ est élevée $(0,94)$. La mortalité actuelle par la pêche est estimée à $64 \%$ de la valeur du $\mathrm{F}_{\text {lim }}$. La probabilité qu'elle soit supérieure à la valeur du $\mathrm{F}_{\text {lim }}$ est de 0,2.

## INTRODUCTION

Ecosystem and Fisheries Management (EFM) has requested identification of reference points for Subarea (SA) 2 + Div.3K and Subdiv. 3Ps American Plaice (Hippoglossoides platessoides) stocks in order to apply the Fisheries and Oceans Canada (DFO) Precautionary Approach (PA) framework (DFO 2009). The limit reference point (LPR) and the upper stock reference will be used to define the Critical, Cautious, and Healthy Zones.

The American Plaice stock in SA2 + Div. 3K was last fully assessed in 2003 and the Subdiv. 3Ps stock was assessed in 2005. This paper provides an update on survey indices and landings (bycatch) for both stocks. Based on fishery and survey data, biomass and removal reference points (upper, limit) are attempted for each stock.

There are number of methods for determining LRPs for fish stocks. These may range from simple empirical techniques to more-involved modeling approaches. Methods were attempted for both stocks, and from this, the status of each stock was provided relative to the reference point given.

## ASSESSMENT UPDATE

## FISHERY

## SA2 + Div. 3K

Landings (from Zonal Interface Format File - ZIFF) increased steadily throughout the 1960s, peaking at 12,686 tin 1970 (Table 1, Fig. 1) and were well below total allowable catch (TAC) quotas for all years since 1982 to 1997. After the declaration of the 200-mile-limit in 1977, catches by non-Canadian fleets were greatly reduced, with the result that the total landings from the stock exceeded 2,000 t on only two occasions after 1981. Reported landings from 1994 to 1999 were less than 30 t per year mostly as by-catch in gillnet fisheries and these are the lowest in the time series. This is due to a large reduction in the TAC (Fig. 1), as well as the moratorium in 1994 on directed fishing (a by-catch TAC remained in place until 1997) and limited fisheries on northern cod (Div. 2J3KL), which, after 1992, essentially eliminated a major source of American Plaice by-catch. The main source of by-catch of American Plaice since 2000 has been in the Greenland Halibut gillnet fishery, and in recent years, the otter trawl fishery for this species. In 2010 and 2011, the total reported landings of American Plaice were $22 t$ and 17 t , respectively. In both years, $97 \%$ of the American Plaice by-catch (in tonnes) came from the Greenland halibut otter trawl fishery, mostly in Div. 3K.

Discards of American Plaice in the shrimp fishery in SA2 + Div. 3K are estimated to be about 13 t to 27 t in the most recent years for which there are results (2006-2009) (Orr 2010). This is based on analysis of observer data collected at sea on the large and small vessel shrimp fleets. These estimates may not be accurate because the minimum measurement of a sample is 1 kg (thus numbers are an overestimate). In addition, shrimp fishing areas do not overlap Northwest Atlantic Fisheries Organization (NAFO) areas exactly. Finally, the timing of the shrimp fishery is from April 1 to March 31. All of these factors will have an impact on the actual estimate of bycatch. The shrimp fishery incidentally takes American Plaice that are less than 30 cm , with large peaks in numbers of some years at 15 cm . Thus the number of small plaice taken in the shrimp fishery is not accurately known.

## Subdiv. 3Ps

Landings (ZIFF) from this stock were highest from 1968 to 1973, exceeding 12,000 t on three occasions in this period (Table 2, Fig. 2). Since 1977 only Canada and France have been involved in this fishery. After the implementation of the 200-mile limit in 1977, landings ranged between $2,000 \mathrm{t}$ to $4,000 \mathrm{t}$ up to 1992. Landings averaged just under $4,000 \mathrm{t}$ during the 1980 s but rapidly declined after 1991. Subsequently there has been a moratorium on direct fishing of American Plaice in Subdiv. 3Ps since September 1993. Since then, there has been only bycatch of American Plaice in other fisheries, which increased substantially since 1995, and was over $1,000 \mathrm{t}$ in each year from 2001 to 2003. However, by-catch has been declining since then and is 402 t and 273 t in 2010 and 2011, respectively.

Data from the Newfoundland reported catch statistics (ZIFF) were examined to determine which fisheries were taking the greatest portion of the American Plaice by-catch. The by-catch of American Plaice is taken in two main fisheries, the directed cod and Witch Flounder fisheries. In 2010, 50\% of the total Newfoundland catch ( t ) came from the Witch Flounder otter trawl fishery and $30 \%$ was taken from the cod gillnet fishery. In 2011, this was somewhat different: $25 \%$ of the by-catch of American Plaice was taken in the Witch Flounder otter trawl fishery and $40 \%$ was taken in the cod gillnet fishery.

## RESEARCH VESSEL (RV) SURVEY

## SA2 + Div. 3K

## Abundance and Biomass

DFO has conducted stratified random bottom trawl surveys in Div. 2G, 2H, 2J, and 3K since the late 1970s, although not annually in Div. 2GH. In 1995, the survey trawl was switched from an Engel 145 Hi-lift trawl with bobbin footgear to a Campelen 1800 shrimp trawl with rockhopper footgear (McCallum and Walsh 1996). The Campelen trawl is more effective in capturing small fish. A comparative fishing experiment was carried out to quantify the differences, and the results are found in Warren (1996). Morgan and Brodie (2000) converted the results of surveys in Div. 2J and 3K from 1978 to 1994 into Campelen-equivalent units for American Plaice. The surveys in Div. 2GH were not converted.

From 1995 to 2011, the fall surveys covered Subarea $2+$ Div. 3KLMNO although the coverage was not comparable in all years, particularly in Div. 2G and inshore Div. 3K. Inshore strata were added in Div. 3K in 1996 and Div. 2H was surveyed in 1996-1999, 2001, 2004, 2006, 2008, and 2010-2011. Div. 2G was last surveyed in 1999.

Biomass and abundance estimates in Div. 2H, although low, have been increasing over the Campelen time series (beginning in 1996) (Table 3, Fig. 3).

Biomass and abundance declined in the late 1980s in Div. 2J3K to a low level in 2002. Since then, both biomass and abundance have increased slightly. Abundance in 2011 is highest since 1994, when the moratorium was instituted (Table 3, Fig. 4). However, current biomass in Div. 2 J 3 K is at $10 \%$ and abundance at $25 \%$ of the average of the mid-eighties.

Ageing of American Plaice in Div. 2J3K has not been updated since 2009. However, length frequency plots from 2009 to 2011 indicate that there is a large peak in numbers of American Plaice at 20 cm in 2011 (Fig. 5).

A comparison of the expanding symbol plots of numbers/tow in Subarea 2 + Div. 3K in 1981 and the most recent years also indicate a decline in abundance (Fig. 6-8). In addition, fewer fish are found on the banks and nearshore in recent years, compared to the 1980s, coinciding with a shift to deeper water that occurred during that time period (Dwyer et al. 2003).

## Mortality

A general linear model was fit to survey data for years 1978-2009 (equivalent to catch curve analysis). The model estimates total mortality $(Z)$ based on the slope of the log of the survey numbers at age for fish in a cohort. Younger American Plaice are not fully selected by the survey gear as illustrated in Fig. 9. Because of this the decision was made to omit data for ages 0 to 6 from the analysis on the basis that these ages are not fully selected to the survey gear.

$$
\log \left(\mathrm{l}_{\mathrm{a}, \mathrm{c}}\right)=\text { int }+ \text { age }+ \text { age }^{*} \mathrm{cohort}_{\mathrm{c}}
$$

Where $\mathrm{I}_{\mathrm{a}, \mathrm{c}}$ is the survey index at age a for cohort ${ }_{\mathrm{c}}$, int is the intercept, age is the age effect (slope) and age*cohort ${ }_{c}$ is the interaction term between age and cohort (separate slope for each cohort) where cohort is a categorical variable.

The equation for the estimate of the mortality rate for a cohort from this model is

$$
\mathrm{Z}_{\mathrm{c}}=\text { age }+ \text { age }^{*} \text { cohort }_{c}
$$

Estimates of $Z_{c}$ are plotted in Fig. 10.
Because cohorts at the beginning and end of the survey series are incomplete these were trimmed from the output before plotting and only data for cohorts 1966 to 1998 are included. Because agedisaggregated survey data for Div. 2J3K has been completed only up to 2009, this means that the 1998 cohort estimate of $Z$ is based only on data for ages 7-11. There was a significant difference between cohorts indicating change in $Z$ over time ( $F=19.02$, df $=39, P<0.0001$ ). $Z$ increased progressively on cohorts that arose from 1970 onwards, reaching a peak of over 1.4 on the 1983 cohort. Thereafter $Z$ showed a decreasing trend reaching a value of $<0.2$ on the most recent cohorts for which there are sufficient information to reliably compute a mortality rate.

## Maturity

Age and length at $50 \%$ maturity were produced from DFO survey data. Estimates were produced by sex and cohort for each population using generalized linear models with a logit link function and binomial error.

There was a significant effect of cohort on proportion mature at age for males $\left(X^{2}=618, \mathrm{df}=34\right.$, $P<0.0001$ ) and females ( $\mathrm{X}^{2}=1838, \mathrm{df}=40, \mathrm{P}<0.0001$ ) and also on proportion mature at length for males ( $X^{2}=945$, df $=36, P<0.0001$ ) and females ( $X^{2}=3377$, df $=41, P<0.0001$ ). In SA2 + Div.3K, age at $50 \%$ maturity (A50) has declined from just under 11 years to around 7 years of age for females (Fig. 11). For males, A50 has declined from around 7 years to just over 4 years of age. There has been some increase in A50 for males of recent cohorts, but such a short term increase has been seen in the time series previously and this will need to continue before it can be
considered a reversal of the decline. Length at $50 \%$ maturity (L50) has also declined for both sexes in SA2 + Div. 3K (Fig. 12). For males, L50 was about 22 cm at the beginning of the time series and for recent cohorts is about 18 cm . Female L50 declined from about 38 cm at the beginning of the time series to about 31 cm recently.

## Spawning Stock Biomass

An index of spawning stock biomass was calculated from the survey by multiplying estimated maturity at age by abundance at age and weight at age in the survey. The spawning stock biomass (SSB) index declined rapidly after 1982 and reached its lowest level in 2003. There was some increase after 2006 but the average of 2007 to 2009 is only $15 \%$ of the average over the 1978 to 1982 period (Fig. 13). Ageing data has not been completed since 2009; hence the time series ends in 2009.

## Recruitment

For SA2 + Div. 3K a relative recruitment index was estimated from DFO survey data using a general linear model applied to ages 3-5, in the same fashion as in Dwyer et al. (2003). Only those cohorts that were observed at least twice in the data were used in the estimation. Relative cohort strengths were estimated using the following model:

$$
\log \left(N_{a, y}\right)=\mu+A_{a}+Y_{y}+\varepsilon_{a, y}
$$

where:

$$
\begin{aligned}
& \mu=\text { intercept } \\
& a=\text { age subscript, age } 3 \text { to } 5 \\
& y=\text { cohort subscript } \\
& N=\text { survey index (abundance in millions) } \\
& A=\text { age effect } \\
& Y=\text { cohort effect } \\
& \varepsilon=\text { error. }
\end{aligned}
$$

There was a significant cohort effect, that is a significant difference in cohort strength ( $F=4.7$, df $=31, \mathrm{P}<0.0001$ ). In SA2 + Div. 3K recruitment declined from the mid-1980s to the late 1990s. There has been an increase in recruitment to approximately $40 \%$ of the average recruitment estimated prior to the mid-1980s (Fig. 14).

## Stock Recruit

As the survey occurs in the fall the SSB index in year y was taken as giving rise to recruits in year $y+1$. A segmented regression was fit to the stock recruit data using the Julious algorithm and maximum likelihood (Julious 2001). The stock recruit data with the fit from the Julious algorithm are shown in Fig. 15. The data do not seem to be well described by the segmented regression. The two methods of fitting the model give very different results with different
estimated break points. However, there is a relationship between recruitment and SSB index with recruitment increasing with SSB index. Generally recruitment is at a lower level when the SSB index is below 70 and this could serve as a LRP for this stock. The 2009 estimate of survey SSB indicates that the stock is at $24 \%$ of the LRP. Estimates are not yet available for 2010 and 2011; however, there has been limited increase in total biomass from the survey and the stock is very unlikely to have increased above $\mathrm{B}_{\text {lim }}$, the biomass limit reference point.

## Subdiv. 3Ps

## Abundance and Biomass

Stratified-random surveys have been conducted by DFO in Subdiv. 3Ps in each year from 1972 to 2011. Coverage prior to 1980 was poor. There were two surveys in 1993, one in February and one in April. Most of the surveys prior to 1993 were in February/March, while those since 1993 have been in April. The data can be split into three time periods based on the trawl used in each period: 1971-1982 was Yankee 36, 1983-1995 was Engel 145, and 1996-2011 was Campelen 1800 (see McCallum and Walsh (1996) for a description of the various trawls). There is a conversion between the second and third survey gears (Morgan et al. 1998) but not the first and third. Only Campelen and Campelen-equivalent units are discussed in this section.

Inshore strata were added in Placentia Bay in 1994 and more were added in Fortune Bay in 1997.

Biomass and abundance indices from 1983 to 2011 are shown in Fig. 16. From the mid-1980s to 1990 there was a large decline in both biomass and abundance indices. Stock size was lowest in the early 1990s. There has been an increase over the 1992-2011 period for both biomass and abundance indices but current biomass is at $25 \%$ and abundance at $52 \%$ of the average of the mid-1980s.

Age data for American Plaice in Subdiv. 3Ps has not been updated since 2009. However, length frequency plots from 2009 to 2011 indicate that there is a large peak in numbers at 10 cm in 2009 (likely age 2) that can be followed to a peak at 20 cm in 2011 (Fig. 17).

American Plaice are distributed (numbers/tow) throughout Subdiv. 3Ps (Fig. 18-20). For comparison, a year in which the index was in Engel units (non-converted) (1983) was provided (Fig. 18). This showed American Plaice catches in that year were distributed throughout the Halibut Channel and onto St. Pierre Bank. In the more recent time period (2007-2011) there tend to be fewer fish in the Halibut Channel and there are now some larger sets on southern St. Pierre Bank (Fig. 19-20). There are also some larger catches in Fortune Bay.

## Mortality

The same general linear model as used in SA2 + Div. 3K was fit to survey data for Subdiv. 3Ps for years 1983 to 2009. Younger American Plaice are not fully selected by the survey gear as illustrated in Fig. 21. Because of this the decision was made to omit data for ages 0 to 6 from the analysis on the basis that these ages are not fully selected. Estimates of $Z_{c}$ are shown in Fig. 22.

Because cohorts at the beginning and end of the survey series are incomplete these were trimmed from the output before plotting and only data for cohorts 1973 to 1998 are included. Because age-disaggregated survey data for Subdiv. 3Ps are only available up to 2009, this
means that the 1998 cohort estimate of $Z$ is based only on data for ages $7-11$. There was a significant difference between cohorts indicating change in $Z$ over time ( $F=11.31$, df = 34, $\mathrm{P}<0.0001$ ). Z increased progressively on cohorts that arose from the mid-1970s onwards, reaching a peak of nearly 1.2 on the 1979 cohort. Thereafter $Z$ showed a generally decreasing trend reaching except for those cohorts of the early 1990s that were subject to the higher catches of the early 2000s. $Z$ has reached values of just below 0.3 on the most recent cohorts for which there are sufficient information to reliably compute a mortality rate (Fig. 22).

## Maturity

Age and length at $50 \%$ maturity were produced from DFO survey data. Estimates were produced by sex and cohort for each population using generalized linear models with a logit link function and binomial error.

American Plaice in Subdiv. 3Ps show a similar decline in A50 and L50 as in SA2 + Div. 3K (Fig. 23 and 24). There was a significant effect of cohort on proportion mature at age for males ( $x^{2}=595, \mathrm{df}=38, \mathrm{p}<0.0001$ ) and females ( $\mathrm{x}^{2}=943$, $\mathrm{df}=40, \mathrm{p}<0.0001$ ) and also on proportion mature at length for males ( $X^{2}=1086, \mathrm{df}=40, \mathrm{p}<0.0001$ ) and females ( $\mathrm{X}^{2}=756$, $\mathrm{df}=41, \mathrm{p}<0.0001$ ). Male A50 has declined from about 7 years to less than 4.5 years, while female A50 has declined from about 11 years to just under 9 years. Male L50 has declined from about 27 cm to less than 19 cm and female L50 has declined from about 40 cm to around 36 cm .

## Spawning Stock Biomass

An index of spawning stock biomass was calculated from the survey by multiplying estimated maturity at age by abundance at age and weight at age in the survey. The SSB index showed a major decline from the mid-1980s to the early 1990s and has shown some increase after 1997. The SSB index from 2007 to 2009 is $30 \%$ of the 1983-1987 average (Fig. 25). Ageing data has not been completed since 2009; hence the time series ends in 2009.

## Recruitment

For Subdiv. 3Ps there was a pattern in the residuals when Campelen and equivalent data were used in the estimation of recruitment, so as in Morgan et al. (2005), a general linear model using Campelen and original Engel data was chosen which eliminated the pattern in the residuals. Only those cohorts that were observed at least twice in the data were used in the estimation. Relative cohort strengths were estimated using the following model:

$$
\log \left(N_{\mathrm{s}, \mathrm{a}, \mathrm{y}}\right)=\mu+\mathrm{Y}_{\mathrm{y}}+(\mathrm{SA})_{\mathrm{s}, \mathrm{a}}+\varepsilon_{\mathrm{s}, \mathrm{a}, \mathrm{y}},
$$

where:

$$
\begin{aligned}
& \mu=\text { intercept } \\
& s=\text { survey subscript, Engel or Campelen } \\
& a=\text { age subscript, age } 2 \text { to } 5 \\
& y=\text { cohort subscript } \\
& N=\text { survey index (abundance in millions) }
\end{aligned}
$$

$\mathrm{Y}=$ cohort effect
SA = Survey * Age effect, and
$\varepsilon=$ error.

There was significant cohort strength variation ( $\mathrm{F}=3.66$, df $=25, \mathrm{p}<0.001$ ). Recruitment declined in Subdiv. 3Ps from 1980 until 1995 (Fig. 26). Since then it has increased fairly steadily to reach levels similar to the beginning of the time series.

## Stock Recruit

A segmented regression was fit to the stock recruit data using the Julious algorithm and maximum likelihood. The stock recruit data with the fit from the Julious algorithm are shown in Fig. 27. Recruitment increases with SSB index and the data are better described by the segmented regression in Subdiv. 3Ps than for SA2 + Div. 3K. The two methods of fitting the model gave the same results with the break point occurring at an SSB index of 24.1.

## CATCH/BIOMASS RATIOS

## SA 2 + Div. 3K

Catch divided by the index of survey biomass $(C / B)$ is a proxy for fishing mortality $(F)$, and the time series of C/B for American Plaice in SA2 + Div. 3K is shown in Fig. 28. Biomass estimates are Campelen equivalents for Div. 2 J and 3 K combined, and the catches are the reported landings data for SA2 + Div. 3K combined. For much of the time period when surveys were available, a substantial part of the commercial landings occurred during the first quarter of the year. Thus the survey estimates of biomass, which were generally from November to December, were taken to represent the biomass on January 1 of the following year. The analysis shows that $C / B$ ratios were all less than $4 \%$, exceeding $3 \%$ on only two occasions. $C / B$ increased after 1999, as landings increased and survey biomass declined, but still remained below 1\% (Fig. 28). The C/B ratios for the past four years have been very low.

## Subdiv. 3Ps

For American Plaice in Subdiv. 3Ps, the C/B ratio was examined from Campelen data from 1983 to 2011 (Fig. 29). C/B ratio increased steadily through the 1980s reaching a peak of 0.31 in 1990. It declined rapidly after that as catches decreased, and dropped to a minimum in 1995. C/B increased until 2002, after which there has been a gradual decline.

The biomass in the 1990 survey of Subdiv. 3Ps was low compared to 1989 and 1991. This may have artificially inflated the estimate of the C/B ratio in that year. If the biomass in 1990 is estimated to be between those of the adjacent years then the C/B ratio in that year would be 0.14 .

## SURPLUS PRODUCTION MODELLING

The Bayesian surplus production modelling was previously used in a recovery potential context for these stocks (Bailey 2012, Morgan et al 2011). Here it is explored in the context of an assessment for these stocks. The starting point was the models applied in the recovery potential assessment. Most of the exploration described here is of the effect of alternative priors on model fit.

The Schaefer (Schaefer 1954) form of a surplus production model used here is:

$$
P_{t}=\left[P_{t-1}+r^{\star} P_{t-1}\left(1-P_{t-1}\right)-C_{t-1} / K\right]{ }^{*} \eta t
$$

where $\mathrm{P}_{\mathrm{t}-1}$ and $\mathrm{C}_{\mathrm{t}-1}$ denote exploitable biomass (as a proportion of carrying capacity, K ) and catch, respectively, for year t-1 (Meyer and Millar 1999a,b). K is the level of stock biomass at equilibrium prior to commencement of a fishery, $r$ is the intrinsic rate of population growth, and $\eta t$ is a random variable describing stochasticity in the population dynamics (process error). The model utilizes biomass proportional to an estimate of K in order to aid mixing of the Markov Chain Monte Carlo (MCMC) samples and to help minimize autocorrelation between each state and K (Meyer and Millar 1999a,b).

An observation equation is used to relate the unobserved biomass, $\mathrm{P}_{\mathrm{t}}$, to the observations that have been made (e.g. through DFO RV surveys), $\mathrm{I}_{\mathrm{t}}$.

$$
\mathrm{I}_{\mathrm{t}}=\mathrm{q}^{*} \mathrm{P}_{\mathrm{t}}{ }^{*} \varepsilon_{\mathrm{t}}
$$

where $q$ is the catchability parameter, $P_{t}$ is an estimate of the biomass proportional to K at time $t$, and $\varepsilon_{t}$ is observation error.

Data explored were as follows (Table 5):
SA2 + Div.3K
(1) Landings - 1960-2010
(2) Canadian RV Survey Indices: Engels Trawl - 1978-1994
(3) Canadian RV Survey Indices: Campelen Trawl - 1995-2010

Subdiv. 3Ps
(1) Landings - 1960-2010
(2) Canadian RV Survey Indices: Engels Trawl - 1972-1995
(3) Canadian RV Survey Indices: Campelen Trawl - 1996-2010

Non-informative or vague priors were used for all parameters as the initial prior distributions (Table 6). Priors on the catchability ( $q$ ) were uniform and broad. Uniform priors for observation error were limited to a lower bound equal to the coefficient of variation (CV) of each survey index. The upper bound was set at 3 times this CV (Swain et al. 2009). Priors on the process error were also broad and uniform.

Often, K is set to the stock biomass in the year prior to the onset of fishing ( $\mathrm{P}_{0}$; see Meyer and Millar 1999a). However, in the models used here, initial stock biomass was not assumed to be the virgin biomass as fishing began on these stocks prior to 1960. $\mathrm{P}_{0}$ was allowed to vary between 0.5 and 1 (i.e. initial biomass was allowed to vary between $\mathrm{K} / 2$ and K ). A lognormal distribution for K was specified here with a mean of 300 ('000t) and a standard deviation of 1000 ('000t) for both stocks. The upper and lower boundaries on K were large.

The prior for $r$ was first based on the expert opinion that it is unlikely to vary greatly from that of cod in Newfoundland waters which has been estimated at 0.26 (Hutchings 1999). Previous work with the models (Bailey 2012) resulted in somewhat different priors for the two stocks. The prior for $r$ was only vaguely informative, utilizing a mean with a very wide lognormal distribution about the mean. For SA2 + Div. 3K the initial prior on $r$ was $\mu=0.25$ and std $=0.32$ and for Subdiv. 3Ps the initial prior on $r$ was set at $\mu=0.15$ and std $=1$. 'For both stocks the posterior for $r$ was restricted to values between 0.0001 and 3 .

Starting with these initial models a series of models were fit where one prior at a time was changed. The model fit (including convergence criteria, residuals for predicted surveys, credible intervals and deviance information criteria (DIC)) and effect on parameter estimates were compared. A new model was then constructed based on these results. For SA2 + Div. 3K most model formulations were run for 200,000 iterations, a burn in of 100,000 iterations, and the results thinned to every $10^{\text {th }}$ iteration. For Subdiv. 3Ps (which converged better than the SA2 + Div. 3K model) most formulations were run with 150,000 iterations, a burn in of 50,000 iterations, and the results thinned to every $10^{\text {th }}$ iteration, although the final run had 300,000 iterations with 3 chains and a burn in of 200,000 iterations, and was thinned at every $20^{\text {th }}$ interation.

## RESULTS

## SA2 + Div. 3K

Parameter estimates for SA2 + Div. 3K were influenced by the choice of prior for $r$. Use of a higher mean and more informative prior for $r$ (mean $=0.4$, std $=0.2, r \sim$ dlnorm ( -1.03 , $4.49)((0.01,2))$ resulted in a much higher estimate of $r$ than for most runs (model 3) (Table 7). Use of the same mean but with a lower precision (model 4) resulted in an estimate more in keeping with the initial run. Any of the priors for $r$ that were tested with a lower precision resulted in $r$ less than 0.1 , mostly near 0.05 . The alternative priors tested for the error on the survey indices resulted in much lower DIC but resulted in a much higher process error relative to the observation error (equal to the observation error). The sensitivity of estimates to the choice of prior for $r$ may mean that the utility of these models is less for SA2 + Div. 3K than for Subdiv. 3Ps (see below). A model was constructed using the tau on the surveys that gave the lowest DIC (from model 10 and 11). This gave a very low DIC (model 13) but there were concerns about the extremely small deviance and with the process error. In addition, the fishery is thought not to have been the main cause of the decline of the stock (Morgan et al. 2002) in which case a production model would not be expected to be a good description of stock history.

As a result of these concerns, the surplus production modeling was not thought to be a good indicator of stock status at this time and no further results are presented here.

## Subdiv. 3Ps

Parameter estimates for Subdiv. 3Ps American plaice were relatively stable, particularly the median for maximum sustainable yield (MSY) which varied from 3,500 to 5,800 tons and for most models was between 4,000 and 5,000 tons (Table 8 ). All formulations show a similar stock history and show that the current stock size is well below the biomass giving maximum sustainable yield ( $\mathrm{B}_{\text {msy }}$ ). However, credible intervals on Fratio ( $\mathrm{F} /$ fishing mortality giving the maximum sustainable yield, $\mathrm{F}_{\text {msy }}$ ) become very large, indicating that this formulation would not be good for assessment purposes. Use of a higher mean and more informative prior for $r$ (mean $=0.4$, std $=0.2, r \sim \operatorname{dlnorm}(-1.03,4.49) I(0.01,2)$ ) resulted in a higher estimate of $r$ than for most runs but did not result in much change in the DIC.

Survey coverage was poor in many years prior to 1980 and the Engel time series is converted from original Yankee trawl units from 1972 to 1980. Residuals revealed a poor model fit to the first part of the Engel time series. However, starting data during the Engel time series from 1983 results in very large credible intervals on Fratio ( $F / F_{\text {msy }}$ ), indicating that this formulation would not be good for assessment purposes (Model 13). Therefore, model 18 (the model with the best combination of convergence statistics, DIC and credible intervals when run on all the data) was run with survey data extending only to 1980. This model formulation used landings from 1960 to 2010, Engel survey data from 1980 to 1995, and Campelen survey data from 1996 to 2010 (with the exception of 2006 when the survey was incomplete). Its priors are given in Table 9 and diagnostics/results in Fig. 30-46. The diagnostics indicate that the model has converged and that process error is low compared to observation error and with little trend over time. Posteriors were updated based on the priors although there may be some influence of the prior for K on its posterior. There should be further exploration of the effect of priors. The survey indices were reasonably predicted (i.e. residuals small and unpatterned). Convergence diagnostics (posterior density of chains, Gelman-Rubin shrink factors, sampler running means, sampler lag autocorrelations) all indicate acceptable convergence.

Production models estimate relative levels of biomass and fishing mortality more precisely than absolute levels. This means that the ratio of biomass to the $\mathrm{B}_{\text {msy }}$ (Bratio) and the ratio of fishing mortality to the $\mathrm{F}_{\text {msy }}$ (Fratio) are more precise than biomass and fishing mortality themselves. For this reason, stock trajectories and reference points are usually reported as these ratios and status determined relative to $F_{\text {msy }}$ and $B_{\text {msy }}$, with the biomass limit reference point set as a percentage of $B_{\text {msy }}$. Consistent with the DFO PA policy (DFO 2009), $\mathrm{B}_{\text {lim }}$ is $40 \% \mathrm{~B}_{\text {msy }}$, the upper stock reference is $80 \% B_{m s y}$ and $F_{\text {lim }}$ (the removals reference point) is $F_{m s y}$. Stock status relative to these reference points was estimated from the model.

Stock size estimated from the surplus production model decreased fairly steadily from the late 1960s to a low in 1993 of less than $10 \%$ of $B_{\text {msy }}$. Biomass has been increasing slowly since 1993; however, current biomass is $50 \%$ of $\mathrm{B}_{\text {lim }}$ and therefore the stock is in the Critical Zone (Fig. 44). Taking uncertainty into account, the probability of being below $\mathrm{B}_{\text {lim }}$ is high (0.94). Fishing mortality reached a peak in 1991 after which it declined for several years. Fishing mortality increased again to above $F_{\text {msy }}$ in the early 2000s when landings exceed $1000 t$ annually. It has since declined and current median fishing mortality is estimated to be $64 \%$ of $\mathrm{F}_{\text {lim }}$ (Fig. 45). The probability of being above $\mathrm{F}_{\text {lim }}$ is relatively low (0.2).

The trajectory of $B / B_{m s y}$ and $F / F_{\text {msy }}$ are plotted together on the proposed $P A$ framework in Fig. 46. When stock biomass was above $B_{\text {msy }}$ and fishing mortality was below $F_{\text {msy }}$ the stock increased. Fishing mortality was above $F_{\text {msy }}$ for many years, resulting in a steady decline in
stock biomass. In the last few years fishing mortality has been below $F_{\text {msy }}$ and stock size has increased, although it remains well below $\mathrm{B}_{\mathrm{lim}}$.

## CONCLUSIONS

Directed fisheries for American Plaice in SA2 + Div. 3K remain under moratorium, and biomass and abundance indices (from RV surveys) are at low levels. It was not possible to use an analytical model of population dynamics for American Plaice in SA2 + Div. 3K at this time. Therefore an empirical reference point based on stock/recruit (S/R) scatter from the RV survey indicate that the most recent estimate of survey SSB (2009) is at $24 \%$ of $\mathrm{B}_{\text {lim }}$.

Similarly fisheries remain under moratorium for American Plaice in Subdiv. 3Ps as well, with an increase in biomass and abundance indices from 1992 to 2011; current biomass is at $25 \%$ and abundance at $52 \%$ of the average of the mid-1980s. A Bayesian Surplus Production Model was carried out using prior information on r and K and catch data from 1960 to 2010 and RV survey data from 1980 to 2012 to produce limit reference points, $B_{\text {lim }}\left(40 \% B_{\text {msy }}\right), 80 \% B_{\text {msy }}$ (an upper stock reference point), and $\mathrm{F}_{\text {lim }}\left(\mathrm{F}_{\text {msy }}\right)$. Stock size estimated from the surplus production model decreased fairly steadily from the late 1960s to a low in 1993; biomass has been increasing slowly since 1993; however, current biomass is $50 \%$ of $\mathrm{B}_{\text {lim }}$ and therefore the stock is in the Critical Zone of the DFO PA framework. Taking uncertainty into account, the probability of being below $\mathrm{B}_{\text {lim }}$ in the most recent year is high (0.94). $\mathrm{F}_{\text {lim }}$ has declined since the early 2000s and current median fishing mortality is estimated to be $64 \%$ of $\mathrm{F}_{\text {lim. }}$. The probability of being above $F_{\text {lim }}$ is relatively low (0.2).

LRPs for American Plaice in SA2 + Div. 3K and Subdiv. 3Ps may be revised as more data become available. Estimates of SSB, recruitment and mortality from DFO surveys were not available beyond 2009 for this meeting due to a backlog in age interpretation of samples.

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Table 1. Total reported catch (t) of American Plaice in SA2 + Div. $3 K$.

| Year | Total Reported Catch (t) | Year | Total Reported Catch (t) |
| :---: | :---: | :---: | :---: |
| 1960 | 16 | 1986 | 3018 |
| 1961 | 67 | 1987 | 1063 |
| 1962 | 64 | 1988 | 953 |
| 1963 | 1421 | 1989 | 4248 |
| 1964 | 3068 | 1990 | 1825 |
| 1965 | 5558 | 1991 | 510 |
| 1966 | 2949 | 1992 | 104 |
| 1967 | 3591 | 1993 | 77 |
| 1968 | 5951 | 1994 | 16 |
| 1969 | 6902 | 1995 | 28 |
| 1970 | 12686 | 1996 | 16 |
| 1971 | 5348 | 1997 | 9 |
| 1972 | 9121 | 1998 | 2 |
| 1973 | 5140 | 1999 | 7 |
| 1974 | 5620 | 2000 | 67 |
| 1975 | 5747 | 2001 | 137 |
| 1976 | 6107 | 2002 | 100 |
| 1977 | 7525 | 2003 | 34 |
| 1978 | 3522 | 2004 | 17 |
| 1979 | 2965 | 2005 | 30 |
| 1980 | 5040 | 2006 | 60 |
| 1981 | 7545 | 2007 | 23 |
| 1982 | 1900 | 2008 | 10 |
| 1983 | 1633 | 2009 | 10 |
| 1984 | 1175 | 2010 | 22 |
| 1985 | 753 | 2011 | 17 |
|  |  |  |  |

Table 2. Total reported catch (t) of American Plaice in Subdiv. 3Ps.

| Year | Total Reported Catch $(\mathrm{t})$ | Year | Total Reported Catch $(\mathrm{t})$ |
| :---: | :---: | :---: | :---: |
| 1960 | 887 | 1986 | 5130 |
| 1961 | 1455 | 1987 | 5331 |
| 1962 | 1024 | 1988 | 4406 |
| 1963 | 754 | 1989 | 2957 |
| 1964 | 1542 | 1990 | 4130 |
| 1965 | 2022 | 1991 | 4395 |
| 1966 | 3406 | 1992 | 2331 |
| 1967 | 4494 | 1993 | 751 |
| 1968 | 14280 | 1994 | 122 |
| 1969 | 6491 | 1995 | 85 |
| 1970 | 12328 | 1996 | 114 |
| 1971 | 7182 | 1997 | 243 |
| 1972 | 6538 | 1998 | 423 |
| 1973 | 14769 | 1999 | 654 |
| 1974 | 6598 | 2000 | 650 |
| 1975 | 4211 | 2001 | 1010 |
| 1976 | 5458 | 2002 | 1128 |
| 1977 | 4605 | 2003 | 1033 |
| 1978 | 3658 | 2004 | 818 |
| 1979 | 3666 | 2005 | 776 |
| 1980 | 2935 | 2006 | 539 |
| 1981 | 3217 | 2007 | 524 |
| 1982 | 2186 | 2008 | 533 |
| 1983 | 1726 | 2009 | 573 |
| 1984 | 2963 | 2010 | 469 |
| 1985 | 4220 | 2011 | 286 |
|  |  |  |  |

Table 3. Biomass (000 t) and abundance (millions) estimates for Div. 2H and Div. 2J3K from fall RV surveys. NC = not converted; NS = not surveyed

|  | Div. 2H |  | Div. 2J3K |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  | Abundanc |  |  |
| Year | Biomass | e | Biomass | Abundanc |
| 1978 | NC | NC | 332.4 | 1711.3 |
| 1979 | NC | NC | 194.7 | 856.8 |
| 1980 | NC | NC | 232.0 | 792.4 |
| 1981 | NC | NC | 254.9 | 788.4 |
| 1982 | NC | NC | 308.1 | 1015.9 |
| 1983 | NC | NC | 305.5 | 885.3 |
| 1984 | NC | NC | 235.7 | 740.6 |
| 1985 | NC | NC | 175.8 | 598.7 |
| 1986 | NC | NC | 165.8 | 617.0 |
| 1987 | NC | NC | 114.9 | 445.9 |
| 1988 | NC | NC | 118.2 | 436.0 |
| 1989 | NC | NC | 127.5 | 533.2 |
| 1990 | NC | NC | 76.2 | 299.8 |
| 1991 | NC | NC | 44.5 | 218.1 |
| 1992 | NC | NC | 22.5 | 136.2 |
| 1993 | NC | NC | 25.2 | 180.8 |
| 1994 | NC | NC | 20.6 | 170.2 |
| 1995 | NC | NC | 13.4 | 127.3 |
| 1996 | 0.2 | 2.3 | 17.1 | 148.3 |
| 1997 | 1.0 | 4.7 | 16.6 | 127.9 |
| 1998 | 0.9 | 8.9 | 19.5 | 125.9 |
| 1999 | 1.1 | 17.2 | 16.6 | 91.4 |
| 2000 | NS | NS | 14.5 | 73.0 |
| 2001 | 0.7 | 8.8 | 12.1 | 56.3 |
| 2002 | NS | NS | 6.8 | 40.3 |
| 2003 | NS | NS | 8.0 | 51.5 |
| 2004 | 1.5 | 20.0 | 9.8 | 71.5 |
| 2005 | NS | NS | 11.5 | 65.9 |
| 2006 | 1.8 | 32.4 | 19.9 | 114.2 |
| 2007 | NS | NS | 31.6 | 144.2 |
| 2008 | 1.6 | 17.4 | 23.1 | 113.8 |
| 2009 | NS | NS | 25.9 | 100.8 |
| 2010 | 3.0 | 20.1 | 20.3 | 101.9 |
| 2011 | 4.1 | 42.1 | 24.9 | 174.9 |
|  |  |  |  |  |
|  |  |  |  |  |
|  |  |  |  |  |

Table 4. Biomass (000 t) and abundance (millions) estimates for Subdiv. 3Ps from spring RV surveys. NS = not surveyed.

|  | Subdiv. 3Ps |  |
| :---: | :---: | :---: |
| Year | Biomass | Abundance |
| 1983 | 119.6 | 519.0 |
| 1984 | 55.5 | 217.3 |
| 1985 | 125.4 | 312.8 |
| 1986 | 73.0 | 293.4 |
| 1987 | 77.9 | 289.0 |
| 1988 | 62.5 | 215.5 |
| 1989 | 38.5 | 149.2 |
| 1990 | 13.2 | 52.3 |
| 1991 | 32.9 | 159.0 |
| 1992 | 16.0 | 54.6 |
| 1993 | 11.3 | 41.5 |
| 1994 | 10.6 | 43.3 |
| 1995 | 12.5 | 49.8 |
| 1996 | 12.4 | 92.1 |
| 1997 | 8.6 | 45.5 |
| 1998 | 14.4 | 52.6 |
| 1999 | 14.6 | 6.2 |
| 2000 | 21.5 | 137.3 |
| 2001 | 18.3 | 82.0 |
| 2002 | 15.9 | 54.8 |
| 2003 | 17.2 | 70.1 |
| 2004 | 14.0 | 68.3 |
| 2005 | 24.2 | 118.5 |
| 2006 | NS | NS |
| 2007 | 22.4 | 107.0 |
| 2008 | 31.2 | 253.3 |
| 2009 | 20.4 | 167.2 |
| 2010 | 22.0 | 182.7 |
| 2011 | 25.0 | 181.5 |

Table 5. Data used in Bayesian surplus production modelling. Note that for Subdiv. 3Ps the accepted model used Engel data starting in 1980.

|  | SA2 + Div. 3K |  |  | Subdiv. 3Ps |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings | Engel | Campelen | Landings | Engel | Campelen |
| 1960 | 0.016 |  |  | 0.887 |  |  |
| 1961 | 0.067 |  |  | 1.455 |  |  |
| 1962 | 0.064 |  |  | 1.024 |  |  |
| 1963 | 1.421 |  |  | 0.754 |  |  |
| 1964 | 3.068 |  |  | 1.542 |  |  |
| 1965 | 5.558 |  |  | 2.022 |  |  |
| 1966 | 2.949 |  |  | 3.406 |  |  |
| 1967 | 3.591 |  |  | 4.494 |  |  |
| 1968 | 5.951 |  |  | 14.28 |  |  |
| 1969 | 6.902 |  |  | 6.491 |  |  |
| 1970 | 12.686 |  |  | 12.328 |  |  |
| 1971 | 5.348 |  |  | 7.182 |  |  |
| 1972 | 9.121 |  |  | 6.538 | 33.8 |  |
| 1973 | 5.140 |  |  | 14.769 | 13.7 |  |
| 1974 | 5.620 |  |  | 6.598 | 13 |  |
| 1975 | 5.747 |  |  | 4.211 | 1.9 |  |
| 1976 | 6.107 |  |  | 5.458 | 37.8 |  |
| 1977 | 7.525 |  |  | 4.605 | 9.1 |  |
| 1978 | 3.522 | 115.20 |  | 3.658 | 3.8 |  |
| 1979 | 2.965 | 68.92 |  | 3.666 | 7.2 |  |
| 1980 | 5.040 | 90.82 |  | 2.935 | 35.8 |  |
| 1981 | 7.545 | 104.48 |  | 3.217 | 26 |  |
| 1982 | 1.900 | 115.07 |  | 2.186 | 39.1 |  |
| 1983 | 1.633 | 126.98 |  | 1.726 | 45.2 |  |
| 1984 | 1.175 | 93.45 |  | 2.963 | 22.5 |  |
| 1985 | 0.753 | 67.58 |  | 4.22 | 64.5 |  |
| 1986 | 3.018 | 67.39 |  | 5.13 | 30.4 |  |
| 1987 | 1.063 | 45.24 |  | 5.331 | 33.9 |  |
| 1988 | 0.953 | 50.68 |  | 4.406 | 27.3 |  |
| 1989 | 4.248 | 46.38 |  | 2.957 | 17 |  |
| 1990 | 1.825 | 24.49 |  | 4.13 | 5.8 |  |
| 1991 | 0.510 | 12.42 |  | 4.395 | 12.1 |  |
| 1992 | 0.104 | 5.52 |  | 2.331 | 6.8 |  |
| 1993 | 0.077 | 5.51 |  | 0.751 | 4.6 |  |
| 1994 | 0.016 | 3.82 |  | 0.122 | 4.2 |  |
| 1995 | 0.028 |  | 13.42 | 0.085 | 3.9 |  |
| 1996 | 0.016 |  | 17.14 | 0.114 |  | 12.4 |
| 1997 | 0.009 |  | 16.63 | 0.243 |  | 8.6 |

Table 5 (cont). Data used in Bayesian surplus production modelling. Note that for Subdiv. 3Ps the accepted model used Engel data starting in 1980.

|  | SA2 + Div. 3K |  | Subdiv 3Ps |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Year | Landings | Engel | Campelen | Landings |  |
| Engel | Campelen |  |  |  |  |
| 1998 | 0.002 |  | 19.49 | 0.423 |  |
| 1999 | 0.007 |  | 16.56 | 0.654 |  |
| 2000 | 0.067 | 14.52 | 0.65 | 14.36 |  |
| 2001 | 0.137 | 12.13 | 1.01 | 21.5 |  |
| 2002 | 0.098 | 6.82 | 1.128 | 18.3 |  |
| 2003 | 0.034 | 8.01 | 1.033 | 15.9 |  |
| 2004 | 0.017 | 9.81 | 0.818 | 17.2 |  |
| 2005 | 0.030 | 11.50 | 0.776 | 14 |  |
| 2006 | 0.060 | 19.92 | 0.539 | 24.2 |  |
| 2007 | 0.023 | 31.59 | 0.524 |  |  |
| 2008 | 0.013 | 23.05 | 0.533 | 22.4 |  |
| 2009 | 0.010 | 25.87 | 0.512 | 31.2 |  |
| 2010 | 0.022 | 20.26 | 0.468 | 20.4 |  |

Table 6: Initial priors for parameters used in surplus production models for SA2 + Div. $3 K$ and Subdiv. 3Ps.

| NAFO | Parameter | Description | Prior Distribution |
| :--- | :--- | :--- | :--- |
| SA2 + | K | Carrying Capacity | normal $(\mu=300 \mathrm{kt}$, |
| Div. 3K |  |  | std=1000kt) |
|  | r | Population growth rate | normal $(\mu=0.25$, |
|  |  |  | $\mathrm{std}=0.32)$ |
|  | logq.eng | Catchability, Canadian Engels Trawl Series | $\mathrm{U}(0,10)$ |
|  | logq.cam | Catchability, Canadian Campelen Trawl Series | $\mathrm{U}(0,10)$ |
|  | Sigma | Process error | $\mathrm{U}(0,10)$ |
|  | tau.eng | Observation error, Canadian Engels Trawl | $\mathrm{U}(0.68,2.03)$ |
|  | tau.cam | Observation error, Canadian Campelen Trawl | $\mathrm{U}(0.41,1.24)$ |
| Subdiv. | K | Carrying Capacity | normal $(\mu=300 \mathrm{kt}$, |
| 3Ps |  |  | std=1000kt) |
|  | r | Population growth rate | normal $(\mu=0.15$, |
|  |  |  | std=1) |
|  | logq.eng | Catchability, Canadian Engels Trawl Series | $\mathrm{U}(0,10)$ |
|  | logq.cam | Catchability, Canadian Campelen Trawl Series | $\mathrm{U}(0,10)$ |
|  | Sigma | Process error | $\mathrm{U}(0,10)$ |
|  | tau.eng | Observation error, Canadian Engels Trawl | $\mathrm{U}(0.79,2.38)$ |
|  | tau.cam | Observation error, Canadian Campelen Trawl | $\mathrm{U}(0.39,1.17)$ |

Table 7. Parameter estimates and Deviance Information Criteria (DIC) for models using different priors for SA2 + Div. $3 K$. Priors for initial model are given in table 6.

| 2+3k |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | r | K | Fmsy | Bmsy | MSY | DIC |
| 1 | Initial model | 0.054(0.015-0.151) | 148.1(102.1-527.1) | 0.027(0.0075-0.07528) | 74.05(51.04-263.5) | 2.114(0.597-8.545) | 245.005 |
| 2 | $\mathrm{r} \sim \mathrm{dlnorm}(-3.81,0.262)$ )(0.0001,3) | 0.013(4.281E-4-0.089) | 171.7(103.5-629.3) | $0.0065(2.11 \mathrm{E}-4-0.045)$ | 85.83(51.75-314.6) | 0.6336(0.021-4.58) | 244.419 |
| 3 | r ~ dlnorm(-1.03,4.49) $(0.01,2)$ | 0.1443(0.0718-0.2838) | 138.0(101.2-450.9) | 0.07213(0.0359-0.1419) | 69.0(50.62-225.5) | 5.281(2.231-18.6) | 238.989 |
| 4 | r $\sim$ dlnorm(-1.91, 0.51 ) (0.0001,3) | 0.0394(0.0053-0.1284) | 153.7(102.4-536.2) | 0.0197(0.0026-0.0642) | 76.85(51.21-268.1) | 1.633(0.228-7.242) | 245.252 |
| 5 | survey q 's dunif( 0,5 ) | 0.054(0.015-0.1475) | 148.4(102.0-527.8) | 0.027(0.0075-0.074) | 74.21(51.0-263.9) | 2.15(0.6004-8.703) | 245.663 |
| 6 | K~dlnorm(6.21,0.3)!(100,1500) | 0.051(0.0145-0.1437) | 171.3(102.9-767.0) | 0.026(0.0073-0.072) | 85.67(51.45-83.5) | 2.308(0.6163-11.17) | 238.589 |
| 7 | K~dInorm( $5.193,0.98)$ )(100,3000) | 0.053(0.0149-0.1413) | 156.6(102.2-508.8) | 0.0260.0074-0.071) | 78.29(51.11-254.4) | 2.156(0.6241-8.444) | 245.069 |
| 8 | K~dlnorm( $4.46,0.40)$ I( 50,3000 ) | 0.058(0.01568-0.161) | 122.9(57.38-406.6) | 0.02912(0.008-0.081) | 61.46(28.69-203.3) | 1.829(0.4762-6.923) | 237.911 |
| 9 | sigma $\sim$ dunif $(0,100)$ | 0.053(0.01496-0.145) | 150.2(102.2-538.5) | 0.027(0.0075-0.0723) | 75.11(51.09-269.3) | 2.141(0.5932-8.704) | 244.25 |
| 10 | tau.eng ${ }^{\sim}$ dunif( $0.2,2.38$ ) | 0.057(0.01538-0.156) | 157.8(102.3-530.9) | 0.02841(0.008-0.0782) | 78.88(51.15-265.5) | $2.428(0.6076-8.817)$ | 204.817 |
| 11 | tau.cam ${ }^{\text {dunif }}(0.2,1.17)$ | 0.056(0.01494-0.151) | 170.2(102.6-638.9) | 0.028(0.007-0.0757) | 85.09(51.31-319.4) | 2.542(0.639-9.893) | 180.414 |
| 12 | base with 3 chains more iterations | $0.054(0.015-0.144)$ | 145.6(101.8-489.6) | 0.027(0.0075-0.07197) | 72.81(50.89-244.8) | 2.044(0.6003-8.345) | 249.215 |
| 13 | base but with tau from 10 and 11 | 0.0569(0.015-0.1537) | 162.3(102.5-535.5) | $0.02847(0.0075-0.07684)$ | 81.15(51.25-267.7) | $2.495(0.6172-8.963)$ | 168.011 |

Table 8. Parameter estimates and Deviance Information Criteria (DIC) for models using different priors for Subdiv. 3Ps. Priors for initial model are given in Table 6.

| 3 Ps |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | r | K | Fmsy | Bmsy | MSY | DIC |
| 1 | Initial model | 0.1401(0.048-0.255) | 114.9(100.5-285.1) | 0.07007(0.024-0.1275) | 57.47(50.23-142.6) | 4.14(1.871-10.17) | 281.783 |
| 2 | r ~ dlnorm(-1.90,1) $(0.01,2)$ | 0.1485(0.076-0.266) | 114.9(100.5-230.2) | $0.07425(0.03816-0.1328)$ | 57.45(50.24-115.1) | 4.357(2.546-9.86) | 282.951 |
| 3 | $\mathrm{r} \sim \operatorname{dln}$ orm(-1.03,4.49) $(0.01,2)$ | 0.1938(0.1194-0.3436) | 114.8(100.5-237.8) | $0.09689(0.05972-0.1718)$ | 57.42(50.24-118.9) | 5.837(3.403-14.27) | 281.401 |
| 4 | survey q's dunif( 0,5 ) | 0.1401(0.048-0.255) | 114.9(100.5-285.1) | $0.07007(0.024-0.1275)$ | 57.47(50.23-142.6) | 4.14(1.871-10.17) | 281.783 |
| 5 | K~dlnorm(6.21,0.3))(100,1500) | 0.1383(0.02993-0.2598) | 122.4(100.8-436.3) | 0.06915(0.01497-0.1299) | 61.22(50.38-218.1) | 4.359(1.535-13.3) | 276.777 |
| 6 | K~dlnorm(4.46,0.40) $(50,3000)$ | 0.1778(0.07753-0.322) | 87.96(54.42-190.2) | 0.08888(0.03876-0.161) | 43.98(27.21-95.12) | 3.923(2.008-7.596) | 280.586 |
| 7 | K~dlnorm(5.193,0.98)I(100,3000) | 0.1416(0.0378-0.2585) | 119.0(100.6-327.4) | 0.07081(0.0189-0.1292) | 59.51(50.32-163.7) | 4.288(1.62-10.81) | 279.325 |
| 8 | sigma ~ dunif( 0,100 ) | $0.1433(0.04208-0.2611)$ | 117.7(100.6-336.3) | 0.07163(0.02104-0.1305) | 58.83(50.3-168.2) | 4.271(1.753-11.03) | 279.357 |
| 9 | tau.cam ${ }^{\text {dunif }}(0.2,2.38)$ | $0.1377(0.0538-0.236)$ | 114.5(100.5-273.7) | 0.06887(0.02692-0.1181) | 57.23(50.25-136.8) | 4.033(2.051-9.086) | 271.138 |
| 10 | tau.eng ${ }^{\sim}$ dunif( $0.2,1.17$ ) | 0.1524(0.04618-0.295) | 118.8(100.6-301.7) | 0.07619(0.02309-0.1475) | 59.42(50.3-150.8) | 4.654(1.815-12.25) | 278.761 |
| 11 | new base with $1(50,1500)$ on K |  |  |  |  |  |  |
|  | K~dlnorm(5.193,0.98) (X=300, std=400) |  |  |  |  |  |  |
|  | tau.cam ${ }^{\text {dunif }}(0.2,2.38)$ |  |  |  |  |  |  |
|  | tau.eng ${ }^{\sim}$ dunif( $\left.0.2,1.17\right)$ | 0.172(0.0754-0.3211) | 92.92(55.64-242.0) | 0.08599(0.0377-0.1606) | 46.46(27.82-121.0) | 3.954(2.062-9.487) | 261.087 |
| 12 | more iterations (200000) | 0.172(0.07297-0.322) | 93.48(56.15-250.8) | $0.08608(0.03648-0.1608)$ | 46.74(28.08-125.4) | 3.988(2.069-9.568) | 261.438 |
| 13 | engels only to 1983 | 0.121(0.0067-0.225) | 114.0(67.0-446.1) | $0.0605(0.003366-0.1123)$ | 56.98(33.5-223.0) | 3.507(0.3854-8.08) | 148.315 |
| 14 | Same as new but K~dlnorm $(4.46,0.4)$ | 0.178(0.07998-0.34) | 87.72(53.67-215.3) | 0.08878(0.03999-0.1702) | 43.86(26.83-107.7) | 3.91(2.033-8.727) | 263.357 |
| 15 | Same as above but also old tau.cam | 0.194(0.0769-0.369) | 83.87(52.66-210.9) | 0.09676(0.03842-0.1843) | 41.94(26.33-105.5) | 4.092(1.943-8.946) | 276.225 |
| 16 | Same as 14 but KI( 35,1500 ) | 0.1812(0.08341-0.3939) | 86.97(43.11-191.8) | 0.09062(0.0417-0.1969) | 43.49(21.55-95.89) | 3.894(1.947-8.028) | 262.139 |
| 17 | same as 14 but 3 chains | 0.177(0.08141-0.3294) | 88.39(53.69-199.0) | 0.08845(0.04071-0.1647) | 44.2(26.84-99.5) | 3.926(2.063-7.865) | 266.337 |
| 18 | same as 17 but Engel data starting in 1980 | $0.1377(0.02761-0.2332)$ | 101.4(63.51-241.7) | $0.06884(0.0138-0.1166)$ | 50.69(31.75-120.9) | 3.576(0.9684-6.25) | 185.166 |

Table 9. Priors for parameters used in 'preferred' formulation of the surplus production model for Subdiv. 3Ps.

| NAFO | Parameter | Description | Prior Distribution |
| :--- | :--- | :--- | :--- |
| Subdiv. | K | Carrying Capacity | normal $(\mu=300 \mathrm{kt}$, |
| 3Ps |  |  | std $=400 \mathrm{kt})$ |
|  | r | Population growth rate | normal $(\mu=0.15$, std=1 $)$ |
|  | logq.eng | Catchability, Canadian Engels Trawl Series | $\mathrm{U}(0,10)$ |
|  | logq.cam | Catchability, Canadian Campelen Trawl Series | $\mathrm{U}(0,10)$ |
|  | Sigma | Process error | $\mathrm{U}(0,10)$ |
|  | tau.eng | Observation error, Canadian Engels Trawl | $\mathrm{U}(0.2,1.17)$ |
|  | tau.cam | Observation error, Canadian Campelen Trawl | $\mathrm{U}(0.2,2.38)$ |



Figure 1. American Plaice landings (t) (1960-2011) and total allowable catch (TAC) (1974-2012) in SA2 + Div. 3K.


Figure 2. American Plaice catches (t) (1960-2011) and total allowable catch (TAC) (1974-2012) in Subdiv. 3Ps.


Figure 3. Biomass (000 t) and abundance (millions) indices of American Plaice from fall surveys, Div. 2 H .


Figure 4. Biomass ( $000 t$ ) and abundance (millions) indices of American Plaice from fall surveys, Div. 2J3K.


Figure 5. Numbers of American Plaice at length for 2009 to 2011 from fall surveys in Div. 2J3K.


Figure 6. Distribution of American Plaice from the fall surveys in Div. 2HJ3K (number/tow) in 1981.


Figure 7a. Distribution of American Plaice from the fall surveys in Div. 2HJ3K (number/tow) from 2006 to 2009.


Figure 7b. Distribution of American Plaice from the fall surveys in Div. 2HJ3K (number/tow) from 2006 to 2009.

```
2010
```



2011


Figure 8. Distribution of American Plaice from the fall surveys in Div. 2HJ3K (number/tow) in 2010 and 2011.


Figure 9. Plot of log survey index against age for ages 0 to 15 over the period 1978-2009 for American Plaice in SA2 + Div. $3 K$.


Figure 10. Estimates of cohort Z for SA2 + Div. 3K American Plaice based on the survey index. Estimates are negated so that higher values indicate increased mortality.


Figure 11. Age at 50\% maturity for male and female American Plaice from SA2 + Div. 3 K.


Figure 12. Length at 50\% maturity for male and female American Plaice in SA2 + Div. 3 K


Figure 13. Index of spawning stock biomass derived from the research vessel surveys of SA2 + Div. 3 K .


Figure 14. Recruitment as relative cohort strength from research vessel data for American Plaice in Div. 2J3K. Estimates are relative to the 2005 cohort.


Figure 15. Recruitment and spawning stock biomass index estimated from the research vessel survey of SA2 + Div. 3K. The solid line gives the results of the segmented regression model fit to these data.


Figure 16. Biomass (000 t) and abundance (millions) indices of American Plaice from spring surveys, Subdiv. 3Ps.


Figure 17. Numbers of American Plaice at length for 2009 to 2011 from spring surveys in Subiv. 3Ps.


Figure 18. Distribution of American Plaice from the spring surveys in Subdiv. 3Ps (number/tow) in 1983.


Figure 19. Distribution of American Plaice from the spring surveys in Subdiv. 3Ps (number/tow) for 2007 (top), 2008 (middle), and 2009 (bottom). Survey was cancelled in 2006.


Figure 20. Distribution of American Plaice from the spring surveys in Subdiv. 3Ps (number/tow) in 2010 (top) and 2011 (bottom).


Figure 21. Plot of log survey index against age for ages 0 to 15 over the period 1983-2009 for Subdiv. 3Ps American Plaice.


Figure 22. Estimates of cohort Z for Subdiv. 3Ps American Plaice based on the survey index. Estimates are negated so that higher values indicate increased mortality.


Figure 23. Age at 50\% maturity for male and female American Plaice from Subdiv. 3Ps.


Figure 24. Length at 50\% maturity for male and female American Plaice in Subdiv. 3Ps.


Figure 25. Index of spawning stock biomass derived from the research vessel surveys of Subdiv. 3Ps.


Figure 26. Recruitment as relative cohort strength from research vessel data for American Plaice in Subdiv. 3Ps. Estimates are relative to the 2005 cohort.


Figure 27. Recruitment and spawning stock biomass index estimated from the research vessel survey of Subdiv. 3Ps. The solid line gives the results of the segmented regression model fit to these data.


Figure 28. Ratio of reported catch to Campelen/Campelen-equivalent unit survey biomass indices for American Plaice in SA2 + Div. 3K.


Figure 29. Ratio of reported catch to Campelen/Campelen-equivalent unit survey biomass indices for American Plaice in Subdiv. 3Ps.


Figure 30. Prior and posterior distributions for sigma, K, and $r$ and the distribution for deviance from the surplus production model for American Plaice in Subdiv. 3Ps.


Figure 31. Prior and posterior distributions for the error and $q$ for the survey indices from the surplus production model for American Plaice in Subdiv. 3Ps.



FMSY


Figure 32. Posterior distributions of $B_{m s y}, M S Y$, and $F_{m s y}$ from the surplus production model for American Plaice in Subdiv. 3Ps.

## 3Ps RV Engels



## 3Ps Engels Residuals



Figure 33. Observed and predicted Engel survey (top) and residuals (bottom) from the surplus production model for American Plaice in Subdiv. 3Ps.

## 3Ps RV Campelen



## 3Ps Campelen Residuals



Figure 34: Observed and predicted Campelen survey (top) and residuals (bottom) from the surplus production model for American Plaice in Subdiv. 3Ps.


Sigma

## Estimated Posterior Density



Figure 35. Posterior distribution from each chain for $r, K$ and sigma from the surplus production model for American Plaice in Subdiv. 3Ps.

## Estimated Posterior Density



Estimated Posterior Density


Figure 36. Posterior distribution from each chain for log $q$ for the Engel survey and log $q$ of the Campelen survey from the surplus production model of American Plaice in Subdiv. 3Ps.


Sigma

## Gelman \& Rubin Shrink Factors



Figure 37. Gelman and Rubin convergence diagnostic for r, K, and Sigma from the surplus production model for American Plaice in Subdiv. 3Ps.

## Gelman \& Rubin Shrink Factors



## Gelman \& Rubin Shrink Factors



Figure 38. Gelman and Rubin convergence diagnostic for $\log q$ for the Engel survey and $\log q$ of the Campelen survey from the surplus production model of American Plaice in Subdiv. 3Ps.


Sigma

## Sampler Running Mean



Figure 39. Running mean of each of 3 chains for $r, K$, and sigma from the surplus production model of American Plaice in Subdiv. 3Ps.

## Sampler Running Mean




Figure 40. Running mean of each of 3 chains for log $q$ for the Engel survey and log $q$ of the Campelen survey from the surplus production model of American Plaice in Subdiv. 3Ps.


Sigma

## Sampler Lag-Autocorrelations



## Sampler Lag-Autocorrelations



## Logq.cam

Figure 42. Autocorrelation for each of 3 chains for log $q$ for the Engel survey and log $q$ of the Campelen survey from the surplus production model of American Plaice in Subdiv. 3Ps.


Figure 43. Median process error with 70 and 95\% credible intervals from the surplus production model for American Plaice in Subdiv. 3Ps.


Figure 44. Ratio of biomass to $B_{\text {msy }}$ from the surplus production model for American Plaice in Subdiv. 3Ps. The median, $70 \%$ and $95 \%$ credible intervals are shown. The red horizontal line is $B_{\text {lim }}$ (where Bratio $=0.4$ ).


Figure 45. Ratio of fishing mortality to $F_{\text {msy }}$ from the surplus production model for American Plaice in Subdiv. 3Ps. The median, $70 \%$ and $95 \%$ credible intervals are shown. The red horizontal line is $F_{\text {lim }}$ (where Fratio=1).


Figure 46. The ratio of $F / F_{m s y}$ and $B / B_{m s y}$ from the surplus production model for American Plaice in Subdiv. 3Ps in relation to stock reference points. The horizontal line is $F_{\text {lim. }}$. The dashed vertical line is $B_{l i m}$, the dashed and dotted vertical line is the suggested upper stock reference and the solid vertical line is where biomass equals $B_{m s y}$.

## APPENDIX 1

## WINBUGS MODEL FOR PRODUCTION MODEL FOR SUBDIV. 3PS AMERICAN PLAICE

```
model
{
# Prior for intrinsic rate of increase(r)
r ~ dlnorm(-3.81,0.262)I(0.0001,3)
# prior distribution of q's
logq.eng~dunif(0,10)
q.eng<-exp(logq.eng)
logq.cam~dunif(0,10)
q.cam<-exp(logq.cam)
# prior distribution of K
K~dlnorm(4.46,0.40)l(50,1500)
# Prior for process error, sigma
sigma ~ dunif(0,10)
isigma2 <- pow(sigma, -2)
# Prior for observation errors, tau.
#tau.eng = Engels series; tau.cam=Campelen series
tau.eng~dunif(0.2,1.17)
itau2.eng <- pow(tau.eng, -2)
tau.cam~dunif(0.2,2.38)
itau2.cam <- pow(tau.cam, -2)
# Prior for initial population size as proportion of K, P[1]. Limited between 0.5*K and K.
Pin~dunif(0.5,1)
Pm[1] <- log(Pin)
P[1] ~ dlnorm(Pm[1], isigma2)I(0.001,5)
#This parameter was used to track process error
P.res[1]<-log(P[1])-Pm[1]
# State equation - SP Model.
for (t in 2:(N)) {
Pm[t] <- log(max(P[t-1] + r*P[t-1]*(1-P[t-1]) - L[t-1]/K, 0.0001))
P[t] ~ dlnorm(Pm[t], isigma2)I(0.001,5)
P.res[t]<-log(P[t])-Pm[t]
}
# Observation equations, Engels and Campelen RV Series
for (t in 21:(36)) {
lengm[t] <- log(q.eng*K * P[t])
leng[t] ~ dlnorm(lengm[t], itau2.eng)
}
```

```
for (t in 37: N) {
Icamm[t] <- log(q.cam*K * P[t])
Icam[t] ~ dlnorm(Icamm[t], itau2.cam)
}
# Output. Using the proportion and K to estimate biomass, B.
for(t in 1:N) {
B[t] <- P[t] *K
#Calculating Fishery Exploitation
F[t]<- L[t]/B[t]
#Bratio calculation
Bratio[t] <- B[t]/BMSY
}
#F Ratio: indicates the ratio of fishing mortality to that estimated for FMSY.
for(t in 1:N) {
Fratio[t] <- F[t]/FMSY
}
#Calculation of Annual Surplus Production one less year than N
for(t in 1:50) {
ASP[t]<-(B[t+1]-B[t])+L[t]
}
# Surplus Production, FMSY, and BMSY calculations for SP Model
MSY <- r*K/4
FMSY<-r/2
BMSY<-K/2
#The section below is used for generating residuals
#generate replicate data sets for Engels and Campelen modeled values
for (i in 21:36){
    leng.rep[i] ~ dlnorm(lengm[i],itau2.eng)
#p.smaller calculates the probability of the modeled value being lower than the actual value
    p.smaller.eng[i] <- step(log(leng[i])-log(leng.rep[i]))
#residuals of log values of replicate data
    res.eng.rep[i] <- log(leng[i])-log(leng.rep[i])
}
for (i in 37:N){
    Icam.rep[i] ~ dlnorm(Icamm[i],itau2.cam)
    p.smaller.cam[i] <- step(log(Icam[i])-log(Icam.rep[i]))
#residuals of log values of replicate data
    res.cam.rep[i] <- log(Icam[i])-log(Icam.rep[i])
}
} ## END
```

