# Boat Electrofishing Survey of the Fish Assemblages in the St. Clair River, Ontario 

A. Edwards, J. Barnucz, and N.E. Mandrak

Fisheries and Oceans Canada
Great Lakes Laboratory for Fisheries and Aquatic Sciences 867 Lakeshore Rd. P.O. Box 5050
Burlington ON L7R 4A6 CANADA

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Fisheries and Oceans Canada
Great Lakes Laboratory for Fisheries and Aquatic Sciences
867 Lakeshore Rd., P.O. Box 5050
Burlington ON L7R 4A6 CANADA

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#### Abstract

Fisheries and Oceans Canada conducted a boat electrofishing survey in the St. Clair River Area of Concern (AOC) in 2004. Sampling replicated, in part, a 1994 Remedial Action Plan (RAP) survey by the Ontario Ministry of Natural Resources. The 2004 survey was conducted primarily to assess changes in fish assemblages and associated Indices of Biotic Integrity (IBI) since the 1994 survey and, secondarily, to examine seasonality changes in fish assemblages. Cluster analysis of annual relative abundance values demonstrated some dissimilarity in the fish assemblage between 1994 and 2004. There was no significant difference between the overall 1994 and 2004 IBI scores; however, there was noticeable variation in individual site IBI scores between the years. To examine seasonal trends in the fish assemblage, DFO sampled the St. Clair River in 2004 during June, July and October. Seasonal trends with respect to relative abundance were observed for many species, such as alewife and rainbow smelt. IBI scores calculated for each month did not differ significantly, although individual site scores demonstrated some variation over the three sampling periods.

Pêches et Océans Canada a mené une étude au moyen d'une embarcation de pêche à l'électricité dans le secteur préoccupant de la rivière Saint-Clair en 2004. L'échantillonnage répétait, en partie, une étude du plan d'assainissement de 1994 effectuée par le ministère des Richesses naturelles de l'Ontario. L'étude de 2004 visait, dans un premier temps, à évaluer les changements dans les assemblages de poissons et les indices d'intégrité biotique (IIB) connexes depuis l'étude de 1994; dans un deuxième temps, elle avait pour but d'examiner les changements de la saisonnalité dans les assemblages de poissons. Une analyse typologique des valeurs annuelles de l'abondance relative a démontré une certaine dissemblance dans l'assemblage de poissons entre 1994 et 2004. On n'a noté aucune différence importante entre les valeurs globales de I'IIB de 1994 et de 2004; on a toutefois observé une variation notable dans les valeurs de l'llB des sites individuels entre les années. Pour examiner les tendances saisonnières dans l'assemblage de poissons, le MPO a échantillonné la rivière Saint-Clair en 2004 durant les mois de juin, juillet et octobre. Des tendances saisonnières afférentes à l'abondance relative ont été observées pour de nombreuses espèces, telles que le gaspareau et l'éperlan. Les valeurs de l'llB calculées pour chaque mois ne différaient pas de façon importante, quoique les valeurs des sites individuelles aient présenté une certaine variation au cours des trois périodes d'échantillonnage.


### 1.0. INTRODUCTION

Sampling large rivers, such as the St. Clair River, presents many logistical and technical challenges due to deep water and strong currents. The St. Clair River runs for 64 km from Lake Huron to Lake St. Clair (GLIN 2005). At its widest point, it is more than a kilometre wide and flow rates average 5710 cms . Although large rivers are more difficult to sample, they are often the most ecologically and economically important systems and are usually the water courses that receive the greatest amount of human impact and degradation (USGS 1997).

Great Lakes Areas of Concern (AOCs) are highly degraded geographic areas within the Great Lakes basin. The St. Clair River AOC extends the entire length of the river from Lake Huron to Lake St. Clair and includes the north shore of Mitchell's Bay on Lake St. Clair, the St. Clair Flats from St. John's Marsh in the west, to the southern tip of Seaway Island (Environment Canada 2004). The International Joint Commission (IJC) Water Quality Board defines an AOC as an area where there is a known impairment of a beneficial water use (MacLennan and Hyatt 1996). An impaired beneficial use is described as an impairment of an environmental feature such as public beaches, drinking water, or fish and wildlife populations that bring economic, sociological and recreational benefits to society (MacLennan and Hyatt 1996). Loss of fish and wildlife habitat was identified as an impaired beneficial use for the St. Clair River AOC, with habitat having been lost to dredging, draining, filling and bulk heading for industrial, urban, agricultural and navigational uses (Dutz 1998). Remedial Action Plans (RAPs) were developed for each AOC to identify specific problems within the AOC and to describe methods for correcting these problems (GLIN 2005).

Fisheries and Oceans Canada (DFO) conducted a boat electrofishing survey of the St. Clair River in 2004. This study replicated, in part, a 1994 RAP survey performed by the Ontario Ministry of Natural Resources (OMNR). The 1994 survey was one monitoring component of the RAP Program within the St. Clair River AOC. The 2004 DFO survey was conducted primarily to assess changes in fish assemblages and associated Indices of Biotic Integrity (IBI) since the 1994 survey and, secondarily, to examine seasonality changes in fish
assemblages. The sites were sampled three times over two day periods during June, July and October. Eight sites were sampled within a 50 km reach of the St. Clair River and were distributed between the Blue Water Bridge, Sarnia and downstream to the confluence of Marshy Creek, immediately upstream of Port Lambton (Figure 1).

### 2.0. METHODS

### 2.1. ELECTROFISHING TECHNIQUES

Electrofishing was performed using a 6.35 m Model SR-20 Smith-Root electrofishing boat equipped with a Model 7.5 kW Smith-Root generator, 7.5 GPP control box, three kick plates and dual foot pedals. Sampling data recorded at each site included capture method, sampling effort, electrofishing settings and a description of the sampling equipment. Two netters retrieved stunned fishes as they appeared and all fishes were transferred from the river into a live-well within the boat. Species were identified, counted and released. Minimum and maximum lengths were recorded for all species captured. Voucher specimens were kept for lab verification at a later date.

### 2.2. SAMPLING OF 2004 SITES

Each sampling site (e.g. SCR04141004001) contained two separate subsites, an upstream site and a downstream site. These sites were denoted with ' $A$ ' for the upstream site and ' $B$ ' for downstream site (e.g. SCR04141004001A and SCR04141004001B) (Appendix 1). Each sampling site consisted of one sampling transect which was sampled twice on the same day. Sampling runs were performed travelling upstream along the OMNR 1994 transects. Upon completion of electrofishing, all fishes were processed before beginning the next sub-site. All sites were sampled between the hours of 0800 h and 1600 h to minimize the influence of diurnal effects on fish movement. Each sampling run was approximately 500 m in length and was sampled for approximately 500 s . For simplicity, each site was numbered one through eight for this report (Appendix 1).

### 2.3. HABITAT DATA COLLECTION

Habitat at each of the 2004 sites was described by recording air temperature, water temperature, conductivity, Secchi depth, stream width, maximum stream depth, distance from shore, maximum sampling depth, flow rate and aquatic vegetation. Habitat data were recorded upon completion of electrofishing at each transect.

### 2.4. ANALYZING THE DATA

### 2.4.1. 1994 vs. 2004

To assess changes in the fish assemblage since the 1994 OMNR survey, the 1994 data were compared to the 2004 data.

To evaluate the health of the aquatic ecosystem the OMNR calculated the Index of Biotic Integrity (IBI) for each 1994 site, using Hamilton's adaptation (MacLennan and Hyatt 1996) (Appendix 2). The OMNR used the same trophic guild classification system (Appendix 3) as Hamilton except in cases where Hamilton had not reported or classified a species, in which case they classified the species according to feeding habits as reported in Scott and Crossman (1973) and Whitehead et al. (1986) (MacLennan and Hyatt 1996).

IBI scores for the 2004 sites were calculated using the same adaptation. DFO used the same trophic classification as the OMNR. New species collected in 2004 and not classified by the OMNR in 1994 were classified according to Coker et al. (2001) (Appendix 3).

IBI scores for 1994 and 2004 data were also re-calculated using a slightly modified IBI scheme (Appendix 4) and an updated trophic guild classification scheme (Appendix 5) based on Coker et al. (2001). The reason for re-calculating the IBI scores was to determine what effect, if any, the above changes would have on the IBI. These new IBI scores were not used to assess the changes in aquatic ecosystem health between 1994 and 2004.

### 2.4.2. June, July and October 2004

The Margalef's Diversity Index (MI) (Green 1979) was calculated to compare species diversity across sites, for each month sampled.

As with the annual data, the health of the aquatic ecosystem over the seasons was assessed using Hamilton's adaptation of the IBI and the trophic classification system used by the OMNR. New species collected in 2004 were classified according to Coker et al. (2001) (Appendix 3). IBI scores for 2004 data were re-calculated using the modified IBI scheme and updated trophic classification scheme mentioned previously. This was done to determine possible effects the changes would have on the IBI. The new IBI scores were not used to assess changes in aquatic ecosystem health between seasons.

### 2.4.3. Statistical Analysis

The data for seasonal and annual sites was not normally distributed; therefore, non-parametric tests were used to test for significance. The MannWhitney U Test was used to analyze differences in the annual IBI scores, and the Kruskal-Wallace Test was performed on the seasonal IBI scores. These tests were performed on both versions of the IBI to determine whether the results differed.

Cluster analysis was performed on seasonal and annual relative abundance values to visually determine how similar the fish assemblages were over these periods.

### 3.0. RESULTS

### 3.1. SAMPLING EFFORT

DFO 2004 sampling efforts yielded a total catch of 2424 fishes, with 664, 740 and 1020 individuals caught in June, July and October, respectively (Table 1). The mean number of fishes captured per site was 83 in June, 92.5 in July and 127.5 in October (Table 1). The minimum number of fishes captured at one site was five (Site 2) in June, ten at Site 8 in July and 43 at Site 8 in October (Table 1, Appendix 6). The maximum number of fishes observed during June sampling was 272 (Site 4), 292 were caught at Site 3 in July, while 310 were caught at Site 1 in October (Table 1, Appendix 6).

Electrofishing sampling effort was comparable between all sites. Mean sampling effort was 32.5 minutes (June), 29.38 minutes (July) and 32.9 minutes (October) of electrofishing time per site (Table 2). Total effort expended among all sites was 259.98 minutes (June), 235.07 minutes (July) and 263.2 minutes (October) (Table 2). Catch per unit effort (CPUE) was 3.28 fishes captured per minute of sampling effort for all three sampling periods (Table 2). The mean CPUE for June sampling was 2.75, while for July it was 3.14. The mean CPUE was highest for October with 3.94 fishes captured per minute of sampling effort (Table 2). Individual site CPUEs for each month are summarized in Appendix 7 and a summary of sampling effort is found in Appendix 8.

### 3.2. FISH ASSEMBLAGE SAMPLING

During the 2004 DFO surveys of the St. Clair River, 36 species were caught in June, and 33 species were caught in July and October (Tables 3 and 4). Ten unique species were seen in June and four unique species were detected in July and October (Table 4). There were 18 species in common between the three sampling times and the total species richness (SR) was 36 (Table 4). Individual site SR values are in Appendix 9. Spotted sucker (Minytrema melanops) (scientific and common names according to Nelson et al. 2004, and listed in Appendix 10), a species designated as Special Concern by the Committee on the Status of Endangered Wildlife in Canada (COSEWIC) (COSEWIC 2004), was detected during the July and October sampling periods (Table 3). Two specimens from the genus Ictiobus, also of Special Concern, were caught in the October sampling period (Table 3).

In June, the MI ranged from 1.37 (Site 6) to 9.04 (Site 4), while in July, Site 7 and Site 8 had the highest and lowest MI scores, respectively. The highest MI in October was observed at Site 7 (7.64) and the lowest at Site 3 (4.45) (Table 5).

2004 IBI - Original OMNR Method
In June, Site 2 had the lowest IBI score (10, Very Poor), while Site 1 had the highest IBI with a score of 30 (Fair to Good). In July, Site 2 had the lowest IBI score (11, Very Poor) and Site 4 had the highest (29, Fair). Site 1 had the
highest IBI score in October, with a score of 26 (Fair), while Sites 5 and 6 tied for the lowest IBI, 17 (Very Poor to Poor) (Table 6, Appendix 11).

2004 IBI - Re-calculated
In June, the IBI score for Site 2 remained the same (10, Very Poor), the lowest of all the sites. Site 4 had the highest IBI score ( 23, Poor). Site 8 had the lowest IBI score ( 9 , Very Poor) in July, and Sites 3 and 4 had the highest IBI score (23, Poor). in October, Site 8 had the lowest IBI score (14, Very Poor) and Site 1 had an IBI score of 23 (Poor), the highest of all the sites (Table 6, Appendix 12).

1994 IBI - Re-calculated
In August 1994, Site 8 had an IBI score of 13 (Very Poor), the lowest of all the sites, and Site 1 had the highest score, 23 (Poor) (Table 7, Appendix 13). The original IBI scores for 1994 sites are presented in Appendix 14.

### 3.3. STATISTICAL ANALYSIS

The IBI scores for June, July and October 2004 did not differ significantly from each other when calculated using the OMNR method ( $\mathrm{H}=0.515_{0.05,2 ;} ;>0.05$ ), or the modified IBI scheme ( $\mathrm{H}=0.369_{0.05,2} ; \mathrm{p}>0.05$ ). Similarly, the $\mathrm{IB} \mid$ scores for 1994 and 2004 did not differ significantly from each other using either IBI method (OMNR method ( $\mathrm{U}=18_{0.05,(2), 8,8} ; \mathrm{p}>0.05$ ); modified method ( $\mathrm{U}=19_{0.05(2), 8,8} ; \mathrm{p}>0.05$ )).

Cluster analysis of seasonal relative abundance values demonstrated that there was no strong seasonal trend in the fish assemblages (Figure 2). Cluster analysis of annual relative abundance values demonstrated that the fish assemblages at the sites were more similar to each other within, rather than between years (Figure 3).

### 3.4. HABITAT SAMPLING

In June, four of the sites had $50 \%$ or less aquatic macrophytes, while the other four had no vegetation at all. Seven sites in July had $45 \%$ or less macrophyte coverage, and Site 1 had no aquatic macrophytes at all. In October, five out of the eight sites had vegetation levels of $30 \%$ or less, two sites had vegetation over at least $70 \%$, while one had no macrophyte growth (Appendix 15). The vast majority of vegetation observed was submergent (Appendix 15).

Many of the sites sampled contained habitats with medium flow rates, and the maximum effective sampling depth at each site was typically 2 m (Appendix 15).

### 4.0. DISCUSSION

### 4.1. HISTORICAL COMPARISON: AUGUST 1994 vs. JULY 2004

In August 1994, the OMNR collected 3720 individuals representing 39 species from eight different sites along the St. Clair River. In July 2004, DFO collected 740 individuals, representing 33 species, from the same eight sites (Tables 8 and 9 ). The 1994 survey had 11 species unique to that time period; whereas, the 2004 survey only had five unique species (Table 9). The average number of fishes captured per site in 1994 was 465 , while in 2004 it was 92.5 (Table 10). Notable differences in the number of individuals of a species caught during each sampling period include: 2562 alewife (Alosa pseudoharengus) in 1994 vs. 1 in 2004; 63 rock bass (Ambloplites rupestris) in 1994 vs. 3 in 2004; 49 largemouth bass (Micropterus salmoides) in 1994 vs. 0 in 2004; 41 common shiner (Luxilus cornutus) in 1994 vs. 0 in 2004 (Appendix 16). This last example may be a case of misidentification as there were no common shiner caught in 2004, but striped shiner (Luxilus chrysocephalus) were collected. It is possible that the common shiner identified in 1994 were actually striped shiner, as the distinction between the two species at the time was fairly new.

The differences in the 1994 and 2004 results may be the result of changes in the fish assemblages or the result of differences in sampling methods. One reason for the difference in the results could be the time of day that the surveys took place. OMNR sampled in the evening from 18:00 h-24:00 h while DFO sampled during the day between 08:00 h-16:00 h. McInerny and Cross (1996) suggested that fishes in clear water detect and avoid electrofishing boats better during the day than at night, reducing the day CPUE. The mean CPUE for the August 1994 sampling period was 11.48 compared to 3.14 in July 2004 (Table 11). The St. Clair River is a clear river with Secchi depths, on average, of at least 2 m (Appendix 15); therefore, it is likely that the electrofishing boat was visible to fishes allowing them to avoid it. Daily patterns in fish movement can also have a major impact on CPUE. Alewife undergo diel vertical migrations,
moving upward in the water column at night to pursue zooplankton (Madenjian et al. 2003), making them more susceptible to capture by boat electrofishing. Another reason for the differences in the total number of fishes caught and CPUE between August 1994 and July 2004 may be the total effort expended over the sampling periods. OMNR spent almost 10 more minutes, on average, per site when compared to DFO 2004 sampling. The total sampling effort for OMNR was 321 minutes; whereas, DFO spent a total of 235.07 minutes electrofishing (Table 11). However, this difference in sampling effort is not enough to explain the very large difference in CPUE for each sampling period. Differences in water temperature between the two periods may have also influenced the CPUE. Unfortunately, water temperature data for the 1994 sampling period are not available; therefore, no comparison can be made.

It is important to note that all species caught in August 1994 were caught during at least one of the three sampling periods in 2004. This indicates that there was no strong sampling bias against any particular species in 2004.

There was no significant difference between the overall 1994 and 2004 IBI scores; however, there was noticeable variation in individual site IBI scores between the years. In 1992, the eight sites sampled were designated as being 'degraded', 'intermediate' or 'unimpaired' by the Ontario Ministry of the Environment (MOE) (MacLennan and Hyatt 1996). Site 8, offshore of the Canadian National Railway (CNR) yards, was chosen to represent an unimpaired fish community as it was upstream of the major industries. However, the $|B|$ score for Site 8 during the 1994 survey was the lowest (16, Very Poor to Poor) of all the sites. In July 2004, the IBI score decreased slightly to 14 (Very Poor), and remains the second lowest (Table 7). The fast current, deep water and a steel breakwall may have created a sampling bias at Site 8 as it cannot be sampled effectively using a boat electrofishing unit. Site 8 was not an appropriate control site, as it is one of the least effectively sampled sites and cannot readily be compared to the other sites.

Site 1 was designated in 1992 as a 'degraded' site by the MOE; however, its IBI score in 1994 was higher (23, Poor) than that of the 'unimpaired' Site 8 (Table 7, Figure 4). It is worth noting that, in 1994, alewife made up
approximately $93 \%$ of the catch at Site 1, which acted to inflate the IBI score. In July 2004, the IBI score for Site 1 decreased by one. Sites 2-7 were designated as 'unimpaired' or 'intermediate'. The IBI scores for Sites 3,4 and 7 remained the same or increased (Site 4) from 1994 to 2004 (Table 7, Figure 4). In 2004, IBIs for Sites 2, 5 and 6 decreased sharply from the 1994 survey, with Site 2 showing the most precipitous drop from 22 (Poor) to 11 (Very Poor) (Table 7, Figure 4). The fast current combined with daytime sampling in 2004 may have played a role in the decrease; however, it is unlikely these are the only contributing factors. Site 6 had the next largest decline in IBI score, from 25 (Fair) to 15 (Very Poor) (Table 7, Figure 4). In 1994, this site was described as having medium amounts of submergent vegetation and extensive amounts of emergent vegetation and a gravel/sand substrate. In 2004, there was very little vegetation and the substrate was $100 \%$ sand. The current, in 1994, was described as medium while in July 2004 it was described as slow. It is possible that sand has been deposited at the site over the intervening years, perhaps making it unsuitable for large amounts of vegetation. This lack of habitat diversity may have contributed to the drop in IBI for Site 6. However, parameters such as flow rate and percent vegetation are somewhat subjective when measured qualitatively, which should be taken into account when looking for trends over time.

### 4.2. INTER-SEASONAL COMPARISON: JUNE, JULY AND OCTOBER 2004

To examine the effects of seasonality on the fish assemblage, DFO sampled the St. Clair River in 2004 at three different times. Seasonal variation in species composition is dependent on the ecological requirements of individual species such as temperature range, habitat and food. When the catch data for the DFO 2004 sampling period were examined, seasonal trends in numbers caught were seen in numerous species (Appendix 6). High numbers of alewife and rainbow smelt (Osmerus mordax) were caught in October, likely due to water temperatures in October averaging closer to the preferred temperature of 11-15 ${ }^{\circ} \mathrm{C}$ for alewife (Brandt et al. 1980) and $7-15^{\circ} \mathrm{C}$ for rainbow smelt (Scott and Crossman 1973). During June and July, it is likely that the alewife and rainbow smelt inhabited the deeper, cooler waters. October was the only month during
the 2004 sampling that brook silverside (Labidesthes sicculus), a warmwater species (Coker et al. 2001), was captured; the majority of which were caught at Site 4, upstream and downstream of the Lambton Generating Station (LGS) (Appendix 17). The downstream segment (outflow of LGS) of this site had the highest temperature ( $17.3^{\circ} \mathrm{C}$ ) of any other site in October (Appendix 15), resulting in brook silverside congregating in higher densities at this location. Twenty-nine smallmouth bass (Micropterus dolomieu) were caught in June (the majority at Site 1), but only eight and five were caught in July and October, respectively (Appendix 6). Smallmouth bass spawn in spring, mainly during the months of May and June (Scott and Crossman 1973); therefore, it is possible that the high numbers seen in June are a result of individuals congregating to spawn. In July, 176 shorthead redhorse (Moxostoma macrolepidotum) were netted, while in June and October only 35 and 15, respectively, were caught (Appendix 6). Aside from the warmer water (average July water temperature was $22.09^{\circ} \mathrm{C}$ ) that the shorthead redhorse prefers (Scott and Crossman 1973), it is unclear as to why there is such a large difference in numbers. Two individuals from the Ictiobus genus were captured only in October 2004. They had subterminal mouths and were not bigmouth buffalo (/ctiobus cyprinellus), a species of Special Concern (COSEWIC 2004). Tissue samples were taken from the Ictiobus specimens and sent for DNA analyses, as distinguishing between the smallmouth buffalo (Ictiobus bubalus) and the black buffalo (Ictiobus niger) (another species of Special Concern (COSEWIC 2004)) is difficult. It is not clear why Ictiobus were not collected in June or July. Another species of Special Concern detected during the 2004 sampling, spotted sucker, has a lirnited distribution in Canada and is found only in the drainages of lakes St. Clair and Erie (COSEWIC 2004). No spotted suckers were seen in June; however, five were caught in July, and 12 in October (Appendix 6). More than half the specimens caught in October were caught at Site 4, at the LGS (Appendix 6), where the water temperature was higher at the station's outflow (Appendix 15).

The CPUE values for June, July and October were comparable. June had the lowest ( $2.75 \mathrm{fish} / \mathrm{min}$ ) and October had the highest ( $3.94 \mathrm{fish} / \mathrm{min}$ ) (Table 2). Site 4 had the highest CPUE in the month of June (11.05) (Appendix 7)
suggesting high diversity; however, $67 \%$ of the catch was comprised of one species (yellow perch (Perca flavescens)). Similar results were observed for Site 3 in July and Site 1 in October, where the CPUE values were high but the total catches were comprised of just a few species (Appendix 6). This supports the findings of Bayley and Austen (2002) that the CPUE cannot accurately describe fish assemblages.

The IBI scores did not differ significantly between the seasons, but individual site scores did show some variation over the seasons. The individual site $|B|$ scores fluctuated over the three month sampling period with no discernible pattern (Table 6, Figure 5, Appendix 11). Site 1, classified as 'degraded' in 1992 by the MOE, had IBI values that were higher than many of the 'unimpaired' or 'intermediate' sites (Table 6, Figure 5, Appendix 11). Site 8, the control site, had IBI scores in June and July of 17 (Very Poor to Poor) and 14 (Very Poor), respectively. The IBI score for Site 8 did increase to 25 (Fair) in October as a result of an increase in generalist feeders, a higher species richness, and a reduction in the number of invasive species, when compared to June and July (Appendix 11).

The MI was calculated at each site for each month sampled to examine the relationship between the number of individuals caught at a site and the number of species caught per site. It is not a comprehensive index like the IBI as it only deals with one component of ecology, species diversity, but is more objective than the IBI. Conversely, the IBI deals with several other components such as trophic guild composition, individual health, abundance, as well as species diversity. In June, Site 4 had the highest MI score (9.04), and Site 6 had the lowest (1.37) (Table 5, Figure 6). Once again, Site 1, the 'degraded' site outperformed many of the 'unimpaired' or 'intermediate' sites, with an MI score of 7.63, the second highest of all the sites in June. Site 7 had the highest Ml score and Site 8, the control site, had the lowest in July. In October, Sites 7 and 3 had the highest and lowest Ml scores, respectively (Table 5, Figure 6).

### 4.3. NOTES ON THE INDEX OF BIOTIC INTEGRITY

The IBI can be a useful tool to assess ecosystem health, if applied correctly (i.e. with relevant metrics that aren't confounding or redundant (Hughes
et al. 1998), and with standardized trophic guild classifications). The drawbacks of indices, such as the IBI, are that they cannot convey causal relationships (Hughes et al. 1998), and they can be difficult to interpret because of either double counting or changes in one variable that mask changes in another (EPA 2003). Hamilton's (1987) adaptation of the IBI, which used the metric 'species richness', was used to interpret the 1994 data. Hughes et al. (1998) suggested that a more appropriate metric would be 'native species richness'. They did not want to confound species richness (a metric that decreases with degradation) with invasive species (a metric that often increases with disturbance) (Hughes et al. 1998). Depending on the number of fishes involved, trophic guild classification can have a major impact on IBI scores. In 1994, an out-of-date trophic guild classification was used in the IBI calculations. For the purpose of comparison, the same classification system was used in this study. However, IBI scores for 1994 and 2004 data were also re-calculated using a slightly modified IBI classification scheme (Appendix 4), as well as a new trophic guild classification system (Appendix 5). The new IBI classification scheme altered the 'naturally spawned salmonid and coregonid species' metric to include only native naturally spawned salmonid and coregonid species. The new trophic classification system was based on Coker et al. (2001). The re-calculated IBI scores for 1994 and 2004 data are in Tables 6 and 7, and Appendices 12 and 13, respectively.

Although there was no overall significant difference between August 1994 and July 2004 IBI scores calculated using the modified IBI scheme, the scores did drop at almost every site for both 1994 and 2004. The IBI scores for the 1994 data were lower at every site except for Site 1, where the score remained the same (Table 7). The overall IBI score for 1994 dropped from 34 (Good) to 27 (Fair). Similar results were seen for the 2004 data (June, July and October). The re-calculated IBI scores did not differ significantly from each other; however, a pattern of decreasing scores was observed. IBI scores at almost every site were lower using the new classification system (Table 6). Overall IBI scores for all sites in June, July and October decreased from 30, 28 and 32 to 22, 23 and 23 , respectively (Table 6). The variation in IBI scores using the two different
methods highlights the impact that the trophic guild metric can have when it is used in the IBI; therefore, it is important to use a classification system that is based on the most current information available.

The IBI classification scheme used by OMNR in 1994 incorporated a metric that counted the number of hybrids, fish(es) with lamprey scars or diseases, as well as invasive species. Only invasive species richness was used to calculate this metric, as data regarding hybrids and diseased/scarred fish were not recorded during either sampling period. Had these data been recorded and included in the IBI calculations, the IBI scores may have been affected for both periods.

Regarding metric and index scoring, Hamilton (1987) scored the metrics as $0,1,3$ or 5 ; whereas, Minns et al. (1994) suggested that an index based on a continuous scoring system of $0.0-100.0$ for the IBI and $0.0-10.0$ for the metrics is more preferable. This method of scoring reduces the variance when metric values $\leq 1$ are scored as different categories, and should make IBIs less variable and more easily understood (Hughes et al. 1998).

The IBI scores at these sites may not give a complete picture of the integrity of the aquatic ecosystem because of sampling bias. Boat electrofishing is most effective at depths of $\leq 2 \mathrm{~m}$. In deeper water, benthic species may not be properly represented in the sample. Also, IBIs for individual months can be affected by seasonal migrations of various species; therefore, it may not be valid to compare IBI scores between months, but to compare them for the same season between years.

Limited time and funds make it hard to effectively sample a river as large as the St. Clair, which requires the use of different gear types and a large number of sample sites to get a comprehensive picture of the fish assemblages. In large rivers, where obtaining representative samples of the fish community can be difficult, it may be useful to use other taxa such as benthic invertebrates that are less mobile than fishes, to evaluate the health of the ecosystem. Ideally, as suggested by Karr (1981), biological monitoring programs should be based on an integrative approach involving the evaluation of more than one major taxa.

### 5.0. RECOMMENDATIONS

True estimates of species richness and indices, such as IBI, using boat electrofishing techniques are difficult. The use of different gear types could allow a more accurate calculation of the IBI, as well as estimating true species richness and describing fish populations (Pugh and Schramm 1998, Colvin 2002, Hughes et al. 2002, Mandrak et al. unpubl. data).

Future sampling of the St. Clair River should incorporate an assortment of active and passive sampling methods. Active sampling methods include boat electrofishing, boat seining ( $1 / 4^{\prime \prime}$ mesh) and manual seining ( $1 / 4^{\prime \prime}$ mesh) (N.E. Mandrak, unpubl. data). Passive sampling methods include hoopnets and trapnets with small diameter mesh ( $1 / 4^{\prime \prime}$ ) (Mandrak et al. unpubl. data). Trawling techniques should be considered within deep riverine habitats (>2 m deep) to seek out benthic fishes (e.g. darters and madtoms). Any further sampling of the St. Clair River should include sampling during both day and night in order to get a more accurate representation of the fish assemblage.

As well as the additional sampling recommended above, boat electrofishing using the same effort as previously undertaken, must be done to examine trends in IBIs over time, including earlier time periods.

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Table 1. Summary of DFO 2004 catch data for the St. Clair River.

| Catch Data | June | July | October | Total |
| :--- | ---: | ---: | ---: | ---: |
| Total Fishes Captured | 664 | 740 | 1020 | $\mathbf{2 4 2 4}$ |
| Mean Number of Fishes <br> Captured/Site (sub-site A+B) | 83 | 92.5 | 127.5 | $\mathbf{1 0 1}$ |
| Minimum Number of Fishes <br> Captured (among sites) | 5 | 10 | 43 |  |
| Maximum Number of Fishes <br> Captured (among sites) | 272 | 292 | 310 |  |
|  |  |  |  |  |

Table 2. Summary of DFO 2004 sampling effort on the St. Clair River.

| Sampling Effort | June | July | October | Average |
| :--- | ---: | ---: | ---: | ---: |
| Mean CPUE (catch/min) | 2.75 | 3.14 | 3.94 | $\mathbf{3 . 2 8}$ |
| Mean Effort/Site (sub-site A+B) <br> (min) | 32.5 | 29.38 | 32.9 | $\mathbf{3 1 . 5 9}$ |
| Total Sampling Effort (min) | 259.98 | 235.07 | 263.2 | $\mathbf{2 5 2 . 7 5}$ |

Table 3. Species captured by DFO in 2004, St. Clair River.

| Legend |  | Present | Absent |  |
| :---: | :---: | :---: | :---: | :---: |
| Scientific Name | Common Name | June | July | October |
| Alosa pseudoharengus | alewife |  |  |  |
| Ambloplites rupestris | rock bass |  |  |  |
| Amia calva | bowfin |  |  |  |
| Aplodinotus grunniens | freshwater drum |  |  |  |
| Carpiodes cyprinus | quillback |  |  |  |
| Catostomus commersonii | white sucker |  |  |  |
| Coregonus sp. | whitefish sp. |  |  |  |
| Cottus bairdii | mottled sculpin |  |  |  |
| Cyprinella spiloptera | spottin shiner |  |  |  |
| Cyprinus carpio | common carp |  |  |  |
| Dorosoma cepedianum | gizzard shad |  |  |  |
| Esox lucius | northern pike |  |  |  |
| Etheostoma caeruleum | rainbow darter |  |  |  |
| Gasterosteus aculeatus | threespine stickleback |  |  |  |
| Hypentelium nigricans | northern hog sucker |  |  |  |
| Ichthyomyzon unicuspis | silver lamprey |  |  |  |
| Ictiobus sp. | buffalo sp. |  |  |  |
| Labidesthes sicculus | brook silverside |  |  |  |
| Lampetra appendix | American brook lamprey |  |  |  |
| Lepisosteus osseus | longnose gar |  |  |  |
| Lepomis cyanellus | green sunfish |  |  |  |
| Lepomis cyanellus $\times$ L. macrochirus | green sunfish/bluegill hybrid |  |  |  |
| Lepomis gibbosus | pumpkinseed |  |  |  |
| Lepomis macrochirus | bluegill |  |  |  |
| Luxilus chrysocephalus | striped shiner |  |  |  |
| Lythrurus umbratilis | redfin shiner |  |  |  |
| Micropterus dolomieu | smallmouth bass |  |  |  |
| Micropterus salmoides | largemouth bass |  |  |  |
| Minytrema melanops | spotted sucker |  |  |  |
| Morone americana | white perch |  |  |  |
| Morone chrysops | white bass |  |  |  |
| Moxostoma anisurum | silver redhorse |  |  |  |
| Moxostoma erythrurum | golden redhorse |  |  |  |
| Moxostoma macrolepidotum | shorthead redhorse |  |  |  |
| Neogobius melanostomus | round goby |  |  |  |
| Nocomis biguttatus | hornyhead chub |  |  |  |
| Notemigonus crysoleucas | golden shiner |  |  |  |
| Notropis atherinoides | emerald shiner |  |  |  |
| Notropis hudsonius | spottail shiner |  |  |  |
| Notropis volucellus | mimic shiner |  |  |  |
| Oncorhynchus mykiss | rainbow trout |  |  |  |
| Oncorhynchus ishawytscha | Chinook salmon |  |  |  |
| Osmerus mordax | rainbow smelt |  |  |  |
| Perca flavescens | yellow perch |  |  |  |
| Percina caprodes | logperch |  |  |  |
| Pimephales notatus | bluntnose minnow |  |  |  |
| Pimephales promelas | fathead minnow |  |  |  |
| Pomoxis nigromaculatus | black crappie |  |  |  |
| Salmo trutta | brown trout |  |  |  |
| Sander vitreus | walleye |  |  |  |
| Semotilus atromaculatus | creek chub |  |  |  |
|  | Total Species Observed | 36 | 33 | 33 |

Table 4. Summary of species collected in the St. Clair River in June, July and October by DFO, 2004.

|  | June | July | October |
| :--- | :---: | :---: | :---: |
| Total Richness | 36 | 33 | 33 |
| Unique Species | 10 | 4 | 4 |
| Common Species | 18 |  |  |
| Total Richness | 36 |  |  |

Table 5. Margalef's Diversity Index scores for DFO 2004 data from the St. Clair River $(\mathrm{MI}=(\mathrm{S}-1) / \log \mathrm{N}$, where S is the number of species and N is the number of individuals of all species).

| Site | June | July | October |
| :---: | ---: | ---: | ---: |
| $\mathbf{1}$ | 7.63 | 6.93 | 5.22 |
| $\mathbf{2}$ | 2.86 | 3.99 | 5.52 |
| $\mathbf{3}$ | 6.52 | 8.92 | 4.45 |
| $\mathbf{4}$ | 9.04 | 8.89 | 7.21 |
| $\mathbf{5}$ | 4.47 | 6.14 | 6.62 |
| $\mathbf{6}$ | 1.37 | 4.10 | 5.02 |
| $\mathbf{7}$ | 5.00 | 12.56 | 7.64 |
| $\mathbf{8}$ | 5.10 | 2.00 | 5.51 |

Table 6. IBI scores for DFO 2004 data from the St. Clair River (IBI calculated two ways. IBI 1 uses Hamilton's adaptation of index and OMNR trophic classification system (Appendix 2, Appendix 3). IBI 2 uses a slightly modified IBI scheme and new trophic classification system based on Coker et al. (2001) (Appendix 4, Appendix 5)).

| IBI 1 |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sites |  |  |  |  |  |  |  |  |  |
| Month | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | All Sites |
| June | 30 <br> Fair to Good | 10 <br> Very Poor | 25 <br> Fair | 23 <br> Poor | 15 <br> Very Poor | 23 <br> Poor | 13 <br> Very Poor | 17 <br> Very Poor to Poor | 30 <br> Fair to Good |
| July | 22 <br> Poor | $\begin{gathered} \hline 11 \\ \text { Very } \\ \text { Poor } \\ \hline \end{gathered}$ | 25 <br> Fair | 29 <br> Fair | $\begin{gathered} 21 \\ \text { Poor } \end{gathered}$ | $\begin{gathered} \hline 15 \\ \text { Very } \\ \text { Poor } \\ \hline \end{gathered}$ | $23$ <br> Poor | $\begin{gathered} 14 \\ \text { Very } \\ \text { Poor } \\ \hline \end{gathered}$ | 28 <br> Fair |
| October | 26 <br> Fair | $19$ <br> Poor | 19 <br> Poor | 27 <br> Fair | 17 <br> Very Poor to Poor | 17 <br> Very Poor to Poor | 21 <br> Poor | 25 <br> Fair | 32 <br> Good |
| Sites |  |  |  |  |  |  |  |  |  |
| Month | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | All Sites |
| June | 18 <br> Poor | 10 <br> Very Poor | 20 <br> Poor | 23 <br> Poor | 14 <br> Very Poor | 14 <br> Very Poor | $12$ <br> Very Poor | 15 <br> Very Poor to Poor | 22 Poor |
| July | $\begin{array}{r} 19 \\ \text { Poor } \\ \hline \end{array}$ | $\begin{gathered} 10 \\ \text { Very } \\ \text { Poor } \\ \hline \end{gathered}$ | $\begin{gathered} 23 \\ \text { Poor } \\ \hline \end{gathered}$ | $\begin{gathered} 23 \\ \text { Poor } \\ \hline \end{gathered}$ | $\begin{array}{r} 18 \\ \text { Poor } \\ \hline \end{array}$ | $\begin{gathered} 10 \\ \text { Very } \\ \text { Poor } \\ \hline \end{gathered}$ | $\begin{array}{r} 18 \\ \text { Poor } \\ \hline \end{array}$ | 9 Very Poor | 23 Poor |
| October | 23 <br> Poor | 17 <br> Very Poor to Poor | 17 <br> Very Poor to Poor | 19 Poor | 15 <br> Very <br> Poor | Very Poor |  | 14 <br> Very Poor | 23 Poor |

Table 7. IBI scores for OMNR 1994 and DFO 2004 data from the St. Clair River (IBI calculated two ways. IBI 1 uses Hamilton's adaptation of index and OMNR trophic classification system (Appendix 2, Appendix 3). IBI 2 uses a slightly modified IBI scheme and new trophic classification system based on Coker et al. (2001) (Appendix 4, Appendix 5)).

| IBI 1 |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Sites |  |  |  |  |  |  |  |  |  |
| Sampling Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | All Sites |
| Aug 1994 | $\begin{gathered} 23 \\ \text { Poor } \\ \hline \end{gathered}$ | $22$ <br> Poor | 25 <br> Fair | 28 <br> Fair | 29 <br> Fair | $\begin{gathered} 25 \\ \text { Very } \end{gathered}$ Poor | 23 <br> Poor | $\begin{gathered} \hline 16 \\ \text { Very } \\ \text { Poor } \\ \hline \end{gathered}$ | $34$ <br> Good |
| July 2004 | $22$ <br> Poor | $\begin{gathered} \hline 11 \\ \text { Very } \\ \text { Poor } \\ \hline \end{gathered}$ | $\begin{aligned} & 25 \\ & \text { Fair } \end{aligned}$ | 29 <br> Fair | $\begin{gathered} 21 \\ \text { Poor } \end{gathered}$ | $\begin{gathered} \hline 15 \\ \text { Very } \\ \text { Poor } \\ \hline \end{gathered}$ | 23 <br> Poor | $\begin{gathered} 14 \\ \text { Very } \end{gathered}$ Poor | $\begin{gathered} 28 \\ \text { Fair } \end{gathered}$ |
| IBI2 |  |  |  |  |  |  |  |  |  |
| Sites |  |  |  |  |  |  |  |  |  |
| Sampling Year | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | All <br> Sites |
| Aug 1994 | 23 <br> Poor | 21 Poor | 21 Poor | 17 <br> Very Poor to Poor | 27 Fair | 21 Poor | 21 Poor | $13$ <br> Very Poor | 27 Fair |
| July 2004 | $\begin{gathered} 19 \\ \text { Poor } \\ \hline \end{gathered}$ | $\begin{gathered} \hline 10 \\ \text { Very } \\ \text { Poor } \\ \hline \end{gathered}$ | $\begin{gathered} \hline 23 \\ \text { Poor } \end{gathered}$ | $23$ <br> Poor | $18$ <br> Poor | $\begin{gathered} \hline 10 \\ \text { Very } \\ \text { Poor } \\ \hline \end{gathered}$ | $18$ <br> Poor | $\begin{gathered} \hline 9 \\ \text { Very } \\ \text { Poor } \\ \hline \end{gathered}$ | 23 Poor |

Table 8. Species captured during the OMNR (August 1994) survey and the DFO (July 2004) survey.

| Legend | Present | Absent |
| :--- | :--- | :--- |


| Scientific Name | Common Name | OMNR | DFO |
| :--- | :--- | :--- | :--- |
| Alosa pseudoharengus | alewife |  |  |
| Ambloplites rupestris | rock bass |  |  |
| Amia calva | bowfin |  |  |
| Aplodinotus grunniens | freshwater drum |  |  |
| Carassius auratus | goldfish |  |  |
| Carpiodes cyprinus | quillback |  |  |
| Catostomus commersonii | white sucker |  |  |
| Cyprinella spiloptera | spotfin shiner |  |  |
| Cyprinus carpio | common carp |  |  |
| Dorosoma cepedianum | gizzard shad |  |  |
| Esox lucius | northern pike |  |  |
| Gasterosteus aculeatus | threespine stickleback |  |  |
| Hypentelium nigricans | northern hog sucker |  |  |
| lchthyomyzon unicuspis | silver lamprey |  |  |
| Labidesthes sicculus | brook silverside |  |  |
| Lepisosteus osseus | longnose gar |  |  |
| Lepomis gibbosus | pumpkinseed |  |  |
| Lepomis macrochirus | bluegill |  |  |
| Luxilus chrysocephalus | striped shiner |  |  |
| Luxilus cornutus | common shiner |  |  |
| Lythrurus umbratilis | redfin shiner |  |  |
| Micropterus dolomieu | smallmouth bass |  |  |
| Micropterus salmoides | largemouth bass |  |  |
| Minytrema melanops | spotted sucker |  |  |
| Morone americana | white perch |  |  |
| Morone chrysops | white bass |  |  |
| Moxostoma anisurum | silver redhorse |  |  |
| Mooxtoma erythurum | golden redhorse |  |  |
| Moxostoma macrolepidotum | shorthead redhorse |  |  |
| Neogobius melanostomus | round goby |  |  |
| Nocomis biguttatus | hornyhead chub |  |  |
| Notemigonus crysoleucas | golden shiner |  |  |
| Notropis atherinoides | emerald shiner |  |  |
| Notropis hudsonius | spottail shiner |  |  |
| Notropis volucellus | mimic shiner |  |  |
| Oncorhynchus mykiss | rainbow trout |  |  |
| Oncorhynchus tshawytscha | Chinook salmon |  |  |
| Osmerus mordax | rainbow smelt |  |  |
| Perca flavescens | yellow perch |  |  |
| Percina caprodes | logperch |  |  |
| Pimephales notatus | bluntnose minnow |  |  |
| Salmo truuta | brown trout |  |  |
| Sander vitreus | walleye |  |  |
| Semotilus atromaculatus | creek chub |  |  |
|  | Total Species Observed | 39 |  |
|  |  |  |  |

Table 9. Summary of species collected by OMNR (August 1994) and DFO (July 2004) in the St. Clair River.

|  | OMNR | DFO |
| :--- | :---: | :---: |
| Total Richness | 39 | 33 |
| Unique Species | 11 | 5 |
| Common Species (1994, 2004) | $\mathbf{2 8}$ |  |
| Total Richness (Unique + <br> Common species) | $\mathbf{4 4}$ |  |

Table 10. Summary of OMNR (August 1994) and DFO (July 2004) catch data, St. Clair River.

| Catch Data | OMNR | DFO |
| :--- | ---: | ---: |
| Total Fishes Captured | 3720 | 740 |
| Mean Number of Fishes <br> Captured/Site | 465 | 92.5 |
| Minimum Number of Fishes <br> Captured (among sites) | 0 | 10 |
| Maximum Number of Fishes <br> Captured (among sites) | 1796 | 292 |

Table 11. Summary of OMNR (August 1994) and DFO (July 2004) sampling effort, St. Clair River.

| Sampling Effort | OMNR | DFO |
| :--- | ---: | ---: |
| Mean CPUE (catch/min) | 11.48 | 3.14 |
| Mean Effort/transect (min) | 40.13 | 29.38 |
| Total Sampling Effort (min) | 321.00 | 235.07 |



Figure 1. DFO 2004 St. Clair River sampling sites.


Figure 2. Cluster analysis based on seasonal relative abundance values. x_monthyy: $\mathrm{x}=$ site number; $\mathrm{y} y=\mathrm{year}$.


Figure 3. Cluster analysis based on annual relative abundance values. x_monthyy: $x=$ site number; $y=y$ year.


Figure 4. IBI scores for August 1994 OMNR and July 2004 DFO data, St. Clair River.


Figure 5. IBI scores for 2004 DFO data, St. Clair River.


Figure 6. MI scores for DFO 2004 data, St. Clair River.

Appendix 1. Summary of DFO 2004 site descriptions, St. Clair River.

| $\begin{aligned} & \text { Un } \\ & \text { Map } \\ & \hline \end{aligned}$ | Fielcl Number | Date | Start Latitucle | Start Longitude | Narrative Locality Description |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | SCR04COA170604001A | 17/06/2004 | 42.92807 | -82.45068 | Downstream of Sarnia; East shore; adjacent to First nations |
|  | SCR04COA170604001B | 17/06/2004 | 42.92441 | -82.45242 | Downstream of chemical plants; East shore@ First nations |
| 2 | SCR04COA170604003A | 17/06/2004 | 42.89474 | -82.46523 | West shore of Stag Island |
|  | SCR04COA170604003B | 17/06/2004 | 42.89474 | -82.46523 | West shore of Stag Island |
| 3 | SCR04COA170604004A | 17/06/2004 | 42.90069 | -82.45824 | Upstream of Corunna; downstream of Talfourd Creek |
|  | SCR04COA170604004B | 17/06/2004 | 42.89725 | -82.45747 | East shore @ Corunna; downstream of Talfourd Creek |
| 4 | SCR04COA160604007A | 16/06/2004 | 42.79572 | -82.47360 | Upstream of Lambton Generating Station |
|  | SCR04COA160604007B | 16/06/2004 | 42.79371 | -82.47223 | Downstream of Lambton Generating Station |
| 5 | SCR04COA160604008A | 16/06/2004 | 42.74936 | -82.46932 | Adjacent and upstream of Clay Creek |
|  | SCR04COA160604008B | 16/06/2004 | 42.74598 | -82.47147 | Downstream of Clay Creek |
| 6 | SCR04COA160604009A | 16/06/2004 | 42.68413 | -82.49736 | Downstream of Fawn Island; East shore |
|  | SCR04COA160604009B | 16/06/2004 | 42.68127 | -82.49893 | Downstream of Fawn Island; East shore |
| 7 | SCR04COA160604010A | 16/06/2004 | 42.69503 | -82.49373 | East shore of Fawn Island |
|  | SCR04COA160604010B | 16/06/2004 | 42.69116 | -82.49630 | Downstream of Fawn Island; East shore |
| 8 | SCR04COA170604011A | 17/06/2004 | 42.98932 | -82.42121 | Upstream Bay Point Sarnia Harbour |
|  | SCR04COA170604011B | 17/06/2004 | 42.98513 | -82.41834 | Downstream of Bay Point Sarnia Harbour |
| 1 | SCR04COA220704001A | 22/07/2004 | 42.92807 | -82.45068 | Downstream of Sarnia; East shore; adjacent to First nations |
|  | SCR04COA220704001B | 22/07/2004 | 42.92441 | -82.45242 | Downstream of chemical plants; East shore © First nations |
| 2 | SCR04COA220704003A | 22/07/2004 | 42.89474 | -82.46523 | West shore of Stag Island |
|  | SCR04COA220704003B | 22/07/2004 | 42.89474 | -82.46523 | West shore of Stag Island |
| 3 | SCR04COA220704004A | 22/07/2004 | 42.90069 | -82.45824 | Upstream of Corunna; downstream of Talfourd Creek |
|  | SCR04COA220704004B | 22/07/2004 | 42.89725 | -82.45747 | East shore © Corunna; downstream of Talfourd Creek |
| 4 | SCR04COA210704007A | 21/07/2004 | 42.79572 | -82.47360 | Upstream of Lambton Generating Station |
|  | SCR04COA210704007B | 21/07/2004 | 42.79371 | -82.47223 | Downstream of Lambton Generating Station |
| 5 | SCR04COA210704008A | 21/07/2004 | 42.75238 | -82.46738 | Adjacent and upstream of Clay Creek |
|  | SCR04COA210704008B | 21/07/2004 | 42.74936 | -82.46932 | Downstream of Clay Creek |
| 6 | SCR04COA210704009A | 21/07/2004 | 42.68413 | -82.49736 | Downstream of Fawn Island; East shore |
|  | SCR04COA210704009B | 21/07/2004 | 42.68127 | -82.49893 | Downstream of Fawn Island; East shore |
| 7 | SCR04COA210704010A | 21/07/2004 | 42.69503 | -82.49373 | East shore of Fawn Island |
|  | SCR04COA210704010B | 21/07/2004 | 42.69116 | -82.49630 | Downstream of Fawn Island; East shore |
| 8 | SCR04COA220704011A | 22/07/2004 | 42.98932 | -82.42121 | Upstream Bay Point Sarnia Harbour |
|  | SCR04COA220704011B | 22/07/2004 | 42.98513 | -82.41834 | Downstream of Bay Point Sarnia Harbour |

Appendix 1. Continued.

| 浬 On Map | Fleld Number | Date | Start Latitude | Start Longitude | Narrative Locality Description |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | SCR04COA141004001A | 14/10/2004 | 42.92807 | -82.45068 | Downstream of Sarnia; East shore; adjacent to First nations |
|  | SCR04COA141004001B | 14/10/2004 | 42.92441 | -82.45242 | Downstream of chemical plants; East shore © First nations |
| 2 | SCR04COA141004003A | 14/10/2004 | 42.89474 | -82.46523 | West shore of Stag Island |
|  | SCR04COA141004003B | 14/10/2004 | 42.89474 | -82.46523 | West shore of Stag Island |
| 3 | SCR04COA141004004A | 14/10/2004 | 42.90069 | -82.45824 | Upstream of Corunna; downstream of Talfourd Creek |
|  | SCR04COA141004004B | 14/10/2004 | 42.89725 | -82.45742 | East shore @ Corunna; downstream of Talfourd Creek |
| 4 | SCR04COA131004007A | 13/10/2004 | 42.79572 | -82.47360 | Upstream of Lambton Generating Station |
|  | SCR04COA131004007B | 13/10/2004 | 42.79371 | -82.47223 | Downstream of Lambton Generating Station |
| 5 | SCR04COA131004008A | 13/10/2004 | 42.74936 | -82.46932 | Adjacent and upstream of Clay Creek |
|  | SCR04COA131004008B | 13/10/2004 | 42.74598 | -82.47147 | Downstream of Clay Creek |
| 6 | SCR04COA131004009A | 13/10/2004 | 42.68413 | -82.49736 | Downstream of Fawn Island; East shore |
|  | SCR04COA131004009B | 13/10/2004 | 42.68127 | -82.49893 | Downstream of Fawn Island; East shore |
| 7 | SCR04COA131004010A | 13/10/2004 | 42.69503 | -82.49373 | East shore of Fawn Island |
|  | SCR04COA131004010B | 13/10/2004 | 42.69116 | -82.49630 | Downstream of Fawn Island; East shore |
| 8 | SCR04COA141004011A | 14/10/2004 | 42.98932 | -82.42121 | Upstream Bay Point Sarnia Harbour |
|  | SCR04COA141004011B | 14/10/2004 | 42.98513 | -82.41834 | Downstream of Bay Point Sarnia Harbour |

Appendix 2. Hamilton's (1987) IBI classification scheme.

| Section | Description | Scoring Criteria |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0 | 1 | 3 | 5 |
| Species Richness and Composition | Number of species collected in each sample (as a \% of total collected in the entire AOC) | 0 | 0-25\% | 26-50\% | > 50\% |
|  | Number of percid species present in each sample area | 0 | 1 | 2 | $\geq 3$ |
|  | Number of naturally-spawned salmonid and coregonid species present in each sample area | 0 | 1 | 2 | $\geq 3$ |
|  | Subtotal: |  |  |  |  |
| Trophic Composition | Proportion of individuals considered specialist/insectivores/planktivores | 0 | < 20\% | 20-40\% | > 40\% |
|  | Proportion of individuals considered generalists | 0 | > 40\% | 20-40\% | < 20\% |
|  | Proportion of individuals considered top piscivores | 0 | < 2\% | 2-5\% | > 5\% |
|  | Subtotal: |  |  |  |  |
| Fish Abundance and Health | Ratio of CPUE in the sample area to mean AOC CPUE (as \%) | - | < 80\% | 80-120\% | > 120\% |
|  | Occurrence of individuals which are hybrids, diseased, have lamprey scars or are invading species | - | > 5\% | 1-5\% | 0 |
|  | Subtotal: |  |  |  |  |
|  | Total: |  |  |  |  |
| Rating System: | $\begin{aligned} & <15=\text { very poor } \\ & 18-23=\text { poor } \\ & 25-29=\text { fair } \\ & 31-34=\text { good } \\ & 37-40=\text { excellent } \end{aligned}$ |  |  |  |  |

Appendix 3. 1994 OMNR trophic guild classifications based on Hamilton (1987), Scott and Crossman (1973) and Whitehead et al. (1986). New species observed in 2004 (in bold) were classified according to Coker et al. (2001).

*invasive species

Appendix 4. Modified IBI scheme.


Appendix 5. 2004 trophic guild classifications based on Coker et al. (2001).

| Generallists |  | Specialists |  | Top Carnivores |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Common Name | Scientific Name | Common Name | Scientific Name | Common Name | Scientific Name |
| rock bass <br> freshwater drum <br> quillback <br> white sucker <br> mottled sculpin <br> spotfin shiner <br> common carp* <br> rainbow darter <br> threespine stickleback <br> northern hog sucker <br> buffalo sp. <br> pumpkinseed <br> bluegill <br> striped shiner <br> redfin shiner <br> spotted sucker <br> silver redhorse <br> golden redhorse <br> shorthead redhorse <br> round goby* <br> hornyhead chub <br> golden shiner <br> emerald shiner <br> spottail shiner <br> mimic shiner <br> rainbow smelt* <br> logperch <br> bluntnose minnow <br> fathead minnow <br> creek chub | Ambloplites rupestris <br> Aplodinotus grunniens <br> Carpiodes cyprinus <br> Catostomus commersonii <br> Cottus bairdiI <br> Cyprinella spiloptera <br> Cyprinus carpio <br> Etheostoma caeruleum <br> Gasterosteus aculeatus <br> Hypentelium nigricans <br> Ictiobus sp. <br> Lepomis gibbosus <br> Lepomis macrochirus <br> Luxilus chrysocephalus <br> Lythrurus umbratiils <br> Minytrema melanops <br> Moxostoma anisurum <br> Moxostoma erythrurum <br> Moxostoma macrolepidotum <br> Neogoblus melanostomus <br> Nocomis biguttatus <br> Notemigonus crysoleucas <br> Notropis atherinoides <br> Notropis hudsonlus <br> Notropis volucellus <br> Osmerus mordax <br> Percina caprodes <br> Pimephales notatus <br> Pimephaies promelas <br> Semotilus atromaculatus | alewife* <br> cisco sp. <br> gizzard shad <br> silver lamprey <br> brook silverside <br> American brook lamprey | Alosa pseudoharengus Coregonus sp. <br> Dorosoma cepedianum Ichthyomyzon unicuspis <br> Labidesthes sicculus <br> Lampetra appendix | bowfin northern pike longnose gar green sunfish smallmouth bass largemouth bass white perch* white bass rainbow trout* Chinook salmon* yellow perch black crappie brown trout* walleye | Amia calva <br> Esox lucius <br> Lepisosteus asseus <br> Lepomis cyanellus <br> Micropterus dolomieu <br> Micropterus salmoides <br> Morone americana <br> Morone chrysops <br> Oncomynchus mykiss <br> Oncomynchus tshawytscha <br> Perca flavescens <br> Pomoxis nigromaculatus <br> Salmo trutta <br> Sander vitreus |

[^0]Appendix 6. Surnmary of species caught by site by DFO in 2004, St. Clair River.

| Sampling Perlod | June |  |  |  |  |  |  |  |  | July |  |  |  |  |  |  |  |  | October |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Site | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | Total | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | Total | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | Total |
| Alosa pseudoharengus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 176 | 26 | 34 | 4 | 4 | 1 | 0 | 0 | 245 |
| Ambloplites rupestris | 28 | 0 | 0 | 0 | 0 | 0 | 1 | 3 | 32 | 0 | 0 | 1 | 0 | 0 | 0 | 2 | 0 | 3 | 0 | 0 | 0 | 3 | 0 | 0 | 2 | 3 | 8 |
| Amia calva | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aplodinotus grunniens | 0 | 0 | 4 | 1 | 0 | 0 | 0 | 0 | 5 | 1 | 0 | 3 | 0 | 4 | 0 | 4 | 0 | 12 | 1 | 1 | 4 | 0 | 2 | 9 | 3 | 0 | 20 |
| Carpiodes cyprinus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Catostomus commersonii | 0 | 0 | 4 | 15 | 2 | 0 | 1 | 0 | 22 | 0 | 5 | 5 | 7 | 1 | 0 | 4 | 0 | 22 | 1 | 9 | 8 | 10 | 7 | 2 | 30 | 2 | 69 |
| Coregonus sp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cottus bairdii | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cyprinella spiloptera | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cyprinus carpio | 0 | 0 | 1 | 28 | 0 | 0 | 1 | 0 | 30 | 1 | 0 | 12 | 4 | 0 | 0 | 0 | 0 | 17 | 0 | 2 | 1 | 5 | 3 | 2 | 7 | 0 | 20 |
| Dorosoma cepedianum | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 2 |
| Esox lucius | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 |
| Etheostoma caeruleum | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gasterostous aculeatus | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hypentelium nigricans | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 3 | 3 | 0 | 0 | 1 | 1 | 0 | 5 | 0 | 10 |
| ichthyomyzon unicuspis | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 2 |
| Ictiobus sp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 2 |
| Labidesthes sicculus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 16 | 1 | 3 | 1 | 0 | 22 |
| Lampetra appendix | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepisostous ossous | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
| Lepomis cyanellus | 2 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| L.cyanellus x L.macrochirus hybria | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis gibbosus | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis macrochirus | 7 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 11 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Luxilus chrysocephalus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 0 | 1 | 0 | 1 | 0 | 6 | 0 | 0 | 0 | 3 | 0 | 1 | 1 | 0 | 5 |
| Lythrurus umbratilis | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Micropterus dolomieu | 16 | 0 | 0 | 1 | 6 | 0 | 1 | 5 | 29 | 1 | 0 | 0 | 1 | 5 | 0 | 1 | 0 | 8 | 4 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 5 |
| Micropterus salmoides | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| Minytrema melanops | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 1 | 0 | 5 | 0 | 0 | 0 | 7 | 0 | 0 | 4 | 1 | 12 |
| Morone americana | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 0 | 0 | 3 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
| Morone chrysops | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 4 | 3 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Moxostoma anisurum | 1 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 3 | 2 | 6 | 4 | 2 | 0 | 17 | 0 | 0 | 0 | 1 | 2 | 0 | 0 | 0 | 3 |
| Moxostoma erythrurum | 0 | 0 | 3 | 3 | 1 | 0 | 0 | 0 | 7 | 1 | 1 | 13 | 0 | 1 | 1 | 2 | 0 | 19 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 |
| Moxostoma macrolepidotum | 1 | 1 | 19 | 10 | 2 | 2 | 0 | 0 | 35 | 48 | 8 | 90 | 9 | 7 | 5 | 9 | 0 | 176 | 0 | 3 | 6 | 0 | 0 | 2 | 2 | 2 | 15 |

Appendix 6. Continued.

| Sampling Period | June |  |  |  |  |  |  |  |  | July |  |  |  |  |  |  |  |  | October |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Site | 1 | 2 | 3. | 4 | 5 | 6 | 7 | 8 | Total | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | Total | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | Total |
| Neogobius melanostomus | 16 | 0 | 5 | 1 | 2 | 0 | 0 | 4 | 28 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Nocomis biguttatus | 6 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 8 | 1 | 1 | 0 | 2 | 0 | 0 | 2 | 0 | 6 | 0 | 1 | 0 | 2 | 0 | 0 | 1 | 0 | 4 |
| Notemigonus crysoleucas | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Notropis atherinoides | 23 | 0 | 20 | 2 | 0 | 0 | 0 | 0 | 45 | 32 | 0 | 76 | 48 | 1 | 11 | 6 | 7 | 181 | 5 | 2 | 11 | 19 | 16 | 28 | 23 | 17 | 121 |
| Notropis hudsonius | 1 | 0 | 8 | 5 | 0 | 12 | 0 | 0 | 26 | 0 | 0 | 2 | 0 | 0 | 0 | 5 | 0 | 7 | 1 | 1 | 0 | 9 | 0 | 0 | 18 | 0 | 29 |
| Notropis volucellus | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 2 | 1 | 0 | 0 | 1 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Oncorhynchus mykiss | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 9 | 10 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 2 |
| Oncorhynchus tshawytscha | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| Osmerus mordax | 0 | 3 | 2 | 0 | 0 | 0 | 0 | 0 | 5 | 3 | 3 | 1 | 1 | 0 | 3 | 0 | 0 | 11 | 69 | 13 | 31 | 3 | 21 | 41 | 4 | 0 | 182 |
| Perca flavescens | 41 | 1 | 62 | 184 | 8 | 15 | 5 | 7 | 323 | 8 | 11 | 53 | 18 | 24 | 25 | 40 | 0 | 179 | 7 | 6 | 8 | 61 | 5 | 8 | 62 | 0 | 157 |
| Percina caprodes | 0 | 0 | 7 | 0 | 0 | 0 | 0 | 0 | 7 | 2 | 0 | 14 | 1 | 10 | 0 | 1 | 0 | 28 | 39 | 0 | 1 | 0 | 1 | 0 | 0 | 14 | 55 |
| Pimephales notatus | 1 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 3 | 3 | 0 | 0 | 1 | 0 | 7 | 0 | 0 | 0 | 20 | 0 | 0 | 0 | 0 | 20 |
| Pimephales promelas | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pomoxis nigromaculatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 1 |
| Salmo trutta | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Sander vitreus | 1 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 2 | 3 | 0 | 0 | 0 | 0 | 5 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Semotilus atromaculatus | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix 7. CPUE values (catch/min) for 2004 DFO St. Clair River sites.

| Site | June | July | October |
| :---: | ---: | ---: | ---: |
| $\mathbf{1}$ | 3.87 | 3.00 | 7.42 |
| $\mathbf{2}$ | 0.22 | 1.44 | 2.87 |
| $\mathbf{3}$ | 4.12 | 9.21 | 3.05 |
| $\mathbf{4}$ | 11.05 | 4.84 | 7.30 |
| $\mathbf{5}$ | 0.74 | 2.19 | 2.13 |
| $\mathbf{6}$ | 0.85 | 1.85 | 3.46 |
| $\mathbf{7}$ | 0.31 | 2.32 | 4.26 |
| $\mathbf{8}$ | 0.85 | 0.30 | 1.02 |

Appendix 8. Summary of DFO 2004 sampling effort and electrofishing settings, St. Clair River (sampling performed using a 6.35 m Model SR-20 Smith-Root dual boom electrofishing boat with a 7.5 kW Smith-Root generator and 7.5 GPP control box).

| $\begin{aligned} & \text { \# On } \\ & \text { Map } \\ & \hline \end{aligned}$ | Field Number | Duration (8) | Electrofishing Settings |
| :---: | :---: | :---: | :---: |
| 1 | SCR04COA170604001A | 1103 | $1000 \mathrm{~V}, 8 \mathrm{mpss}, 60 \mathrm{~Hz}$ @ 40\% |
|  | SCR04COA170604001B | 1207 | $1000 \mathrm{~V}, 7 \mathrm{amps}, 60 \mathrm{~Hz}$ @ $40 \%$ |
| 2 | SCR04COA170604003A | 800 | $1000 \mathrm{~V}, 7.5 \mathrm{amps}, 60 \mathrm{~Hz}$ @ $80 \%$ |
|  | SCR04COA170604003B | 597 | $1000 \mathrm{~V}, 7.5 \mathrm{mpss}, 60 \mathrm{~Hz}$ @ $80 \%$ |
| 3 | SCR04COA170604004A | 1033 | $1000 \mathrm{~V}, 7.5 \mathrm{amps}, 60 \mathrm{~Hz}$ @ $40 \%$ |
|  | SCR04COA170604004B | 1004 | $1000 \mathrm{~V}, 7.5 \mathrm{amps}, 60 \mathrm{~Hz}$ @ $40 \%$ |
| 4 | SCR04COA160604007A | 718 | $1000 \mathrm{~V}, 8 \mathrm{amps}, 60 \mathrm{~Hz}$ @ $40 \%$ |
|  | SCR04COA160604007B | 759 | $1000 \mathrm{~V}, 8 \mathrm{mpss}, 60 \mathrm{~Hz}$ @ $40 \%$ |
| 5 | SCR04COA160604008A | 896 | $1000 \mathrm{~V}, 8 \mathrm{mpss}, 60 \mathrm{~Hz}$ @ 40\% |
|  | SCR04COA160604008B | 896 | $1000 \mathrm{~V}, 8 \mathrm{mmps}, 60 \mathrm{~Hz}$ @ $40 \%$ |
| 6 | SCR04COA160604009A | 1101 | $1000 \mathrm{~V}, 7 \mathrm{mmps}, 60 \mathrm{~Hz}$ @ $40 \%$ |
|  | SCR04COA160604009B | 936 | $1000 \mathrm{~V}, 7 \mathrm{mpss}, 60 \mathrm{~Hz}$ @ 40\% |
| 7 | SCR04COA160604010A | 1031 | $1000 \mathrm{~V}, 8 \mathrm{amps}, 60 \mathrm{~Hz}$ @ $40 \%$ |
|  | SCR04COA160604010B | 901 | $1000 \mathrm{~V}, 8 \mathrm{mps}, 60 \mathrm{~Hz}$ @ $40 \%$ |
| 8 | SCR04COA170604011A | 1208 | $1000 \mathrm{~V}, 7 \mathrm{mpss}, 60 \mathrm{~Hz}$ @ $40 \%$ |
|  | SCR04COA170604011B | 1409 | $1000 \mathrm{~V}, 7 \mathrm{mpss}, 60 \mathrm{~Hz}$ @ $40 \%$ |
| 1 | SCR04COA220704001A | 1020 | $1000 \mathrm{~V}, 5.5 \mathrm{mpss}, 30 \mathrm{~Hz}$ @ $55 \%$ |
|  | SCR04COA220704001B | 1079 | $1000 \mathrm{~V}, 7 \mathrm{mpss}, 60 \mathrm{~Hz}$ @ 35\% |
| 2 | SCR04COA220704003A | 645 | $1000 \mathrm{~V}, 7 \mathrm{mmps}, 60 \mathrm{~Hz}$ @ $40 \%$ |
|  | SCR04COA220704003B | 687 | $1000 \mathrm{~V}, 6.5 \mathrm{amps}, 60 \mathrm{~Hz}$ @ $80 \%$ |
| 3 | SCR04COA220704004A | 843 | $1000 \mathrm{~V}, 7 \mathrm{mpss}, 60 \mathrm{~Hz}$ @ 35\% |
|  | SCR04COA220704004B | 1060 | $1000 \mathrm{~V}, 7 \mathrm{mpss}, 60 \mathrm{~Hz}$ @ $40 \%$ |
| 4 | SCR04COA210704007A | 572 | $1000 \mathrm{~V}, 7 \mathrm{mpss}, 60 \mathrm{~Hz}$ @ 60\% |
|  | SCR04COA210704007B | 742 | $1000 \mathrm{~V}, 7 \mathrm{amps}, 30 \mathrm{~Hz}$ @ 60\% |
| 5 | SCR04COA210704008A | 868 | $1000 \mathrm{~V}, 7 \mathrm{mpss}, 30 \mathrm{~Hz}$ @ 60\% |
|  | SCR04COA210704008B | 830 | $1000 \mathrm{~V}, 7 \mathrm{mps}, 30 \mathrm{~Hz}$ @ 60\% |
| 6 | SCR04COA210704009A | 826 | $1000 \mathrm{~V}, 6.5 \mathrm{mps}, 30 \mathrm{~Hz} \mathrm{©} 60 \%$ |
|  | SCR04COA210704009B | 828 | $1000 \mathrm{~V}, 6.2 \mathrm{mps}, 30 \mathrm{~Hz}$ @ $60 \%$ |
| 7 | SCR04COA210704010A | 1036 | $1000 \mathrm{~V}, 8 \mathrm{mpss}, 60 \mathrm{~Hz}$ @ 40\% |
|  | SCR04COA210704010B | 1080 | $1000 \mathrm{~V}, 8 \mathrm{mps}, 60 \mathrm{~Hz}$ @ $40 \%$ |
| 8 | SCR04COA220704011A | 994 | $1000 \mathrm{~V}, 7 \mathrm{mmps}, 60 \mathrm{~Hz}$ @ 40\% |
|  | SCR04COA220704011B | 994 | $1000 \mathrm{~V}, 5.5 \mathrm{mps}, 30 \mathrm{~Hz}$ (8) 55\% |
| 1 | SCR04COA141004001A | 1161 | $1000 \mathrm{~V}, 5.5 \mathrm{amps}, 30 \mathrm{~Hz}$ @ 60\% |
|  | SCR04COA141004001B | 1345 | $1000 \mathrm{~V}, 5.5 \mathrm{amps}, 30 \mathrm{~Hz}$ @ 60\% |
| 2 | SCR04COA141004003A | 709 | $1000 \mathrm{~V}, 5.5 \mathrm{amps}, 30 \mathrm{~Hz}$ @ $80 \%$ |
|  | SCR04COA141004003B | 709 | $1000 \mathrm{~V}, 5.5 \mathrm{mps}, 30 \mathrm{~Hz}$ @ $80 \%$ |
| 3 | SCR04COA141004004A | 942 | $1000 \mathrm{~V}, 5.7 \mathrm{mpss}, 30 \mathrm{~Hz}$ @ 60\% |
|  | SCR04COA141004004B | 1120 | $1000 \mathrm{~V}, 5.5 \mathrm{amps}, 30 \mathrm{~Hz}$ @ $60 \%$ |
| 4 | SCR04COA131004007A | 610 | $1000 \mathrm{~V}, 5.5 \mathrm{amps}, 30 \mathrm{~Hz}$ @ 60\% |
|  | SCR04COA131004007B | 755 | 1000V, $6 \mathrm{amps}, 30 \mathrm{~Hz}$ @ 60\% |
| 5 | SCR04COA131004008A | 894 | $1000 \mathrm{~V}, 5.5 \mathrm{mps}, 30 \mathrm{~Hz}$ @ $50 \%$ |
|  | SCR04COA131004008B | 938 | $1000 \mathrm{~V}, 5.5 \mathrm{mpss}, 30 \mathrm{~Hz}$ @ $60 \%$ |
| 6 | SCR04COA131004009A | 863 | $1000 \mathrm{~V}, 5 \mathrm{mps}, 30 \mathrm{~Hz}$ @ 60\% |
|  | SCR04COA131004009B | 838 | $1000 \mathrm{~V}, 5.5 \mathrm{amps}, 30 \mathrm{~Hz}$ @ $60 \%$ |
| 7 | SCR04COA131004010A | 1274 | $1000 \mathrm{~V}, 5.5 \mathrm{mps}, 30 \mathrm{~Hz}$ © 60\% |
|  | SCR04COA131004010B | 1094 | $1000 \mathrm{~V}, 5 \mathrm{mpss}, 30 \mathrm{~Hz}$ @ 60\% |
| 8 | SCR04COA141004011A | 1109 | $1000 \mathrm{~V}, 5.5 \mathrm{amps}, 30 \mathrm{~Hz}$ @ $80 \%$ |
|  | SCR04COA141004011B | 1431 | $1000 \mathrm{~V}, 5.5 \mathrm{mpss}, 30 \mathrm{~Hz}$ @ $60 \%$ |

Appendix 9. Species richness per site for DFO 2004 sampling, St. Clair River.

| Site | June | July | October |
| :---: | ---: | ---: | ---: |
| $\mathbf{1}$ | 17 | 15 | 14 |
| 2 | 3 | 7 | 11 |
| 3 | 15 | 23 | 10 |
| 4 | 23 | 19 | 17 |
| 5 | 7 | 12 | 13 |
| 6 | 3 | 8 | 11 |
| 7 | 6 | 16 | 18 |
| 8 | 9 | 3 | 10 |

Appendix 10. Scientific and cornmon names of species collected in the St. Clair River based on Neison et al. (2004).

| Scientific Name | Common Name |
| :--- | :--- |
| Alosa pseudoharengus | alewife |
| Ambloplites rupestris | rock bass |
| Amia calva | bowfin |
| Aplodinotus grunniens | freshwater drum |
| Carassius auratus | quilfish |
| Carpiodes cyprinus | white sucker |
| Catostomus commersonii | mottled sculpin |
| Cottus bairdii | spottin shiner |
| Cyprinella spiloptera | common carp |
| Cyprinus carpio | gizzard shad |
| Dorosoma cepedianum | northern pike |
| Esox lucius | rainbow darter |
| Etheostoma caeruleum | threespine stickleback |
| Gasterosteus aculeatus | northern hog sucker |
| Hypentelium nigricans | silver lamprey |
| lchthyomyzon unicuspis | brook silverside |
| Labidesthes sicculus | American brook lamprey |
| Lampetra appendix | longnose gar |
| Lepisosteus osseus | green sunfish |
| Lepomis cyanellus | pumpkinseed |
| Lepomis gibbosus | bluegill |
| Lepomis macrochirus | striped shiner |
| Luxilus chrysocephalus | common shiner |
| Luxilus comutus | redfin shiner |
| Lythrurus umbratilis | smallmouth bass |
| Micropterus dolomieu | largemouth bass |
| Micropterus salmoides | spotted sucker |
| Minytrema melanops | white perch |
| Morone americana | white bass |
| Morone chrysops | silver redhorse |
| Moxostoma anisurum | golden redhorse |
| Moxostoma erthrurum | shorthead redhorse |
| Moxostoma macrolepidotum | round goby |
| Neogobius melanostomus | hornyhead chub |
| Nocomis biguttatus | golden shiner |
| Notemigonus crysoleucas | emerald shiner |
| Notropis atherinoides | spottail shiner |
| Notropis hudsonius | mimic shiner |
| Notropis volucellus | rainbow trout |
| Oncorhynchus mykiss | chinook salmon |
| Oncorhynchus tshawytscha | rainbow smelt |
| Osmerus mordax | yellow perch |
| Perca flavescens | logperch |
| Percina caprodes | bluntnose minnow |
| Pimephales notatus | brown trout |
| Pimephales promelas | Pomoxis nigromaculatus |
| Salmo trutta | Sander vitreus | | Semotilus atromaculatus |
| :--- |

Appendix 11. IBI for eight sites in the St. Clair River sampled in June, July and October 2004. IBI calculated using Hamilton's (1987) adaptation of the index as proposed by Karr (1981).

| June | Site |  |  |  |  |  |  |  | All Sites |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |  |
| \# of Sp. at Each Site as \% of Total Sp. \# | 47.22 | 8.33 | 41.67 | 63.89 | 19.44 | 8.33 | 16.67 | 25.00 | 100.00 |
| Score | 3 | 1 | 3 | 5 | 1 | 1 | 1 | 3 | 5 |
| Number of Percid Sp. | 2 | 1 | 3 | 2 | 2 | 1 | 1 | 1 | 4 |
| Score | 3 | 1 | 5 | 3 | 3 | 1 | 1 | 1 | 5 |
| \# of Naturally-Spawned Salmonid/Coregonid Spp. at Each Site | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 4 |
| Score | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 5 |
| \% Individuals Generalists | 12.08 | 80.00 | 25.71 | 25.00 | 22.72 | 6.90 | 30.00 | 0 | 20.48 |
| Score | 5 | 1 | 3 | 3 | 3 | 5 | 3 | 0 | 3 |
| \% Individuals Specialists | 47.65 | 0 | 28.57 | 4.41 | 13.63 | 41.38 | 10 | 37.84 | 23.04 |
| Score | 5 | 0 | 3 | 1 | 1 | 5 | 1 | 3 | 3 |
| \% Individuals Top Carnivores | 40.27 | 20.00 | 45.71 | 70.59 | 63.63 | 51.72 | 60.00 | 62.16 | 56.47 |
| Score | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 |
| (Site CPUE/mean CPUE) $\times 100$ | 140.72 | 7.82 | 149.82 | 401.82 | 26.80 | 31.05 | 11.31 | 30.84 | 100 |
| Score | 5 | 1 | 5 | 5 | 1 | 1 | 1 | 1 | 3 |
| \% Individuals - Invading Species | 12.75 | 60.00 | 5.71 | 10.66 | 9.09 | 0 | 10.00 | 37.84 | 11.44 |
| Score | 1 | 1 | 1 | 1 | 1 | 5 | 1 | 1 | 1 |
| Index of Biotic Integrity (IBI) | 30 | 10 | 25 | 23 | 15 | 23 | 13 | 19 | 30 |
| Rating | Fair to Good | Very Poor | Fair | Poor | Very Poor | Poor | Very Poor | Poor | Fair to Good |

Appendix 11. Continued.

| July | Site |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | $\begin{gathered} \text { All } \\ \text { Sites } \end{gathered}$ |
| \# of Sp. at Each Site as \% of Total Sp. \# | 45.5 | 21.2 | 69.7 | 57.6 | 36.4 | 24.2 | 48.5 | 9.09 | 100 |
| Score | 3 | 1 | 5 | 5 | 3 | 1 | 3 | 1 | 5 |
| Number of Percid Sp. | 2 | 1 | 3 | 3 | 2 | 1 | 2 | 0 | 3 |
| Score | 3 | 1 | 5 | 5 | 3 | 1 | 3 | 0 | 5 |
| \# of Naturally-Spawned Salmonid/Coregonid Spp. at Each Site | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| Score | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| \% Individuals Generalists | 53.30 | 56.25 | 45.21 | 24.53 | 33.87 | 27.45 | 29.27 | 0.00 | 39.32 |
| Score | 1 | 1 | 1 | 3 | 3 | 3 | 3 | 0 | 3 |
| \% Individuals Specialists | 34.29 | 9.38 | 35.27 | 52.83 | 19.35 | 23.53 | 20.73 | 80.00 | 33.38 |
| Score | 3 | 1 | 3 | 5 | 1 | 3 | 3 | 5 | 3 |
| \% Individuals Top Carnivores | 12.38 | 34.37 | 19.52 | 22.64 | 46.77 | 49.01 | 50.00 | 20.00 | 27.30 |
| Score | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 |
| (Site CPUE/mean CPUE)x $100$ | 95.54 | 45.86 | 293.31 | 154.14 | 69.74 | 58.92 | 73.89 | 9.59 | 100 |
| Score | 3 | 1 | 5 | 5 | 1 | 1 | 1 | 1 | 3 |
| \% Individuals - Invading Species | 4.76 | 9.37 | 5.14 | 5.66 | 0 | 7.84 | 0 | 20.00 | 4.73 |
| Score | 3 | 1 | 1 | 1 | 5 | 1 | 5 | 1 | 3 |
| Index of Biotic Integrity (IBI) | 22 | 11 | 25 | 29 | 21 | 15 | 23 | 14 | 28 |
| Rating | Poor | Very Poor | Fair | Fair | Poor | Very poor | Poor | Poor | Fair |

Appendix 11. Continued.

| October | Site |  |  |  |  |  |  |  | All Sites |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |  |
| \# of Sp. at Each Site as \% of Total Sp. \# | 42.42 | 33.33 | 30.3 | 51.51 | 39.39 | 33.33 | 54.54 | 30.3 | 100 |
| Score | 3 | 3 | 3 | 5 | 3 | 3 | 5 | 3 | 5 |
| Number of Percid Sp. | 2 | 2 | 2 | 1 | 2 | 1 | 1 | 1 | 3 |
| Score | 3 | 3 | 3 | 1 | 3 | 1 | 1 | 1 | 5 |
| \# of Naturally-Spawned Salmonid/Coregonid Spp. at Each Site | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 3 |
| Score | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 5 |
| \% Individuals Generalists | 23.87 | 44.62 | 47.62 | 17.47 | 55.38 | 58.16 | 33.33 | 11.63 | 32.94 |
| Score | 3 | 1 | 1 | 5 | 1 | 1 | 3 | 5 | 3 |
| \% Individuals Specialists | 71.94 | 44.62 | 44.76 | 45.78 | 33.85 | 33.67 | 28.57 | 81.40 | 50.29 |
| Score | 5 | 5 | 5 | 5 | 3 | 3 | 3 | 5 | 5 |
| \% Individuals Top Carnivores | 4.19 | 10.77 | 7.62 | 36.75 | 10.77 | 8.16 | 38.10 | 6.98 | 16.76 |
| Score | 3 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 |
| (Site CPUE/mean CPUE)x $100$ | 189.28 | 70.15 | 77.8 | 186.22 | 54.34 | 88.23 | 108.67 | 26.02 | 100 |
| Score | 5 | 1 | 1 | 5 | 1 | 3 | 3 | 1 | 3 |
| \% Individuals - Invading Species | 80.00 | 63.08 | 62.86 | 7.83 | 43.08 | 44.9 | 6.54 | 4.65 | 44.41 |
| Score | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 3 | 1 |
| Index of Biotic Integrity (IBI) | 26 | 19 | 19 | 27 | 17 | 17 | 21 | 25 | 32 |
| Rating | Fair | Poor | Poor | Fair | Very Poor to Poor | Very Poor to Poor | Poor | Fair | Good |

Appendix 12. New IBI scores for 2004 data calculated using a modified IBI scheme and an updated trophic guild classification system based on Coker et al. (2001).

| June | Site |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | All Sites |
| \# of Sp. at Each Site as \% of Total Sp. \# | 47.22 | 8.33 | 41.67 | 63.89 | 19.44 | 8.33 | 16.67 | 25.00 | 100.00 |
| Score | 3 | 1 | 3 | 5 | 1 | 1 | 1 | 3 | 5 |
| Number of Percid Sp. | 2 | 1 | 3 | 2 | 2 | 1 | 1 | 1 | 4 |
| Score | 3 | 1 | 5 | 3 | 3 | 1 | 1 | 1 | 5 |
| \# of Native Naturally- <br> Spawned <br> Salmonid/Coregonid Spp. at Each Site | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| Score | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| \% Individuals Generalists | 58.39 | 80.00 | 53.57 | 29.04 | 36.36 | 48.27 | 40.00 | 21.62 | 42.01 |
| Score | 1 | 1 | 1 | 3 | 3 | 1 | 3 | 3 | 1 |
| \% Individuals Specialists | 0 | 0 | 0 | 0.37 | 0 | 0 | 0 | 0 | 1.05 |
| Score | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
| \% Individuals Top Carnivores | 41.61 | 20.00 | 46.43 | 70.59 | 63.64 | 51.72 | 60.00 | 62.16 | 56.93 |
| Score | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 |
| (Site CPUE/mean CPUE)x 100 | 140.72 | 7.82 | 149.82 | 401.82 | 26.80 | 31.05 | 11.31 | 30.84 | 100 |
| Score | 5 | 1 | 5 | 5 | 1 | 1 | 1 | 1 | 3 |
| \% Individuals - Invading <br> Species/Hybrids | 14.09 | 60.00 | 5.71 | 10.66 | 9.09 | 0 | 10.00 | 37.84 | 11.75 |
| Score | 1 | 1 | 1 | 1 | 1 | 5 | 1 | 1 | 1 |
| Index of Biotic Integrity (IBI) | 18 | 10 | 20 | 23 | 14 | 14 | 12 | 15 | 22 |
| Rating | Poor | Very Poor | Poor | Poor | Very <br> Poor | Very Poor | Very <br> Poor | Very Poor to Poor | Poor |

Appendix 12. Continued.

| July | Site |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | $\begin{gathered} \text { All } \\ \text { Sites } \end{gathered}$ |
| \# of Sp. at Each Site as \% of Total Sp. \# | 45.5 | 21.2 | 69.7 | 57.6 | 36.4 | 24.2 | 48.5 | 9.09 | 100 |
| Score | 3 | 1 | 5 | 5 | 3 | 1 | 3 | 1 | 5 |
| Number of Percid Sp. | 2 | 1 | 3 | 3 | 2 | 1 | 2 | 0 | 3 |
| Score | 3 | 1 | 5 | 5 | 3 | 1 | 3 | 0 | 5 |
| \# of Native Naturally- <br> Spawned <br> Salmonid/Coregonid Spp. at <br> Each Site | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Score | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| \% Individuals Generalists | 85.71 | 65.63 | 79.45 | 75.47 | 51.61 | 49.02 | 50.00 | 80.00 | 71.46 |
| Score | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| \% Individuals Specialists | 1.9 | 0 | 0.68 | 0.94 | 0 | 0 | 0 | 0.00 | 0.67 |
| Score | 1 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 1 |
| \% Individuals Top Carnivores | 12.38 | 34.38 | 19.86 | 23.58 | 48.39 | 50.98 | 50.00 | 20.00 | 27.84 |
| Score | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 |
| (Site CPUE/mean CPUE)x 100 | 95.54 | 45.86 | 293.31 | 154.14 | 69.74 | 58.92 | 73.89 | 9.59 | 100 |
| Score | 3 | 1 | 5 | 5 | 1 | 1 | 1 | 1 | 3 |
| \% Individuals - Invading Species | 4.76 | 9.37 | 5.14 | 5.66 | 0 | 7.84 | 0 | 30.00 | 4.86 |
| Score | 3 | 1 | 1 | 1 | 5 | 1 | 5 | 1 | 3 |
| Index of Biotic Integrity (IBI) | 19 | 10 | 23 | 23 | 18 | 10 | 18 | 9 | 23 |
| Rating | Poor | Very Poor | Poor | Poor | Poor | Very <br> Poor | Poor | Very Poor | Poor |

Appendix 12. Continued.

| October | Site |  |  |  |  |  |  |  | All Sites |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 |  |
| \# of Sp. at Each Site as \% of Total Sp. \# | 42.42 | 33.33 | 30.3 | 51.51 | 39.39 | 33.33 | 54.54 | 30.3 | 100 |
| Score | 3 | 3 | 3 | 5 | 3 | 3 | 5 | 3 | 5 |
| Number of Percid Sp. | 2 | 2 | 2 | 1 | 2 | 1 | 1 | 1 | 3 |
| Score | 3 | 3 | 3 | 1 | 3 | 1 | 1 | 1 | 5 |
| \# of Native NaturallySpawned Salmonid/Coregonid Spp. at Each Site | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Score | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| \% Individuals Generalists | 38.71 | 49.23 | 60 | 50 | 81.54 | 87.75 | 59.52 | 93.02 | 56.57 |
| Score | 3 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| \% Individuals Specialists | 57.1 | 40 | 32.38 | 12.65 | 7.69 | 4.08 | 2.38 | 0.00 | 26.57 |
| Score | 5 | 3 | 3 | 1 | 1 | 1 | 1 | 0 | 3 |
| \% Individuals Top Carnivores | 4.19 | 10.77 | 7.62 | 37.35 | 10.77 | 8.16 | 38.09 | 6.98 | 16.86 |
| Score | 3 | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 5 |
| (Site CPUE/mean CPUE)x $100$ | 189.28 | 70.15 | 77.8 | 186.22 | 54.34 | 88.23 | 108.67 | 26.02 | 100 |
| Score | 5 | 1 | 1 | 5 | 1 | 3 | 3 | 1 | 3 |
| \% Individuals - Invading Species | 80.00 | 63.08 | 62.86 | 7.83 | 43.08 | 44.9 | 6.54 | 4.65 | 44.41 |
| Score | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 3 | 1 |
| Index of Biotic Integrity (IBI) | 23 | 17 | 17 | 19 | 15 | 15 | 17 | 14 | 23 |
| Rating | Poor | Very Poor to Poor | Very Poor to Poor | Poor | Very Poor | Very Poor | Very Poor to Poor | Very Poor | Poor |

Appendix 13. New IBI scores for August 1994 data based on a modified IBI scheme, and an updated trophic guild classification system based on Coker et al. (2001).

|  | Site |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | All Sites |
| \# of Sp. at Each Site as \% of Total Sp. \# | 41 | 28.2 | 64.1 | 59 | 56.4 | 53.8 | 38.5 | 30.80 | 100.00 |
| Score | 3 | 3 | 5 | 5 | 5 | 5 | 3 | 3 | 5 |
| Number of Percid Sp. | 2 | 2 | 3 | 2 | 3 | 3 | 3 | 1 | 3 |
| Score <br> \# of Native Naturally- <br> Spawned <br> Salmonid/Coregonid | 3 | ${ }^{3}$ | 5 | ${ }^{3}$ | 5 | 5 | 5 | 1 | 5 |
| Spp. at Each Site | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Score | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| \% Individuals Generalists | 5.86 | 7.57 | 56.21 | 58.58 | 34.02 | 51.79 | 36.12 | 75.41 | 23.55 |
| Score | 5 | 5 | 1 | 1 | 3 | 1 | 3 | 1 | 3 |
| \% Individuals Specialists | 93.2 | 88.38 | 25.94 | 7.07 | 55.12 | 30.28 | 59.91 | 13.11 | 69.09 |
| Score | 5 | 5 | 3 | 1 | 5 | 3 | 5 | 1 | 5 |
| \% Individuals Top Carnivores | 0.93 | 4.04 | 17.84 | 34.34 | 10.86 | 17.93 | 3.96 | 11.47 | 7.36 |
| Score | 1 | 3 | 5 | 5 | 5 | 5 | 3 | 5 | 5 |
| (Site CPUE/mean CPUE) 100 | 414.4 | 42.6 | 79.6 | 42.6 | 104.90 | 54 | 48.8 | 13.1 | 100 |
| Score | 5 | 1 | 1 | 1 | 3 | 1 | 1 | 1 | 3 |
| \% Individuals Invading Species | 93.46 | 91.91 | 38.11 | 18.18 | 56.76 | 32.67 | 62.55 | 36.06 | 72.12 |
| Score Index of Biotic Integrity (IBI) | 1 23 | 1 21 | 1 21 | 1 17 | 1 27 | 1 21 | 1 21 | 1 13 | 1 27 |
| Rating | Poor | Poor | Poor | Very Poor to Poor | Fair | Poor | Poor | Very <br> Poor | Fair |

Appendix 14. Original IBI scores for August 1994, from MacLennan and Hyatt (1996).

| Site |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | All Sites |
| \# of Sp. at Each Site as \% of Total Sp. \# | 41 | 28.2 | 64.1 | 59 | 56.4 | 53.8 | 38.5 | 30.80 | 100.00 |
| Score | 3 | 3 | 5 | 5 | 5 | 5 | 3 | 3 | 5 |
| Number of Percid Sp. | 2 | 2 | 3 | 2 | 3 | 3 | 3 | 1 | 3 |
| Score | 3 | 3 | 5 | 3 | 5 | 5 | 5 | 1 | 5 |
| \# of NaturallySpawned Salmonid/Coregonid Spp. at Each Site | 0 | 1 | 0 | 2 | 0 | 0 | 0 | 1 | 3 |
| Score | 0 | 1 | 0 | 3 | 0 | 0 | 0 | 1 | 5 |
| \% Individuals Generalists | 4.8 | 5.6 | 20.5 | 14.1 | 8 | 23.1 | 9.3 | 65.6 | 9.8 |
| Score | 5 | 5 | 3 | 5 | 5 | 3 | 5 | 1 | 5 |
| \% Individuals Specialists | 94.2 | 90.4 | 64.6 | 52 | 81.6 | 60.2 | 86.8 | 24.6 | 83.3 |
| Score | 5 | 5 | 5 | 5 | 5 | 5 | 5 | 3 | 5 |
| \% Individuals Top Carnivores | 0.9 | 4 | 14.9 | 33.8 | 10.5 | 16.7 | 4 | 9.8 | 6.9 |
| Score | 1 | 3 | 5 | 5 | 5 | 5 | 3 | 5 | 5 |
| (Site CPUE/mean CPUE) $\times 100$ | 414.4 | 42.6 | 79.6 | 42.6 | 104.90 | 54 | 48.8 | 13.1 | 100 |
| Score | 5 | 1 | 1 | 1 | 3 | 1 | 1 | 1 | 3 |
| \% Individuals Invading Species | 93.4 | 91.9 | 38.1 | 18.2 | 56.8 | 32.7 | 62.6 | 32.8 | 72 |
| Score | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 | 1 |
| Index of Biotic Integrity (IBI) | 23 | 22 | 25 | 28 | 29 | 25 | 23 | 16 | 34 |
| Rating | Poor | Poor | Fair | Fair | Fair | Fair | Poor | Very Poor to Poor | Good |

Appendix 15. Summary of DFO 2004 habitat data, St. Clair River.

| \# <br> Map | Field Number | Air Temp (C) | Water Temp (C) | Secchi Depth (m) | Conductivity ( $\mu \mathrm{S}$ ) | Stream Width (m) | Max Stream Depth (m) | Distance From Shore (m) | Meix <br> Samp <br> Depth <br> (m) | Flow Rate | Aquatic Veg Typei | (\%) | Aquatic Veg Type2 | (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | SCR04COA170604001A | 23.5 | 14.7 | N/A | 227 | 500 | 10 | 2 | 2.5 | Medium | None | 100 |  |  |
|  | SCR04COA170604001B | 22.5 | 15.9 | N/A | 227 | 500 | 9 | 2 | 2.5 | Medium | None | 100 |  |  |
| 2 | SCR04COA170604003A | 18.7 | 15.5 | N/A | 212 | 500 | 9 | 50 | 2 | Medium | Submergent | 30 | None | 70 |
|  | SCR04COA170604003B | 18.7 | 15.2 | N/A | 223 | 500 | 8 | 50 | 2 | Medium | Submergent | 50 | None | 50 |
| 3 | SCR04COA170604004A | 24.1 | 17.5 | N/A | 233 | 500 | 9 | 10 | 2 | Slow/Medium | Submergent | 50 | None | 50 |
|  | SCR04COA170604004B | 21.9 | 16.8 | N/A | 233 | 500 | 9 | 10 |  | Medium | Submergent | 50 | None | 50 |
| 4 | SCR04COA160604007A | 18 | 18 | N/A | 210 | 500 | 7 | 10 | 2 | Medium | Myriophylum | 40 | None | 60 |
|  | SCR04COA160604007B | 22.1 | 18.2 | N/A | 217 | 500 | 7 | 10 | 2 | Medium | Myriophylum | 50 | None | 50 |
| 5 | SCR04COA160604008A | 26.1 | 16.9 | N/A | 220 | 500 | 8 | 30 | 2 | Medium | Submergent | 20 | None | 80 |
|  | SCR04COA160604008B | 26.1 | 16.9 | N/A | 220 | 500 | 8 | 30 | 2 | Medium | Emergent | 50 | None | 50 |
| 6 | SCR04COA160604009A | 23.8 | 16.8 | N/A | 222 | 500 | 8 | 50 | 2 | Medium | None | 100 |  |  |
|  | SCR04COA160604009B | 25.3 | 17 | N/A | 204 | 500 | 8 | 50 | 2 | Medium | None | 100 |  |  |
| 7 | SCR04COA160604010A | 21.5 | 16 | N/A | 212 | 500 | 8 | 10 | 2 | Medium | None | 100 |  |  |
|  | SCR04COA160604010B | 21.8 | 16 | N/A | 212 | 500 | 8 | 30 | 3 | N/A | N/A | N/A |  |  |
| 8 | SCR04COA170604011A | N/A | N/A | N/A | N/A | 400 | 10 | 2 | 2 | Medium | None | 100 |  |  |
|  | SCR04COA170604011B | 21.8 | 15.2 | N/A | 220 | 350 | 9 | 10 | 2 | N/A | None | 100 |  |  |
| 1 | SCR04COA220704001A | 26.1 | 21.5 | >3 | 224 | 700 | 12 | 2 | 2 | Medium | None | 100 |  |  |
|  | SCR04COA220704001B | 23.9 | 20.3 | >3 | 221 | 700 | 12 | 2 | 2 | Fast | None | 100 |  |  |
| 2 | SCR04COA220704003A | 21.5 | 19.8 | >3 | 223 | 700 | 10 | 80 | 2 | Fast | Submergent | 40 | None | 60 |
|  | SCR04COA220704003B | 22.4 | 19.8 | $>3$ | 225 | 700 | 10 | 70 | 2 | Medium | Submergent | 40 | None | 60 |
| 3 | SCR04COA220704004A | 23.7 | 20.5 | >3 | 223 | 800 | 10 | 20 | 2 | Slow | Submergent | 50 | None | 50 |
|  | SCR04COA220704004B | 22 | 21.9 | >3 | 230 | 800 | 10 | 10 | 1.33 | Slow | Submergent | 40 | None | 60 |
| 4 | SCR04COA210704007A | 25.8 | 29.2 | N/A | 236 | 600 | 15 | 20 | 2 | Medium | Submergent | 30 | None | 70 |
|  | SCR04COA210704007B | 25.5 | 25 | N/A | 232 | 600 | 15 | 20 | 2 | Medium | Emergent | 40 | None | 60 |
| 5 | SCR04COA210704008A | 25.2 | 21.9 | N/A | 224 | 750 | 10 | 100 | 2 | N/A | None | 80 | Submergent | 20 |
|  | SCR04COA210704008B | 27.6 | 22.6 | N/A | 214 | 600 | 12 | 100 | 2 | N/A | Submergent | 30 | None | 70 |
| 6 | SCR04COA210704009A | 26.2 | 22.2 | >3 | 220 | 600 | 12 | 75 | 2 | Slow | Submergent | 30 | None | 70 |
|  | SCR04COA210704009B | 26.4 | 21.7 | $>3$ | 223 | 600 | 12 | 75 | 2 | Slow | None | 100 |  |  |
| 7 | SCR04COA210704010A | 29.8 | N/A | $>3$ | N/A | 250 | 15 | 10 | 2 | Slow | Submergent | 60 | None | 40 |
|  | SCR04COA210704010B | 26.9 | 22.4 | $>3$ | 212 | 600 | 16 | 150 | 2 | Medium | Submergent | 30 | None | 70 |

Appendix 15. Continued.

| \# <br> On <br> Map | Field Number | Alr Temp (C) | Water Temp (C) | Secchi Depth (m) | Conductivity | Stream Width (m) | Max Stream Depth (m) | Distance From Shore (m) | Max <br> Samp <br> Depth <br> (m) | Flow Rate | Aquatic Veg Type1 | (\%) | Aquatic Veg Type2 | (\%) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 8 | SCR04COA220704011A | 27.1 | 21.2 | $>3$ | 214 | 600 | 12 | 2 | 2 | Medium | None | 100 |  |  |
|  | SCR04COA220704011B | 26.8 | 21.8 | >3 | 218 | 700 | 12 | 20 | 2 | Medium | Submergent | 30 | None | 70 |
| 1 | SCR04COA141004001A | N/A | 14.6 | N/A | 207 | 400 | 15 | 2 | 2 | Medium | Submergent | 5 | None | 95 |
|  | SCR04COA141004001B | 12.3 | 14.2 | N/A | 207 | N/A | N/A | N/A | N/A | N/A | Submergent | 5 | None | 95 |
| 2 | SCR04COA141004003A | N/A | 14.1 | $>2$ | 203 | 400 | 15 | 150 | 2 | Medium | Submergent | 20 | None | 80 |
|  | SCR04COA141004003B | 11.9 | 14 | $>2$ | 203 | 400 | 12 | 120 | 2 | Medium | Submergent | 40 | None | 60 |
| 3 | SCR04COA141004004A | 11.9 | 14.5 | $>2$ | 210 | 450 | 15 | 30 | 2 | Medium | Submergent | 30 | None | 70 |
|  | SCR04COA141004004B | N/A | 16.6 | $>2$ | 218 | 400 | 15 | 30 | 2 | Slow | Submergent | 10 | None | 90 |
| 4 | SCR04COA131004007A | 17 | 15.5 | 2.2 | 201 | 400 | 15 | 30 | 2.5 | Medium | Submergent | 70 | None | 30 |
|  | SCR04COA131004007B | 11 | 17.3 | 2.3 | 210 | 400 | 15 | 20 | 2 | Medium | Submergent | 70 | None | 30 |
| 5 | SCR04COA131004008A | N/A | 15.9 | 2.3 | 157 | 400 | 15 | 100 | 2 | Medium | Submergent | 30 | None | 70 |
|  | SCR04COA131004008B | 18 | 16 | $>1.2$ | 198 | 400 | 15 | 50 | 2 | Slow | Submergent | 20 | None | 80 |
| 6 | SCR04COA131004009A | 20 | 16.2 | $>2$ | 195 | 400 | 15 | 60 | 2 | Medium | Submergent | 30 | None | 70 |
|  | SCR04COA131004009B | 20 | 18 | >2.3 | 190 | 400 | 15 | 30 | 2 | Slow | Submergent | 30 | None | 70 |
| 7 | SCR04COA131004010A | 19 | 15.4 | >2 | 198 | 400 | 15 | 5 | 2 | Slow | Submergent | 100 |  |  |
|  | SCR04COA131004010B | 21 | 15.9 | 2.4 | 197 | 400 | 15 | 150 | 2 | N/A | Submergent | 50 | None | 50 |
| 8 | SCR04COA141004011A | N/A | N/A | $>2$ | N/A | 300 | 15 | 2 | 2 | Medium | None | 100 |  |  |
|  | SCR04COA141004011B | 12.5 | 13.9 | $>2$ | 204 | 400 | 15 | 2 | 2 | Medium | None | 100 |  |  |

Appendix 16. Summary of species caught by site during OMNR (August 1994) and DFO (July 2004) sampling, St. Clair River.

|  | OMINR 1994 |  |  |  |  |  |  |  |  | DFO 2004 |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Site | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | Total | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | Total |
| Alosa pseudoharengus | 1796 | 175 | 96 | 14 | 268 | 70 | 136 | 7 | 2562 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
| Ambloplites rupestris | 14 | 1 | 23 | 5 | 6 | 7 | 3 | 4 | 63 | 0 | 0 | 1 | 0 | 0 | 0 | 2 | 0 | 3 |
| Amia calva | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 1 | 3 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 |
| Aplodinotus grunniens | 4 | 1 | 6 | 0 | 13 | 1 | 0 | 0 | 25 | 1 | 0 | 3 | 0 | 4 | 0 | 4 | 0 | 12 |
| Carassius auratus | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Carpiodos cyprinus | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
| Catostomus commersonii | 9 | 3 | 2 | 2 | 14 | 8 | 2 | 0 | 40 | 0 | 5 | 5 | 7 | 1 | 0 | 4 | 0 | 22 |
| Cyprinella spiloptera | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
| Cyprinus carpio | 3 | 3 | 21 | 13 | 4 | 9 | 6 | 1 | 60 | 1 | 0 | 12 | 4 | 0 | 0 | 0 | 0 | 17 |
| Dorosoma cepedianum | 0 | 0 | 0 | 0 | 1 | 5 | 0 | 1 | 7 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 2 |
| Esox lucius | 1 | 0 | 0 | 3 | 2 | 0 | 1 | 0 | 7 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gasterosteus aculeatus | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 |
| Hypentelium nigricans | 2 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 1 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 3 |
| Ichthyomyzon unicuspis | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 2 |
| Labidesthes sicculus | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepisosteus osseus | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 3 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
| Lepomis gibbosus | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis macrochirus | 0 | 0 | 7 | 2 | 0 | 0 | 0 | 0 | 9 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Luxilus chrysocephalus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 0 | 1 | 0 | 1 | 0 | 6 |
| Luxilus comutus | 0 | 0 | 3 | 2 | 2 | 32 | 1 | 1 | 41 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lythrurus umbratilis | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
| Micropterus dolomieu | 4 | 0 | 15 | 0 | 0 | 0 | 2 | 0 | 21 | 1 | 0 | 0 | 1 | 5 | 0 | 1 | 0 | 8 |
| Micropterus salmoides | 0 | 0 | 1 | 37 | 5 | 6 | 0 | 0 | 49 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Minytrema melanops | 0 | 0 | 1 | 0 | 0 | 2 | 5 | 0 | 8 | 0 | 0 | 4 | 0 | 0 | 0 | 1 | 0 | 5 |
| Morone americana | 0 | 0 | 11 | 1 | 1 | 2 | 0 | 0 | 15 | 0 | 0 | 1 | 1 | 0 |  | 0 | 0 | 3 |
| Morone chrysops | 0 | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 3 | 3 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 6 |
| Moxostoma anisurum | 4 | 0 | 1 | 3 | 0 | 2 | 0 | 0 | 10 | 0 | 0 | 3 | 2 | 6 | 4 | 2 | 0 | 17 |
| Moxostoma erythrurum | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 1 | 13 | 0 | 1 | 1 | 2 | 0 | 19 |
| Moxostoma macrolepidotum | 71 | 3 | 13 | 2 | 1 | 2 | 4 | 11 | 107 | 48 | 8 | 90 | 9 | 7 | 5 | 9 | 0 | 176 |
| Neogobius melanostomus | 1 | 0 | 10 | 6 | 2 | 1 | 0 | 0 | 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Nocomis biguttatus | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 2 | 1 | 1 | 0 | 2 | 0 | 0 | 2 | 0 | 6 |
| Notemigonus crysoleucas | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 |
| Notropis atherinoides | 3 | 0 | 50 | 3 | 44 | 29 | 54 | 1 | 184 | 32 | 0 | 76 | 48 | 1 | 11 | 6 | 7 | 181 |
| Notropis hudsonius | 1 | 0 | 7 | 38 | 30 | 26 | 3 | 0 | 105 | 0 | 0 | 2 | 0 | 0 | 0 | 5 | 0 | 7 |
| Notropis volucellus | 0 | 0 | 0 | 2 | 18 | 0 | 0 | 0 | 20 | 0 | 0 | 2 | 1 | 0 | 0 | 1 | 0 | 4 |
| Oncorhynchus mykiss | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 3 |
| Oncorhynchus tshawytscha | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Osmerus mordax | 0 | 1 | 3 | 0 | 0 | 0 | 0 | 11 | 15 | 3 | 3 | 1 | 1 | 0 | 3 | 0 | 0 | 11 |
| Perca flavescens | 8 | 1 | 35 | 22 | 39 | 32 | 4 | 0 | 141 | 8 | 11 | 53 | 18 | 24 | 25 | 40 | 0 | 179 |
| Percina caprodes | 0 | 0 | 1 | 8 | 14 | 1 | 1 | 0 | 25 | 2 | 0 | 14 | 1 | 10 | 0 | 1 | 0 | 28 |
| Pimephales notatus | 0 | 3 | 29 | 15 | 5 | 8 | 0 | 0 | 60 | 0 | 0 | 3 | 3 | 0 | 0 | 1 | 0 | 7 |
| Salmo trutta | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sander vitreus | 5 | 4 | 1 | 0 | 5 | 4 | 2 | 5 | 26 | 0 | 0 | 2 | 3 | 0 | 0 | 0 | 0 | 5 |
| Semotilus atromaculatus | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

Appendix 17. Species catch summary at each sub-site, October 2004, St. Clair River.

| Specles List | \%Composition | Total | 18 | 16 | 2a | 2b | 3a | 36 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alosa psoudoharengus | 24.01960784 | 245 | 53 | 123 | 26 | 0 | 25 | 9 |
| Ambloplites rupestris | 0.784313725 | 8 | 0 | 0 | 0 | 0 | 0 | 0 |
| Amia calva | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aplodinotus grunniens | 1.960784314 | 20 | 0 | 1 | 0 | 1 | 1 | 3 |
| Carpiodos cyprinus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Catostomus commersonii | 6.764705882 | 69 | 1 | 0 | 0 | 9 | 1 | 7 |
| Coregonus sp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cottus bairdii | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cyprinella spiloptera | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cyprinus carpio | 1.960784314 | 20 | 0 | 0 | 1 | 1 | 1 | 0 |
| Dorosoma cepodianum | 0.196078431 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| Esox lucius | 0.098039216 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ethoostoma caeruieum | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gasterostous aculeatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hypentellum nigricans | 0.980392157 | 10 | 0 | 3 | 0 | 0 | 0 | 0 |
| Ichthyomyzon unicuspis | 0.196078431 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ictiobus sp. | 0.196078431 | 2 | 0 | 0 | 0 | 0 | 1 | 0 |
| Labidesthes sicculus | 2.156862745 | 22 | 0 | 1 | 0 | 0 | 0 | 0 |
| Lampetra appendix | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepisostous ossous | 0.098039216 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepornis cyanellus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis cyanellus/macrochinus hybrid | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis gibbosus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis macrochirus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Luxilus chrysocephalus | 0.490196078 | 5 | 0 | 0 | 0 | 0 | 0 | 10 |
| Lythrurus umbratilis | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Micropterus dolomieu | 0.490196078 | 5 | 1 | 3 | 0 | 0 | 0 | 10 |
| Micropterus salmoides | 0.098039216 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Minytrema melanops | 1.176470588 | 12 | 0 | 0 | 0 | 0 | 0 | 0 |
| Morone americana | 0.098039216 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Morone chrysops | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10 |
| Moxostoma anisurum | 0.294117647 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| Moxostoma erythrurum | 0.098039216 | 1 | 0 | 0 | 0 | 0 | 0 | 10 |
| Moxostoma macrolepidotum | 1.470588235 | 15 | 0 | 0 | 3 | 0 | 5 | 1 |
| Neogobius melanostomus | 0.098039216 | 1 | 0 | 1 | 0 | 0 | 0 | 10 |
| Nocomis biguttatus | 0.392156863 | 4 | 0 | 0 | 0 | 1 | 0 | 0 |
| Notemigonus crysoleucas | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Notropls atherinoides | 11.8627451 | 121 | 1 | 4 | 0 | 2 | 6 | 5 |
| Notropis hudsonius | 2.843137255 | 29 | 0 | 1 | 0 | 1 | 0 | 1 |
| Notropis volucellus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Oncortynchus mykiss | 0.196078431 | 2 | 1 | 0 | 0 | 0 | 0 | 0 |
| Oncortynchus tshawytscha | 0.098039216 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Osmerus mordax | 17.84313725 | 182 | 44 | 25 | 3 | 10 | 14 | 17 |
| Perca flavescens | 15.39215686 | 157 | 6 | 1 | 1 | 5 | 1 | 7 |
| Percina caprodes | 5.392156863 | 55 | 10 | 29) | 0 | 0 | 1 | 0 |
| Pimephales notatus | 1.960784314 | 20 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pimephales promelas | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pomoxis nigromaculatus | 0.098039216 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmo trita | 0.098039216 | 1 | 0 | 1 | 0 | 0 | 0 | 0 |
| Sander vitrous | 0.098039216 | 1 | 0 | 0 | 1 | 0 | 0 | 0 |
| Semotilus atromaculatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total Catch |  | 1020 | 117 | 193 | 35 | 30 | 56 | 48 |

Appendix 17. Continued.

| Species List | 4 a | 4b | 5a | 5b | 6a | 6b | 7a | 7b | 89 | 8b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alosa pseudoharengus | 0 | 4 | 3 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| Ambloplites rupestris | 0 | 3 | 0 | 0 | 0 | 0 | 2 | 0 | 3 | 0 |
| Amia calva | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aplodinotus grunniens | 0 | 0 | 2 | 0 | 0 | 9 | 3 | 0 | 0 | 0 |
| Carpiodes cyprinus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Catostomus commersonii | 5 | 5 | 4 | 3 | 2 | 0 | 22 | 8 | 0 | 2 |
| Coregonus sp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cottus bairdii | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cyprinella spiloptera | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cyprinus carpio | 1 | 4 | 3 | 0 | 1 | 1 | 6 | 1 | 0 | 0 |
| Dorosoma cepedianum | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 |
| Esox lucius | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Ethoostoma caeruleum | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gasterostous aculeatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hypentelium nigricans | 1 | 0 | 0 | 1 | 0 | 0 | 5 | 0 | 0 | 0 |
| Ichthyomyzon unicuspis | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Ictiobus sp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Labidesthes sicculus | 1 | 15 | 1 | 0 | 3 | 0 | 1 | 0 | 0 | 0 |
| Lampetra appendix | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepisosteus osseus | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis cyanellus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis cyanellus/macrochirus hybrid | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis gibbosus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis macrochirus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Luxilus chrysocephalus | 0 | 3 | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 |
| Lythrurus umbratilis | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Micropterus dolomieu | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Micropterus salmoides | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Minytrema melanops | 6 | 1 | 0 | 0 | 0 | 0 | 3 | 1 | 0 | 1 |
| Morone americana | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Morone chrysops | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Moxostoma anisurum | 0 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Moxostoma enthrurum | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| Moxostoma macrolepidotum | 0 | 0 | 0 | 0 | 2 | 0 | 1 | 1 | 0 | 2 |
| Neogobius melanostomus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Nocomis biguttatus | 0 | 2 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Notemigonus crysoleucas | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Notropis atherinoides | 2 | 17 | 3 | 13 | 13 | 15 | 17 | 6 | 9 | 8 |
| Notropis hudsonius | 0 | 9 | 0 | 0 | 0 | 0 | 12 | 6 | 0 | 0 |
| Notropis volucellus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Oncorhynchus mykiss | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Oncorhynchus tshawytscha | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Osmerus mordax | 0 | 3 | 12 | 9 | 28 | 13 | 0 | 4 | 0 | 0 |
| Perca flavescens | 17 | 44. | 1 | 4 | 6 | 2 | 43 | 19 | 0 | 0 |
| Percina caprodes | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 10 | 4 |
| Pimephales notatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pimephales promelas | 0 | 20 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pomoxis nigromaculatus | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Salmo trutta | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sander vitreus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Semotilus atromaculatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total | 35 | 131 | 31 | 34 | 57 | 41 | 122 | 46 | 22 | 21 |

Appendix 18. Species catch summary at each sub-site, July 2004, St. Clair River.

| Species List | \%Composition | Total | 1 a | 16 | 2a | 2b | 3 a | 3b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alosa pseudoharengus | 0.135135135 | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
| Ambloplites rupestris | 0.405405405 | 3 | 0 | 0 | 0 | 0 | 0 | 1 |
| Amia calva | 0.135135135 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aplodinotus grunnions | 1.621621622 | 12 | 1 | 0 | 0 | 0 | 2 | 1 |
| Carplodes cyprinus | 0.135135135 | 1 | 0 | 0 | 0 | 0 | 1 | 0 |
| Catostomus commersonii | 2.972972973 | 22 | 0 | 0 | 2 | 3 | 2 | 3 |
| Coregonus sp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cottus bairdll | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cyprinella spiloptera | 0.135135135 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cyprinus carpio | 2.297297297 | 17 | 0 | 1 | 0 | 0 | 10 | 2 |
| Dorosoma cepedianum | 0.27027027 | 2 | 0 | 1 | 0 | 0 | 1 | 0 |
| Esox lucius | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Etheostoma caeruleum | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gasterosteus aculeatus | 0.135135135 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hypentelium nigricans | 0.405405405 | 3 | 0 | 1 | 0 | 0 | 0 | 0 |
| Ichthyomyzon unicuspis | 0.27027027 | 2 | 0 | 1 | 0 | 0 | 0 | 0 |
| Ictiobus sp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Labldesthes slcculus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lampetra appendix | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepisostous osseus | 0.135135135 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis cyanollus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis cyanellus/macrochinus hybrid | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis gibbosus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis macrochinus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Luxilus chrysocephalus | 0.810810811 | 6 | 0 | 0 | 3 | 0 | 0 | 1 |
| Lythrurus umbratlis | 0.135135135 | 1 | 0 | 0 | 0 | 0 | 0 | 1 |
| Microptorus dolomieu | 1.081081081 | 8 | 0 | 1 | 0 | 0 | 0 | 0 |
| Micropterus salmoides | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Minytrama melanops | 0.675675676 | 5 | 0 | 0 | 0 | 0 | 1 | 3 |
| Morone americana | 0.405405405 | 3 | 0 | 0 | 0 | 0 | 1 | 0 |
| Morone chrysops | 0.810810811 | 6 | 0 | 3 | 0 | 0 | 0 | 2 |
| Moxostome anisurum | 2.297297297 | 17 | 0 | 0 | 0 | 0 | 2 | 1 |
| Moxostoma erythrurum | 2.567567568 | 19 | 1 | 0 | 0 | 1 | 0 | 13 |
| Moxostoma macrolepidotum | 23.78378378 | 176 | 11 | 37 | 0 | 8 | 43. | 47 |
| Neogobius melanostomus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Nocomis biguttatus | 0.810810811 | 6 | 0 | 1 | 1 | 0 | 0 | 0 |
| Notemigonus crysoleucas | 0.135135135 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Notropis atherincides | 24.45945946 | 181 | 1 | 31 | 0 | 0 | 64. | 12 |
| Notropis hudsonlus | 0.945945946 | 7 | 0 | 0 | 0 | 0 | 0 | 2 |
| Notropis volucellus | 0.540540541 | 4 | 0 | 0 | 0 | 0 | 1 | 1 |
| Oncorhynchus mykiss | 0.405405405 | 3 | 0 | 1 | 0 | 0 | 0 | 0 |
| Oncomynchus tshawytscha | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Osmerus mordax | 1.486486486 | 11 | 0 | 3 | 3 | 0 | 1. | 0 |
| Perca flavescens | 24.18918919 | 179 | 7 | 1 | 10 | 1 | 31 | 22 |
| Percina caprodes | 3.783783784 | 28. | 1 | 1 | 0 | 0 | 2 | 12 |
| Pimephales notatus | 0.945945946 | 7 | 0 | 0 | 0 | 0 | 0 | 3 |
| Pimephales promelas | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pomoxis nigromaculatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmo trutta | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sander vitreus | 0.675675676 | 5 | 0 | 0 | 0 | 0 | 2 | 0 |
| Semotilus atromaculatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total Catch |  | 740 | 22 | 83 | 19 | 13 | 164 | 120 |

## Appendix 18. Continued.

| Species List | 49 | 4b | 5a | 5b | 68 | 6b | 7a | 7b | 80 | 8b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alosa psoudoharengus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ambloplites rupestris | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 |
| Amla calva | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aplodinotus grunniens | 0 | 0 | 1 | 3 | 0 | 0 | 4 | 0 | 0 | 0 |
| Carpiodes cyprinus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Catostomus commersonii | 5 | 2 | 1 | 0 | 0 | 0 | 3 | 1 | 0 | 0 |
| Coregonus sp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cottus bairdii | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cyprinella spiloptera | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cyprinus cappio | 2 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Dorosoma cepedianum | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Esox lucius | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Etheostoma caeruleum | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gasterostous acuieatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| Hypentelium nigricans | 0 | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |
| Ichthyomyzon unicuspis. | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ictiobus sp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Labidesthes sicculus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lampetra appendix | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepisosteus asseus | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis cyanellus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis cyanellus/macrochirus hybrid | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis gibbosus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis macrochirus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Luxilus chrysocephalus | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Lythrurus umbratilis | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Micropterus dolomiou | 0 | 1 | 3 | 2 | 0 | 0 | 0 | 1 | 0 | 0 |
| Micropterus salmoides | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Minytrema melanops | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Morone americana | 0 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Morone chrysops | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Moxostoma anisurum | 0 | 2 | 6 | 0 | 2 | 2 | 1 | 1 | 0 | 0 |
| Moxostoma erythrurum | 0 | 0 | 1 | 0 | 0 | 1 | 1 | 1 | 0 | 0 |
| Moxostoma macrolepidotum | 4 | 5 | 4 | 3 | 1 | 4 | 4 | 5 | 0 | 0 |
| Neogobius melanostomus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Nocomis biguttatus | 0 | 2 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 |
| Notemigonus crysoleucas | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Notropis atherinoides | 0 | 48 | 1 | 0 | 11 | 0 | 4 | 2 | 1 | 6 |
| Notropis hudsonius | 0 | 0 | 0 | 0 | 0 | 0 | 5 | 0 | 0 | 0 |
| Notropis volucellus | 0 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Oncorhynchus mykiss | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 |
| Oncortynchus tshawytscha | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Osmerus mordax | 1 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 |
| Perca flavescens | 6 | 12 | 20 | 4 | 16 | 9 | 38 | 2 | 0 | 0 |
| Percina caprodes | 1 | 0 | 1 | 9 | 0 | 0 | 1 | 0 | 0 | 0 |
| Pimephales notatus | 0 | 3 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Pimephales promelas | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pomoxis nigromaculatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmo trutta | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sander vitreus. | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Semotilus atromaculatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total | 22 | 84 | 41 | 21 | 33 | 18 | 69 | 13 | 4 | 6 |

Appendix 19. Species catch summary at each sub-site, June 2004, St. Clair River.

| Spacies List | \% Composition | Total | 1a | 1b | 2a | 2b | $3{ }^{\text {3 }}$ | 3b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alosa psoudoharengus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ambloplites rupestris | 4.819277108 | 32 | 19 | 9 | 0 | 0 | 0 | 0 |
| Amia calva | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aplodinotus grunnians | 0.753012048 | 5 | 0 | 0 | 0 | 0 | 2 | 2 |
| Carpiodos cyprinus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Catostomus commersonii | 3.313253012 | 22 | 0 | 0 | 0 | 0 | 1 | 3 |
| Coregonus sp. | 0.903614458 | 6 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cottus bairdii | 0.15060241 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cyprinella spiloptera | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cyprinus carpio | 4.518072289 | 30 | 0 | 0 | 0 | 0 | 1 | 0 |
| Dorosoma copodianum | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Esox lucius | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ethoostoma caeruloum | 0.15060241 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gasterostous aculeatus | 0.301204819 | 2 | 2 | 0 | 0 | 0 | 0 | 0 |
| Hypentellum nigricans | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| ichthyomyzon unicuspis | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ictiobus sp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Labidesthes sicculus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lampetra appendix | 0.15060241 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepisostaus ossaus | 0.15060241 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis cyanellus | 0.451807229 | 3 | 2 | 0 | 0 | 0 | 0 | 1 |
| Lepomis cyanellus/macrochirus hytorid | 0.15060241 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| Lepomis gibbosus | 0.15060241 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis macrochinus | 1.656626506 | 11 | 2 | 5 | 0 | 0 | 0 | 0 |
| Luxilus chrysocephalus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lythrurus umbratilis | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Micropterus dolomieu | 4.36746988 | 29 | 8 | 8 | 0 | 0 | 0 | 0 |
| Micropterus salmoides | 0.301204819 | 2 | 0 | 0 | 0 | 0 | 0 | 0 |
| Minytrema melanops | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Morone americana | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Morone chrysops | 0.602409639 | 4 | 0 | 0 | 0 | 0 | 0 | 0 |
| Moxostoma anisurum | 0.451807229 | 3 | 0 | 1 | 0 | 0 | 1 | 0 |
| Moxostoma erythrurum | 1.054216867 | 7 | 0 | 0 | 0 | 0 | 0 | 3 |
| Moxostoma macrolepidotum | 5.271084337 | 35 | 0 | 1 | 0 | 1 | 11 | 8 |
| Neogobius melanostomus | 4.21686747 | 28 | 10 | 6 | 0 | 0 | 4 | 1 |
| Nocomis biguttatus | 1.204819277 | 8 | 2 | 4 | 0 | 0 | 1 | 0 |
| Notemigonus crysoleucas | 0.15060241 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Notropis atherinoides | 6.777108434 | 45 | 23 | 0 | 0 | 0 | 19 | 1 |
| Notropis hudsonius | 3.915662651 | 26 | 0 | 1 | 0 | 0 | 6 | 2 |
| Notropis volucellus | 0.15060241 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Oncormynchus mykiss | 1.506024096 | 10 | 1 | 0 | 0 | 0 | 0 | 0 |
| Oncorhynchus tshawytscha | 0.15060241 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Osmerus mordax | 0.753012048 | 5 | 0 | 0 | 2 | 1 | 2 | 0 |
| Perca flavescens | 48.64457831 | 323 | 11 | 30 | 1 | 0 | 27 | 35 |
| Percina caprodes | 1.054216867 | 7 | 0 | 0 | 0 | 0 | 1 | 6 |
| Pimephales notatus | 0.451807229 | 3 | 1 | 0 | 0 | 0 | 0 | 0 |
| Pimephales promelas | 0.451807229 | 3 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pomoxis nigromaculatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmo trutla | 0.15060241 | 1 | 0 | 1 | 0 | 0 | 0 | 0 |
| Sander vitraus | 0.602409639 | 4 | 0 | 1 | 0 | 0 | 0 | 2 |
| Somotilus atromaculatus | 0.15060241 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total Catch |  | 664 | 82 | 67 | 3 | 2 | 76 | 64 |

Appendix 19. Continued.

| Species List | 49 | 4b | 5a | 5b | 6a | 6b | 7a | 7b | 8 a | 8b |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Alosa psoudoharengus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ambloplites rupestris | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 3 |
| Amia calva | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Aplodinotus grunniens | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Carpiodes cyprinus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Catostomus commersonii | 0 | 15 | 2 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Coregonus sp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 |
| Cottus bairdif | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| Cyprinella spiloptera | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cyprinus carpio | 0 | 28 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Dorosoma cepedianum | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Esox luclus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Etheostoma caeruloum | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| Gasterosteus aculeatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hypentelium nigricans | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ichthyomyzon unicuspis | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Ictobus sp. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Labidesthes sicculus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lampetra appendix | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepisosteus osseus | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis cyanellus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis cyanellus/macrochinus hybrid | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis gibbosus | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lepomis macrochirus | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Luxllus chrysocephalus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Lythrurus umbratilis | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Micropterus dolomieu | 0 | 1 | 5 | 1 | 0 | 0 | 1 | 0 | 0 | 5 |
| Micropterus salmoides | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Minytrema melanops | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Morone americana | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Morone chrysops | 0 | 4 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Moxostoma anisurum | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Moxostoma erythrurum | 0 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Moxostoma macrolepidotum | 0 | 10 | 1 | 1 | 2 | 0 | 0 | 0 | 0 | 0 |
| Neogobius melanostomus | 1 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 2 | 2 |
| Nocomis biguttatus | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Notemigonus crysoleucas | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Notropis atherinoides | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Notropis hudsonius | 0 | 5 | 0 | 0 | 11 | 1 | 0 | 0 | 0 | 0 |
| Notropis volucellus | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Oncortynchus mykiss | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 9 |
| Oncorlynchus tshawyscha | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |
| Osmerus mordax | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Perca flavescens | 7 | 177 | 2 | 6 | 12 | 3 | 4 | 1 | 0 | 7 |
| Percina caprodes | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pimephales notatus | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pimephales promelas | 0 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Pomoxis nigromaculatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Salmo trutta | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Sander vitreus | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Semotilus atromaculatus | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Total | 9 | 263 | 11 | 11 | 25 | 4 | 9 | 1 | 3 | 34 |


[^0]:    *invasive species

