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Flatfish Stock Assessments for the west coast of Canada for 1999 and Recommended Yield Options for 2000

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#### Abstract

-Abstract-

Interim assessments were prepared for important stocks of flatfish caught in the B.C. trawl fishery. A summary of landing statistics including the observations from the 1997 fishery is presented for all of these. New data was available for only two stocks, Hecate Strait rock sole and Hecate Strait English sole. The catch-age analyses for those two stocks have been updated. There is no change in any of the yield recommendations from last year.

Landings for Area 3C-D and 5C-E Dover sole increased in 1998. Landings for the 3C-D stock were near the low end of the recommended yield range identified last year while 1998 landings for the 5C-E stock were near the upper limit of the recommended yield range identified last year.

Rock sole biomass in Areas 5C-D in 1998 declined slightly from 1997 but remains above the long-term average for the last 50 years. The estimate of fishing mortality for the stock in 1998, was $\mathrm{F}=0.13$, below $\mathrm{F}_{0.1}(\mathrm{~F}=0.22)$. With fixed exploitation at this level there is a 0.91 probability that the stock will maintain its spawning biomass and ensure the recruitment necessary to sustain itself in the future.

English sole biomass in Areas 5C-D in 1998 declined from 1997 to a level near the long-term average for the last 50 years. The estimate of fishing mortality for the stock in 1998 was $\mathrm{F}=0.16$, below $\mathrm{F}_{0.1}(\mathrm{~F}=0.25)$. With fixed exploitation at this level there is a 0.88 probability that the stock will maintain its spawning biomass and ensure the recruitment necessary to sustain itself in the future.


## -Résumé-

Des évaluations provisoires ont été effectuées pour d'importants stocks de poissons plats faisant l'objet de la pêche au chalut en C.-B. Un sommaire des statistiques sur les débarquements, y compris les observations touchant la pêche de 1997, est présenté pour tous ces stocks. De nouvelles données n'ont été obtenues que pour deux stocks, celui de la fausse limande du détroit d'Hecate et celui du carlottin anglais du détroit d'Hecate. L'analyse des captures selon l'âge a été mise à jour pour ces deux stocks. Aucune modification n'est proposée pour les recommandations de rendement, par rapport à l'an dernier.

Les débarquements de limande-sole en provenance des zones 3C-D et 5C-E ont augmenté en 1998. Les débarquements du stock de 3C-D approchaient la limite inférieure de la gamme de rendement définie l'an dernier tandis que ceux de 1998 du stock 5C-E étaient près de la limite supérieure recommandée identifiée l'an dernier.

La biomasse de fausse limande des zones 5C-D en 1998 a diminué légèrement par rapport à 1997 mais demeure supérieure à la moyenne à long terme des 50 dernières années. La mortalité par pêche estimée du stock en $1998(\mathrm{~F}=0,13)$ était inférieure au niveau $\mathrm{F}_{0.1}(\mathrm{~F}=0,22)$. Un taux d'exploitation fixe à ce niveau correspond à une probabilité de 0,91 que le stock
conserve sa biomasse de géniteurs et garantit le recrutement nécessaire au maintien du stock dans le futur.

La biomasse du carlottin anglais dans les zones 5C-D en 1998 a diminué par rapport à 1997 et atteint un niveau approchant la moyenne à long terme des 50 dernières années. La mortalité par pêche estimée du stock en $1998(\mathrm{~F}=0,16)$ en deçà de la valeur $\mathrm{F}_{0.1}(\mathrm{~F}=0,25)$. Un taux d'exploitation fixe à ce niveau correspond à une probabilité de 0,88 que le stock conserve sa biomasse de géniteurs et garantit le recrutement nécessaire au maintien du stock dans le futur.

### 1.0.1. General Introduction

This year interim assessments have been prepared for Hecate Strait Rock sole (Lepidopsetta bilineata) and English sole(Parophrys vetulus) and Area 3CD and 5CDE Dover sole (Microstomus pacificus). Landing statistics have been updated to include information from the 1998 fishery. Landings, $25 \%$ qualified effort and median CPUE are presented for Area 3CD Dover sole and Area 5C-E Dover sole and Hecate Strait Rock sole and English sole. In addition, the catch-age analysis has been updated for the assessments of Hecate Strait Rock sole and Hecate Strait English sole.

The groundfish trawl fishery has changed significantly in recent years. Changes in the management of groundfish fisheries including observer coverage, vessel quotas and changes in vessel catching coefficients over time have nullified the comparison of fishery CPUE over time. As well, the hyperstability of this index has been documented for many situations. Fishery CPUE often does not provide a signal of stock decline until the stock has been depleted (Hillborn and Walters 1992, Richards and Schnute 1986). Accordingly, CPUE is not used as the basis of assessment for any of the cases presented here. The median statistic has been presented for each case as a gross indicator of abundance for select periods of time where there was no regulatory effect and differences in fleet catching coefficients were negligible. In the presence of skewed observations, both the mean of ratios and the ratio of means of CPUE perform badly as they are sensitive to a small number of outliers. The median or $50 \%$ trimmed mean provides a robust alternative to the two former statistics (Fargo and Kronlund 1997).

For Hecate Strait Rock sole and English sole the state space model of Schnute and Richards (1995) was applied to a time series of age composition data to estimate stock biomass and recruitment. Yield for these cases was determined using the $25^{\text {th }}$ and $50^{\text {th }}$ percentiles of the $95 \%$ confidence region for the 1998 biomass estimate using the fishing mortality reference point $\mathrm{F}_{0.1}$. Yield options for 1999 and 2000 are summarised in Table 1.1.

### 1.1. Coastwide

Yield options are not proposed for flatfish species on a coastwide basis.

### 1.2. Strait of Georgia

Yield options are not proposed for flatfish for this region.

### 1.3. West Coast of Vancouver Island (Areas 3C and 3D)

### 1.3.1 Area 3CD Dover sole

### 1.3.1.1 Area 3CD Introduction

Significant commercial quantities of Dover sole (Microstomus pacificus) occur along the Pacific coast from California to Alaska. Dover sole abundance has been shown to decrease with increasing latitude (Westrheim et al. 1992). Results of U.S. tagging studies indicate that a number of individual stocks exist along the Pacific coast and that there is minimal intermingling of the adults among stocks (Westrheim et al. 1992). However, the larvae of this species undergo a prolonged pelagic phase offshore that can last as long as two years. Thus, the larvae of different stocks probably intermingle extensively. The population off the West Coast of Vancouver Island is probably a discrete stock. Dover sole become vulnerable to the commercial trawl fishery at about 5 years of age but are not fully recruited until age 7-8 (Fargo and Workman 1995). Adults undertake a bathymetric migration from shallow (140-200 m) to deep (400-800 m) water for spawning (Westrheim et al. 1992). They spawn over a six month season (DecemberMay) and spawning is age specific with older fish spawning earlier than younger fish (Hunter et al. 1992). The maximum age for Area 3CD Dover sole estimated from biological samples collected to date is 49 years (Fargo and Workman 1995).

### 1.3.1.2 Area 3CD Management History

The Area 3CD Dover sole fishery was unregulated prior to 1992. In 1992, a $20,000 \mathrm{lb}(9 \mathrm{t})$ trip limit was invoked after $70 \%$ of the quota was caught. Since 1992, variable trip limits less than $50,000 \mathrm{lbs}$ ( 23 tons) have been used to manage the fishery. During the period from 1988 to 1996, trips where less than $50,000 \mathrm{lb}(23 \mathrm{t})$ of Dover sole were landed accounted for $60 \%-95 \%$ of the total landings from this area. Trips greater than $50,000 \mathrm{lb}(23 \mathrm{t})$ were not permitted after 1993 and in 1996 the trawl fleet went to IVQs. Landing statistics from the commercial fishery have been updated for this assessment using data from the 1997 fishery.

### 1.3.1.3 Area 3CD Commercial Catch and Effort Data

Landing statistics for Area 3CD Dover sole are presented in Table 1.2. CPUE and effort statistics for 1996 through 1998 are not directly comparable to those for previous years because the estimates are determined from observations made by at-sea observers while estimates for previous years were determined from logbook information recorded by the vessel captains (Fargo and Kronlund 1997). Landings of Area 3CD Dover sole in 1998 were 1092 t compared to

788 t in 1997. Effort was similar for 1997 and 1998, 1876 h and 1829 h , respectively. Median CPUE increased to $0.349 \mathrm{t} / \mathrm{h}$ in 1998 from $0.263 \mathrm{t} / \mathrm{h}$ in 1997.

### 1.3.1.4. Stock status

Evidence from a biomass survey conducted in 1995 (Fargo and Workman 1995) suggests that this stock is fully exploited. There were no biological samples collected from the 1997 and 1998 fisheries so no new analysis has been done. Landings in 1998 were near the low end of the yield range recommended last year. As well, fishing effort in 1998 remained near the level observed in 1997.

### 1.3.1.5.1 Recommendations and yield options

There is no change in the yield recommendations from last year for this stock.

## Low-risk yield option:

A yield of 1000 t is sustainable based on analysis of data collected during a biomass survey of this area conducted in 1995.

## High-risk yield option:

Yields above 1500 t observed from 1991 to 1995 are associated with a significant decline in the commercial CPUE index. A yield of 1500 t can be considered as an upper limit of the sustainable range for this stock.

### 1.5. HECATE STRAIT (Areas 5C and 5D)

### 1.5.1. Rock sole - Hecate Strait

### 1.5.1.1. Introduction

Stock delineation work of Ketchen (1982) and Fargo and Westrheim (1987) indicates that there are probably several stocks of Rock sole in Hecate Strait. However, these stocks are treated as a unit for assessment and management. An age composition data series is available for Rock sole caught in the fishery taking place at Two Peaks and Butterworth fishing grounds at the north end of the Strait. Past work has suggested that both density-dependent and density-independent factors regulate the abundance of this species. Spawning biomass and ocean temperature at the time of spawning are two significant determinants of recruitment. Low recruitment has been associated with low spawning biomass and warm ocean temperatures during larval development (Forrester and Thomson 1969, Fargo and McKinnell 1989).

Recruitment for these stocks has fluctuated over time with the last significant increase occurring during the late 1980s and early 1990s. For this assessment a fishery update is provided which includes landing statistics from the 1998 fishery. The age composition data series has been updated with samples from the 1998 fishery and that analysis forms the basis for this assessment.

### 1.5.1.2. Landing statistics

Landing statistics for Rock sole in Hecate Strait are presented in Table 1.3. Annual statistics for the 1945-98 period are calculated directly from data observations. No detailed data records exist prior to 1954 and the index of Forrester and Thomson (1969) has been used.

Landings in 1998 declined to 576 t from 677 t in 1997while effort in 1998 declined to 2395 h from 2667 h in 1997. Median CPUE decreased slightly to $0.182 \mathrm{t} / \mathrm{h}$ in 1998 from $0.191 \mathrm{t} / \mathrm{h}$ in 1997. Since the early 1980s there has been little contrast in the commercial CPUE series. Area-specific trip limits have influenced the fishing patterns of the fleet in the past. In 1996, individual vessel quotas were applied along with at-sea observers on all option A vessels. This has resulted in more comprehensive fishery data but prevents direct comparison of these data with data collected in previous years.

### 1.5.1.3 Catch-age Analysis

The Rock sole age composition time series was updated with biological data collected during the 1998 trawl fishery. Only samples collected from Minor Area 4 (Major Area 8) between 1945 and 1998 were used for the catch-age analysis. The range of ages used for the catch-age analysis was 4 to $12+$ with the last age group representing fish aged 12 years or older. Three-year-olds are not fully recruited and fish 12 and older were grouped together because of differences in the ageing technique. Otolith surface readings (1945-72) under-estimate the ages of older fish (beginning at about age 12) compared to readings determined from otolith burnt cross-sections (1973-98).

The catch-age model of Schnute and Richards (1995) was used for this assessment. This model is similar to most other catch-age models (Fournier and Archibald 1982, Methot 1989) but does differ in the specification of the model error structure. Parameters in the model likelihood include standard deviations $\sigma 1, \tau 1$, and $\tau 2$, corresponding to the error in the recruitment, biomass index and proportions at age, respectively. The variance ratio $\rho=\sigma_{1}{ }^{2} /\left(\tau_{1}{ }^{2}+\tau_{2}{ }^{2}\right)$ must be specified in the likelihood calculation, analogous to emphasis factors in the stock synthesis model of Methot (1989, 1990).

The model sensitivity to changing values for $\rho$ was examined in the previous assessment (Fargo and Kronlund 1997). Briefly, increasing or decreasing $\rho$ explicitly increases or decreases the standard deviation for the recruitment index. Broader confidence intervals for the model estimates of recruitment occurred with high values for $\rho$ and narrower intervals occurred for low values of $\rho$. Changing the value of $\rho(7 \geq \rho \leq 9)$ had virtually no effect on the biomass index and there were no changes in the overall trends for the population estimates. Fargo
(1995) examined the sensitivity of the model to changing values for M . He found that the best fit occurred with $M$ fixed at 0.20 to 0.25 . The lower value for $M$ was chosen to provide more conservative estimates of biomass. The natural mortality rate was not treated as a parameter to be estimated by the model. It was fixed at 0.20 and remained constant over time for this analysis. Work by Richards et al. (1997) indicated broader confidence intervals for population estimates when M was treated as a parameter to be estimated. The best fit of the model, as indicated by the model likelihood statistic, occurred with $\rho=0.7$ and $M$ fixed at 0.20 . This is the same configuration that was used in the last assessment (Fargo and Kronlund 1997, Fargo 1998). Details of the model are provided in Appendix A.1.

Input data included landed catch, proportions at age in the catch (numbers), mean weight at age, maturity at age and CPUE estimates for Rock sole (adults) obtained from the Hecate Strait surveys conducted between 1984 and 1998. Mean weight at age was computed for each age group in each year from age-length data using an allometric length-weight relationship derived for this stock. Since 1984 a synoptic survey has been conducted every two years in this area. The survey CPUE data for Rock sole was used to tune the catch-age model.

Biomass and recruitment trajectories from the model are presented in Figure 1.1. Between 1980 and 1995, exploitable biomass for this stock increased to the highest level recorded in the last 50 years. Biomass in 1998 is higher than the long-term average for the 1945 to 98 period. The estimate of exploitable biomass in 1998 , $\mathrm{B}_{98}$, was $4854 \mathrm{t} \pm 1303 \mathrm{t}$ (for the $95 \%$ confidence interval). Between the late 1980s and early 1990s recruitment increased to the highest level on record. Between 1993 and 1995 recruitment declined significantly but has remained relatively stable for the last three years. The uncertainty in both population estimates is highest for the later years, a function of the low number of observations used to determine them. A retrospective analysis performed last year indicated that the terminal estimate of biomass varied by $\pm 30 \%$ over a span of four years. After five years it had stabilized (Fargo 1998).

The model estimate of fishing mortality for the stock in 1998 was 0.13 , about the same level as estimated in last year's assessment (Fargo 1998). Although the current rate of fishing appears to be relatively low, observations from several more years are needed before the terminal population estimates and fishing rate for the stock can be considered accurate.

### 1.5.1.4 Fishing rate reference points

A class of F reference points associated with recruitment overfishing was determined from survival ratios using the population estimates from catch-age analysis (Shepherd, 1982, Gabriel et al. 1989). The strategy here is to avoid stock collapse by managing fishing effort to maintain a stable spawning biomass expressed as spawning stock biomass per recruit (SSB/R) (Patterson 1992). The SSB/R index was used to evaluate the spawning potential of Hecate Strait Rock sole over a range of fishing mortality rates. The SSB/R calculations are analogous to those in yield per recruit analysis (Gabriel et al. 1989) and when combined with population estimates from catch-age analysis provide a biological reference for maintenance or rebuilding of spawning stock biomass. Rock sole $\mathrm{SSB} / \mathrm{R}$ ratios were computed over a range of fishing rates (Gabriel et al. 1989) as per yield per recruit calculations. Then, SSB/R was
computed for each year using the population estimates of the catch-age analysis (historical series). The percentiles of the $\mathrm{SSB} / \mathrm{R}$ historical series were used to identify fishing mortality reference points on the $\mathrm{SSB} / \mathrm{R} /$ fishing mortality curve (Figure 1.2). Such targets are the $\mathrm{F}_{\text {low }}$, $\mathrm{F}_{\text {med }}$ and $\mathrm{F}_{\text {high }}$ reference points. These are the fishing mortalities at which the historical data on recruitment suggest that the stock has an over $90 \%$ a $50 \%$ or a worse than $10 \%$ chance of maintaining its spawning stock biomass (Patterson 1992).

The following fishing mortality reference points were estimated in last year's assessment of Hecate Strait Rock sole: $\mathrm{F}_{0.1}=0.22, \mathrm{~F}_{\text {med }}=0.37, \mathrm{~F}_{\text {max }}=0.57 \mathrm{~F}_{\text {high }}=0.63$ and $\mathrm{F}_{\text {low }}=0.16$ (Fargo 1998). Fishing at a rate equivalent to $\mathrm{F}_{\max }$ while permitting the maximum yield to be obtained from a stock, has resulted in stock depletion in the past (FAO 1995) and is clearly not relevant for management purposes other than in the short-term. Similarly $\mathrm{F}_{\text {high }}$ corresponds to a fishing rate that will significantly lower the spawning stock biomass and eventually lead to recruitment overfishing. The target rate $\mathrm{F}_{\text {med }}$ was intended to serve as an indicator for recruitment overfishing (Sissenwine and Shepherd 1987). That is with the fishing rate below this level the stock will maintain or increase its spawning stock biomass. With the fishing rate above this level the spawning stock biomass will decrease, increasing the probability of poor recruitment which will lead to recruitment overfishing in the long-term. The target fishing mortality reference point $\mathrm{F}_{0.1}$ has been advocated by many as a reference point, below $\mathrm{F}_{\text {med }}$, to prevent recruitment overfishing (FAO 1995). $\mathrm{F}_{\text {low }}$ corresponds to a target fishing mortality rate that is consistent with a precautionary management strategy. That is, with the fishing rate at or below this level the spawning stock biomass should increase so that stock rebuilding can take place.

### 1.5.1.5 Stock status

The results of the catch-age analysis indicate a significant increase in stock biomass occurred between the late 1980s and early 1990s due to strong year-classes produced in the mid 1980s. By the mid 1990s, however, significant declines in recruitment and biomass had occurred. Rock sole CPUE from research trawl surveys conducted in Hecate Strait has declined since the early 1990s as well (Fargo 1998). The El Nino event along the B.C. coast in 1996 and 1997 should produce unfavourable temperature conditions for Rock sole egg development and larvae survival (Forrester and Thomson 1969, Fargo and McKinnell 1989). Thus, recruitment and yield for this stock may continue to decline over the next several years.

### 1.5.1.6 Recommendations and Yield Options

The fishing mortality reference point $\mathrm{F}_{0.1}$, $(\mathrm{F}=0.22$ ), was used to estimate sustainable yield for the stock. The $\mathrm{F}_{0.1}$ yield was calculated for the $25^{\text {th }}, 50^{\text {th }}$ and $75^{\text {th }}$ percentiles of the $95 \%$ confidence region for $\mathrm{B}_{98}( \pm 1329 \mathrm{t}) ; 3957 \mathrm{t}, 4854 \mathrm{t}$ and 5750 t , respectively. The $\mathrm{F}_{0.1}$ yield corresponding to these estimates is $779 \mathrm{t}, 956 \mathrm{t}$ and 1133 t , respectively. The yield recommendations are unchanged from last year.

Low risk yield option -- A yield of 800 t , is the low-risk sustainable option for these stocks.

High risk yield option -- A yield of 1000 t , is the high-risk sustainable option for these stocks.
1.5.2. English sole - Hecate Strait

### 1.5.2.1. Introduction

Stock delineation studies conducted by Ketchen (1956) and Fargo et al. (1984) indicate that a single stock of English sole is resident in Hecate Strait. The stock was near the pristine level in the 1940s, declined after large removals in the early 1950s and has remained fairly stable since the late 1960s (Fargo 1998). Both density dependent and density independent factors exert an influence on the abundance of this stock (Fargo 1994). A high rate of Ekman transport through the Strait during the spawning period results in the loss of eggs and larvae due to advection and is associated with subsequent low recruitment. Low recruitment also results when the stock spawning biomass is lower than $30 \%$ of the pristine level (Fargo 1994). The stock has produced strong year-classes about once a decade with the latest increase in recruitment occurring in the early 1990s. The age of recruitment is 4 years for females although they are not fully recruited until age 5 . Only a small proportion of males attain commercial size. Length at $50 \%$ maturity, $L_{50}$, is $25.5 \mathrm{~cm}(3 \mathrm{y})$ for males and $35.1 \mathrm{~cm}(4 \mathrm{y})$ for females (Foucher et al. 1989). The contribution of strong year-classes to the fishery usually lasts about 4 or 5 years. The series of annual landing statistics has been updated using data from the 1998 fishery. The age composition data series has been updated with data from samples collected from the 1998 fishery as well. Analysis of the age composition data is the basis for this assessment.

### 1.5.2.2. Landing statistics

Managers used area-specific quotas as a catch limitation tool in the past. These have undoubtedly had an influence on the fishing patterns of the fleet. In 1996 individual vessel quotas were invoked and at-sea observers were required for all option A vessels. This has resulted in the collection of more comprehensive data but prevents direct comparison with landing statistics for previous years. Annual landing statistics are presented in Table 1.4. Statistics for 1954-98 are calculated directly from data obtained from vessel skipper logs and observer logs. No detailed records are available prior to 1954 and the indices of Ketchen (1980) has been used. English sole landings decreased to 492 t in 1998 from 554 t in 1997 while effort decreased substantially to 725 h from 1286 h over the same period. CPUE in 1998 increased to $0.261 \mathrm{t} / \mathrm{h}$ from $0.227 \mathrm{t} / \mathrm{h}$ in 1997.

### 1.5.2.3.1 Catch-age analysis

The catch-age model of Schnute and Richards (1995) was used for this assessment. See section 1.5.1.3 for information about the model configuration and Appendix A.1. for details of the model. The age composition data series for this stock was updated with age determinations for samples collected during the 1998 trawl fishery. The range of ages used for catch-age analysis was 4 to $12+$ with the last age group representing fish aged 12 years or
older. Fish are fully recruited by age 4 while fish of 12 years and older were grouped together because of differences in the ageing technique used over time. Otolith surface readings (194572) under-estimate the ages of older fish (beginning at about age 12) compared to readings made from otolith burnt cross-sections (1973-98). The full compliment of years where data was available, 1944-98, was analysed. Input data for the model included landed catch, proportions at age in the catch weight at age and annual CPUE estimates for English sole from the Hecate Strait research trawl surveys conducted every two years between 1984 and 1998. The CPUE index was used to tune the catch-age model.

Biomass and recruitment trajectories from the model are presented in Figure 1.3. Between 1950 and the mid 1960s the biomass of this stock declined significantly. Between the mid 1960s and the mid 1970s it rebuilt and it has fluctuated without trend since that time. The estimate of exploitable biomass in 1997, $\mathrm{B}_{98}$, was $3005 \mathrm{t} \pm 366 \mathrm{t}$, (for the $95 \%$ confidence interval). Biomass in 1998 was near the long-term average for the last 50 years. A retrospective analysis performed last year indicated that the terminal estimate of biomass varied by $\pm 20 \%$ over a span of four years but after five years it had stabilized (Fargo 1998). Recruitment increased to the highest level on record in the early 1990s and has declined steadily since that time although the values for 1997 and 1998 are very similar. As in the case of rock sole the uncertainty in the population estimates is greatest for the later years. The model estimate of fishing mortality for the stock in 1998 was 0.16 , the same as that in 1997.

### 1.5.2.4. Fishing rate reference points

Spawning stock biomass per recruit analysis was applied to the data for Hecate Strait English sole. The spawning stock biomass per recruit ratios for English sole were computed over a range of fishing rates (Gabriel et al. 1989) to produce the yield curve in Figure 1.4. The percent quantiles of the historical series of spawning stock biomass per recruit were used to identify fishing mortality reference points on the curve. Such targets are the $\mathrm{F}_{\text {low }}, \mathrm{F}_{\text {med }}$ and $\mathrm{F}_{\text {high }}$ reference points. These are the fishing mortalities at which the historical data on recruitment suggest that the stock has an over $90 \%$ a $50 \%$ or a worse than $10 \%$ chance of maintaining its spawning stock biomass (Patterson 1992).

The target fishing mortality references points estimated for this stock in last year's assessment were: $\mathrm{F}_{0.1}=0.25, \mathrm{~F}_{\text {med }}=0.28, \mathrm{~F}_{\text {low }}=0.11, \mathrm{~F}_{\text {high }}=0.50$ and $\mathrm{F}_{\text {max }}=0.83$ (Fargo 1998). These represent exploitation rates than can be used by managers as fixed strategies for achieving sustainable yield or stock rebuilding. Fishing at a rate equivalent to $\mathrm{F}_{\max }$ while permitting the maximum yield to be obtained from a stock, has resulted in stock depletion in the past (FAO 1995) and is clearly not relevant for management purposes other than in the short-term. Similarly $\mathrm{F}_{\text {high }}$ corresponds to a fishing rate that will significantly lower the spawning stock biomass and eventually lead to recruitment overfishing. The target rate $\mathrm{F}_{\text {med }}$ was intended to serve as an indicator for recruitment overfishing (Sissenwine and Shepherd 1987). That is with the fishing rate below this level the stock will maintain or increase its spawning stock biomass. With the fishing rate above this level the spawning stock biomass will decrease, increasing the probability of poor recruitment which will lead to recruitment overfishing in the long-term. The target fishing mortality reference point $\mathrm{F}_{0.1}$ has been advocated by many as a reference point,
below $\mathrm{F}_{\text {med }}$, to prevent recruitment overfishing (FAO 1995). $\mathrm{F}_{\text {low }}$ corresponds to a target fishing mortality rate that is consistent with a precautionary management strategy. That is, with the fishing rate at or below this level the spawning stock biomass should increase so that stock rebuilding can take place.

### 1.5.2.5 Stock status

The distribution of 1998 biomass estimate, $\mathrm{B}_{98}, 3005 \mathrm{t}$ was used to estimate yield using the reference point of $\mathrm{F}_{0.1}(\mathrm{~F}=0.25)$. The $\mathrm{F}_{0.1}$ yield was calculated for the $25^{\text {th }}, 50$ th and $75^{\text {th }}$ percentiles of the $95 \%$ confidence region for $B_{98}( \pm 366 \mathrm{t}) ; 2753 \mathrm{t}, 3005 \mathrm{t}$ and 3257 t , respectively. The corresponding yields were $608 \mathrm{t}, 664 \mathrm{t}$, and 720 t respectively. There is no change in yield recommendations from last year.

### 1.5.2.6. Recommendations and Yield Options

Low risk yield option -- A yield of 500 t t , is the low-risk option.
High risk yield option -- A yield of 600 t , is the high-risk option.

### 1.5.3 Dover sole (Areas 5C-E)

### 1.5.3.1 Introduction

The fishery for Dover sole in Areas 5C-E takes place in northern Hecate Strait at 100 to 160 m depths between May and October, and off the west coast of the Queen Charlotte Islands at 400 to 800 m depths from December to April. The seasonal shift in the fishery is related to the bathymetric spawning migration for the species. The fishery off the west coast of the Queen Charlotte Islands takes place on a spawning population. Dover sole begin to recruit to the fishery at 5 years of age but are not fully recruited until age 7-8. The Dover sole fishery in area 5C-E was unregulated prior to 1981 when annual quotas were first applied.

### 1.5.3.2 Area 5CDE Landing Statistics

As in other cases the CPUE index for this fishery deteriorated in the 1980s due to the combined effect of changes in management and vessel efficiency. Also, after 1996 annual statistics are not directly comparable to those in previous years because of the implementation of the at-sea observer program and individual vessel quotas (IVQs). Landing statistics for Dover sole from the Area 5C-E trawl fishery for 1970-98 are presented in Table 1.5. Landings increased to 1043 t in 1998 from 714 t in 1997 while effort decreased to 1520 h in 1998 from 1563 h in 1997 and CPUE increased to $0.422 \mathrm{t} / \mathrm{h}$ in 1998 from $0.326 \mathrm{t} / \mathrm{h}$ in 1998.

Estimates of the total mortality rate, Z, estimated from age composition data in 1996-97 ranged between 0.15 to 0.27 (Fargo 1998). No age composition data were available for the 1998 fishery. Estimates of the rate of fishing mortality for the stock in 1996 and 1997 ranged from 0.10 to 0.15 while $\mathrm{F}_{0.1}$ for the stock is 0.13 (Fargo 1998). The level of effort in 1998 was similar to that in 1997. Thus, the yield recommendations for the 2000 fishery remain the same as those for the 1999 fishery.

### 1.5.3.4 Recommendations and Yield Options

Low risk yield option:
A yield of 800 t , equivalent to the MSY estimated using surplus production analysis (Fargo 1998).

High risk yield option:
A yield of $1200 t$ is suggested as an upper limit for yield for the area 5CDE Dover sole stock using results from a length-based analysis (Fargo 1991).

## References

Alverson, D.L. 1960. A study of annual seasonal bathymetric catch patterns for commercially important groundfishes of the Pacific Northwest Coast of North America. Pac. Mar. Fish. Comm. Bull. 4: 66p.

Anthony, V.C. 1982. The calculation of $\mathrm{F}_{0.1}$; a plea for standardization. NAFO SCR Doc. 82/VI/64 Ser. No. N557: 15p.

Breen, P.A. and M. Stocker. 1993. Evaluating the consequences of constant catch levels on the red rock lobster, Jasus edwardsii, population of New Zealand. Proceedings of the International Symposium on Management Strategies for Exploited Fish Populations, Alaska Sea Grant College Program. AK-SG-93-02.

Castillo, G.C., H.W. Li and J.T. Golden. 1994. Environmentally induced recruitment variation in Petrale sole, Eopsetta jordani. Fishery Bulletin 92: 481-493.

Clark, W.G. 1991. Groundfish exploitation rates based on life history parameters. Can. J. Fish. Aquat. Sci. 48: 734-750.

FAO 1995. Precautionary approach to fisheries. Guidelines on the precautionary approach to capture fisheries and species introductions. Elaborated by the Technical Consultation on the Precautionary Approach to Capture Fisheries. Lysekil, Sweden. June 6-13 1995. FAO Fisheries Technical Paper No. 350, Part 1. Rome FAO. 52 p.

Fargo, J. 1991. Flatfish. pp. 71-116. In Fargo, J. and B.M. Leaman [Eds.] Groundfish stock assessments for the west coast of Canada in 1990 and recommended yield options for 1991. Can. Tech. Rep. Fish. Aquat. Sci. No. 1778: 320 p.

Fargo, J. 1994. Examining recruitment relationships for Hecate Strait English sole (Pleuronectes vetulus). Neth. J. Sea Research 32(3/4):385-397.

Fargo, J. 1995. Flatfish. pp. 160-222. In Stocker, M. and J. Fargo [Ed.] Groundfish stock assessments for the west coast of Canada in 1994 and recommended yield options for 1995. Can. Tech. Rep. Fish. Aquat. Sci. 2069: 440 p.

Fargo, J. Flatfish. 1998. Flatfish Stock Assessments for the west coast of Canada for 1997 and recommended yield options for 1998. Canadian Stock Assessment Secretariat Research Document 97/36. 55p.

Fargo, J. and S. McKinnell 1989. Effects of temperature and stock size on year-class production for Rock sole (Lepidopsetta bilineata) in northern Hecate Strait, British Columbia, p. 327-333. In R.J. and G.A. Mc Farlane [ed.] Effects of ocean variability on recruitment
and an evaluation of parameters used in stock assessment models. Can. Spec. Publ. Fish. Aquat. Sci. 108.

Fargo, J., R.P. Foucher, S.C. Shields and D. Ross. 1984. English sole tagging in Hecate Strait, R/V G.B. REED, June 6-24, 1983. Can. Data Rep. Fish. Aquat. Sci. No. 427: iii + 49p.

Fargo, J. and A.R. Kronlund. 1997. Flatfish stock assessments for the west coast of Canada for 1996 and recommended yield options for 1997. Can. Tech. Rep. Fish. Aquat. Sci. 2149: 124 p.

Fargo, J. and S.J. Westrheim. 1987. Results, through 1985, of the Rock sole (Lepidopsetta bilineata) tagging experiments in Hecate Strait (British Columbia) during April-May 1982 with regard to stock delineation. Can. MS. Rep. Fish. Aquat. Sci. 1912: 51p.

Fargo, J. and G.D. Workman. 1995. Results of the Dover sole (Microstomus Pacificus) biomass survey conducted off the west coast of Vancouver Island February 13-27, 1995. Can. MS Rep. Fish. Aquat. Sci. 2340:73p.

Forrester, C.R. and J. A. Thomson. 1969. Population studies on the Rock sole (Lepidopsetta bilineata) of northern Hecate Strait British Columbia. Fish. Res. Board Can. 108:104 p.

Foucher, R.P., A.V. Tyler, J. Fargo, and G.E. Gillespie. 1989. Reproductive biology of Pacific cod and English sole from the cruise of the F/V Blue Waters to Hecate Strait, January 30 to February 11, 1989. Can. MS. Rep. Fish. Aquat. Sci. 2026:85 p.

Gabriel, W.L., M.P. Sissenwine and W.J. Overholtz. 1989. Analysis of spawning stock biomass per recruit: an example for Georges Bank haddock. North Am. J. Fish. Mgmt. 9: 383391.

Fournier, D. and C.P. Archibald. 1982. A general theory for analyzing catch at age data. Can. Fish. Aquat. Sci. 39: 1195-1207.

Hart, J.L. 1973. Pacific fishes of Canada. Fish. Res. Board Can. Bulletin No. 180: 740 p.
Hilborn, R. and C.J. Walters. 1992. Quantitative fisheries stock assessment: choice dynamics and uncertainty. Chapman and Hall, New York. xv +570 p.

Hunter, J.R., B.J. Macewicz, N. Chan-huei Lo and C.A. Kimbrell. 1992. Fecundity, spawning and maturity of female Dover sole, Microstomus Pacificus, with an evaluation of assumptions and precision. Fishery Bulletin 90: 101-128.

Ketchen, K.S. 1956. Factors influencing the survival of the lemon sole (Parophrys vetulus) in Hecate Strait, British Columbia. J. Fish. Res. Bd. Canada 13(5): pp. 647-694.

Ketchen, K.S. 1980. Assessment of groundfish stocks off the west coast of Canada. (1979). Can. Data Rep. Fish. Aquat. Sci. 185: 213p.

Ketchen, K.S. 1982. Stock delineation and growth of Rock sole (Lepidopsetta bilineata) as indicated by tagging in British Columbia waters, 1944-66. Can. MS Rep. Fish. Aquat. Sci. 1683: 41p.

Ketchen, K.S. and C.R. Forrester. 1966. Population dynamics of the Petrale sole, (Eopsetta jordani), in waters off western Canada. Fish. Res. Board Can. Bull. 153: 95 p.

Methot, R.D. 1989. Synthetic estimates of historical abundance and mortality for northern anchovy, p. 68-82. In: E.F. Edwards and B. Megrey (eds.). 1989. Mathematical analyses of fish stock dynamics: review and current applications. Am. Fish. Soc. Symp. 6.

Methot, R.D. 1990. Synthesis model: an adaptable framework for analysis of diverse stock assessment data. Int. Pac. Fish. Comm. Bull. 50: 259-277.

Pedersen, M.G. 1975a. Movements and growth of Petrale sole (Eopsetta jordani) tagged off Washington and southwest Vancouver Island. J. Fish. Board Can. 32: 2169-2177.

Pedersen, M.G. 1975b. Recent investigations of Petrale sole off Washington and British Columbia. Wash. Dept. Fish. Tech. Rep. 17: 72p.

Patterson, K. 1992. Fisheries for small pelagic species: an empirical approach to management targets. Reviews in Fish Biology and Fisheries, 2: 321-338.

Richards, L.J. and N. Olsen. Slope rockfishes. PSARC Working paper G96-9.
Richards, L.J. and J.T. Schnute. 1992. Statistical models for estimating CPUE from catch and effort data. Can. J. Fish. Aquat. Sci. 49:7 pp. 1315-1327.

Perry, R.I., M. Stocker and J. Fargo. 1994. Environmental effects on the distributions of groundfish in Hecate Strait, British Columbia. Can. J. Fish. Aquat. Sci. 51: 1401-1409.

Quinn, T.J. 11. 1985. Catch-per-unit-effort: a statistical model for Pacific halibut (Hippoglossus stenolepis). Can. J. Fish. Aquat. Sci. 42: 1423-1429.

Richards, L.J. 1988. Inshore rockfish. In: Fargo, J. and A.V. Tyler (Eds.). Groundfish stock assessments for the west coast of Canada in 1987 and recommended yield options for 1988. Can. Tech. Rep. Fish. Aquat. Sci. 1617: 273-294.

Richards, L.J. and J. T. Schnute. 1986. An experimental and statistical approach to the question: Is CPUE an index of abundance? Can. J. Fish. Aquat. Sci. 43: 1214-1227.

Richards, L.J., J.T. Schnute and N. Olsen. 1997. Visualising catch-age analysis: a case study. Can. J. Fish. Aquat. Sci. 54: 1646-1658.

Schnute, J.T. 1994. A general framework for developing sequential fisheries models. Can. J. Fish. Aquat. Sci. 51: 1676-1688.

Schnute, J.T. and L.J. Richards. 1995. The influence of error on population estimates from catchage models. Can. J. Fish. Aquat. Sci. 52:2063-2077.

Shepherd, J.G. 1982. A versatile new stock-recruitment relationship of fisheries and construction of sustainable yield curves. J. Cons. Int. Explor. Mer 40:67-75.

Sissenwine, M.P. and J.G. Shepherd. 1987. An alternative perspective on recruitment overfishing and the biological reference points. Can. J. Fish. Aquat. Sci. 44: 913-18.

Stocker, M. and J. Fargo (Eds). 1995. Groundfish stock assessments for the west coast of Canada in 1994 and recommended yield options for 1995. Can. Tech. Rep. Fish. Aquat. Sci. 2069: 440 p .

Turnock, J., M. Wilkins, M. Saelens and B. Lauth. 1994. Status of west coast Petrale sole in 1994. Status of the Pacific coast groundfish fishery through 1994 and recommended acceptable biological catches for 1995. Pacific Management Council. Vol. 1: p. C1-C62.

Turnock, J., M. Wilkins, M. Saelens and B. Lauth. 1995. Status of west coast Dover sole in 1994. Pacific Fishery Management Council, Portland, Oregon.

Walters, C.J. and A.M. Parma. 1995. Fixed exploitation rate strategies for coping with the effects of climate change. Can. J. Fish. Aquat. Sci. 53: 148-158.

Westrheim, S.J., W.H. Barrs, E.K. Pikitch and L.F. Quirollo. 1992. Stock delineation of Dover sole in the California-British Columbia region based on tagging studies conducted during 1948-79. North American Journal of Fisheries Management 12:172-181.

Table 1.1. Yield options for British Columbia flatfish species/stocks
1999-00

| Species | Area | 1999 |  | 2000 |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | low risk | high risk | low risk | high risk |
| Dover sole | Area 3C-D | 1000 t | 1500 t | 1000 t | 1500 t |
|  | Area 5C-E | 800 t | 1200 t | 800 t | 1200 t |
| Rock sole | Area 5C-D | 800 t | 900 t | 800 t | 900 t |
| English sole | Area 5C-D | 500 t | 600 t | 500 t | 600 t |

Table 1.2 Annual landing statistics for the Area 3CD Dover sole trawl fishery, 1980-98.

| Year | Landings <br> $(\mathrm{t})$ | Effort $^{\mathrm{a}}$ <br> $(\mathrm{h})$ | CPUE $^{\mathrm{b}}$ <br> $(\mathrm{t} / \mathrm{h})$ |
| :---: | ---: | ---: | ---: |
| 1980 | 184 | 306 | 0.556 |
| 1981 | 171 | 461 | 0.339 |
| 1982 | 129 | 281 | 0.361 |
| 1983 | 22 | 84 | 0.389 |
| 1984 | 24 | 79 | 0.256 |
| 1985 | 3 | 9 | 0.280 |
| 1986 | 2 | 8 | 0.321 |
| 1987 | 1 | 4 | 0.143 |
| 1988 | 371 | 620 | 0.426 |
| 1989 | 1115 | 1754 | 0.415 |
| 1990 | 1122 | 1882 | 0.402 |
| 1991 | 1222 | 2572 | 0.316 |
| 1992 | 1382 | 3034 | 0.357 |
| 1993 | 1785 | 4459 | 0.318 |
| 1994 | 1492 | 4626 | 0.267 |
| 1995 | 1630 | 5352 | 0.259 |
| 1996 | 1083 | 2318 | 0.229 |
| 1997 | 788 | 1876 | 0.263 |
| 1998 | 1092 | 1829 | 0.349 |

[^0]Table 1.3. Canada-U.S. landing statistics for Hecate Strait Rock sole, 1945-98.

| Year | Landings ( t ) | Effort (h) ${ }^{\text {a }}$ | CPUE (t/h) ${ }^{\text {b }}$ |  |
| :---: | :---: | :---: | :---: | :---: |
|  | 45 | 121 | 434 | 0.279 |
|  | 46 | 410 | 2228 | 0.184 |
|  | 47 | 1181 | 1946 | 0.607 |
|  | 48 | 901 | 1753 | 0.514 |
|  | 49 | 657 | 1352 | 0.486 |
|  | 50 | 784 | 1452 | 0.540 |
|  | 51 | 1024 | 944 | 1.085 |
|  | 52 | 2292 | 2014 | 1.138 |
|  | 53 | 779 | 1227 | 0.635 |
|  | 54 | 926 | 840 | 0.938 |
|  | 55 | 1560 | 1558 | 0.680 |
|  | 56 | 1160 | 1484 | 0.644 |
|  | 57 | 1151 | 2019 | 0.443 |
|  | 58 | 1256 | 1331 | 0.650 |
|  | 59 | 416 | 636 | 0.403 |
|  | 60 | 1127 | 1100 | 0.680 |
|  | 61 | 744 | 694 | 0.900 |
|  | 62 | 829 | 849 | 0.735 |
|  | 63 | 881 | 735 | 0.737 |
|  | 64 | 743 | 835 | 0.531 |
|  | 65 | 879 | 629 | 0.545 |
|  | 66 | 2544 | 2491 | 0.598 |
|  | 67 | 2162 | 2324 | 0.511 |
|  | 68 | 2366 | 4209 | 0.386 |
|  | 69 | 1461 | 4485 | 0.314 |
|  | 70 | 1403 | 3660 | 0.326 |
|  | 71 | 1503 | 3587 | 0.255 |
|  | 72 | 515 | 650 | 0.337 |
|  | 73 | 507 | 619 | 0.435 |
|  | 74 | 622 | 603 | 0.475 |
|  | 75 | 1204 | 1912 | 0.360 |
|  | 76 | 1438 | 1830 | 0.402 |
|  | 77 | 846 | 1896 | 0.285 |
|  | 78 | 874 | 1662 | 0.336 |
|  | 79 | 1313 | 1943 | 0.330 |
|  | 80 | 977 | 2420 | 0.254 |
|  | 81 | 584 | 806 | 0.287 |
|  | 82 | 291 | 841 | 0.209 |
|  | 83 | 247 | 499 | 0.286 |
|  | 84 | 188 | 573 | 0.188 |
|  | 85 | 112 | 276 | 0.242 |
|  | 86 | 219 | 470 | 0.345 |
|  | 87 | 536 | 577 | 0.389 |
|  | 88 | 1402 | 2520 | 0.410 |
|  | 89 | 1422 | 3757 | 0.288 |
|  | 90 | 1519 | 3948 | 0.319 |
|  | 91 | 2666 | 6552 | 0.295 |
|  | 92 | 2226 | 5777 | 0.289 |
|  | 93 | 2080 | 5851 | 0.301 |
|  | 94 | 1384 | 4282 | 0.275 |
|  | 95 | 1294 | 3538 | 0.322 |
|  | 96 | 670 | 2336 | 0.207 |
|  | 97 | 677 | 2667 | 0.191 |
|  | 98 | 576 | 2395 | 0.182 |

[^1]Table 1.4. Canada-U.S. landing statistics for Hecate Strait English sole, 1944-98.

| Year | Landings ( t ) | Effort (h) ${ }^{\text {a }} \quad$ CPUE (t/h) ${ }^{\text {b }}$ |  |
| :---: | :---: | :---: | :---: |
| 44 | 152 | 215 | 0.707 |
| 45 | 304 | 365 | 0.832 |
| 46 | 470 | 809 | 0.581 |
| 47 | 350 | 538 | 0.651 |
| 48 | 937 | 2740 | 0.342 |
| 49 | 795 | 1893 | 0.420 |
| 50 | 2622 | 4910 | 0.534 |
| 51 | 1024 | 2142 | 0.478 |
| 52 | 1347 | 3293 | 0.409 |
| 53 | 871 | 2084 | 0.418 |
| 54 | 455 | 563 | 0.362 |
| 55 | 875 | 744 | 0.401 |
| 56 | 956 | 1344 | 0.349 |
| 57 | 552 | 640 | 0.244 |
| 58 | 693 | 617 | 0.337 |
| 59 | 940 | 772 | 0.315 |
| 60 | 1147 | 1058 | 0.333 |
| 61 | 871 | 1615 | 0.298 |
| 62 | 459 | 903 | 0.247 |
| 63 | 408 | 568 | 0.207 |
| 64 | 436 | 441 | 0.272 |
| 65 | 414 | 326 | 0.317 |
| 66 | 362 | 354 | 0.302 |
| 67 | 534 | 535 | 0.411 |
| 68 | 671 | 844 | 0.302 |
| 69 | 819 | 1314 | 0.390 |
| 70 | 1002 | 2042 | 0.312 |
| 71 | 488 | 1585 | 0.192 |
| 72 | 371 | 550 | 0.230 |
| 73 | 667 | 514 | 0.411 |
| 74 | 500 | 519 | 0.519 |
| 75 | 938 | 1015 | 0.466 |
| 76 | 1133 | 1627 | 0.275 |
| 77 | 1179 | 2201 | 0.310 |
| 78 | 559 | 944 | 0.246 |
| 79 | 864 | 980 | 0.337 |
| 80 | 995 | 1105 | 0.327 |
| 81 | 1327 | 2149 | 0.249 |
| 82 | 428 | 1062 | 0.219 |
| 83 | 430 | 834 | 0.240 |
| 84 | 658 | 1129 | 0.290 |
| 85 | 585 | 1520 | 0.226 |
| 86 | 335 | 469 | 0.365 |
| 87 | 630 | 396 | 0.347 |
| 88 | 688 | 540 | 0.493 |
| 89 | 826 | 925 | 0.385 |
| 90 | 992 | 1335 | 0.383 |
| 91 | 913 | 940 | 0.308 |
| 92 | 987 | 1602 | 0.307 |
| 93 | 1421 | 2636 | 0.295 |
| 94 | 1000 | 1860 | 0.343 |
| 95 | 1190 | 2321 | 0.320 |
| 96 | 455 | 570 | 0.310 |
| 97 | 554 | 1286 | 0.227 |
| 98 | 492 | 725 | 0.261 |

[^2]Table 1.5. Canada-U.S. landing statistics for Dover sole, Areas 5C-E, 1970-98.

|  |  |  |  |
| :---: | ---: | ---: | :---: |
| Year | Landings (t) | Effort (h) | CPUE (t/h) ${ }^{\mathrm{b}}$ |
|  |  |  |  |
| 70 | 965 | 1324 | 0.590 |
| 71 | 903 | 1367 | 0.556 |
| 72 | 922 | 1495 | 0.543 |
| 73 | 768 | 910 | 0.679 |
| 74 | 767 | 878 | 0.687 |
| 75 | 882 | 1135 | 0.573 |
| 76 | 1022 | 1465 | 0.440 |
| 77 | 577 | 900 | 0.319 |
| 78 | 483 | 650 | 0.497 |
| 79 | 697 | 1057 | 0.333 |
| 80 | 807 | 724 | 0.416 |
| 81 | 840 | 1079 | 0.428 |
| 82 | 512 | 894 | 0.433 |
| 83 | 693 | 544 | 0.568 |
| 84 | 953 | 1526 | 0.448 |
| 85 | 830 | 1039 | 0.485 |
| 86 | 1040 | 931 | 0.562 |
| 87 | 503 | 432 | 0.549 |
| 88 | 649 | 652 | 0.594 |
| 89 | 696 | 775 | 0.567 |
| 90 | 787 | 1181 | 0.542 |
| 91 | 649 | 1041 | 0.428 |
| 92 | 883 | 1444 | 0.381 |
| 93 | 1508 | 2767 | 0.414 |
| 94 | 1418 | 3117 | 0.371 |
| 95 | 1587 | 4220 | 0.320 |
| 96 | 1133 | 2245 | 0.308 |
| 97 | 714 | 1563 | 0.326 |
| 98 | 1043 | 1520 | 0.422 |
|  |  |  |  |

[^3]
## Appendix. A. 1 Schnute and Richards (1995) Catch-age model

The catch-age model used for the assessments of Hecate Strait rock and English soles is an application of the state space model developed by Schnute and Richards (1995). The model attempts to reconstruct the population history from known controls and observations. In this context, the catch biomass acts as a known control on the population dynamics. Observations, including proportions at age in the catch and a biomass index from survey CPUE values, describe the current state of the system. The model relates the observations, measured with error, to unknown numbers of fish in the population.

Table A. 1 contains a deterministic version of the model, with notation described in Table A.2. Equations in Table A. 1 are tailored for each application. In particular, the Schnute-Richards model is based on numbers of fish; we use known weights, $w_{a t}$ of fish of age $a$ at time $t$ and the maturity ogive $m_{a}$ to determine the spawning biomass $S_{t}$ and exploitable population biomass $B_{t}$.

Similar to other stochastic catch-age models, our analysis contains a separability assumption. The two parameters $\alpha$ and $\beta_{l}$ describe a selectivity function which is timeindependent and asymptotic with age. The quantity $\beta_{a}$ in equation (A.2) denotes the proportion of age $a$ fish that are vulnerable to the fishery.

Other quantities in the parameter vector $\Theta$ are the natural mortality rate $M$, the survey catchability $q$, and the time series of recruitments $R_{t}$. We treat the recruitments as parameters to be estimated from the data; our analysis does not contain an explicit stockrecruitment function.

The prediction equations (A.14) and (A.15) relate quantities $\bar{I}_{t}$ and $\bar{p}_{a t}$ obtained from the model dynamics to observations $I_{t}$ and $p_{a t}$ of survey CPUE and age proportions, respectively. (We use the convention of a bar over a quantity to denote a prediction for that quantity.) We assume in (A.14) that the survey CPUE indexes the population biomass after half of the annual catch has been removed. The catchability $q$ converts units of population biomass into units of CPUE. Although the relationship (A.14) could be made age-specific, age composition data are not available for the early surveys. The predicted age proportions in the catch are obtained from the underlying population age structure in equation (A.4).

Schnute and Richards (1995) specify stochastic counterparts of the deterministic equations (Table A.1), model residuals, and the model likelihood function. They impose three sources of error: (1) autoregressive lognormal process error among the recruitments $R_{t} ;$ (2) lognormal error in CPUE; and (3) multivariate logistic error in the observed proportions $p_{a t}$. These error structures lead to residual functions

$$
\begin{aligned}
& \xi_{t}=\log I_{t}-\log \bar{I}_{t} \\
& \eta_{a t}=\log p_{a t}-\log \bar{p}_{a t}-\frac{1}{A} \sum_{a=1}^{A}\left[\log p_{a t}-\log \bar{p}_{a t}\right]
\end{aligned}
$$

that describe model relationships between predictions and observations of survey CPUE and age proportions, respectively.

The likelihood for this catch-age model conforms to the errors-in-variables paradigm (Schnute 1994); apparent variations in abundance can be explained through high process error $\sigma$ in recruitment or high measurement error $\tau$ in CPUE. Schnute and Richards (1995) resolve this ambiguity by fixing the model variance ratio, $\rho=\frac{\sigma^{2}}{\sigma^{2}+\tau^{2}}$, between the recruitment variance and the total variance $\left(\sigma^{2}+\tau^{2}\right)$.

For the catch-age analysis, we fix the variance ratio $\rho=0.7$, a value that represents moderate levels of error in both recruitment and survey CPUE. Similar stock reconstructions were obtained for a range of reasonable choices of $\rho$ in preliminary model runs. We also employ a fixed natural mortality rate of $M=0.2$. Age classes in the model range from recruits to the fishery at age 4 to an accumulator age class for age 12 and older. To reduce the influence on the model likelihood of very small age proportion observations (obtained from a small number of fish), we group consecutive age classes such that $p_{a t} \geq 0.02$ for each age $a$ and time $t$ (Richards et al. 1997).

The model was implemented using AD Model Builder software (Otter Research Ltd. 1994). Standard errors for the model parameters and other quantities were obtained from the model hessian matrix. These allow calculation of symmetric confidence intervals, assuming that the parameter estimates have a multivariate normal distribution. In particular, we used AD Model Builder to compute standard errors for log recruitment, log spawner biomass and log exploitable biomass.

Table A.1. Deterministic catch-age model. Calculations begin with the parameter vector $\boldsymbol{\Theta}$ and proceed recursively to define all states and observations.

## Parameters

(A.1) $\Theta=\left(\alpha, \beta_{1}, M, q,\{R\}_{t=2-A}^{T}\right)$

Selectivity
(A.2)
$\beta_{a}=1-\left(1-\beta_{1}\right)\left(\frac{A-a}{A-1}\right)^{\alpha}$
State moments
(A.3) $P_{t}=\sum_{a=1}^{A} \beta_{a} N_{a t}$
(A.4) $u_{a t}=\beta_{a} N_{a t} / P_{t}$
(A.5) $B_{t}=\sum_{a=1}^{A} \beta_{a} w_{a t} N_{a t}$
(A.6) $S_{t}=\sum_{a=1}^{A} m_{a} w_{a t} N_{a t}$
(A.7) $C_{t}=D_{t} / \sum_{a=1}^{A} u_{a t} w_{a t}$
(A.8) $F_{t}=\log \left(\frac{P_{t}}{P_{t}-C_{t}}\right)$

Initial states
(A.9) $N_{a 1}=R_{2-a} e^{-M(a-1)} ; 1 \leq \mathrm{a}<\mathrm{A}$
(A.10) $N_{A 1}=R_{2-A}\left(\frac{e^{-M(A-1)}}{1-e^{-M}}\right)$

State Dynamics
(A.11) $N_{1 t}=R_{t}$
(A.12) $N_{a t}=e^{-M}\left[N_{a-1, t-1}-u_{a-1, t-1} C_{t-1}\right] ; 2 \leq \mathrm{a}<\mathrm{A}$
(A.13) $N_{A t}=e^{-M}\left[N_{A-1, t-1}+N_{A, t-1}-\left(u_{A-1, t-1}+u_{A, t-1}\right) C_{t-1}\right]$

Predicted Observations
(A.14) $\bar{I}_{t}=q\left(B_{t}-0.5 D_{t}\right)$
(A.15) $\bar{p}_{a t}=u_{a t} ; 2 \leq \mathrm{a} \leq \mathrm{A}$

Appendix table 1.1. Description of the notation for the input data, parameters, and other calculated model quantities in Table A.1.

| Symbol | Description |
| :---: | :---: |
|  | Index quantities |
|  | $a$ age-class from 1 to $A$ |
|  | $t$ year from 1 to $T$ |
|  | Input data |
|  | $D_{t}$ observed catch biomass in year $t$ |
|  | $I_{t}$ observed survey CPUE in year $t$ |
|  | $m_{a}$ proportion of age-class $a$ fish which are mature |
|  | $p_{a t}$ observed proportion of age-class $a$ fish in the year $t$ catch |
|  | $w_{a t}$ weight of age-class $a$ fish in year $t$ |
|  | Parameters |
|  | $\Theta$ parameter vector |
|  | $\alpha$ selectivity slope parameter |
|  | $\beta_{1}$ selectivity of age-class 1 |
|  | $M$ natural mortality rate |
|  | $q$ catchability for survey CPUE |
|  | $R_{t}$ age-class 1 recruitment in year $t$ |
|  | Calculated quantities |
|  | $\beta_{a}$ selectivity for age-class $a$ |
|  | $B_{t}$ exploitable population biomass at the start of year $t$ |
|  | $C_{t}$ catch number in year $t$ |
|  | $F_{t}$ fishing mortality rate in year $t$ |
|  | $N_{a t}$ number of age-class $a$ fish at the start of year $t$ |
|  | $P_{t}$ exploitable population numbers at the start of year $t$ |
|  | $S_{t}$ spawning stock biomass at the start of year $t$ |
|  | $u_{a t}$ exploitable proportion of age-class $a$ fish in year $t$ catch |

Appendix table 1.2 Values for biological statistics used in the yield per recruit and spawning stock biomass per recruit analysis.

| Parameter or | Rock sole | English sole |
| :--- | :--- | :--- |
| calculated value |  |  |



Where:
$K, L \infty, W \infty$ and $t_{o}$ are von Bertalanffy growth curve coefficients
M is the instantaneous rate of natural mortality
w is the mean weight at age in grams
1 is the mean length at age in centimetres
p is the proportion mature at age
and : j indexes age groups 4-12+


Figure 1.1. Biomass and recruitment trajectories (1945-98) from the catch-age analysis for Hecate Strait Rock sole. The vertical bars represent the $95 \%$ confidence limits for individual estimates.


Figure 1.2. Results from the spawning stock biomass per recruit analysis for Hecate Strait Rock sole. The solid line indicates the fishing mortality rate where the stock has a $50 \%$ chance of maintaining its spawning biomass ( $\mathrm{F}_{\mathrm{med}}$ ). The dotted lines indicate the fishing rates associated with the $10^{\%}$ and $90^{\%}$ quantiles ( $\mathrm{F}_{\text {low }}$ and $\mathrm{F}_{\text {high }}$, respectively) of the $\mathrm{SSB} / \mathrm{R}$ index computed from the rock sole catch-age analysis.

Hecate Strait English sole catch-age analysis


Hecate Strait English sole catch-age analysis


Figure 1.3. Biomass and recruitment trajectories (1945-98) from the catch-age analysis for Hecate Strait English sole. The vertical bars represent the $95 \%$ confidence limits for individual estimates.


Figure 1.4. Results from the spawning stock biomass per recruit analysis for Hecate Strait English sole. The solid line indicates the fishing mortality rate where the stock has a $50 \%$ chance of maintaining its spawning biomass ( $\mathrm{F}_{\mathrm{med}}$ ). The dotted lines indicate the fishing rates associated with the $10 \%$ and $90 \%$ quantiles ( $\mathrm{F}_{\text {low }}$ and $\mathrm{F}_{\text {high }}$, respectively) of the $\mathrm{SSB} / \mathrm{R}$ index computed from the population estimates from catch-age analysis.


[^0]:    ${ }^{\text {a }}$ Annual effort for $25 \%$ qualified landings
    ${ }^{\mathrm{b}}$ Median CPUE for $25 \%$ qualified landings

[^1]:    ${ }^{\text {a }}$ Annual effort for $25 \%$ qualified landings.
    ${ }^{\mathrm{b}}$ Median CPUE for $25 \%$ qualified landings

[^2]:    ${ }^{\text {a }}$ Annual effort for 25\% qualified landings.
    ${ }^{\text {c }}$ Median CPUE for $25 \%$ qualified landings

[^3]:    ${ }^{\text {a }}$ Annual effort for $25 \%$ qualified landings.
    ${ }^{\mathrm{b}}$ Median CPUE for $25 \%$ qualified landings

