Geochemistry of snow around the copper smelter at Rouyn-Noranda, Quebec: comparison of 1998 and 2001 surveys

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Abstract: Two snow surveys were conducted (1998 and 2001) in the region surrounding the copper smelter at Rouyn-Noranda, Quebec.

Total loading rates of metals per year (ng/cm²/a) were determined for a large suite of elements, of which 13 (Cu, Pb, Zn, Cd, As, Sb, S, Ag, Ni, Al, Mg, Fe, Mn) are reported on here. The spatial distributions of loading rates of smelter-derived metals from both survey years show a bull's-eye pattern centred on the smelter, skewed northeast and southeast of the smelter as a consequence of the prevailing wind directions. Most element patterns can be divided into two parts, a proximal part close to the smelter with high loading rates dominated by deposition of smelter-emitted metals and a distal part in which loading rates approach an ambient background level and metals are predominantly from non-smelter sources. The radius of the area obviously affected by metal emissions is usually about 50 km. The differences in deposition rates for smelter-derived metals (Cu, Pb, Zn, As, Cd) between the two sampling years may be explained in part by changes in reported emissions between 1998 and 2001.

All samples were thawed and filtered. Dissolved and particulate fractions were analyzed separately. The proportion of total metal in dissolved form provides an indication of potential bioavailability. It differs among elements, between years for the same element (due to changes in filter size), and, for some elements, with distance from the smelter.

Résumé : Deux études portant sur la neige ont été respectivement menées en 1998 et en 2001 dans la région de la fonderie de Rouyn-Noranda (Québec).

On a calculé les taux de charge totaux annuels (ng/cm²/a) d'un grand nombre d'éléments, dont 13 (Cu, Pb, Zn, Cd, As, Sb, S, Ag, Ni, Al, Mg, Fe et Mn) font l'objet du présent article. Du point de vue spatial, les taux de charge en métaux émis par la fonderie et établis pendant les deux études susmentionnées sont répartis concentriquement depuis la fonderie, mais en biais vers le nord-est et le sud-est en raison des vents prédominants. La plupart des éléments ont pu être répartis en deux groupes, un à proximité de la fonderie, où les taux de charge élevés se rattachent surtout à des métaux émis par la fonderie, et un autre à une plus grande distance de la fonderie, où les taux de charge se rapprochent des concentrations de fond et se rattachent principalement à des métaux qui ne sont pas issus de la fonderie. Le rayon de la zone où il est évident que des émissions de métaux ont été déposées mesure dans l'ensemble environ 50 km. Les différences de taux de dépôt de métaux émis par la fonderie (Cu, Pb, Zn, As et Cd) relevées entre les deux années d'échantillonnage peuvent s'expliquer en partie par des variations d'émissions rapportées de 1998 à 2001.

Tous les échantillons ont été dégelés et filtrés. Les fractions dissoutes et particulaires ont fait l'objet d'analyses distinctes. La proportion de métaux dissous est un indice de leur biodisponibilité potentielle. Elle varie selon les éléments, d'une année à une autre pour un élément donné (en raison de l'utilisation de filtres de différentes dimensions) et, dans le cas de certains éléments, selon la distance par rapport à la fonderie.

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BACKGROUND

In certain environments, snow is an excellent sampling medium for studies of metals and other substances transported in the atmosphere. In many parts of Canada, precipitation falls as snow during the winter, usually November to April. Barring the occurrence of melting periods, the resulting snowpack records a sequence of precipitation events (wet deposition) and dry atmospheric fallout (dry deposition), thereby providing a sample of atmospheric deposition integrated over the time from first snow until the time of collection. Consequently, chemical and mineralogical investigations of regional snowpack (and ice) are becoming widely used in regions with sufficiently cold winters as a medium to characterize emissions from industrial, urban, and other sources (e.g. Barrie and Vet, 1984; Dick and Peel, 1985; Shewchuk, 1985; Wolff and Peel, 1985; Chan and Lusis, 1986; Philips et al., 1986; Gorzelska, 1989; Grosch and Georgii, 1989; Landsberger et al., 1989; Jickells et al., 1992; Wolff, 1992; Lazareva et al., 1993; Malakov and Senilov, 1993; Hinkley, 1994; Ayras et al., 1995; Niskavaara et al., 1996; Reimann et al., 1996; Viklander, 1996; Hinkley et al., 1997; Viskari et al., 1997; Yakhnin et al., 1997; Gregurek et al., 1998; Hudson and Golding, 1998; Šakalys et al., 1999; Ingersoll, 2000; Kaasik et al., 2000; Rosman et al., 2000; Sherrel et al., 2000; Simonetti et al., 2000a, b). The chemical composition of snow changes both spatially and temporally due to various factors such as the local sources of material emitted to the atmosphere (including anthropogenic and natural sources), weather conditions, topography, forest cover, redistribution by wind, and others (Colbeck, 1981; Wolff, 1992; Reimann et al., 1996). In order to determine the extent of contamination of the environment around a point source, it is necessary to determine the ambient background of atmospheric fallout (i.e. the loading rate in the absence of the smelter) in the study area. Once this ambient background is established, the spatial patterns of chemical elements deposited in snow (corrected for ambient background) provide information about the influence of the point source, the distance of transport of emissions deposited during the winter months, the loading rates of metals with distance and direction from the smelter, and the availability of smelter-derived metals in readily soluble form.

As part of the Geological Survey of Canada's Metals in the Environment (GSC MITE) initiative, two snow surveys (1998 and 2001) were completed around the Horne smelter, a copper smelter that has been operating since 1927 in the Quebec town of Rouyn-Noranda (Fig. 1). Their purpose was to characterize the chemical footprint of smelter emissions transported by the atmosphere. The winter 1998 snow survey successfully defined the size of the footprint around the Horne smelter and provided a comprehensive picture of metal distribution in the snow — how metal levels change with distance from the source and what processes controlled metal deposition (Telmer et al., 2004). The 1998 sampling distribution was restricted mainly to samples along three radial traverses (northwest, northeast, and south), and the farthest sample was 50 km from the smelter. This left some uncertainties about the spatial representivity of the survey. Would a second survey provide a similar pattern? Is 50 km a great enough distance to reach ambient background levels? If samples were not restricted to radial traverses, would the radial deposition pattern observed in the 1998 survey change shape?

In addition to changes in the sampling pattern, some minor changes were made to the site selection and laboratory protocol used in 1998. The 2001 sites were restricted mainly to locations on frozen lake surfaces to avoid the possibility of contamination from the soil surface and 'wicking' of metals from the ground upward into the snow by capillary action. The 1998 samples were thawed and filtered using 0.45 μm filter paper, whereas the filter size for the 2001 samples was 0.1 μm . Thus the proportion of 'soluble' to 'particulate' fractions differs to some degree between the two years due to this change in filter size, and this must be considered in the interpretation.

MATERIALS AND METHODS

Regional snowpack

Two snow surveys (1998, 2001) involved collecting 160 samples over approximately 75 000 km² of the area surrounding the Horne smelter (Fig. 1). The sample density was higher close to the smelter and decreased with increasing radial distance from Rouyn-Noranda. All samples were collected before mid-March to avoid the spring melt. Ideally, sampling in mid-March in this region of Quebec provides a measure of total wintertime accumulation. Details of both surveys have been documented in Kliza et al. (2000, 2002).

In 1998, snow was collected between March 10 and 12 within a 50 km radial area around the smelter (Fig. 1). A total of 82 sites were sampled and at every tenth site, a field duplicate was collected, giving a total of 93 samples. Where practical, sampling was done on a 9 km² grid close to the smelter, and along three radial traverses, with a sample spacing of about 3 km. Samples were collected by helicopter or truck. Sites were not restricted to lake surfaces or type of location, except that sample locations were always at least 25 m from a road.

The snowpack in the 1998 survey incorporates about 129 days of atmospheric metal accumulation from November 5 to March 12. The mean snow depth was 81 cm, and the snow was typically dry and heterogeneous, and composed of two types of snow. Generally, the top section was made up of soft 'fluffy' snow whereas the bottom section was composed of hard granular snow crystals. A total of 183.7 mm of snow precipitation was reported at the Rouyn-Noranda airport weather station, with rare days of wet precipitation mixed with snow not accumulating more than 2 mm. The wind direction for these four winter months was dominantly to the south and southeast as demonstrated by the wind rose for that time period (Fig. 2).

The 2001 snow survey used a different radial sampling scheme designed to fill in some gaps in the 1998 coverage, and in particular to provide information about the area farther than 50 km from the smelter. A total of 57 sites were sampled, out to a distance of approximately 275 km from Rouyn-Noranda (Fig. 1). The samples were concentrated in

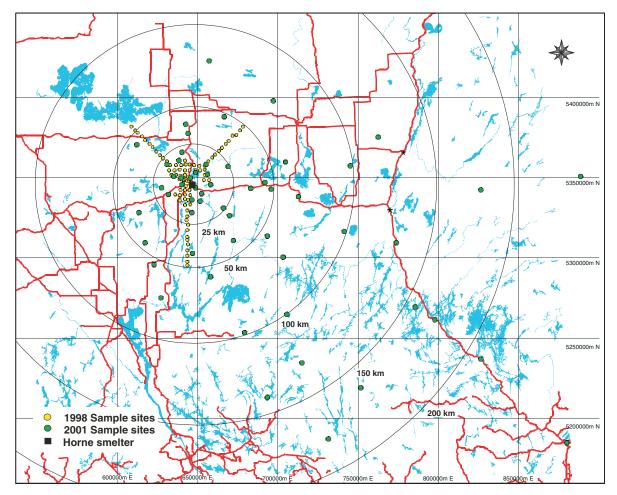


Figure 1. Location map showing sample sites for 1998 and 2001 surveys. Note that the 1998 survey was restricted to a maximum distance of 50 km from the Horne smelter and concentrated on three radial traverses, whereas the 2001 survey extended to 275 km from the smelter and provided more coverage east and southeast of the smelter.

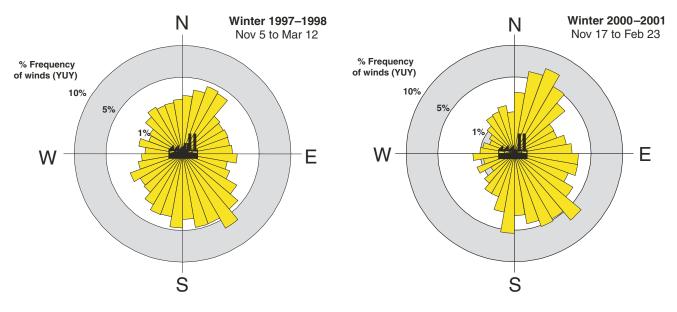


Figure 2. Wind roses for winter 1997–1998 and 2000–2001. The length of each 10° sector shows the proportion of time that wind was blowing toward that direction. The roses were drawn using the program by Baas (2000). Data from Rouyn-Noranda airport (YUY).

the southeast quadrant in the dominant wind direction for the winter months, as established from historical wind records. Most samples were collected by helicopter, using preselected sites on the surfaces of relatively small frozen lakes, thus eliminating contamination by soil and minimizing effects of ground vegetation. Any slush existing at the lake—snow interface was excluded. At most stations, samples were taken about 10 m from the landing point in undisturbed, relatively clean snow. At 11 of the 57 sites, a duplicate sample was taken, giving a total of 67 samples.

The 2001 snow was dry and light and visually homogenous. Pack depth varied from 19 to 96 cm, averaging 64 cm depth. A total of 142.5 mm of snow was reported at the Rouyn-Noranda airport. Snowpack samples represented three months (about 99 days) of snow accumulation. For this period (November 17 to February 23), temperatures remained at or below freezing with no melting events. Figure 2 illustrates that the dominant winds were from the northwest (37.8 %) and southwest (33.5%).

Sample handling and processing methods

Procedures for sampling snow are described in detail in Kliza et al. (2000, 2002). At each sample site, snow depth was measured with a plastic measuring stick. Samples were collected using an 8.26 cm outer diameter, clear Lexan (polycarbonate) tube approximately 1 m in length. The inner diameter of the tubes changed slightly between the two surveys (1998 = 8.9)cm, 2001 = 8.8 cm), and this difference was considered in computing deposition loading rate per unit area of ground. The snow tube was cleaned at each site by dipping it into the snowpack three times to remove any contaminating dust or moisture. The tube was then pressed through the snow to the ice of the frozen lake surface, forcing the snow up into the tube. For samples taken on a soil surface, the tube was stopped approximately 3 cm from the surface, the depth predetermined by a preliminary probe. A polycarbonate plate was used to plug the bottom of the core tube so as to ensure that no sample was lost during withdrawal. Each snow sample was put into a clear, 0.1 mm thick polyethylene bag measuring 30 cm by 46 cm. The plastic bag was then folded over, rolled, and sealed with white plastic electrical ties to make the bag airtight and prevent contamination during storage. Sample bags were packed around with snow in Styrofoam coolers for transport to the laboratory and then transferred to a freezer unit.

All processing was carried out in a 'Class 100' clean room at the Geological Survey of Canada in Ottawa. Samples were brought out to thaw approximately 24 h before the day of preparation. The outside of each sample bag was thoroughly washed with deionized water and checked for any punctures or cracks. Samples were weighed to determine the total mass of snow. Each sample bag was placed into another unused bag in order to contain the snow meltwater in case the sample bag leaked. To begin melting, samples were placed on a countertop until slushy (approximately 6 h). Partly melted samples were subsequently moved to a refrigerator to continue melting at temperatures of approximately 4°C for 12 h.

Sample preparation involved the separation of dissolved and suspended particulate matter using a vacuum filter apparatus. The 1998 samples were filtered through preweighed Durapore® membrane filters at the conventional boundary (0.45 µm) for dissolved material. The 2001 samples were filtered through stacked 0.45 and 0.1 um Durapore® membrane filters to ensure the collection of fine atmospheric dust, metals bound in colloids, metal organic complexes, inorganic complexes, and free inorganic ions. The membranes were stacked to prevent clogging of the 0.1 µm filter by straining off the coarse fraction first. The stacked membranes were dried and analyzed in combination, as the sample was too small to analyze the 0.1 to 0.45 µm fraction separately. The volume of filtrate was recorded to provide a measure of the amount of snow in the sample. The filtered meltwater was separated into aliquots. One aliquot was preserved by acidification with 0.4% HNO₃. The filter membranes were placed in a covered Petrie dish to dry, then weighed. Tables 1 and 2 summarize the list of elements, instrumentation, and detection

The filtered meltwater contains the fraction of the deposited material that occurs in dissolved, rather than particulate, form, and therefore provides an estimate of the most readily available metals (potentially important for impact studies). However, the particulate fraction may also contain metals in a relatively available form (particularly in the 0.45 to 2 μm , for example), so the total loading values (combining the two fractions) may give a more reliable picture of metal deposition for use in impact studies. The total load provides an estimate of the total dry and wet deposition of material (including aerosol particles) accumulated since the first snowfall of the winter.

Table 1. List of elements analyzed in meltwater, showing analytical method and detection limit.

Symbol	Element	Method	Units	Detection
,				limit
Ag	Silver	ICP-MS	ppb	0.05 ₁ , 0.005 ₂
Al	Aluminium	ICP-MS	ppb	2 _{1,2}
As	Arsenic	ICP-MS	ppb	0.1 _{1,2}
Ca	Calcium	ICP-ES	ppm	0.02
Cd	Cadmium	ICP-MS	ppb	$0.05_{1}, 0.02_{2}$
Cr	Chromium	ICP-MS	ppb	0.1 _{1.2}
Cu	Copper	ICP-MS	ppb	0.1,2
Fe	Iron	ICP-MS	ppb	5 _{1,2}
Mg	Magnesium	ICP-ES	ppm	0.005 _{1.2}
Mn	Manganese	ICP-MS	ppb	0.1 _{1,2}
Na	Sodium	ICP-ES	ppb	50 _{1,2}
Ni	Nickel	ICP-MS	ppb	0.2 _{1.2}
Pb	Lead	ICP-MS	ppb	0.01 _{1.2}
S	Sulphur	ICP-ES	ppm	$0.15_1, 0.05_2$
Sb	Antimony	ICP-MS	ppb	0.01 _{1.2}
Zn	Zinc	ICP-MS	ppb	0.5 _{1,2}
1				

¹1998 snow samples (<0.45 μm)

 $^{^2}$ 2001 snow samples (<0.1 μ m)

Table 2. List of elements analyzed in particulate fraction, showing analytical method and detection limit.

Symbol	Element	Method	Units	Detection
				limit
Ag	Silver	ICP-MS	ppb	0.001 ₁ , 0.0001 ₂
Al	Aluminium	ICP-MS	ppb	0.04 ₁ , 0.1 ₂
As	Arsenic	Hydride-ICP-MS ₁ ,	ppb	0.004 ₁ , 0.2 ₂
		ICP-MS ₂		
Ca	Calcium	ICP-ES	ppb	0.4 _{1.2}
Cd	Cadmium	ICP-MS	ppb	0.001 ₁ , 0.0004 ₂
Cr	Chromium	ICP-ES	ppb	0.02 _{1.2}
Cu	Copper	ICP-MS	ppb	0.002 _{1,2}
Fe	Iron	ICP-ES	ppb	0.1 _{1,2}
Mg	Magnesium	ICP-ES	ppb	0.1 _{1,2}
Mn	Manganese	ICP-MS	ppb	0.002 _{1.2}
Na	Sodium	ICP-ES	ppb	1 _{1,2}
Ni	Nickel	ICP-MS	ppb	0.004 _{1.2}
Pb	Lead	ICP-MS	ppb	0.0002 _{1.2}
S	Sulphur	ICP-ES	ppb	1 _{1,2}
Sb	Antimony	ICP-MS	ppb	$0.0002_1, 0.002_2$
Zn	Zinc	ICP-MS	ppb	0.01 _{1,2}
¹ 1998 snow	samples (> 0.4	l5 μm)		
² 2001 snow	samples (> 0.1	um)		

Laboratory methods

Table 1 lists the elements determined by inductively coupled plasma (ICP) emission and mass spectrometric methods on filtered meltwater, after preservation with 0.4% nitric acid. Detection limits for some elements differ a little between the two years. Table 2 lists the elements determined by the same methods, after dissolution of the filtered particulates with a hydrofluoric acid-aqua regia mix. Again, detection limits for some elements differ between survey years. Dissolved organic carbon values were determined for the filtered but unpreserved aliquots and pH was determined on unfiltered samples. These results are not shown here, but are available in Kliza et al. (2000, 2002).

Concentration data (ppb) were recalculated to total mass of metal deposited per unit area of ground per year (ng/cm²/a) using the sampling tube cross-section area and sample volume and annualized assuming that the deposition rate is constant over a year at each location. In this way, the two years can be compared even though the period of deposition represented by the samples, and the tube cross-sections, differ between years. Combined total-load data are given for each year (Table 3, 4), with a '% soluble' field following each element field to indicate the proportion of the total that is either less than 0.45 μm (1998) or less than 0.01 μm (2001).

RESULTS AND DISCUSSION

The results are summarized for four aspects of the work, as follows: 1) maps showing spatial patterns observed in the metal-loading data; 2) summary statistics from fitting models to the data for both years; 3) comparison of metal concentration values with those reported in the literature and their classification according to a published scheme; and 4) information on metal solubility and how this changes with distance from the smelter.

Spatial and multivariate patterns

Figure 3 shows a series of maps for the 1998 and 2001 surveys by element, with the sample locations superimposed. Spatial interpolation was carried out using an inverse-distance weighting method (an option in Vertical MapperTM; an add-on program to MapInfo® GIS; see Northwood Geoscience, 1999, p. 24–28), to convert the element values known at the sample locations to a continuous geochemical surface. These surfaces were then colour-coded using a colour ramp ranging from blue (low values) to red (high values), controlled by percentiles of deposition rate for each element, as shown in Figure 3. Thus, the 90th percentile is shown approximately by the transition from yellow to green and the 50th percentile is approximately at the blue-green transition. Although the interpolation error on these maps differs from place to place depending mainly on the local sample density, the maps give a graphical representation of the main spatial patterns present in the data. Dot plot maps (requiring no spatial interpolation) are also available in Kliza et al. (2000, 2002).

Although a radial bull's-eye pattern centred on Rouyn-Noranda is prominent for most elements (the top 10% of the data loading rates usually occur within 10 km of the smelter), some exceptions exist. For example, many elements have elevated rates about 40 km east of the smelter where a sample was taken close to the main east-west road (*see* for example, Zn values in the 2001 survey in Figure 3). This sample was taken close to a road and was almost certainly contaminated by road dust. The interpolation procedure produced a large 'high' on the geochemical surface simply because there were no other nearby samples to constrain the estimation procedure. Had interpolation been carried out by kriging, the kriging variance in this area would have indicated great uncertainty, because the interpolated values were affected by a single point only.

The effect of prevailing wind direction on these maps is not as great as might be expected. There is some suggestion on certain maps (Sb for example) of a southeastward elongation of the contours. However, this is uncertain and may be controlled at least in part by the uneven sampling distribution pattern

As described by Telmer et al. (2004), three spatial patterns describe the distribution of elements in snow around the Horne smelter. First, a strong to moderate smelter-centric pattern exists for some metals that normally occur in low concentrations in snow (e.g. Pb, Cu, As) and are known to be at significant levels in the smelter emissions. A second pattern presents little or no relationship to the smelter, but is governed more by other sources of variation such as weathering of rocks, dust, road dust, urban emissions, highway emissions, and others (e.g. Al, Mg). A third pattern represents a mixture of both factors. This inference is supported partly by single element patterns, but also by multi-element patterns derived by multivariate statistical analysis.

A principal-components analysis of 13 elements in the 2001 data (log transformed to stabilize the variance) shows that the first two principal components define two major factors (Fig. 4). Samples can be considered in terms of mixtures

Table 3. Summary of analytical results for 1998 samples, expressed as total (soluble + particulate) loading rate per year. % soluble column allows data to be broken down into soluble and particulate fractions.

Sample	Distance from smelter	Ag 1998		1898		As 1998		Cd 1998		Cu 988	Fe 1998		Mg 1998		Mn 1998		IN 1998	- #	Pb 1998	- 5	S 866	Sb 1998	_	Zn 1998	
Units	km	ng/cm²/a % solubl	ng/cm²/a ble		% ng/c	ng/cm²/a % soluble	ng/cm ²	a %	ng/cm ²	la % soluble	ng/cm²/a sc	% oluble	ng/cm²/a so	% ng/ luble	ng/cm²/a % solubl	ng/cm²/a	n²/a % solubi	ng/cm ²	a % soluble	ng/cm²/a	a % soluble	ng/cm²/a	% oluble	ng/cm²/a	% soluble
RN98-08	1.1	252.20 0.4	1 316	165 2.	5.				8449		69173	0.5		5.4						63017	55.6	896.3	1.3	20172	34.0
RN98-09	4.1	130.32 0.8					38.9	9 56.7	22998	3 38.8	26242	0.4	1734	24.6	199 31.1		411.7 16.5	8992	45.8	10213	32.4	385.5	10.9	5081	27.1
HN98-03	2.3	42.96 11.2 107.36 9.6		4447 4. 6870 7.	7.4	2697 71.1 4335 43.2			3462		10666	2.2 0.9		87.9 6.0						30596	76.1	188.6	36.8	4045 11519	52.4
RN98-02	2.3								2243	-	18043	9.0		9.0						21726	58.0	498.3	18.3	8381	39.5
RN98-07	2.4	76.33 10.5		_		1277 51.8			1742.		7469	3.2		8.6						8228	70.5	224.9	15.9	2702	55.5
RN98-01	2.4	55.48 15.5							945		7411	1.1		0.63						5138	64.5	275.7	17.1	2462	53.6
HN98-05	0.0	60.31 12.9		9/0/ 5	0.0	1693 60.0			184		9858	7.7		51.6						47760	7.00	7.867	9.8	3046	24.04 10.04.03
RN98-35	3.0								842			15.9		6.1						15459	92.0	139.2	27.0	1712	69.6
RN98-34	3.5								1209			2.0		0.4						19119	88.1	611.5	19.1	4737	66.3
RN98-82	4.6	82.45 7.3				884 62.6			1072			5.6		2.7						19036	7.78	412.6	21.8	4084	67.3
RN98-79	4.7	35.89 21.5		4265 4.	4.0	936 58.5			685			5.0		90.0						10287	88.4	167.2	23.3	2288	64.1
HN98-06	ъ. т. ъ. с								748			J.5		9.13						9064	82.58 03.53 03.03	1/2.1	d. F2 d. 6	1302	53.9
RN98-33	6.1					383 79.5			203(1.1		5.7						8337	8.06	37.5	34.5	712	71.0
RN98-78	6.2				5.8				347			10.3		4.6						11604	93.5	138.4	25.5	1438	8.79
RN98-80	7.5	22.66 15.2				534 75.1			393.			10.4		1.9						9029	6.06	0.66	32.3	1303	6.69
RN98-52	7.5								340			3.9		38.1						13813	89.2	150.1	20.7	1663	62.9
HN98-27	7.6	22.46 27.0		3791 5.	5.1	788 77.4			502			14.4		0.4						15603	93.0	146.8	28.4	1626	63.7
BN98-32	. 60				-				1548			11.1		0 00						17811	95.6	32.0	37.8	659	76.2
RN98-77	8.7	6.30 17.5				206 80.6			127;			4.6		7.7						17013	2.96	33.9	36.3	712	4.7
RN98-53	0.6								213			2.4		11.8						24404	53.4	95.9	27.1	3908	18.2
RN98-28	9.5	6.18 17.9		3078 4.	6.4	250 78.4			000			10.8		18.4						13876	95.0	29.5	36.1	479	68.2
RN98-76	9.0								97			15.1		4.04						6610	90.00 0.00	9.08	0.75	476	0.07
RN98-74	10.3					405 79.0			246			15.1		9.6						12879	95.6	61.0	43.6	1043	78.4
RN98-54	10.4								87			5.0		7.2						11808	93.3	27.2	33.4	468	59.1
RN98-73	10.5								227.			6.6		38.4						21139	95.2	88.8	41.7	1285	71.5
RN98-56	10.8			2664 3.	3.6	172 76.6			62			9.6		33.6						13162	95.1	24.6	41.4	477	68.2
RN98-25	10.9	3.48 31.8				395 /8.4			226			L./L		- ά. α.						10153	9.49 6.60	93.9	38.3	93/ 93/	c:0/
RN98-29	1.1				5.3				86			11.1		8.4						11890	94.2	17.8	42.7	467	0.99
RN98-51	11.3				7.0				123			16.1		9.5						13457	6.96	28.9	45.9	754	83.4
RN98-24	1.3	7.74 14.3		2307 5.		381 83.2		86.8	170	7 73.4		13.3		8 8						11059	94.9	42.7 65 p	43.1	737	72.0
RN98-75	11.6				5.0				152(13.7		1.0						17787	97.1	39.7	52.1	886	82.5
RN98-49	12.6				5.2				100			9.4		34.5						10183	86.5	46.8	36.6	733	49.9
RN98-55	12.7				1.1				20			7.6		5.1						3769	87.9	26.1	38.3	346	67.9
RN98-58	13.2	3.91 28.3		2290 5. 1982 5.	5.5	121 74.0			85 G			6.1		.1.5 -1.5						10294	94.4 6.15	20.8	34.7	354	7.0.7
RN98-30	13.2				6.1				58			7.5		6.9						15254	97.1	11.5	42.1	283	69.5
RN98-72	13.3				4.7				134			13.5		14.7						9352	94.6	43.9	38.1	672	70.1
RN98-26	13.3			1333 3.	8.3				31			4.6		40.4						3584	92.5	9.1	21.3	129	54.5
RN98-48	14.3	3.20 34.6			5.6	118 79.1			118			7.1		0.0						8912	95.4	20.3	36.4	315	67.4
RN98-22	15.0				6.3				50			18.6		8.8						7689	92.8	9.7	43.9	225	67.5
RN98-70	16.8				4.7	170 83.6			69			13.4		6.4						9601	94.9	25.5	55.3	447	76.2
RN98-21	17.8				7.5				43			17.0		5.4	55 82.1					10014	96.4	8.3	46.0	229	72.8
HN98-47	17.9	2.79 39.7		2352 2.	0.5				3 3			L.7 7.3		ი.ი. ი.ი.	22 68.5					10539	96.2	14.1	37.3	185	0.19
RN98-20	20.8	51.			. 00	34 74.9			28	7 50.8		17.0	417 4	5.1	46 76.4	Ξ.	1.0 40.1			5604	94.5	4.5	40.5	165	0.69
RN98-46	21.1				5.5			92.4	25	9.69 C		8.1	862 (5.4	500 96.1	7	0.8 40.7	379	67.5	14103	98.1	11.5	41.7	253	79.8
RN98-68	22.4	- 1			5.7	- 1	-	1 81.5	55	7 57.3	- 1	12.9	2326	4.4		2	1.1 21.0	1068	60.1	18108	96.3	21.7	61.3	400	75.1

Sample	Distance	Aa		IA	As		В		Cu	Fe		Ma	2	ء ا			Pb		S		Sp		Zn
	from smelter	1998		1998	1998		1998		1998	1998		1998	16	1998	1998	98	1998		1998		1998		1998
laite	a.	ng/cm²/a %	e e	ng/cm²/a % soluble	ng/cm²/a	n %	ng/cm²/a % solu	% ng/cm²	cm²/a % soluble	ng/cm²/a	- e	ng/cm²/a % soluble	ng/cm²/a	% E	ng/cm²/a	soluble	ng/cm²/a	soluble	ng/cm²/a % solut	ı ə	ng/cm²/a % soluble	_	ng/cm²/a % soluble
																			Ί				
RN98-19	23.3		59.8	1493 6.8	58	75.5			226 51.2	1089 17	17.3	421 56.1	48		12.6	34.9	132	76.2	8903 97.1	Ψ.	4.0 44.2		
RN98-45	23.6	4.37	25.3	2127 4.5	46	75.3	3.5 83.	83.1	166 52.6	1256 8.	80	2886 84.3	945	92.5	11.9	37.1	256	58.2	14916 96.8	89.	9.3 27.	_	151 65
RN98-67	25.5	3.13 3	35.3	2407 3.2	54	78.5	3.7 70		289 54.3	1566 11	11.2	616 53.2	108		18.7	23.6	419	64.0	11079 97.2	2	9.4 50.7		
RN98-18	26.6	2.25 4	49.2	1849 5.7	52	80.1	7.0 74	74.8	102 50.5	1106 17	17.5	1521 84.6	149	93.0	21.5	71.8	80	68.9	3609 91.	89.	3.5 56.7		123 79.3
RN98-44	26.8		59.8	2056 4.5	35	75.8	4.4 75	75.2	159 64.2		Ε.	•	63	77.5	9.8	45.2	195	71.7	11337 97.2	5	5.0 49.9		
RN98-66	28.4	2.88 3	38.4	3232 3.9	130	2.06	8.8 87	87.8	270 59.4	3025 43	43.3	709 48.6	121	84.5	20.9	51.7	466	6.07	15154 97.8	ω.	10.0 61.9		
RN98-43	29.6	1.72 6	64.4	2009 4.8	39	8.92	2.7 83	83.3	140 62.4	1263 8.	80		57	79.2	10.4	42.3	181	63.0	12284 96.9	6.	5.9 48.6		
RN98-17	30.0		79.0		12	71.0	1.7 75	75.0	73 54.1				27	74.3	9.5	47.8	49	61.2	3504 94.6	9.	1.5 54.9		
RN98-65	32.5		47.4		103	92.7			196 60.4		2.		71	72.9	19.8	49.8	305	66.5	9594 96.9	6.	6.7 62.5		
RN98-42	32.6	1.73 6	64.0	2543 3.3	4	80.7	2.7 86.				9.	397 27.9	50	71.7	11.1	39.7	167	67.5	15339 98.0	0.	5.0 42.		
RN98-16	32.9		78.5	1757 2.5	12	75.2		83.2	65 41.7	1051 10	10.5	697 72.0	68	84.7	9.4	46.7	51	62.0	3548 93.4	4.	1.3 55.5		63 70
RN98-64	34.5	2.42	45.7	3374 4.9	09	84.8	6.0 83		201 60.9		6.		366	88.7	16.2	27.3	373	61.4	12421 95.3	ω.	9.0 66.		
RN98-41	35.4		45.6	2431 5.9	64	71.1		75.9	136 52.4	1602 6.	o.	865 60.7	719	93.3	15.8	28.0	299	58.5	13928 96.8	89.	7.6 43.5		
RN98-15	36.4	1.35 83	82.2		18	48.6	1.5 74.	74.9	80 59.0				31	79.3	11.1	39.9	9/	83.3	3496 94.8	ω.			
RN98-63	37.2		55.6	2826 3.3	36	82.6			_		Ψ.		68	79.2	14.2	31.0	252	72.7	12233 97.6	9.	5.1 64.8		
RN98-14	39.0	1.35 8	81.9	1952 4.6	16	81.3		80.5	62 56.9		ci.		42	76.2	11.9	37.0	71	74.2	3582 92.5	5.	1.6 59.		
RN98-40	39.1		73.6	1907 4.9	37	87.1	3.3 33.	33.1	58 52.1	_	ci.		57	79.7	9.1	48.7	102	9.07	11717 97.8	89.	2.6 51.9		
RN98-39	40.9		67.9	1664 7.3	40	87.4			76 69.2		.3		106	90.2	9.0	49.2	143	9.08		.5	3.8 62.2		129 86
RN98-62	40.9		62.0		38	86.2	3.0 67.	6.79			.7	430 51.5	46	77.4	12.2	36.3	200	80.9		ı.	4.3 61.9		
RN98-13	42.7	1.31	84.2	1636 6.0	13	81.7			48 57.9		1.7	628 69.9	273	92.8	11.8	37.5	09	74.5	7233 96.5	i.	1.4 61.1		
RN98-61	43.7	2.10 5	52.6	3781 10.7	39	80.9				_	Ε.		70	75.0	15.6	28.4	232	74.9		4.	5.5 60.	_	169 81
RN98-38	44.5	1.43 7	77.4	2285 5.6	30	83.2	3.0 87	87.8	80 71.8	1177 9.	4	496 49.9	53	80.0	10.2	43.2	140	6.77	16471 98.5	5.	3.6 56.3		
RN98-12	44.5	1.40 7	6.87	1825 7.5	15	9.62	7.9 34	34.3	71 68.6	1300 24	24.4	518 58.2	103	90.4	13.2	33.4	18	82.4	3558 93.1	Ψ.	1.9 58.3		
RN98-60	47.6	1.68 6	62.9	3558 3.0	36	84.2	4.0 76		93 63.2		6.3		250	90.5	13.0	33.9	182	72.5	13233 98.1	Ψ.	4.8 66.3		111 80.9
RN98-37	47.6	_	84.6		17	84.4	1.5 73.		_	911 12	12.1	650 73.0	109	91.7	7.9	55.9	72	86.5	7805 96.5	ıci.	1.8 63.4		
RN98-11	47.8	1.34 8:	82.8	1509 10.2	10	75.0	1.5 75.	75.8	71 68.6		23.2		149	93.2	11.5	38.5	98	89.3	3598 92.	Ψ.	2.1 62.5		
RN98-59	50.0		66.1	3218 3.3	35	80.2	4.6 91	91.2	103 47.6	1977 13	13.4	1227 67.8	296	96.2	10.4	42.5	158	66.3	10874 97.	Ψ.	4.2 47.9		

of two end members — a smelter end member and a geological end member. The elements lie along a general trend between these two extremes, as shown on a plot of PC-1 versus PC-2 (Fig. 4). For example, As, Sb, Cu, Ag, Pb, and Cd lie close to the smelter end member, whereas Al, Mg, Mn, and Fe are associated with the geological end member. Sulphur, zinc, and nickel lie somewhere between these two groups, with some influence from both factors. Sulphur appears to lie off the main trend, possibly indicating a somewhat different chemical behaviour. This interpretation is similar to the one made by Telmer et al. (2004) on the basis of the 1998 snow survey and confirms that the 2001 data behave comparably to the 1998 data.

Summary statistics

Because of the strong radial pattern in the element maps, two-dimensional plots of each element versus distance from the smelter provide good summaries of the spatial distributions and models fitted to the data values give useful summary statistics about key parameters that help to characterize the distance of transport of emissions from the smelter. For example, the plot for copper (Fig. 5a) shows that Cu values decrease rapidly by three orders of magnitude within a distance of about 40 km from the smelter. Note that the 1998 and 2001 surveys show good agreement where they both overlap (within 50 km from the smelter), but the 2001 values taken beyond 50 km indicate that deposition rates continue to decrease and therefore the ambient background levels from the 1998 data are somewhat biased upward.

Background values shown on the plot and 'distance to background' values are derived from a model fitted by nonlinear least squares to the data (Bonham-Carter and McMartin, 1997; Bonham-Carter and Kettles, 2001). The model satisfies the relationship

$$y = \beta_0 + \beta_1 2^{\frac{-\chi}{\lambda}} + \varepsilon \tag{1}$$

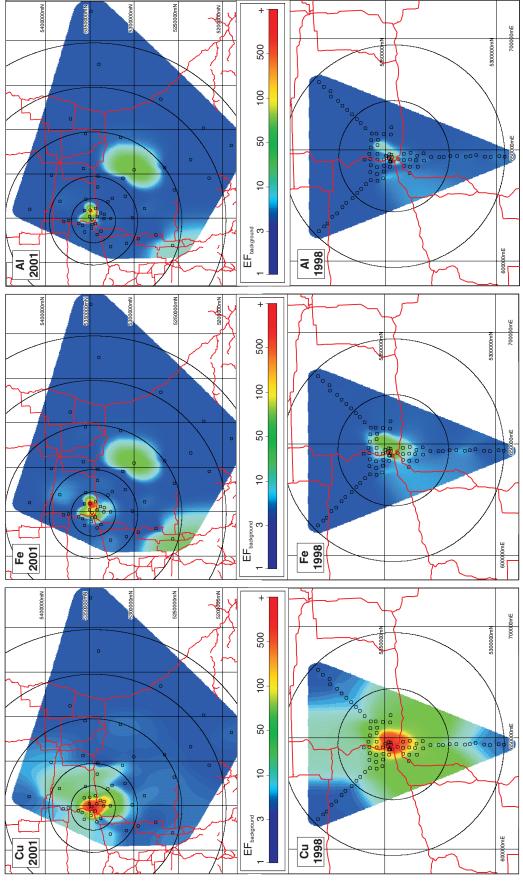
where y is the natural logarithm of the metal deposition rate (µg/cm²/a; please note that Tables 3 and 4 and Figure 3 show metal loading rates in ng/cm²/a, whereas the modelling and statistical results are reported in µg/cm²/a), x is the distance from smelter (km), β_0 is the natural log of the deposition rate at an infinite distance from the smelter, ($\beta_0 + \beta_1$) is the natural log deposition rate very close to the smelter (at x = 0), λ is the 'half distance' at which the natural log deposition rate has dropped to ($\beta_0 + \beta_1$)/2, and ϵ is an error term. By fitting this model to the data, estimates of the three parameters and their standard errors can be obtained. The estimated metal concentration, \hat{c} , is then

$$\hat{c} = \exp(y) = \exp(\beta_0 + \beta_1 2^{\frac{-x}{\lambda}}). \tag{2}$$

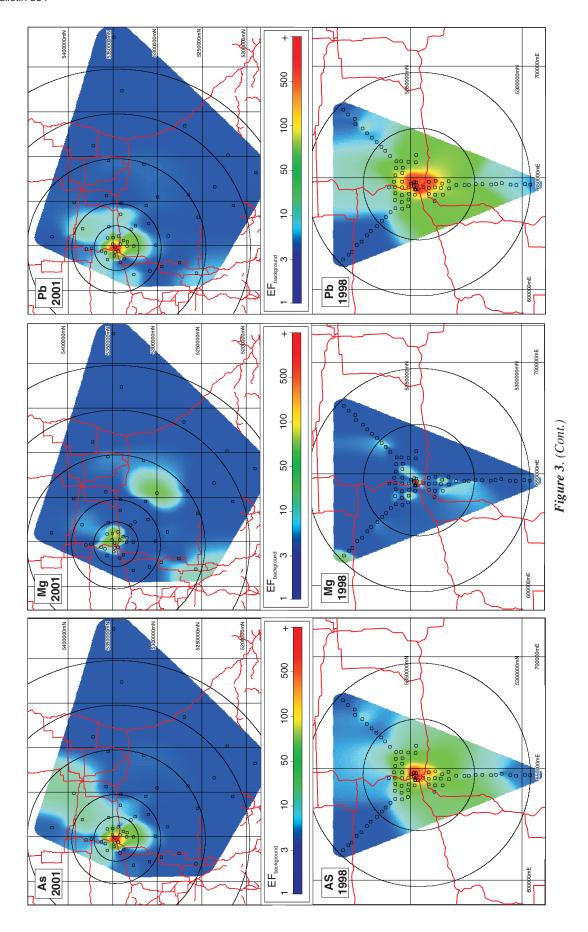
The ambient background deposition rate toward which the model tends with increasing distance from the smelter is $\exp(\beta_0)$ µg/cm²/a. The distance to background can be arbitrarily defined as the distance (x_1) km at which the modelled value reaches the background value plus one standard error of

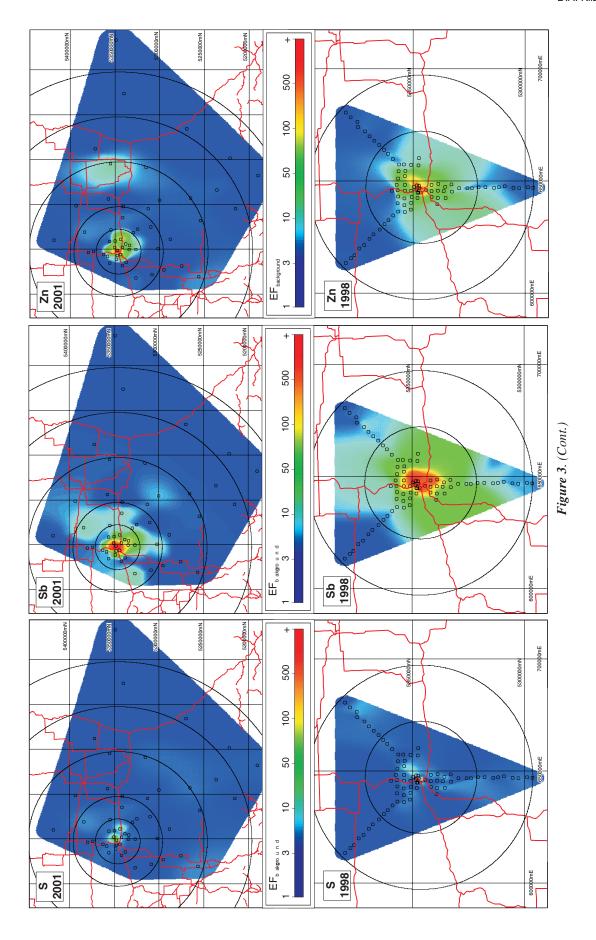
Table 4. Summary of analytical resluts for 2001 samples, expressed as total (soluble + particulate) loading rate per year.

2001 2001
km ng/cm²ta % ng/cm²ta % ng/cm²ta % ng/cm²ta % ng/cm²ta soluble soluble soluble soluble
1.7 12029 6.4 3044 35.1 175.9 90.6
19.6 3917 13.8 855 75.1 40.5 78.5
30.66 14.9 56648 3.4 632 44.3 51.3
2.7 2620 4.7 428 69.5 34.7 71.8
10.90 38.4 2184 22.9 238 84.7 18.8 95.3
12.88 35.9 2258 23.6 420 80.8 32.7 95.2
25.1 1934 15.7 26 80.8 5.9 94.9
2.58 6.1 2605 2.4 21 36.4 17.4 96.4
1 9.25 15.9 2643 15.9 292 83.3 20.0 91.3
42.5 2065 16.6 10 58.3 4.5 93.8 14.7 2102 37.2 42 64.0 5.9 93.5
0.81 44.6 2778 15.1 49 86.8 5.5 92.7
15.1 1652 8.7 9 66.7 1.7 87.8 166 2422 196 22 75.0 32 88.5
3.94 20.6 3005 17.6 154 81.9 18.7 96.1
1.17 31.1 2440 13.5 43 83.9 3.7 86.8
25.4 1939 11.8 21 73.0 2.4 91.5
0.63 25.0 1773 17.5 25 74.4 7.6 67.3
0.54 49.6 1368 20.9 25 56.0 2.9
39.3 Z/5/ 10.7 19 60.8 6.9 95.1 44.2 2194 27.4 30 83.9 5.7 96.6
1.13 15.4 2851 21.7 59 88.2 10.6 69.2
0.33 28.9 1801 15.9 23 83.3 4.1 96.5
46.9 1001 33.5 10 68.7 15.2 E4.1 2602 112 12 64.9 27
10.48 1.6 49675 4.4 37 21.0 60.2 19.1
0.18 30.8 1143 39.7 13 83.6 1.4
0.21 39.6 1954 19.4 5 33.3 3.5
28.0 1674 17.0 17 78.7 4.3
0.41 30.3 2889 12.9 12 60.0 5.3
1.73 24.8 8634 7.5 17 75.6 12.4
30.3 1929 14.6 61 92.4 4.0 36.8 1828 25.5 8 63.0 4.2
0.40 34.6 1765 16.5 17 66.7 3.8
0.38 35.2 33006 3.8 12 56.5 8.2
0.2/ 4/.0 /69/ 9.9 11 52.4 4.9
15.8 2419 14.5 14 74.4 6.3
0.40 33.9 3063 22.3 18 69.7 11.9
65.2 1922 29.1 14 73.2 7.6
0.18 44.2 11.72 20.4 / 50.5 3.6
0.23 62.0 1982 33.2 9 33.3 6.2
42.5 1749 15.0 9 33.3 3.2
0.09 71.5 992 26.5 3 50.0 1.7
0.18 60.2 1416 19.9 6 33.3 3.4
0.24 53.2 10760 8.3 8 33.3 67.8
49.1 1290 33.2 8 63.0 8.4 40.3 2219 25.7 10 60.0 2.9
0.70 51.5 136308 0.9 8 33.3 4.5
0.15 51.6 1274 28.0 5 33.3 3.9
9.3 26798 5.6 8 33.3 13.8



tions to illustrate general patterns. In general, most elements show a strong smelter influence, with minor differences between years. A circular bull's-eye pattern predominates, moderated somewhat by the influence of the wind. A small group of samples taken along a road east of the smelter in the 2001 data set are characterized by high values, probably an effect of road dust. These samples have been removed from some profile plots showing fitted model (Fig. 5). **Figure 3.** Maps showing element deposition rates in ng/cm²/a, coloured by percentile values as indicated in the legend. Values were interpolated from sample loca-





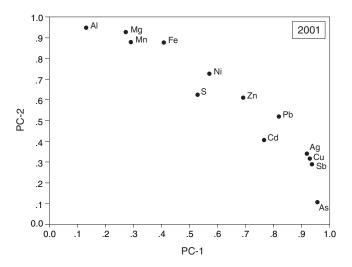


Figure 4. Principal components scores based on an analysis of elements from 2001 snow data. PC-1 is dominated by smelter metals and PC-2, by elements associated with weathering of silicate minerals.

the background (because this is an exponential model, the model value never actually reaches the background value until at an infinite distance from the smelter, but of course it approaches background much closer). The expression for this value is given by

$$x_1 = -\lambda \log_2 \frac{SE(\beta_0)}{\beta_1}$$
 (3)

where $SE(\beta_0)$ is the standard error of β_0 obtained from the fitting process. The value of x_1 is not the final distance travelled by elements from the smelter; rather, it is a relative measure for comparing element behaviour and is the distance beyond which the variability in the background effectively makes it statistically difficult to separate elements from a smelter from other sources of elements in the atmosphere. Undoubtedly a significant proportion of the smelter loading travels farther than this distance, as discussed by Bonham-Carter et al. (2005).

Table 5 summarizes the values of the model parameters generated by fitting equation (1) to the 1998 and 2001 data for each element. Figure 5 provides plots of element-loading versus distance with both years on the same graph, and models for both years superimposed. In both the table and plots, the samples close to the road running east of Rouyn-Noranda and suspected of being contaminated by road dust, have been excluded from the analysis (sample numbers RN01-50, RN01-51, RN01-52, RN01-53, RN01-54, RN01-55, and RN01-56). Table 6 summarizes the background and distance parameters, with the elements in the order of increasing distance of transport, based on 1998 data. A column for each year also shows the median values for samples collected within 50 km of the smelter. A comparison of 1998 and 2001 medians (more than 50 km), background, and distance values is shown as three columns of ratios.

The following paragraphs summarize the results for each element, beginning with the smelter-related elements. Note that in Table 6, the most 'smelter-centric' elements (Cu, Pb, Sb, As, Zn) are at the bottom of the table, because the travel distance (size of smelter footprint) is greatest for these elements. Elements such as Mg, Al, S, Fe, and to some extent Ni, Ag, and Cd have a smaller footprint size and the influence of the smelter gets 'lost' in background at shorter distances from the smelter.

Because the median values within 50 km may be affected by the different spatial distribution of samples between the two years, an inset in Figure 5 shows a plot of the median metal-loading rates for 10 km intervals, starting from the smelter and increasing to 50 km, between 1998 and 2001.

Copper

Perhaps not surprisingly, copper (Fig. 5a) has the largest smelter footprint (approximately 50 km for 1998, approximately 58 km for 2001), consistent with the fact that the Horne is a copper smelter (although lead emissions are much higher than copper emissions — see Table 7). Copper background value is approximately 0.067 μ g/cm²/a (1998) but is lower at approximately 0.044 μ g/cm²/a for 2001. This is because the distal samples taken in the 2001 survey at distances of more than 50 km from the smelter show that the background value is lower than what would be obtained from samples taken only out to 50 km from the smelter (as was the case in 1998). In fact, the data for samples taken within 50 km of the smelter are remarkably similar in both years, as is confirmed on the inset medians plot for Cu.

Lead

Lead (Fig. 5b) has the next largest smelter footprint (approximately 50 km for 1998, approximately 46 km for 2001). Major differences occur between the two years, with 1998 levels being higher than 2001 levels at every distance. For example, the median Pb loading rate for samples closer than 50 km from the smelter is $1.060~\mu g/cm^2/a$ in 1998 and $0.103~\mu g/cm^2/a$ in 2001, or 10 times lower in 2001. The inset medians plot shows that this difference is systematic, with consistently higher 1998 medians over each 10 km interval. The reason for this is uncertain. Lead emissions were 150 tonnes/a in 1998 and 65.3 tonnes/a in 2001, or 2.5 times lower in 2001; however, this difference appears to be insufficient to explain the differences in lead content in snow between the two years.

Antimony

The two Sb curves (Fig. 5c) give similar estimates of background ($0.0022~\mu g/cm^2/a$ vs. $0.0019~\mu g/cm^2/a$) and similar footprint radii (approximately 47 and 41 km), but the median values out to 50 km indicate a systematically larger loading in 1998. Emissions data were not available for this metal, but the reduction in loading rates in values is likely related to a decrease in emissions.

Table 5. Summary of model parameters fitted to element data for the 1998 and 2001 snow surveys.

Element	1	2	3	4	5	6	7	8	9	10
/Year	β_0	SE(β ₀)	β1	SE(β ₁)	λ	SE(λ)	$\exp(\beta_0 + \beta_1)$	exp(β ₀)	X _I	Res
unit					km	km	μg/cm²/a	μg/cm²/a	km	
Cu-1998	-2.699	0.144	6.710	0.180	8.96	0.65	55.2020	0.0673	49.6	0.432
Cu-2001	-3.146	0.197	6.635	0.629	11.46	1.90	33.4817	0.0430	58.2	0.837
Pb-1998	-2.552	0.258	5.935	0.252	10.92	1.42	29.4593	0.0779	49.4	0.601
Pb-2001	-3.547	0.162	5.752	0.648	9.23	1.62	9.0703	0.0288	47.6	0.750
Zn-1998	-2.282	0.137	4.775	0.207	7.76	0.82	12.0975	0.1021	39.7	0.477
Zn-2001	-2.265	0.158	4.801	0.848	6.79	1.62	12.6291	0.1038	33.5	0.809
Cd-1998	-5.894	0.162	3.728	0.231	8.08	1.26	0.1146	0.0028	36.6	0.541
Cd-2001	-5.356	0.116	4.109	0.772	5.39	1.27	0.2874	0.0047	27.7	0.629
As-1998	-3.846	0.173	5.522	0.220	8.83	0.94	5.3441	0.0214	44.1	0.527
As-2001	-4.449	0.169	5.844	0.756	8.27	1.58	4.0350	0.0117	42.3	0.816
Sb-1998	-6.113	0.221	6.020	0.242	9.87	1.15	0.9112	0.0022	47.1	0.587
Sb-2001	-6.266	0.175	6.564	0.820	7.88	1.42	1.3472	0.0019	41.2	0.858
Al-1998	0.792	0.061	1.734	0.272	3.11	0.76	12.5034	2.2078	15.0	0.354
Al-2001	0.713	0.170	1.311	0.740	8.54	7.21	7.5685	2.0401	25.1	0.815
Fe-1998	0.308	0.080	3.379	0.260	3.99	0.52	39.9249	1.3607	21.6	0.420
Fe-2001	0.043	0.198	3.710	0.978	7.45	2.75	42.6488	1.0439	31.5	0.988
Mg-1998	-0.353	0.113	1.180	0.364	4.04	2.14	2.2864	0.7026	13.7	0.594
Mg-2001	-0.563	0.177	2.371	0.636	10.30	4.56	6.0982	0.5695	38.5	0.789
S-1998	2.150	0.143	0.680	0.239	7.16	5.77	16.9455	8.5849	16.1	0.535
S-2001	1.977	0.088	0.977	0.311	10.39	5.50	19.1825	7.2210	36.1	0.389

Column 1. Background parameter (β_0) fitted in equation (1), units are in natural logarithms ($\mu g/cm^2/a$).

Column 2. Standard error of β_0 obtained by fitting, units same as in column 1.

Column 3. Source parameter (β_1) fitted in equation (1), units same as in column 1.

Column 4. Standard error of β_1 obtained by fitting, units same as in column 1.

Column 5. Half-distance parameter (λ) fitted in equation (1).

Column 6. Standard error of λ lobtained by fitting.

Column 7. Loading rate (μ g/cm²/a) close to smelter, i.e. modelled value at distance = 0 km from smelter.

Column 8. Ambient background loading rate ($\mu g/cm^2/a$) far away from smelter, i.e. modelled value at distance = ∞ km from smelter.

Column 9. Distance, x, (km), at which modelled loading rate is within one standard error of ambient background loading rate.

Column 10. Residual standard error, with 79 degrees of freedom (1998) or 47 degrees of freedom (2001) as measure of fit, units same as in column 1.

Table 6. Median values less than 50 km from smelter, ambient background levels, and distance to background (x_j) summary, with elements sorted by increasing distance based on 1998 data.

		1998			2001		Ratio	: 1998 value/200	1 value
Element	Median <50 km	Background exp(β_0)	Distance x ₁	Median <50 km	Background exp(β_0)	Distance x ₁	Median <50 km	Background exp(β ₀)	Distance x_1
	(μg	/cm²/a)	(km)	(μց	_J /cm² /a)	(km)			•
Mg	0.767	0.7026	13.7	0.805	0.5695	35.1	0.95	1.19	0.39
Al	2.583	2.2078	15.0	2.605	2.0401	22.4	0.99	1.06	0.67
S	11.471	8.5849	16.1	10.976	7.2210	31.4	1.05	1.16	0.51
Fe	1.964	1360.7	21.6	1.986	1.0439	29.3	0.99	1.26	0.74
Ni	0.0195	0.0121	24.9	0.0203	-	-	0.96	-	-
Ag	0.0047	0.0016	35.7	0.0011	-	-	4.3	-	-
Cd	0.0088	0.0028	36.6	0.0076	0.047	27.7	1.2	0.60	1.31
Zn	0.457	0.1021	39.7	0.197	0.1038	32.9	2.3	0.97	1.21
As	0.152	0.0214	44.1	0.037	0.0117	42.8	4.1	1.83	1.03
Sb	0.0238	0.0022	47.1	0.0071	0.0019	41.5	3.4	1.15	1.13
Pb	1.006	0.0779	49.4	0.103	0.0288	45.8	9.8	2.60	1.08
Cu	0.661	0.0673	49.6	0.275	0.0430	57.8	2.4	1.52	0.86
*Model did	not converg	e for Ni, Ag for 200	01 or for Mn fo	r either vear.				•	•

Table 7. Summary of annual emissions from the Horne smelter from 1994 to 2001. Data supplied by Noranda, Inc.

	1994	1995	1996	1997	1998	1999	2000	2001
	(tonnes/a)							
Feed	853 551	881 025	934 187	896 863	881 283	829 017	787 696	840 233
SO ₂	154 000	172 000	148 000	144 000	112 500	94 000	90 000	
Particulates	770	940	1200	870	920	720	620	595
Pb	260	340	300	197	150	100	80	65.3
Zn	62	85	85	55	39	23	19	17.8
Cd	3	3	2	1.4	2.4	1.6	2.2	2.5
As	19	29	63	55	78	64	59	97.9
Cu	70	58	65	50	70	69	59	41.7

Arsenic

Median As values (Fig. 5d) within 50 km of the smelter are about four times higher in 1998 than in 2001 (0.152 μ g/cm²/a for 1998 versus 0.037 μ g/cm²/a for 2001), although the footprint radius is about the same for both years (approximately 44 and 43 km). The inset medians plot shows that the higher values in 1998 occur at least up to the 40 to 50 km distance interval. The background estimate is definitely higher for 1998, although only by a factor of 1.8. The emission tonnages for this element actually increased over the sampling period, from 78 tonnes/a in 1998 to 98 tonnes/a in 2001, counter to the trends shown in the snow data. Annual arsenic emissions have likely been insufficiently resolved over time to be sure whether the drop shown in the snow data can be accounted for by the change in emission rates.

Zinc

Zinc values (Fig. 5e) in snow appear to be similar for both sampling years, as shown from the two curves and the background levels. However, the median values for distances less than 50 km are $0.457~\mu g/cm^2/a$ in 1998 and $0.197~\mu g/cm^2/a$ in 2001, a decrease by a factor of more than two. The medians plot shows that the higher 1998 Zn values occur within the first 20 km; beyond this point the differences are small. Notice that the footprint radius is somewhat less in 2001 (drop from approximately 40 km to approximately 33 km). Over the interval 1998 to 2001, zinc emissions decreased from 39 tonnes/a to 18 tonnes/a, consistent with the drop in median values from snow.

Cadmium

The median loading rates for Cd (Fig. 5f) in samples taken within 50 km of the smelter are similar in both sampling years (0.0089 $\mu g/cm^2/a$ in 1998 and 0.0076 $\mu g/cm^2/a$ in 2001), although background levels increased over this time interval from 0.0028 $\mu g/cm^2/a$ to 0.0047 $\mu g/cm^2/a$; this can also be seen in the medians plot. The distal samples have a higher cadmium content in 2001 than would be expected from the 1998 data. Emission figures indicate essentially the same values for both years (2.4 tonnes in 1998, 2.5 tonnes in 2001). Increased background values could be due to an increase in cadmium in other air masses that controlled the ambient background, although no independent evidence exists to support this.

Silver

Median Ag values at less than 50 km from the smelter (not shown in Fig. 5) decreased from $0.0047 \,\mu\text{g/cm}^2/\text{a}$ in 1998 to $0.0011 \,\mu\text{g/cm}^2/\text{a}$ in 2001.

Nickel

Median Ni values (not shown on Fig. 5) remained essentially unchanged in 1998 and 2001.

Iron

Although background values for Fe (Fig. 5g) were lower in 2001 than in 1998, the median value at distances under 50 km from the smelter remained similar in both sampling years.

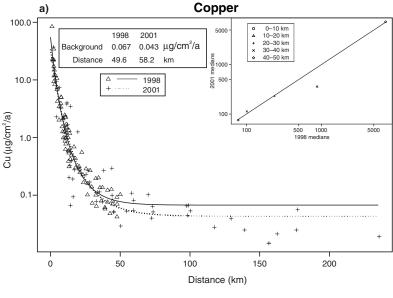
Sulphur

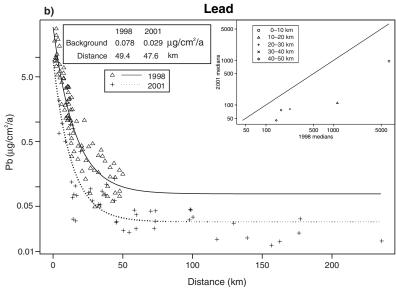
For some reason, S values between 25 and 50 km from the smelter are bimodal in the 1998 data, with modes of about 2 μ g/cm²/a and 15 μ g/cm²/a (Fig. 5h). Sulphur is erratic and its presence is not readily explained in terms of a smelter origin.

Aluminium

The Al data show a similar pattern and level for both 1998 and 2001 (Fig. 5i).

In summary, it is interesting to note (Table 6, final column) that the 1998:2001 footprint ratio is generally larger for the smelter-centric elements than for the non-smelter elements. The only exception is Cu. The loading rates for Cu, Pb, Sb, As, Zn, and Cd are all higher within 50 km of the smelter, and these elements have a somewhat larger footprint radius than Ni, Fe, S, Al, and Mg. Background levels are generally higher in the 1998 data than in the 2001 data (except for Cd and Zn), mainly because sampling was restricted to a 50 km radius in 1998, which proved to be too close to the smelter to obtain an unbiased estimate of background values.





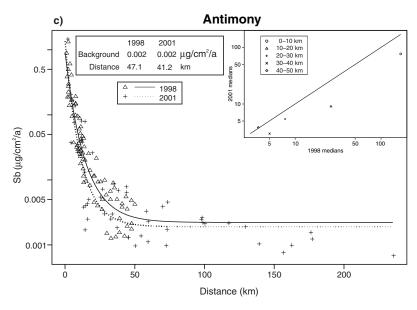
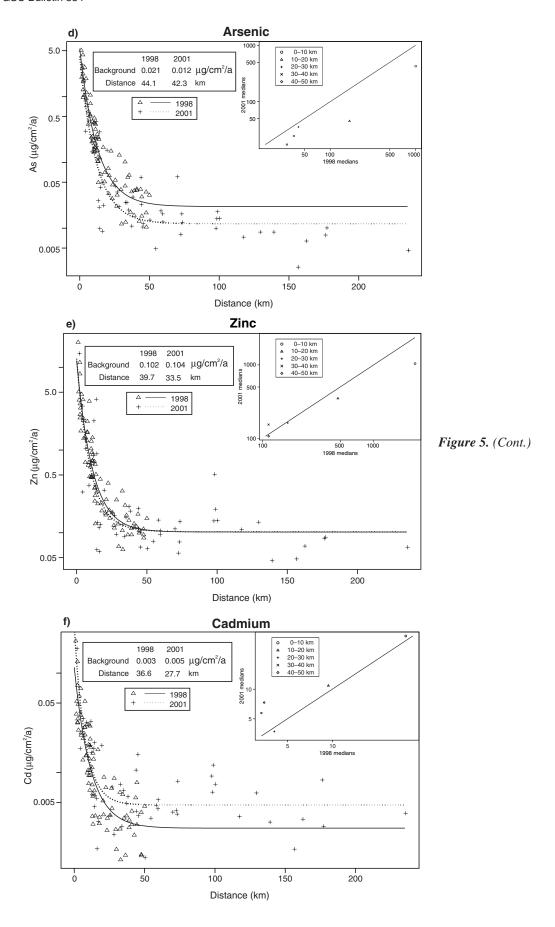
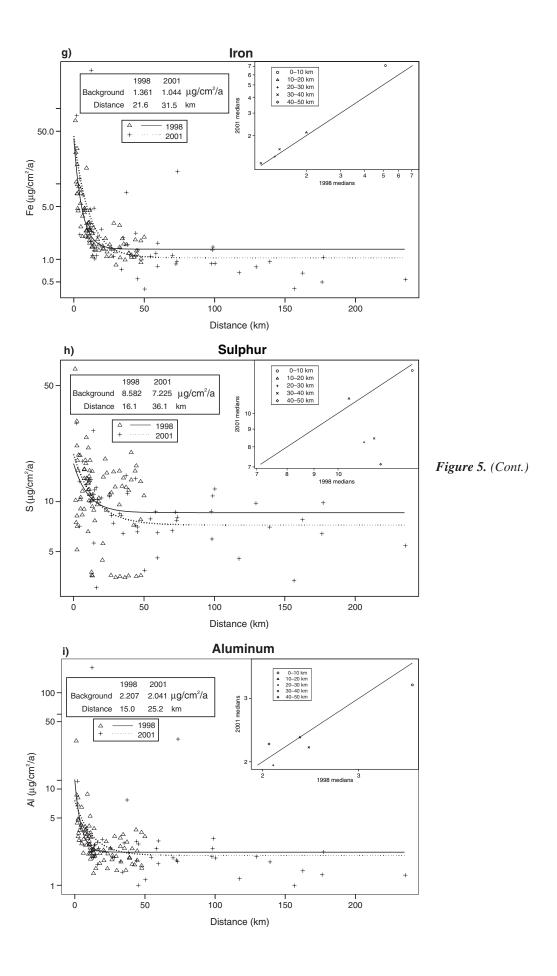


Figure 5.

Plots of metal-loading versus distance for elements in snow. Note the log scale for the y-axis. *The model (equation (1)) fitted to the 1998 data is* shown as a solid line and to the 2001 data, as a dotted line. Loading units are expressed as µg/cm²/a (cf. the $ng/cm^2/a$ units in Tables 3 and 4). 'Background' values are fitted ambient background values of the model at an infinite distance from the smelter. 'Background distance' values (equation (2)) reflect the distance at which the model value first reaches a level within 1 standard error of background. This distance is referred to as the 'footprint radius' and is not to be confused with the limit of transport of the metal. Inset diagrams summarize median metal levels at five distance intervals between 1998 and 2001 sampling years, and facilitate a comparison of the surveys. The straight line is for a 1:1 correlation.





Comparison of Rouyn-Noranda metal concentration values from snow with values reported in the literature

It is instructive to see how the snow chemistry around Rouyn-Noranda compares with results reported from other parts of the world. Table 8 summarizes metal levels from a variety of snow surveys and provides a comparison with values from this study. Note that the comparison is on the basis of concentrations, not loading rates, because of the insufficient number of published loading rates. Metal concentration levels in bulk precipitation were grouped by Galloway et al. (1982) into three classes, i.e. urban, rural, and remote, as shown in Table 8 and Figure 6.

In general, the class ranges of Galloway et al. (1982) overlap. For example, although Cu values can range from 6.8 to 120 ppb in his urban class, rural samples can range from 0.4 to 150 ppb, i.e. some rural values can even exceed the urban range. Similarly, remote samples can contain as much as 0.85 ppb or as little as 0.035 ppb, again demonstrating an overlap with the rural class.

In general, the Rouyn-Noranda snows have metal levels higher than the range typical of remote sites. On the other hand, the range of metal levels around Rouyn-Noranda are for the most part typical of those of the urban and remote classes of Galloway et al. (1982). Copper values within 25 km of the smelter are typical of urban values, except for samples from within about 12 km of the smelter, where levels can be higher than expected for an urban setting. Lead levels out to about 50 km from the smelter are typical of urban values, with most values exceeding the urban maximum, even close to the smelter. Cadmium levels in samples within 20 km of Rouyn-Noranda are also typical of an urban environment; beyond this distance, they are typical of rural environments. Zinc levels exceed the urban range within 10 km of the smelter, but are typical of rural values beyond 50 km.

We conclude that within 10 to 15 km from the smelter, some metal levels (Cu, Zn) in the Rouyn-Noranda surveys exceed values typical of precipitation in urban environments, being more characteristic of industrial sites such as smelters in the Kola Peninsula (Reimann and de Caritat, 1998) and the Belledune smelter in New Brunswick (Pilgrim and Hughes, 1994).

Metal solubility

One of the important questions for studies of metals being deposited around smelters is the proportion of the metal deposited that is bioavailable. Bioavailability is a complex and difficult factor to measure and is beyond the scope of this study. However, it is possible to examine the relative proportions of dissolved and particulate material for each element, because each sample was thawed and filtered, with the filtrate and residue being analyzed separately.

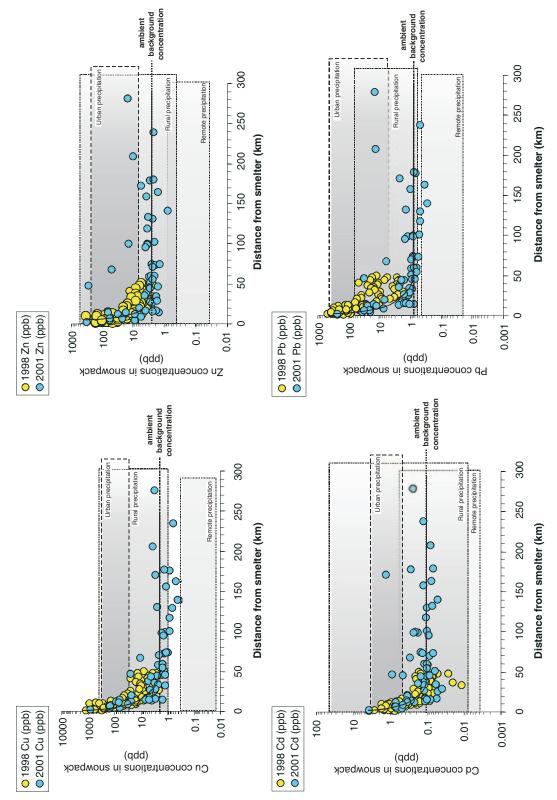
Figure 7 shows graphs of the dissolved/(dissolved + particulate) ratio for 12 elements as a function of distance from the smelter. A novel presentation of the solubility information is made by Telmer et al. (2004), using a more complete suite of elements than the one summarized here. The 2001 data were obtained using a finer filter than in 1998 (0.1 μm instead of 0.45 μm), so, as a general rule, one would expect a higher proportion of material to show up in the filtrate in the 1998 data, and therefore the 'solubility', as the term is used here, would normally be higher for the 1998 samples than the 2001 samples. This does not always occur, however. Median solubility ratios are summarized by year in Table 9.

In general, the solubility–distance curves have several different characteristics, depending on the element.

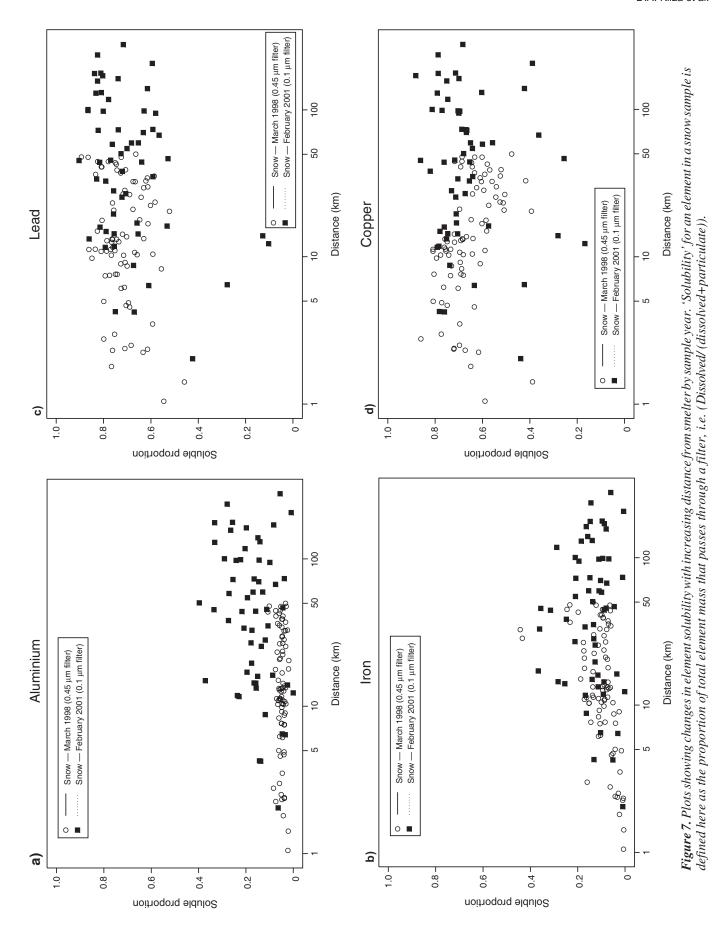
- 1. Elements that are very insoluble, with little change with distance from the smelter. Aluminum (Fig. 7a) is the least soluble element reported here, although for some reason the 2001 samples were often more soluble than the 1998 samples, despite the finer filter used for 2001 samples. Iron (Fig. 7b) is also mostly in the particulate fraction (i.e. very insoluble), with no obvious differences between years. There is a suggestion of solubility increasing with distance from the smelter, but this is not pronounced. Both Fe and Al usually have at least 80% of the total element in particulate form.
- 2. Elements that are moderately soluble, with little systematic change in solubility with increasing distance from the smelter; the solubility is usually within a characteristic range. This includes Ni (not shown), Pb (Fig. 7c), Cu (Fig. 7d), and As (Fig. 7e), with Ni being the least soluble of this group and As the most soluble.
- 3. *Magnesium* (Fig. 7f) behaves differently in that solubility can occur across a wide range of values from about 20 to 90%, with no obvious change with distance from the smelter.
- 4. Elements that become more soluble with distance from the smelter. These include Ag (Fig. 7g), Sb (Fig. 7h), Zn (Fig. 7i), and possibly Mn (Fig. 7j). The 1998 data show Mn solubility changing from about 40% close to the smelter to 90% at 50 km. However, solubility in the 2001 data at distances greater than 50 km appears to decrease again, or at least to become much more variable. For both 1998 and 2001, Ag shows a marked trend, with solubility in 2001 being less than in 1998, particularly at great distances from the smelter. Antimony and zinc (although noisier) show a similar trend.
- 5. Elements that are very soluble, with only a weak change in solubility with distance. Sulphur (Fig. 7k) is the best example of this (a few samples within about 5 km of the smelter have solubilities in the 40 to 80% range, but beyond 10 km, more than 90% of sulphur is in solution). Cadmium (Fig. 7l) also is highly soluble at all distances.

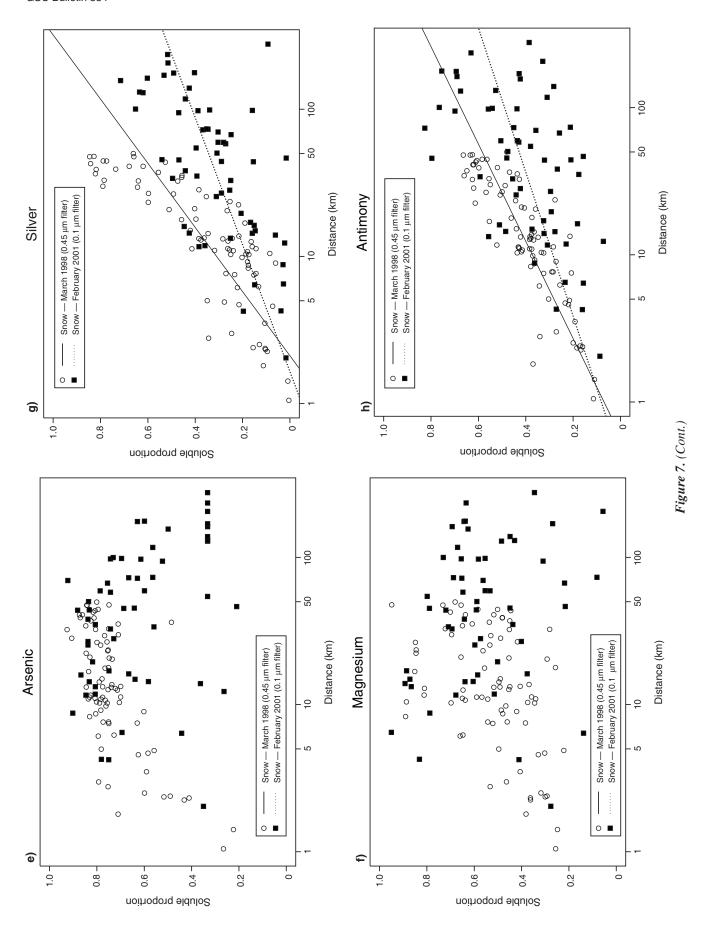
Table 8. Concentration values of metals in snow from various sources, including the classification of Galloway et al. (1982).

Reference	Sample type	Location	Designation	Zn ppb¹	Pb ppb	Cu	iN	Cd
Bulk precipitation								
Galloway et al. (1982)	precipitation (wet &		urban	18–280	5.4–147	6.8–120	2-114	0.48-2.3
Galloway et al. (1982)	dry) precipitation (wet &		rural	(34) 1–311	(44) 0.6–64	(41) 0.4–150	0.6–48	(0.7) 0.008–46
Galloway et al. (1982)	dry) precipitation (wet &		remote	(36) 0.016–0.32	(12) 0.02–0.41	(5.4) 0.035–0.85	(2.4) <dl< td=""><td>(0.5) 0.004–0.639</td></dl<>	(0.5) 0.004–0.639
	dry)			(0.22)	(0.09)	(0.06)	, OL	(0.008)
Snow			Industrial and urban	ban				
this study	snowpack	10 km from Rouyn-Noranda, Quebec	industrial, urban	10.9	7.98	31.5		0.28
Niza et al. (2000) Reimann and De Caritat (1998)	snowpack	To km from Houyn-Noranda, Quebec Kola Peninsula. Russia	industriai, urban industrial	4.6.7	48.87 8.43	40. Io 690.50	853	0.30 315.095
Pilgrim and Hughes (1994)	snowpack	Belledune, New Brunswick	industrial	108.9	648.76	0.70	ç	12.52
GOIZEISKA (1909)	SHOWPACK	ilidvik (average), ivaliavat	ווממאוומו	† - - - -	08.873	9.	7	0.0
this study	snowpack	15 km from Bouyn-Noranda, Quebec	urban	0 0	3.28	11.66		0.17
Niza et al. (2000) Jonasson (1973)	snowpack fresh snow	i o kii iioiii nouyii-noranda, Quebec Ottawa, Ontario	urban	9.6 18–192	55-410	15–69	17–21	0.1-1
		==	-	(25)	(83)	(19)	(19)	(0.7)
Jeffries and Snyder (1981)	snowpack	Sudbury, Ontario	urban	1	0 0	52.00		0 0 0
omoneul et al. (1990)	STOWPACK	outside of Moritreal, Quebec	urban	(18)	(0.59)	(0.92)		(0.084)
			Rural					
this study	snowpack	25 km from Rouyn-Noranda, Quebec	rural	3.2	1.34	3.6		0.12
Kliza et al. (2000)	snowpack	25 km from Rouyn-Noranda, Quebec	rural	4.6	6.99	4.80		0.09
Simonetti et al. (1996)	meltwater snowpack	western Quebec	rural	2.1–5.4	0.52–2.54	0.39-4.01		0.018-0.097
Simonetti et al. (1996) Barrie and Vet (1984)	menwater showpack	Larriotrie (55 km east of Norarida) Val-d'Or region, western Quebec	רעים	4	2.44 1.86–7.15	4.01	09.0	0.92
			İ		(2.11)	(1.27)	(09.0)	
Jonasson (1973)	snowpack	eastern Ontario	rural	10	. 80			
			Remote					
this study	snowpack	>200 km from Rouyn-Noranda, Quebec	remote	2.6	92.0	1.2		0.11
Kliza et al. (2000)	snowpack	>200 km from Rouyn-Noranda, Quebec	remote	2.8	2.13	1.80		0.05
() () () () () () () () () ()	snowpack	Kola Peninsula, Hussia	remote	3.52	0.75	0.34	0.22	0.04
Jickells et al. (1992)	sriowpack fresh fallen snow	Scottish Highlands	remote	0.3–63	0.0	0.4–5	C2-6:0	
)		(4.2)		(0.7)		
Gorzelska (1989)	fresh fallen snow	20 km from Inuvik, Nunavut	remote	0.8	0.63	0.26	0.18	0.012
Gorlach and Boutron (1992)	recent snow	Antarctic	extreme remote	0.0042	0.0054	0.01		0.00031
Davidson et al. (1981)	recent snow	Greenland	extreme remote	0.21	0.14	0.04	0.05	0.011
Boutron et al. (1993)	fresh snow	Dye 3, Greenland	extreme remote	0.016-0.300	0.013-2.7	0.013-2.65		0.0002-0.0145
Wolff and Peel (1985)	fresh snow	Dye 3, Greenland	extreme remote	0.008-0.048	0.005-0.09	0.002-0.015		0.0002-0.0013
DL = detection limit Darks ner hillion (nnh) are the unite equivalent to unit of theward snow	the or male of the state of the	awad enow						
	S cyalvalor in the part of the	2000 C						



Ranges of remote, urban, and rural concentrations in wet deposition based on a world-wide survey by Galloway Figure 6. Concentrations (ppb or µg/L) of Cu, Pb, Zn, and Cd plotted against distance from the smelter (km). et al. (1982) are superimposed.





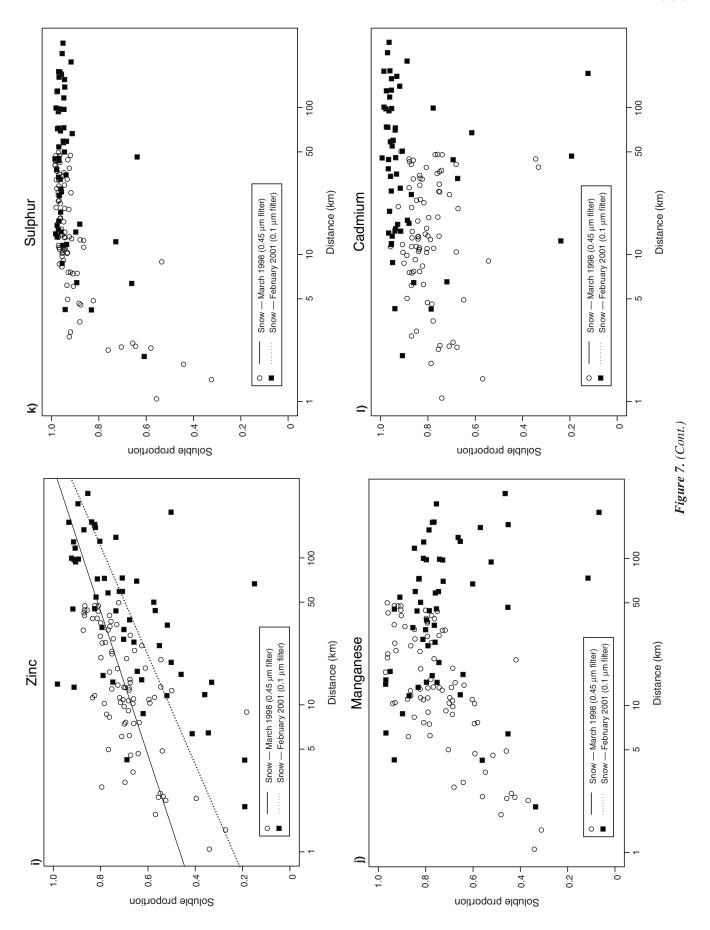


Table 9. Median values of the soluble:total ratio for 1998 and 2001, in order of increasing value, based on 1998 values. Values approaching 1 indicate very high solubility; values approaching 0 indicate very low solubility.

Element	1998	2001
Al	0.05	0.14
Fe	0.01	0.17
Ag	0.23	0.39
Ni	0.35	0.30
Sb	0.32	0.32
Mg	0.41	0.51
Cu	0.42	0.70
Zn	0.40	0.68
Pb	0.78	0.66
As	0.73	0.63
Mn	0.72	0.73
Cd	0.88	0.95
s	0.95	0.94

CONCLUSIONS

In general, the 2001 snow data confirm the conclusions drawn from the data collected in 1998 as discussed by Telmer et al. (2004).

- 1. Metals emitted from the smelter show a large, approximately circular, footprint in snow around the Horne smelter at Rouyn-Noranda. After subtracting the effects of ambient background levels of element deposition from the atmosphere, the influence of the smelter can be recognized reliably only to about 50 km from Rouyn-Noranda. This is not the maximum distance travelled by smelter emissions; rather, it is the distance at which the influence of the smelter emissions, after deposition, can be distinguished reliably from other sources of material in the atmosphere. As shown by Bonham-Carter et al. (2005) from a comparison of loading rates between snow, peat, and soil, the amount of metal in the geochemical anomaly around the smelter is insufficient to account for the known emission tonnages. This suggests that some smelter emissions travel well beyond the obvious smelter footprint, but become so dilute that the influence of the smelter cannot be seen on the ground beyond 50 km.
- 2. Levels of metals in snow around the Rouyn-Noranda smelter as compared to published studies from other parts of the world suggest that metals concentrations are higher than a range of values typical for snow in an urban environment only out to a maximum distance of 15 and 20 km from the smelter.
- 3. Estimates of solubility show different behaviour between elements. Some elements such as S and Cd are highly soluble and their impact on the environment may be greater than less soluble elements such as Fe and Al. Of the toxic metals being emitted by the smelter, about 90% of Cd in

snow is soluble, 65 to 75% of Pb and As are soluble, and Cu is more variable, with annual medians ranging from 42% to 70% soluble.

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