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# DIRECT CONTRIBUTION OF RAIN WATER TO LOADING OF URBAN STORM RUNOFF

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#### ABSTRACT

Rainwater and runoff samples were systematically collected and analyzed for 30 storm events in a light industrial urban catchment. Analysis of the water samples showed that rainwater supplied practically all the dissolved nitrogen compounds in the observed runoff. Rainwater also supplied significant amounts of nickel and copper and a portion of zinc, total phosphorus and total carbon. Rainwater is not an important source of major ions. The antecedent conditions of the preceding period had little linear relationship with rainwater and runoff quality.

## RESUMÉ

Des échantillons d'eau de pluie et de ruissellement ont été recueillis dans un bassin hydrographique d'une zone urbaine d'industrie légère pour 30 événements de tempête et ont été analysés systématiquement. L'analyse des échantillons d'eau a démontré que l'eau de pluie renfermait également une importante quantité de nickel et de cuivre ainsi que du zinc, du phosphore total et du zinc total. L'eau de pluie n'est pas une importante source d'ions majeurs. Les conditions qui avaient prévalu au cours de la période précédente avaient une bien faible relation linéaire avec la qualité de l'eau de pluie et de ruissellement.

#### MANAGEMENT PERSPECTIVE

It is known that urban runoff can contain significant amounts of pollution loads. This report shows that a significant portion of certain pollutants is derived from the rainwater and not from the land. This information is useful for water quality modelling of urban runoff.

T. Milne Dick Chief Hydraulics Division

### PERSPECTIVE - GESTION

L'on sait que les eaux de ruissellement urbain peuvent renfermer d'importantes charges polluantes. Ce rapport démontre qu'une bonne part de certains polluants provient de l'eau de pluie, et non pas de la terre. Ce renseignement est utile dans la modélisation de la qualité de l'eau du ruissellement urbain.

Le chef.

T. Milne Dick Division de l'hydraulique

## TABLE OF CONTENTS

		Page
ABS	TRACT	i
MAN	AGEMENT PERSPECTIVE	ii
1.0	INTRODUCTION	1
2.0	STUDY AREA	2
3.0		2
	3.1 Water Samples Collection	2
	3.1.1 Rainfall monitor station	2.
	3.1.2 Runoff monitor station	3
	3.2 Rainwater and Runoff Samples Handling and	
	Laboratory Analysis	3
4.0	SELECTION OF CHEMICAL CONSTITUENTS AND ITS	
	CONCENTRATION DETECTION LIMIT	4
5.0	DATA TREATMENT	5
	5.1 Rainwater Sequential Data	5
	5.2 Runoff Sequential Data	6
	5.3 Volume-Weighted Mean Concentration of Constituent .	6
	5.4 Loading Rates	7
	5.4.1 Rainwater loading rate	7
	5.4.2 Runoff loading rate	7
6.0	SUMMARY OF RESULTS	8
	6.1 Selection of Events	8
	6.2 Statistic of Constituent Concentration	9
	6.3 Average Contaminant Loads of Rainwater and Runoff .	9
7.0	DISCUSSION	9
8.0	CONCLUSION	15
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	RENCES	
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IGUR	RES	

### 1.0 INTRODUCTION

This report is part of the Hydraulics Division study "Effect of Urban Land Use on Runoff Quantity and Quality". One of the goals of this study is to evaluate the contribution of rainwater to storm runoff quality. An earlier report by Ng (1982) presented some results of variations of rainwater chemistry and a subsequent report by Ng and Marsalek (1983) dealt with the effects of rainwater composition on urban runoff quality. Both reports found that rainwater contributed substantial portions of the pollutant loading in urban storm runoff, particularly the constituent of nitrogen compounds.

Although the studies dealing with pollutants and contamination sources are numerous, the contribution of pollutants from rainwater to urban storm runoff was studied in only a few cases. Black (1980) observed high concentration of ammonia nitrogen in surface runoff pollutants from a parking lot. He concluded that the source of ammonia nitrogen was apparently from rainfall. Wilber et al. (1975) found that lead and zinc loading in rainwater accounted for 89% of the total in a multiple land uses urban catchment of Lodi, New Jersey. They hypothesized that the lead and zinc originated from precipitation. Recently, Halverson, et al. (1984) studied the contaminant load in precipitation and urban runoff in a non-industrial urban area, and determined the direct contribution of material in the wet deposition. They found high ranking of nitrogen compounds in urban runoff which was input from precipitation.

It appears from various studies that rainwater contributes significant quantities of certain pollutants to the runoff loadings. Following such findings, it was interesting to evaluate the rainwater contribution to the runoff pollutant loadings in a systematical approach.

The report which follows deals primarily with the contribution of rainwater to the pollutants loading of urban storm runoff from field data collected during the 1982 and 1983 field seasons. An extensive field data set containing 19 dissolved chemical constituents were thoroughly evaluated in this study.

#### 2.0 STUDY AREA

The study site is shown in Figure 1, which is referred to as the Blair Road Catchment. A brief description on the catchment follows. The land use in the area is classified as light industrial with a total contributing area of 10.3 ha. The catchment imperviousness is determined as 69%. The catchment is served by storm sewer. The sewer pipes are made of concrete and vary in diameter from .61 m to .69 m at the outfall. The sewer systems are a simple layout with one main pipe and one lateral. The catchment is well established. During the study period, there was no construction activity.

The types of industries located in the study area are: food processing, sheet metal, steel fabrication, engine oil refinery, woodworking, wholesale offices and warehouses. The catchment is fairly well maintained and clean.

## 3.0 METHODS AND OBSERVATIONS

## 3.1 Water Samples Collection

The Blair Road Catchment has been instrumented for collections of rainfall, runoff data including sampling street surface deposits and road runoff.

The data collection systems consisted of a rainfall monitoring station and a runoff monitoring station.

## 3.1.1 Rainfall monitoring station

The rainfall monitoring station served to collect both rainwater samples and rainfall depths. For rainwater samples collection, a specially designed rainwater sampler was used (Ng, Boucher and Dolanjski, 1981). The rainwater sampler collected sequential rainwater samples, dry deposits and recorded the rainfall depths.

A brief description of the features of the sampler follows.

The sampler is equipped with a lid covering the rainwater collector (500 mm x 500 mm) funnel during dry periods. At the start of a rain event, the moisture sensing grids activated a control motor to drive the lid from the collector funnel and place it over the dry collector. Thus the rainwater samples collected in this manner are wet deposits from precipitation events. When the rain stops, the lid will move back to cover the rainwater collector by means of the electrically heated sensing grids sensor. The dry collector now opens to collect dry deposits.

A total of 23 sequential rainwater samples of 500 ml each can be collected during a rain event. The sequential rainwater samples are collected according to a preselected time interval.

## 3.1.2 Runoff Monitoring Station

The runoff monitoring station consisted of an automatic sampler and a flow meter. The sampler is a Sigmamotor Model 6201 Automatic Wastewater Sampler. The sampler collected up to 24 sequential samples of 500 ml each at a selected constant time interval. The constant time interval can be selected from 0 to 99 minutes. However, the time interval used in this study was in the range from 5 to 15 minutes. A calibrated weir was used in conjunction with a stilling well and a float-type water level recorder for flow measurements.

## 3.2 Rainwater and Runoff Samples Handling and Laboratory Analysis

A procedure was established to handle both rainwater and runoff samples after a storm event.

At the end of each precipitation event, rainwater and runoff samples were removed from the sampling devices. For rainwater sampling station, bottles with rainwater collected were removed and replaced with cleaned bottles for the next storm event. It should be noted in here that during a precipitation event certain sampling bottles may collect a

full jar of rainwater and some bottles can be without sample in the sequence, because the uncertainties of precipitation nature and sampling technique. Details have been discussed by Ng (1982). For runoff sampling sation, all 24 samples were removed from the sampler and replaced with a cleaned set of bottles for the next runoff event. Again, it should be noted that runoff samples can always collect a full jar of sample up to 24 bottles during a runoff event, because the sampler is activated by a selected height of water level in a stilling well. As long as the level in the stilling well holds, the automatic sampler will continue to sample at selected constant time until the number of samples in the sequence are exhausted.

The samples were transported to the laboratory as soon as possible, usually within 12 hours after the event except when the storm period ended on a weekend or on a statutary holiday.

In the laboratory, the samples were transferred from field containers into laboratory bottles, splitted and preserved according to the amount of water required for analysis of a constituent. The procedures recommended by the Water Quality Branch, IWD, Ontario Region (1978) were used. The analytical methods for individual constituent are given in another reference by Water Quality Branch, IWD, Ottawa (1979 and 1981 update).

## 4.0 SELECTION OF CHEMICAL CONSTITUENTS AND ITS CONCENTRATION DETECTION LIMIT

Three groups of chemical constituents were selected for this study. These constituents are most commonly related to urban storm runoff and the atmospheric fallout pollution. A list of these substances and the detection limits for analyses are presented in Table 1.

#### 5.0 DATA TREATMENT

## 5.1 Rainwater Sequential Data

As noted before, sequential rainwater samples collected during a precipitation event at constant time intervals were often insufficient (if less than 500 ml) for laboratory analysis of all the selected chemical constituents because of the nature of precipitation and sampling technique. The sampling technique used in this study was to collect sequential rainwater sample based on a fixed time interval during a rain event. Thus the volume of rainwater collected in this manner could be a full sample bottle or none, or in between. depended on the rainfall intensity. The collecton of a full sample (500 ml) corresponded to the rainfall depth of 2 mm (Ng et al., 1981). If a sample collected is not sufficient for chemical analysis, several samples in the sequential order are composited to bring the amount of the sample to 500 ml. Similarly, if individual sample or a composited sample happen to be insufficient to meet the chemical analysis requirement, doubled distilled water is added to bring the sample to the required amount. Consequently, the analytical results of a constituent is corrected by a dilution factor. This dilution factor varied by the amount of distilled water being added. The dilution factor is calculated by the following expression:

$$DF = \frac{V_s}{(V_s - V_d)} \tag{1}$$

where

DF is the dilution factor

 $V_S$  is the total volume of the sample

 $V_{f d}$  is the distilled water volume being added to the sample.

The dilution factor for correction of analytical results of rainwater samples are presented in Table 2. If DF = 1.000, indicated that no dilution was made for that sample.

## 5.2 Runoff Sequential Data

Sequential runoff samples collected during the runoff event were very often sufficient to meet the amount required for analysis of the selected constituent. The reason for this had been described in section 3.1.2. Thus correction on the analytical results is not needed.

## 5.3 <u>Volume-Weighted Mean Concentration of Constituent</u>

For the purpose of computing the loading of each of the selected constituents for a storm event, the mean concentration of each constituent is calculated from analytical results of sequential samples. The calculation is based on volume-weighted procedure which are applied to both rainwater and runoff constituents. The calculation is expressed as

$$\bar{c} = \frac{\sum c_i V_i}{\sum V_i}$$
 (2)

where  $\bar{c}$  is the mean concentration of a constituent

C<sub>i</sub> is the concentration of a constituent at time t,

 $V_i$  is the volume between time intervals during the event,  $\Sigma$   $V_i$  is the sum of volume.

and i is a subscript referred to sequential sample number.

## 5.4 Loading Rates

## 5.4.1 Rainwater

After the mean concentration of each constituent has been computed, the loading rates can be computed for rainwater. The unit area loading calculation is shown below:

$$L_{p} = A \times P \times \overline{C}_{p}$$
 (3)

where

 $L_{\rm p}$  is the rainwater loading rate in mg/m² of a constituent A is the conversion factor of Blair Road catchment area  $(10^6 \ell/{\rm mm}/10.3 \times 10^4 {\rm m}^2)$ 

P is the "effective rainfall depth" in mm, and

 $\bar{c}_{p}$  is the volume-weighted mean concentration in mg/l.

Note that "effective rainfall" defined in here is that portion of rainfall depth which produced the equivalent depth of measured runoff. Further detail will be given in the Discussion section regarding "effective rainfall". The measured rainfall and the measured runoff relationship is plotted in Figure 2. As seen from Figure 2, a 1 mm of rainfall yields an average of 0.55 mm of measured runoff in Blair Road Catchment. This average value is not used in the loading rate calculation since the interest for loading rate calculation is in individual event. Instead, the "effective rainfall depth" is used to calculate the loading rate. That is, P is numerically equal to the runoff depth R.

## 5.4.2 Runoff

The loading rate for runoff was calculated in the same manner as the rainwater. The unit area loading rate calculation is shown below:

$$L_{R} = A \times R \times \bar{C}_{R} \tag{4}$$

where  $L_{\rm p}$  is the runoff loading rate in mg/m<sup>2</sup>

A is given as above,

R is the runoff depth in mm, and

 $\overline{\mathsf{c}}_\mathsf{R}$  is the volume-weighted mean concentration in mg/l.

It should be noted that during dry period, there is very little or no flow at the measuring gauge of the sewer. Therefore, the runoff volume measured during rain period is not subject to the base flow effect.

## 6.0 SUMMARY OF RESULTS

## 6.1 <u>Selection of Events</u>

Rainwater and runoff samples along with rainfall and runoff records were collected for 49 storm events in a light industrial urban catchment between June 1982 and December 1983. Among the 49 events some had incomplete data.

In this study, events with at least one complete set of either nutrients, or major ions, or total metals data were selected. selected events along with rainfall and runoff depths, antecedent dry periods, and ratio of runoff to rainfall depths are listed in Table 3. Such selected events will enable a systematic computations on event loadings for both rainwater and runoff. So, after chemical analysis for each event, the volume-weighted mean concentration of each constituent for both rainwater and runoff was computed according to equations (1) Subsequently, the loading rate of each constituent of the sampled events were computed by equations (3) and (4) for rainwater and runoff respectively. In conjunction with the computation, it was noted that the concentration of some constituents was frequently below the detection limit. A summary of the occurrence of concentration above detection levels is presented in Table 4. In this case,

concentration of a constituent is found below detection level, it is assumed to equal to detection level. Thus the loading rate of an event can be computed. Results of loading rates computation are presented in Tables 5a, 5b and 5c. Accordingly, using the loading rates shown in Tables 5a, 5b and 5c, the ratio of rainwater load to runoff load is calculated. The results are presented in Tables 6a, 6b and 6c. Included in Tables 6a, 6b, and 6c, under Column A is the ratio of precipitation (by total depth of rain) loading to runoff loading.

## 6.2 <u>Statistics of Constituent Concentration</u>

Because of the tremendous variability in concentration, direct comparison of the same constituent between events may not be feasible. Subsequently, some basic statistics about the constituent concentration for all the events studied are determined, and are presented in Table 7. The basic statistics shown in Table 7 are minimum, maximum, and mean concentration along with the standard deviation from the mean for rainwater and runoff constituents.

## 6.3 Average Contaminant Loads of Rainwater and Runoff

By using Tables 5a, 5b and 5c and Tables 6a, 6b, and 6c, the average values and ratio of loading can be calculated. The results are presented in Table 8.

#### 7.0 DISCUSSION

#### Effective Rainfall:

A very common method for calculating the loading of a constituent in a precipitation event is based on the product of precipitation volume and its concentration. If this precipitation load (calculated by total volume of rain of an event) is compared with the runoff load

(calculated by measured runoff volume) of the same event, it could overestimate the contribution from the precipitation load. The runoff volume measured in a sewered urban catchment resulting from a rainfall event is only a portion of the rainfall volume. The other portion of the rainfall volume is lost in the forms as infiltration, surface detention, evaporation, wetting the ground surface, and a portion may be bypassed out of the sewer system. Therefore it seemed appropriate for the purpose of this study to use the effective rainfall volume to calculate the rainwater loading of a constituent for each rainfall event.

#### Rainwater Loads:

Results of rainwater loading for each constituent of each event are shown in Tables 5a, 5b and 5c. The level of contamination varies widely within the event and interevents. The average contaminant loads among all the events studied range from highest of sulphate to the lowest of Cadmium with the loading rates of  $23.549~\text{mg/m}^2$  to  $0.025~\text{mg/m}^2$  as shown in Table 8. Most of the loads vary over one and usually two, orders of magnitude. The level of variation of the nutrient group was in the order of same magnitude. The exception is phosphorus which was found in small amounts in the rainwater and the same result was reported by Halverson, et al. (1984). The nitrogen compounds appear to be quite consistent for 1982 and 1983 events and are relatively high. Similar result was reported by Randall et al. (1978).

Of the major ions loading, sulphate was found to be the highest in both years, and both sodium and potassium were found to be the lowest for both years of 1982 and 1983. A fact attributed to low loading rate in rainwater was believed that most storms were origined from inland regions or localized which carried rain with only minor concentrations of sodium as compared with that found in sea coast regions (Kennedy, et al. 1979). The major ions group in rainwater was extensively studied by Kennedy et al. (1979). However, very little information on rainwater loading rates has been found in literature.

Hence, it precludes comparison of the results of this study with other workers.

Among the trace metals studied, the loading of zinc, copper, nickel and lead were found to be predominently from rainwater. Cadmium, chromium and barium were usually found in considerably smaller quantities. This fact is also reflected by the percentage of samples with concentrations above the limit as presented in Table 4. The loading of zinc was found to be the highest of the trace metals group. The findings of the high zinc content in rainwater supports the observations by Wilber et al. (1975).

Regression analysis was carried out to determine significant of linear relationship between precipitation quality and storm size and the length of the antecedent dry period for all the constituents. The results showed that no significant linear relationship between antecedent dry period or the storm size. The highest coefficient of determination between the antecedent dry period was .346 for total carbon and the lowest was -.378 for chromium. fact, the coefficients of correlation for all the trace metals were negative. Similarly, there was no significant linear relationship between storm size and rainwater chemistry. It is apparent that the rainwater chemistry in an individual storm is not affected by either the antecedent dry period or the storm size. The poor relationship between rainwater chemistry and both of the antecedent dry period and storm size may suggest that rainwater chemistry in individual storm is the result of a rainout but not a washout processes in the air mass. These rainout and washout processes are beyond the scope of this study.

#### Runoff Loads:

Runoff loads also showed a wide variation of each contaminant in runoff water. The loads of sulphate and calcium were significantly higher than the other constituents. The nitrogen compounds were pronounced which were in the same order of magnitude as in rainwater.

Among all the constituents, the major ions group showed significantly higher loading rates, with the exception of potassium which is an order of magnitude less in the group. Sulphate and calcium loads were significantly dominant. Such finding supports the investigation in a non-industrial urban area reported by Halverson, et al. (1984).

The loading rate of zinc was significant among the trace metals. The loading rates of cadmium and chromium were only a small quantity, which were almost equal the same loading rates of precipitation.

The length of the antecedent dry period, was correlated with the runoff quality for all the constituents investigated. Regression analysis showed that there was no significant linear relationship between runoff quality and antecedent dry period. The highest and the lowest correlation coefficients were .622 and -.358 for ammonia and cadmium respectively. Model simulation of the urban storm runoff chemistry using accumulations and washoff equations (The Urban Drainage Committee, 1980. Alley and Smith, 1981, and Huber, et al., 1982) assumes that all pollutants are eroded from surface after a storm, and that the contribution of pollutants to runoff of the same event from rainwater is neglected. This assumption could not be confirmed here since the antecedent conditions tested had little linear relationship with runoff pollutants.

Ratio of Rainwater Load to Runoff Load:

The direct contribution of rainwater to pollutants of urban runoff may be viewed as the ratio between rainwater loads divided by runoff loads, since there are no standard appraoch and no rigorous method available to lead to a direct result. The ratio computed for each constituent of each event for both rainwater and runoff is shown in Tables 6a, 6b and 6c. The average ratio for each constituent and in different years of data is shown in Table 8, and plotted on bar charts

as shown in Figures 3 and 4. Figures 3 and 4 give a better view for comparison between the constituent group both in magnitudes and in Note that the average ratio is not identical to a different years. ratio as calculated for average loads as shown in Table 8. The overall average of the ratio between rainwater loads and runoff loads shown a wide ranges of ratio from the lowest of .060 (calcium and the highest of 2.552 (ammonia). It is worthy to mention that the total wet and dry deposits from atmospheric input are important in runoff loading according to Miller and Mattraw, 1982, and the precipitation input alone is an important factor as well. But precipitation loads are not equally important for all constituents to urban runoff. From the results of this study, the nitrogen compounds in the runoff seemed to be entirely derived from rainwater but the major ions appeared to be contributed from surface sources, as indicated by the ratios shown in Table 8. Most of the nitrogen loading ratios are greater than 1. All the major ions loading ratio are less than 1. The major ions group with the highest ratio found is only .221 for sulphate and the lowest ratio found is calcium being .060.

In the trace metal group, copper and nickel had ratios greater than 1. It appears that these two substances in runoff were input from rainwater.

It is noted that the occurrence of detection level of copper and nickel in rainwater (Table 4) are higher than the occurrences of detection level in runoff. The ratios of the rest of the trace metal group are within 0.5 to .75. There are difficulties encountered in justifying the contributions of cadmium, chromium, lead and barium in rainwater because of the uncertainties of their occurrences in the detection limit, as shown in Table 4. Their highest occurrence above detection levels were slightly higher than 50% (rainwater). These substances were believed present in both rainwater and runoff although they were frequently below the detection limits. In addition, trace metals are non-degradable which may precipitate out of solutions under favourable pH conditons, be absorbed by clay particles or bound by the

hydrous oxides by other metals such as iron, aluminum and manganese. The magnitude of percentage of occurrence above detection level of a substance in rainwater is not necessary to produce the same magnitude of the ratio of rainwater load to runoff load as shown in Table 4 and in Table 8. In the case of zinc, it had 100% occurrence of concentration above detection limit in both rain and runoff waters. But the ratios of the rainwater loading to runoff loading was accounted only about 61%.

Figure 3 reveals two interesting aspects. The first is that runoff loading rates of the major ions group were predominately higher than the rainwater loading rates. The second aspect is that both loading rates of rainwater and runoff contaminants had no substantial difference in both years. The only exception was that loading rate of sulphate had about 33% higher in 1983 as compared to 1982.

On Figure 4 the ratio of rainwater loading rate to the runoff loading rate showed nitrogen compounds (nitrate + nitrite, ammonia, and kjedahl), copper and nickel had the ratio greater than 1 in both years of 1982 and 1983. An exception was nitrate + nitrite which was slightly less than 1 (.947) in 1983.

Finally, the levels of direct contribution of rainwater to quality of runoff could be ranked according to the ratios of loading rates between rainwater and runoff presented in Table 8. The ranking is as follows:

Nitrogen compounds, copper and nickel - very significant
Lead and zinc - less significant
Total phosphorus - less significant
Total carbon - less significant
Major ions - least significant
Cadmium, chromium and barium - uncertain

#### 8.0 CONCLUSION

Investigation of constituents in individual rainfall - runoff events showed that rainwater contributed most of the nitrogen compounds observed in runoff. Rainwater also contributed significant amount of copper and nickel to runoff, and fair amount of zinc, lead, and certain amount of cadmium, chromium and barium. Rainwater contributed very little amount of major ions to runoff.

The length of antecedent dry periods and storm size showed little or no linear correlation with runoff quality as well as rainwater quality.

Models of urban runoff pollutants should consider the rainwater composition, and in particular, the atmospheric contributions of nitrogen compounds should be included.

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TABLE 1. Constituents Studied and Detection Limits

Constituent	Detection Limit (mg/l)
Nitrate + Nitrite N	.005
Ammonia N	.001
Total Kjeldahl N	.010
Dissolved Carbon	.200
Total Phosphorus P	.001
Calcium	.200
Magnesium	.100
Sodium	.100
Potassium	.100
Sulphate	1.000
Chloride	.100
Barium	.100
Cadmium	.010
Chromium	.010
Copper	.010
Lead	.010
Nickel	.010
Zinc	.010

TABLE 2. The Dilution Factor for Correction of Analytical Results of Rain Samples

Event	Storm Dates			Seque	ntial Or	der of S	ample		
No.	1982	1	2	3	4	5	6	7	Average
8	June 1	1.000	1.000	1.000	1.000				1.000
14	June 21	1.220	1.111	1.389					1.240
15	June 23	1.000	1.000	1.000	1.515				1.129
17	June 29	1.000	1.389		. •				1.195
18	June 30	1.000	1.000	1.000	1.000				1.000
20	July 17	1.000	1.000	1.000	1.000	1.220			1.044
21	July 28	1.000	1.000	1.000	1.000	,			1.000
23	Aug. 2	1.000	1.000	1.000	1.000	1.316	<u> </u>		1.063
25	Aug. 8	1.000	1.000	1.000	1.000	1.282	1.000		1.007
29	Aug. 25	1.000	1.000	1.429	1.515	1.000	1.000	1.351	1.185
34	Sept 18	1.429	1.136	1.351	1.000	1.471			1.189
41	Sept 28	1.087	1.087	1.000	1.000	1.471			1.129
42	Oct. 7	2.083	1.000	1.000	1.000		ļ		1.271
43	Oct. 7	1.000	1.111	1.000	1.000				1.028
48	Nov. 1-2	1.000	1.000	1.000	1.000	1.000			1.000
49	Nov. 2	1.087	1.000	1.351	1.000	1.000	1.000	1.000	1.050
	1983								
9	Apr. 15	1.075	1.087	1.000	1.000				1.041
11	Apr. 29	1.000	1.000	1.000	1.000				1.000
17	May 14	1.515	1.111	1.000	1.000				1.157
18	May 19	1.000	1.064	1.000					1.021
21	May 30	1.000	1.429	1.000			·		1.143
25	June 27	1.200	1.000	1.000	1.000				1.050
27	July 4	1.000	1.000	1.000	1.000				1.000
34	Aug. 11*	1.000	1.000	1.000	1.220	1.000	1.000		1.037
35	Aug. 11*	1.000	1.000	1.000	1.000	1.000	1.400		1.067
36	Aug. 22	1.000	1.000	1.000	1.000	1.000	1.000		1.100
39	Aug. 31	1.000	1.000	1.000					1.000
49	Oct. 12	1.000	1.220	1.000	1.000	1.000	1.000		1.104
51	Oct. 24	1.000	1.000	1.312			]	·	1.000
58	Nov. 24	1.000	1.000	1.000	1.000	1			1.000
59	Nov. 28	1.000	1.000	1.000			Ì		

TABLE 3. Storm Dates, Rainfall and Runoff Depths and Antecedent Dry Period of Selected Storm Events

Event I.D.	Storm Date	Rainfall Depth (mm)	Runoff Depth (mm)	Antecedent Dry Period (days)	Ratio of Runoff/Rainfall Depth
	1982				
8	June 1	7.90	1.89	1 1	.239
14	June 21	6.10	1.44	6	.236
15	June 23	7.90	1.71	1	.217
17	June 29	7.90	1.86	1	.235
18	June 30	3.3	0.73	1	.221
20	July 17	18.00	3.65	12	.203
21	Júly 28	6.60	1.32	11	.200
23	Aug. 2	16.00	3.47	5	.217
25	Aug. 8	13.20	3.63	4	.275
29	Aug. 25	77.50	15.65	1	.202
34	Sept 18	12.20	2.79	2	.229
41	Sept 28	31.20	6.74	3	.216
42	Oct. 7	3.30	0.89	10	.270
43	Oct. 7	12.40	2.91	10	.235
48	Nov. 1-2	23.60	5.20	12	.220
49	Nov. 2	10.90	2.37	1	.217
	1983				, ·
9	Apr. 15	9.90	2.82	3	.285
11	Apr. 29	4.10	1.20	2	.293
17	May 14	4.60	0.97	7	.211
18	May 19	23.90	5.75	4	.241
21	May 30	14.70	3.36	4	.229
25	June 27	32.80	6.76	21	.206
27	July 4	18.00	3.64	1	.202
34	Aug. 11*	42.90	8.97	2	.209
35	Aug. 11*	42.90	8.97	2	.209
36	Aug. 22	10.20	2.09	11	.205
39	Aug. 31	18.00	4.85	1	.269
49	Oct. 12	17.30	4.71	1	.272
51	Oct. 24	21.60	4.71	10	.218
58	Nov. 24	5.60	1.14	3	.204
59	Nov. 28 ent with two	18.30	3.96	14	.216

TABLE 4. Occurrence of Constituent Concentration Above Detection Limit in Rainwater and Runoff Samples

Gnoup	Constituent	Occurrence of above Detection	
Group	constituent	Rainwater	Runoff ,
Nutrient	Nitrate + Nitrite Ammonia Kjedahlas N Total Carbon Total Phosphorus	100 100 100 96.88 100	100 100 100 100 100
Major Ions	Calcium Magnesium Sodium Potassium Sulphate Chloride	98.94 98.94 78.35 90.82 100 92.86	100 100 100 100 100 100
Trace Metals	Cadmium Copper Chromium Lead Nickel Zinc Barium	56.58 90.00 38.16 47.50 91.25 100 3.80	8.67 82.14 44.90 48.21 51.28 100 12.57

Contaminant Load (mg/m $^2$ ) of Rahmwater and Runoff Events by Catchment Surface TABLE 5a.

al norus	~	.044 .044 .378 .273 .436	. 292 . 303 . 111 . 453 . 337 . 655 . 956	.629 .312 * .814 .632 .611 .883 .557 .479 .361 .361
Total Phosphorus	a.	.029 .067 .047 .006	.040 .039 .048 .035 .014	.105 .020 .284 .284 .252 .078 .038 .052
Carbon	~	6.609 2.332 7.670 5.827 38.449	16.917 18.186 15.033 14.373 13.542 13.229 18.549 6.377	5.857 *- 7.877 *- 18.280 9.750 60.810 35.212 33.784 20.629 27.436 13.896 8.660 1.454 7.887
Total (	۵	4.036 3.347 .992 .696 3.485	5.789 4.797 6.990 9.571 2.385 4.085 27.254	1.667 1.407 * 7.973 14.941 74.284 1.955 5.251 1.306 1.673 2.318 1.893 3.241
Je N	84.	1.048 .262 1.387 .805 2.807	2.430 2.162 .813 2.504 2.992 2.471 4.248 1.293	1.628 * 1.622 * 3.288 1.868 1.868 2.750 2.750 2.168 2.168 2.267
Kjedahle	ď	1.383 1.709 .694 .315 1.972	2.566 4.785 1.106 2.863 1.039 1.234 19.774	1.924 4.429 4.282 4.752 4.752 2.528 1.301 1.805 1.22 2.027
ia	R	.705 .080 .532 .550 .774	1.023 .551 .186 1.214 1.618 1.197 2.477	.577 .931 * 1.331 .858 1.536 1.514 .973 .879 .879
Ammonia	م	1.174 1.341 .576 .217 1.512	1.990 1.417 1.417 2.067 2.067 847 .890 15.382	1.353 1.978 * 2.560 1.791 12.809 3.755 2.166 .705 1.607 1.607
ite + te	R	1.514 .412 1.423 1.059 2.485	2.350 1.667 3.161 1.848 3.004 2.834 3.825 1.587	2.195 *894 2.227 2.731 27.559 5.632 5.378 3.834 2.946 3.292 3.292 3.292
Nitrate Nitrite	d	1.658 1.979 .330 .138 1.605	2.546 1.954 1.754 .993 1.519 1.147 12.366	2.664 2.008 2.708 2.790 1.185 1.629 1.597 2.103
Date	1982	June 1 June 23 June 30 June 30 July 17		Apr. 15 Apr. 29 May 14 May 30 June 27 Aug. 11 Aug. 22 Aug. 22 Oct. 24 Nov. 28
Event		20 18 18 18 18	4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4	111 122 133 33 33 34 51 51 53 53

Runoff Rainwater

No data

TABLE 5b. Contaminant Load (mg/m²) of Rainwater and Runoff Events by Catchment Surface

oride	~	24.056 7.200 21.913 9.301 24.851 3.298 14.201 8.812 4.603 18.284 6.145 10.425 14.939 6.737	26.059 14.672 4.249 11.647 20.023 19.010 20.913 8.924 10.176 9.554 5.327 6.823 9.923
Chlor	۵	.676 1.008 1.188 1.188 .932 .932 .507 .793 .713 .713 .713 .713	1.962 1.001 1.001 1.001 1.761 1.761 1.255 1.255 1.136 1.136 1.136
hate	æ	132.024 34.680 74.473 67.384 104.181 30.526 97.768 20.760 90.220 44.608 85.049 138.492	73.203 37.685 16.619 107.422 112.615 198.804 140.760 77.767 120.484 152.290 84.660 39.419
Sulphat	م	18.346 9.411 33.001 9.765 12.900 1.188 16.839 28.690 8.949 17.835 10.260 15.031 62.872 5.437	16.821 12.399 4.898 23.476 20.893 162.103 12.945 23.343 9.738 5.439 14.875 6.720 1.885
ssium	ж	2.186 1.601 2.273 2.088 3.389 2.088 2.381 2.040 2.040 2.139 3.739 5.527 2.381	2.864 2.440 1.087 3.847 2.670 7.051 5.956 1.972 3.532 2.733 1.376 2.493
Pota	۵	.171 .380 .076 .969 .233 .238 .505 .489 .297 .434 .141	.447 .251 .126 .126 .911 .911 .256 .349 .080 .251 .169 .407
i.	œ	13.145 7.165 11.539 14.864 2.909 8.659 9.177 3.687 4.028 7.517 4.162	17.801 9.437 2.961 14.500 8.480 13.689 18.896 7.107 7.107 7.684
Sodi	م.	.378 .439 .017 .186 .185 .199 .246 .244 .532 .320	1.575 .537 .453 2.525 1.807 2.555 .951 .648 1.60 1.130 1.621 1.429 1.28
sium	œ	7.353 7.736 9.426 5.748 15.004 3.928 8.459 11.436 4.438 9.532 21.805	12.683 5.251 2.999 20.248 12.896 23.891 27.047 25.452 17.894 15.406 10.992 5.483
Magne	Ь	.588 .588 .168 1.723 1.175 .211 .736 .855 1.135 .216 .532 .532	.958 1.469 .618 3.189 2.832 7.718 1.469 .601 .669 .936 .936 .741 1.407
ium	R	78.831 46.868 46.297 90.495 73.634 106.228 31.963 96.780 48.591 91.910 136.062 53.295	70.768 48.033 21.617 118.823 89.547 167.078 169.376 163.147 64.660 111.801 113.447 72.886 31.533 80.908
Calcium	Ь	1.776 2.805 1.099 9.542 3.589 1.518 6.087 4.251 2.046 1.911 1.208 15.786	3.484 5.765 1.920 12.934 10.010 2.23.395 4.710 2.285 2.260 1.656 3.763
Date	1982	June 1 June 23 June 23 June 30 July 17 July 28 Aug. 8 Sept 18 Sept 18 Oct. 7 Oct. 7 Nov. 2 1983	Apr. 15 Apr. 29 May 14 May 30 June 27 Aug. 11 Aug. 21 Oct. 24 Nov. 24
Event No.		8 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4	911119 121124 121121 121121 121121 13112 131

Runoff Ú œ Rainwater

No data

ij

TABLE 5c. Contaminant Load (mg/m²) of Rainwater and Runoff Events by Catchment Surface

		079 017 036 * * 234 182 322 358 358 358 524 358	362 1155 1155 003 007 871 871 871 8623 6623
Ē	~		ं नं नं
Barium	ď	.038 .017 * .1183 .130 .146 .280 .674 .089 .143	215 .040 .040 .097 1.591 .117 .1187 .209
ပ္	~	.410 .037 .032 * .267 .309 .782 .782 .782 .782 .1.617 5.802	1.129 602 * * * 1.370 1.370 1.372 1.366 1.966
Zinc	Ь	.127 .032 .332 .308 .567 .165 .180 .1413 .207 .350	.559 
[a]	~	.036 .025 .025 .025 .047 .297 .179 .302 .302 .302 .209	
Nickel	م	214 .032 * .206 .206 .139 .261 .277 .079 .062	323 .016 * * .161 .137 .253 .107 .167 .152
P	~	316 230 * 230 * 234 * 263 248 .215 .215 .215 .215 .215 .215 .215 .228	.248 * * * * .527 .527 .583 .583 .338 .323 .253
Lead	ط	.091 .068 .010 * .183 .140 .278 4.067 .147 .337 .124 .083	.151 .020 * * * .080 .083 .093 .209 .152 .152
ium	R	.037 .010 .018  .052 .078 .078 .072 .076 .046 .052	.036 .036 .096 .007 .007 .0027 .0627 .063
Chromium	Ь	.023 .014 .009 .048 .038 .038 .038 .014 .010	.022 .004 * * * .010 .012 .012 .042 .010
er	æ	.058 .040 * .040 .061 .074 .074 .119 .119 .195	.181 * * * * .124 .140 .111 .105 .105 .105
Copper	d	.093 .016 .016 .183 .079 .079 .081 .097 .093	
mium	R	.019 .010 .010 .018 .023 .023 .072 .072 .095 .095	.033 .015 .087 .087 .062 .062 .061 .061
Cadm	ď	.022 .014 .009 .018 .020 .038 .057 .010 .010	043 004 004 0050 0013 0013 0010 0010 0010
	Ċ	11 11 12 12 13 13 13 13 13 14 15 17 17 17 17 17 17 17 17 17 17 17 17 17	115 22 33 33 31 111 111 112 22 24 28
Date	1982	June June June July July July Aug. Sept Oct. Oct.	Apr. May May June June June Aug. Aug. Oct. Nov.
Event		20 20 20 20 20 20 20 20 20 40 40 40 40 40 40 40 40 40 40 40 40 40	11 11 12 13 13 13 13 13 13 13 13 13 13 13 13 13

P = Rainwater R = Runoff \* = No data

TABLE 6a. Ratio of Rainwater Load to Runoff Load

1 orus	8	.066 .124 .124 .021 .113 .136 .130 .107 .105 .005	.166 .065 .349 * 2.865 .285 .517 .164 .104 .126 .360
Total Phosphorus	Ø	. 277 6.470 . 569 . 095 . 556 . 629 . 473 . 390 . 022 . 066	. 584 . 223 1. 447 * 13.098 1. 366 2. 474 . 799 . 387 . 463 1. 914 . 334
Carbon	В		.285 .179 .436 1.533 1.222 .056 .155 .061 .167 .219
Total	٧	2.555 6.081 .596 .540 .447 1.577 3.089 3.089 6.679 1.590	.999 .610 .610 .266 .744 .309 .227 .227 .1.003 1.903
las N	В	1.320 500 .391 .702 1.056 2.213 1.144 1.144 .499 4.654	1.181 1.551 1.347 2.292 1.737 1.665 .919 .364 .515 .833 .890
Kjedahlas	A	5.523 2.305 1.769 3.460 4.865 8.047 5.294 1.286 2.125 1.850	4.145 5.293 5.293 10.008 8.434 7.967 1.776 1.915 3.061 4.084 2.275 4.140
ıja	8	1.665 1.083 1.083 1.955 1.946 2.570 4.409 1.703 .744 6.210	2.344 2.124 1.923 2.088 2.272 2.244 1.783 1.833 3.411 1.694
Ammonia	A	6.966 70.941 4.993 1.787 9.629 8.967 9.345 1.939 3.164 28.226 3.094	8.225 7.248 7.979 9.116 11.027 11.696 8.533 3.157 6.737 15.648 8.302 10.386
te + te	8	1.095 4.808 .232 .131 .160 .546 .537 .537 .506 .506 .506 .506	1.214 2.245 1.838 1.022 .711 .632 .425 .425 .710 .710 .764
Nitrate Nitrite	A	4.583 20.371 1.068 3.181 4.994 4.217 2.420 2.488 1.722 1.722 1.722 1.908	4.259 7.663 7.663 7.626 2.146 3.402 3.023 2.073 1.827 4.154 3.259 3.744
Date	1982	June 1 June 23 June 23 June 30 July 17 Aug. 2 Aug. 8 Sept 18 Sept 28 Oct. 7 Nov. 2 Nov. 2	Apr. 15 May 19 May 30 June 27 Aug. 11 Aug. 22 Aug. 31 Oct. 12 Oct. 24 Nov. 28
Event		81 18 23 23 24 43 43 43 43 43 43 43 43 43 43 43 43 43	11 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1

A = Calculated by total rainfall depth; B = Calculated by effective rainfall depth \* = no data

TABLE 6b. Ratio of Rainwater Load to Runoff Load

June 1         1.094         0.23         3.34         0.07         1.20         0.029         3.37           June 21         .284         .067         .322         .076         .260         .061         1.000           June 21         .284         .067         .322         .076         .260         .061         1.000           June 23         .108         .067         .322         .078         .018         .007         .002         .153           July 17         .195         .040         .386         .078         .018         .001         .1975           July 17         .195         .040         .386         .078         .014         .034         .1975           Aug. 2         .381         .083         .401         .087         .106         .023         .977           Aug. 2         .381         .089         .051         .143         .039         .144         .120         .087         .087         .087         .252         .054         .120         .087         .056         .054         .124         .087         .057         .057         .057         .057         .057         .057         .056         .054         .144	Event Date No.	Calcium	E	Magnesium	sium	Sodium	ium	Potassium	sium	Sulp	Sulphate	Chloride	ide
1. 094       .023       .334       .070       .120       .029         23       .108       .322       .076       .260       .061         23       .108       .023       .082       .018       .007       .002         30       .877       .206       1.275       .300       .144       .034         17       .195       .040       .386       .078       .048       .010         2       .381       .083       .401       .087       .106       .023         .146       .040       .272       .075       .097       .027         .18       .395       .091       1.270       .291       .145       .033         .28       .098       .021       .069       .015       .064       .014         .7       .146       .039       .143       .024       .181       .043         .14       .039       .144       .120       .489       .132         .146       .039       .015       .015       .064       .014         .2       .056       .013       .1476       .325       1.000       .220         .2       .058       .013       .0	1982	Ø	_ω	<b>V</b>	89	А	В	A	В	A	മ	A	æ
21       .284       .067       .322       .076       .260       .061       1         23       .108       .023       .082       .018       .007       .002         30       .877       .206       1.275       .300       .144       .034         17       .195       .040       .386       .078       .048       .010         28       .282       .056       .269       .054       .118       .064         18       .395       .091       1.270       .291       .145       .033         28       .098       .021       .069       .015       .064       .014         7       .046       .027       .075       .097       .027         18       .395       .091       1.270       .291       .145       .033         28       .098       .021       .069       .015       .064       .014         7       .046       .039       .444       .120       .489       .132         7       .056       .013       .051       .011       .172       .034         16       .058       .013       .051       .051       .134       .053 </td <td></td> <td>.094</td> <td>.023</td> <td>.334</td> <td>070.</td> <td>.120</td> <td>.029</td> <td>.327</td> <td>.078</td> <td>.581</td> <td>.139</td> <td>.118</td> <td>.028</td>		.094	.023	.334	070.	.120	.029	.327	.078	.581	.139	.118	.028
23       .108       .023       .082       .018       .007       .002         30       .877       .206       1.275       .300       .144       .034       .1         28       .282       .056       .269       .054       .318       .064       .010         28       .282       .056       .269       .057       .097       .027         18       .395       .091       1.270       .291       .145       .033         28       .098       .021       .069       .015       .064       .014         7       .056       .013       .143       .034       .181       .043         7       .056       .013       .143       .034       .181       .043         1-2       .527       .116       .1476       .325       1.000       .220         2       .058       .013       .051       .011       .172       .037         3       .058       .013       .051       .194       .057         44       .120       .265       .076       .310       .089         2       .058       .013       .051       .154       .153		.284	.067	.322	9/0.	.260	.061	1.000	.236	1.150	.271	.593	.140
30       .877       .206       1.275       .300       .144       .034       1         28       .282       .056       .269       .054       .318       .064       .010       2         28       .282       .056       .269       .054       .318       .064       .010       2         38       .146       .040       .272       .075       .097       .027       .097       .027       .097       .027       .097       .027       .097       .027       .099       .015       .064       .014       .014       .014       .027       .095       .015       .064       .014       .027       .095       .013       .144       .120       .489       .132       .049       .132       .034       .181       .043       .181       .043       .181       .043       .181       .043       .181       .043       .181       .043       .181       .043       .181       .043       .181       .044       .120       .489       .132       .034       .181       .043       .181       .044       .184       .181       .044       .184       .181       .044       .184       .184       .054       .184       .054		.108	.023	.082	.018	.007	.002	.153	.033	2.042	.443	.863	.187
17       .195       .040       .386       .078       .048       .010       2         28       .282       .056       .269       .054       .318       .064       .010         8       .282       .056       .269       .054       .318       .064       .023         8       .146       .040       .272       .075       .097       .027         18       .395       .091       1.270       .291       .145       .033         28       .098       .021       .069       .015       .064       .014         7       .146       .039       .444       .120       .489       .132         1-2       .527       .116       1.476       .325       1.000       .220       1         2       .056       .013       .051       .011       .172       .037       .1         3       .057       .013       .051       .011       .172       .037       .1         4       .120       .955       .280       .194       .057       .1       .1       .1         14       .421       .089       .976       .206       .724       .153       .1		.877	.206	1.275	300	.144	.034	1.975	464	.617	.145	.544	.128
28       .282       .056       .269       .054       .318       .064         2       .381       .083       .401       .087       .106       .023         8       .146       .040       .272       .075       .097       .027         18       .395       .091       1.270       .291       .145       .033         28       .098       .021       .069       .015       .064       .014         7       .146       .039       .044       .120       .489       .132         1-2       .527       .116       1.476       .325       1.000       .220       1.32         1-2       .527       .013       .051       .011       .172       .037         2       .058       .013       .051       .194       .057         3       .410       .120       .265       .280       .194       .057         14       .421       .089       .976       .206       .724       .153         14       .421       .089       .976       .206       .724       .153         2       .488       .112       .959       .220       .961       .198		.195	.040	.386	.078	.048	.010	2.954	009.	909	.123	.089	.018
2 .381 .083 .401 .087 .106 .023 8 .146 .040 .272 .075 .097 .027 18 .395 .091 1.270 .291 .145 .033 28 .098 .021 .069 .015 .064 .014 7 .146 .039 .444 .120 .489 .132 1-2 .527 .116 1.476 .325 1.000 .220 2 .058 .013 .051 .011 .172 .037 15 .173 .049 .265 .076 .310 .089 2 .410 .120 .955 .280 .194 .057 14 .421 .089 .976 .206 .724 .153 2 .488 .112 .959 .220 .931 .213 2 .680 .140 1.568 .323 .961 .198 11 .067 .014 .113 .024 .164 .034 2 .075 .020 .222 .061 1.049 .285 2 .158 .022 .222 .061 1.049 .285 2 .158 .022 .222 .061 1.049 .285 2 .124 .025 .216 .044 .135 .028		.282	.056	.269	.054	.318	.064	.880	.176	.195	.039	.400	080
8       .146       .040       .272       .075       .097       .027         18       .395       .091       1.270       .291       .145       .033         28       .098       .021       .069       .015       .064       .014         7       .146       .039       .444       .120       .489       .132         1-2       .527       .116       1.476       .325       1.000       .220       1         2       .058       .013       .051       .011       .172       .037         3       .40       .120       .265       .076       .310       .089         4       .421       .089       .265       .280       .194       .057         14       .421       .089       .265       .280       .194       .057         14       .421       .089       .976       .206       .724       .153         2       .680       .112       .959       .220       .931       .174       .1         3       .488       .112       .959       .220       .951       .158       .032         11       .067       .014       .113       .0		.381	89.	.401	.087	106	.023	.977	.212	.720	.156	.355	.077
18       .395       .091       1.270       .291       .145       .033         28       .098       .021       .069       .015       .064       .014         7       .146       .039       .444       .120       .489       .132         7       .056       .013       .143       .034       .181       .043         1-2       .527       .116       1.476       .325       1.000       .220       1         2       .058       .013       .051       .011       .172       .037         3       .40       .120       .265       .076       .310       .089         4       .421       .089       .265       .280       .194       .057         14       .421       .089       .976       .206       .724       .153         2       .488       .112       .959       .220       .931       .213         3       .488       .112       .959       .220       .931       .194         11       .067       .014       .113       .024       .158       .032         12       .073       .020       .222       .061       1.049 <td< td=""><td></td><td>.146</td><td>.040</td><td>.272</td><td>.075</td><td>.097</td><td>.027</td><td>.524</td><td>.144</td><td></td><td>.294</td><td>.385</td><td>106</td></td<>		.146	.040	.272	.075	.097	.027	.524	.144		.294	.385	106
28 .098 .021 .069 .015 .064 .014 7 .146 .039 .444 .120 .489 .132 1-2 .527 .116 1.476 .325 1.000 .220 2 .058 .013 .051 .011 .172 .037 3 15 .173 .049 .265 .076 .310 .089 14 .421 .089 .976 .266 .194 .057 19 .452 .109 .654 .158 .723 .174 11 .133 .028 .260 .054 .333 .070 11 .067 .014 .113 .024 .164 .034 22 .158 .032 .245 .050 .158 .032 12 .073 .020 .222 .061 1.049 .285 24 .124 .025 .216 .044 .135 .028 28 .215 .047 .400 .086 .372 .080		.395	.091	1.270	.291	.145	.033	.636	.146	1.883	431	476	109
7		860.	.021	690.	.015	.064	.014	.442	.095		198	201	043
7       .056       .013       .143       .034       .181       .043         1-2       .527       .116       1.476       .325       1.000       .220       1         2       .058       .013       .051       .011       .172       .037         3       .058       .013       .051       .011       .172       .037         15       .173       .049       .265       .076       .310       .089         14       .421       .089       .976       .280       .194       .057         19       .452       .109       .654       .158       .723       .174       .1         27       .680       .112       .959       .220       .931       .213       .1         27       .680       .140       1.568       .323       .961       .198       .1         27       .680       .140       1.568       .323       .961       .198       .1         28       .158       .028       .260       .054       .164       .034         29       .158       .020       .158       .050       .158       .251         24       .124       .0		.146	.039	.444	.120	.489	.132	.243	990	.852	230	430	116
1-2       .527       .116       1.476       .325       1.000       .220       1         2       .058       .013       .051       .011       .172       .037       1         3       .058       .013       .051       .011       .172       .037       1         15       .173       .049       .265       .076       .310       .089         29       .410       .120       .955       .280       .194       .057         14       .421       .089       .976       .206       .724       .153         19       .452       .109       .654       .158       .723       .174       .1         27       .680       .140      568       .323       .961       .198       .1         27       .680       .140       1.568       .323       .961       .198       .1         27       .680       .140       1.568       .323       .961       .198       .1         11       .067       .014       .113       .024       .164       .034         28       .158       .020       .222       .061       1.049       .285         29<		•026	.013	.143	.034	.181	.043	106	025	.770	18	415	860
2 .058 .013 .051 .011 .172 .037  15 .173 .049 .265 .076 .310 .089  16 .421 .089 .976 .206 .724 .153  19 .452 .109 .654 .158 .723 .174  27 .680 .140 1.568 .323 .961 .198  21 .067 .014 .113 .024 .164 .034  22 .158 .032 .245 .050 .158 .032  24 .104 .023 .194 .042 1.554 .339  24 .124 .025 .216 .044 .135 .028  28 .215 .047 .400 .086 .372 .080	_	.527	.116	1.476	.325	1.000	.220	1.293	.285	2.064	454	1.738	382
3         15       .173       .049       .265       .076       .310       .089         29       .410       .120       .955       .280       .194       .057         14       .421       .089       .976       .206       .724       .153         19       .452       .109       .654       .158       .723       .174       .13         20       .488       .112       .959       .220       .931       .213       .13         27       .680       .140       1.568       .323       .961       .198       .1         27       .680       .140       1.568       .323       .961       .198       .1         27       .680       .140       1.568       .323       .961       .198       .1         28       .133       .028       .260       .054       .333       .070       .158         29       .075       .014       .113       .024       .159       .285         24       .124       .023       .216       .044       .135       .028         215       .047       .400       .086       .372       .080       1		.058	.013	.051	.011	.172	.037	.233	.050		.139	.450	.098
15 .173 .049 .265 .076 .310 .089 29 .410 .120 .955 .280 .194 .057 14 .421 .089 .976 .206 .724 .153 19 .452 .109 .654 .158 .723 .174 27 .680 .140 1.568 .323 .961 .198 11 .133 .028 .260 .054 .333 .070 11 .067 .014 .113 .024 .164 .034 22 .158 .032 .245 .050 .158 .032 24 .124 .025 .216 .044 .135 .028 28 .215 .047 .400 .086 .372 .080	000												
15       .173       .049       .265       .076       .310       .089         29       .410       .120       .955       .280       .194       .057         14       .421       .089       .976       .206       .724       .153         19       .452       .109       .654       .158       .723       .174       .1         27       .488       .112       .959       .220       .931       .213       .1         27       .680       .140       1.568       .323       .961       .198       .1         11       .133       .028       .260       .054       .333       .070         11       .067       .014       .113       .024       .164       .034         22       .029       .245       .050       .158       .032         12       .073       .020       .139       .037       .591       .159         24       .104       .023       .194       .042       .135       .028         24       .124       .025       .216       .044       .135       .028         24       .124       .025       .216       .086 <t< td=""><td>1983</td><td></td><td></td><td></td><td>····</td><td></td><td></td><td></td><td></td><td></td><td></td><td></td><td>,</td></t<>	1983				····								,
29       .410       .120       .955       .280       .194       .057         14       .421       .089       .976       .206       .724       .153         19       .452       .109       .654       .158       .723       .174       .1         20       .488       .112       .959       .220       .931       .213       .1         27       .680       .140       1.568       .323       .961       .198       .1         11       .133       .028       .260       .054       .333       .070         11       .067       .014       .113       .024       .164       .034         22       .158       .032       .245       .050       .158       .032         31       .075       .020       .139       .037       .591       .159         24       .104       .023       .194       .042       1.554       .339         28       .215       .047       .400       .086       .372       .080       1			.049	.265	920.	.310	680	.548	.156	908	.230	.264	.075
14       .421       .089       .976       .206       .724       .153         19       .452       .109       .654       .158       .723       .174       1         20       .488       .112       .959       .220       .931       .213       1         27       .680       .140       1.568       .323       .961       .198       1         11       .133       .028       .260       .054       .333       .070         11       .067       .014       .113       .024       .164       .034         22       .158       .032       .245       .050       .158       .032         31       .075       .020       .139       .037       .591       .159         12       .073       .020       .222       .061       1.049       .285         24       .104       .023       .194       .042       1.35       .028         24       .124       .025       .216       .044       .135       .028         28       .215       .047       .372       .080       1			.120	.955	.280	.194	.057	.351	.103	1.123	329	.233	990
19       .452       .109       .654       .158       .723       .174       1         30       .488       .112       .959       .220       .931       .213       1         27       .680       .140       1.568       .323       .961       .198       1         11       .133       .028       .260       .054       .333       .070         11       .067       .014       .113       .024       .164       .034         22       .158       .032       .245       .050       .158       .032         31       .075       .020       .139       .037       .591       .159         12       .073       .020       .222       .061       1.049       .285         24       .104       .023       .194       .042       1.554       .339         24       .124       .025       .216       .044       .135       .028         28       .215       .047       .372       .080       1			680.	926.	.206	.724	.153	.547	.115	1.397	.295	.450	.095
30     .488     .112     .959     .220     .931     .213     1       27     .680     .140     1.568     .323     .961     .198     1       11     .133     .028     .260     .054     .333     .070       11     .067     .014     .113     .024     .164     .034       22     .158     .032     .245     .050     .158     .032       31     .075     .020     .139     .037     .591     .159       24     .104     .023     .194     .042     1.554     .339       28     .215     .047     .400     .086     .372     .080     1			109	.654	.158	.723	.174	1.242	.299	.907	.219	900	.216
27     .680     .140     1.568     .323     .961     .198     1       11     .133     .028     .260     .054     .333     .070       11     .067     .014     .113     .024     .164     .034       22     .158     .032     .245     .050     .158     .032       31     .075     .020     .139     .037     .591     .159       12     .073     .020     .222     .061     1.049     .285       24     .104     .023     .194     .042     1.554     .339       28     .215     .047     .400     .086     .372     .080     1			.112	.959	.220	.931	.213	1.491	.341	.810	.186	099.	.151
11     .133     .028     .260     .054     .333     .070       11     .067     .014     .113     .024     .164     .034       22     .158     .032     .245     .050     .158     .032       31     .075     .020     .139     .037     .591     .159       12     .073     .020     .222     .061     1.049     .285       24     .124     .025     .216     .044     .135     .028       28     .215     .047     .400     .086     .372     .080     1			.140	1.568	.323	.961	.198	1.037	.214	4.019	.828	.968	.199
11       .067       .014       .113       .024       .164       .034         22       .158       .032       .245       .050       .158       .032         31       .075       .020       .139       .037       .591       .159         12       .073       .020       .222       .061       1.049       .285         24       .104       .023       .194       .042       1.554       .339         24       .124       .025       .216       .044       .135       .028         28       .215       .047       .400       .086       .372       .080       1			.028	.260	.054	.333	.070	.222	.047	.312	.065	1.390	.291
22     .158     .032     .245     .050     .158     .032       31     .075     .020     .139     .037     .591     .159       12     .073     .020     .222     .061     1.049     .285       24     .104     .023     .194     .042     1.554     .339       24     .124     .025     .216     .044     .135     .028       28     .215     .047     .400     .086     .372     .080     1			.014	.113	.024	.164	.034	.281	.059	.794	.166	.813	.170
31     .0/5     .020     .139     .037     .591     .159       12     .073     .020     .222     .061     1.049     .285       24     .104     .023     .194     .042     1.554     .339       24     .124     .025     .216     .044     .135     .028       28     .215     .047     .400     .086     .372     .080     1			.032	.245	.050	.158	.032	.198	.041	.611	.125	.176	.036
12     .0/3     .020     .222     .061     1.049     .285       24     .104     .023     .194     .042     1.554     .339       24     .124     .025     .216     .044     .135     .028       28     .215     .047     .400     .086     .372     .080     1			020	.139	.037	.591	.159	.264	.071	.168	.045	.459	.123
24 .104 .023 .194 .042 1.554 .339 24 .124 .025 .216 .044 .135 .028 28 .215 .047 .400 .086 .372 .080 1			020	.222	.061	1.049	-282	.228	790.	.359	860.	.614	.167
28 .215 .047 .400 .086 .372 .080 1			.023	.194	.042	1.554	.339	.832	.181	.364	.079	.978	.213
28 .215 .047 .400 .086 .372 .080 1			.025	.216	-044	.135	.028	.235	.048	.235	.048	.242	.049
			.047	.400	980.	.372	080.	1.380	.298	1.115	.241	.738	.159

A = Calculated by total rainfall depth; B = Calculated by effective rainfall depth

TABLE 6c. Ratio of Rainwater Load to Runoff Load

T	1		
5	<u>~</u>	.480 .714 .779 .779 .716 .716 .781 .781 .781 .781 .781	.593 .259 .111 .753 .474 .175 .175 .671
Barium	A	2.009 7.264 2.191 3.299 3.299 1.646 4.128 3.412 5.960 0.959 3.681	2.082 .884 .539 3.726 2.268 2.130 2.495 1.273
	<u></u>	.308 .098 .098 .219 .243 .243 .503 .900 .060	. 153 . 153 . 174 1. 174 1. 174 . 407 . 505 . 587 . 120
Zinc	A	1.290 12.671 .454 1.007 2.972 1.204 2.197 8.365 *	1.738 .521 1.037 6.689 5.615 1.947 2.465 3.520 2.150
1	80	5.951 6.477 1.278 4.401 2.148 1.811 2.569 1.169 1.386 1.386 325	4.451 1.036 .730 2.557 .681 1.993 1.342 1.196
Nicke	Ø	24.900 27.446 5.889 21.678 9.898 6.586 5.104 4.252 1.265 1.265	15.616 3.536 3.544 12.660 5.010 9.724 4.990 4.398
ъ	В	288 466 043 779 530 121 1.176 684 778 223 100 166	.606 .100 .153 .527 .417 .172 .266 .658
Lea	A	1.207 6.212 .198 3.839 2.444 4.076 5.822 2.985 3.602 3.602 .757 2.457	2.127 .340 .740 2.611 1.994 1.296 2.445 1.726
i um	В	.619 .475 .937 1.001 .529 .894 .316 .198 .163	.593 .259 .101 .449 .474 .259 .437 .671
Chromium	А	2.589 6.016 2.191 4.618 1.902 5.847 2.309 4.137 1.172 1.172 3.234	2.082 .884 .488 2.223 2.268 1.239 2.495 1.273
er	æ	1.604 10.151 .406 3.001 1.173 5.904 2.228 2.028 .944 .476 .262	5.219 .453 .611 2.078 3.669 4.107 2.756 1.201 1.591
Coppe	<b>d</b>	6.710 43.012 1.872 14.785 5.403 5.588 9.387 3.495 2.027 1.192 2.760	18.311 1.547 2.963 10.289 17.557 19.650 13.442 4.466 7.298
mn .	<u>α</u>	1.111 1.420 .475 .779 1.083 1.644 1.061 1.091 1.024 .296 .215	1.300 .259 .256 .1.056 .500 .175 .494 .671
Cadmium	∢	4.647 6.016 2.191 3.839 4.989 5.251 5.251 3.793 1.258 4.059	4.561 .884 1.435 5.227 2.394 2.407 2.495 1.273
Date	1982	une und	1983 Apr. 15 June 27 July 4 Aug. 11 Aug. 22 Aug. 31 Oct. 24
Event No.		114 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4	519 225 233 334 549 549 549 549 549 549 549 549 549 54

A = Calculated by total rainfall depth; B = Calculated by effective rainfall depth; \* = No data

TABLE 7. Statistics of Constituent Concentration (mg/l) observed in Rainwater and Runoff for 1982 - 1983 Storm Events Studied

· ·	Minimu	um	Max imur	n	Meai	1	Standard [	Deviation
	Rainwater	Runoff	Rainwater	Runoff	Rainwater	Runoff	Rainwater	Runoff
Nitrate + Nitrite	.147	.214	3.240	7.650	.756	.757	.579	.690
Ammonia	.044	.020	4.038	1.350	.585	.302	.653	.292
Kjedahlas N	.107	.132	5.180	2.790	.827	.572	.820	.489
Total Carbon	<.200	.991	19.000	11.900	1.753	3.824	2.255	2.580
Total Phosphorus	.001	.024	2.180	1.608	.036	.122	.058	.087
Calcium	<.050	8.930	6.930	56.100	1.378	19.742	1.134	7.256
Manganese	<.010	1.090	5.200	11.400	.392	2.924	.364	.830
Sodium	<.020	.698	.950	11.000	.231	2.441	.185	1.496
Potassium	<.020	.083	1.230	1.880	.124	.755	.083	.355
Sulphate	1.123	5.800	60.400	93.900	6.106	21.687	5.499	9.979
Chloride	<.100	<.100	7.700	21.900	.505	3.627	.451	2.645
Cadmium	<.010	<.010	.040	.020	.011	.010	.004	.002
Copper	<.010	<.010	.090	.090	.068	.029	.077	.023
Chrominum	<.010	<.010	.040	.050	.014	.013	.021	.007
Lead	<.050	<.050	.160	.490	.105	.095	.298	.101
Nickel	<.030	<.010	.200	.070	.049	.026	.034	.014
Zinc	<.010	<.010	.590	1.789	.229	.140	.216	.224
Barium	<.050	<.050	.100	.200	.052	.031	.032	.031

TABLE 8. The Average Contaminant Load of Rainwater and Runoff and the Average Ratio of Rainwater to Runoff of the Studied Storm Events

		Average Loading (mg/m <sup>2</sup> )				· · · · · · · · · · · · · · · · · · ·		
Group	Constituent	1982		1983		Ratio of Rainwater to Runoff		
		Rainwater	Runoff	Rainwater	Runoff	1982	1983	All Events
Nutrient	Nitrate + Nitrite	2.202	2.013	3.225	4.799	1.139	.947	1.043
	Ammonia	2.122	.879	2.407	1.198	3.124	1.979	2.552
	Kjedahl as N	3.074	1.940	3.283	2.548	1.625	1.127	1.376
	Total Carbon	5.819	13.853	9.104	19.349	.494	.391	.442
·	Total Phosphorus	.066	.434	.253	.505	.506	.455	.481
Major Ions	Calcium	3.942	69.273	5.524	80.259	.059	.059	.060
	Magnesium	1.103	8.840	1.658	14.947	.112	.119	.115
	Sodium	.353	8.433	1.153	17.801	.052	.137	.096
	Potassium	.409	2.574	.480	3.270	.186	.145	.166
	Sulphate	17.895	76.097	23.549	101.16	.232	.211	.221
	Chloride	1.365	12.485	2.030	12.558	.115	.144	.129
	Cadmium	.031	.035	.025	.060	.913	.530	.746
	Copper	.131	.097	.287	.130	2.302	2.201	2.258
	Chromium	.030	.052	.043	.061	.749	.379	.554
	Lead	.142	.356	.118	.359	.607	.356	.498
	Nickel	.143	.118	.155	.130	2.286	1.573	1.976
	Zinc	.286	1.064	.570	1.172	.705	.593	.609
	Barium	.172	.279	.206	.619	.735	.424	.602

Note that the average ratio is not identical to a ratio as calculated from average loads.

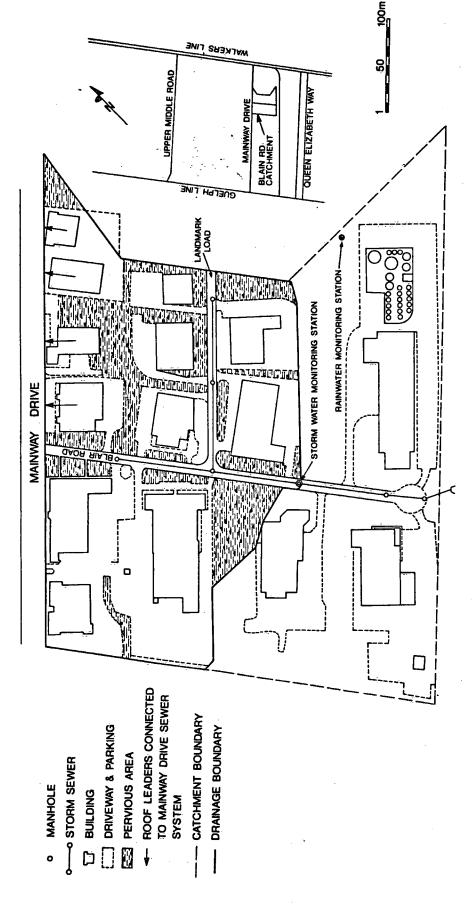
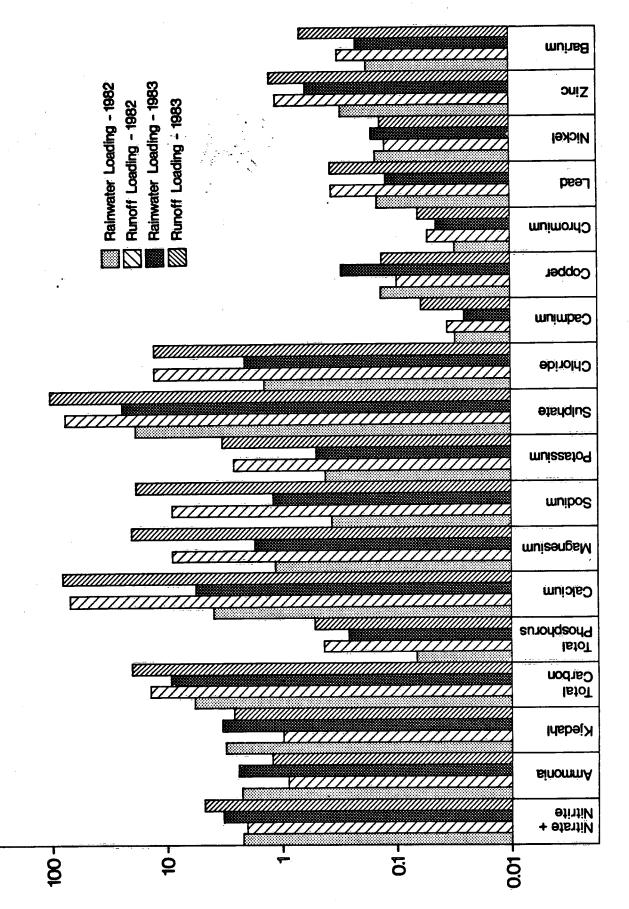


Figure 1 BLAIR ROAD CATCHIMENT BURLINGTON, ONTARIO

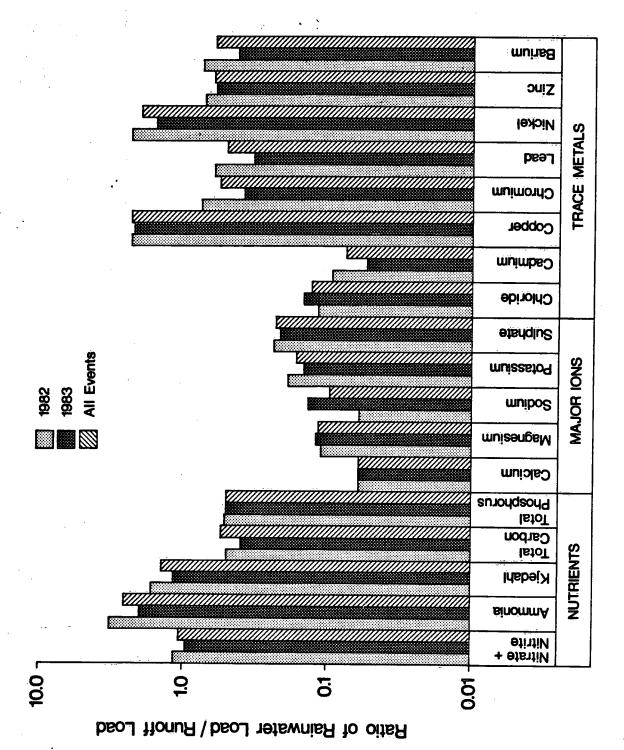
FIGURE 2. RAINFALL AND RUNOFF RELATIONSHIP

ω.

FIGURE



Average Contaminant Loading of Rainwater and Runoff (mg/m²)



igure 4 Ratio of rainwater load to runoff load.