# WINTER AND SPRING MIGRATION ECOLOGY <br> OF BLACK BRANT (BRANTA BERNICLA NIGRICANS) IN THE STRAIT OF GEORGIA, BRITISH COLUMBIA 

Katherine Hagmeier

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#### Abstract

A population may be defined as a group of organisms living in a particular space at a particular time. Understanding processes that affect population change is essential for conservation and management of many populations, but depends on the spatial and temporal scale within which the population is examined. The Pacific Flyway Brant Population in North America contains both winter resident and spring transient populations at different times in an annual cycle. Furthermore, there are two distinct taxa within the Flyway population: Black Brant (Branta bernicla nigricans) which comprise $>90 \%$ of the Flyway population and Grey-bellied Brant, one of four distinct stocks of Light-bellied Brant (Branta bernicla hrota), which comprise the remainder of the population.

A small population of Brant winters in the Boundary Bay and Roberts Bank area of the Fraser River delta in southern British Columbia, Canada. This population has increased rapidly since the 1990's due to a combination of annual productivity and immigration. The largest source of immigration is from Grey-bellied Brant, which traditionally winter in northern Washington State, USA, 60 km south of the Fraser River delta. Increased integrated management of the Boundary Bay-Roberts Bank wintering Brant and Brant wintering in Washington State is recommended. Wintering Brant in Boundary Bay and Roberts Bank are subjected to high levels of disturbance. The frequency of disturbance was $2.17 / \mathrm{h}$ in Boundary Bay and $1.32 / \mathrm{h}$ in Roberts Bank. Brant spent approximately $10 \%$ of their time responding to disturbance and this amounted to an estimated $45 \%$ of their daily energy expenditure. There is potential for wintering Brant to abandon the area if the levels of disturbance remain high or increase. The most common source of disturbance was Bald Eagles, hence reducing disturbance will be challenging for managers. During spring migration, transient populations of Brant stage in two key areas in the Strait of Georgia, British Columbia. These are Boundary Bay and Roberts Bank and ParksvilleQualicum on the east coast of Vancouver Island, British Columbia. To estimate the total number of Brant using these two areas, weekly counts combined with capture-mark-recapture analysis using Program Mark were used to develop models describing spring migration patterns. These models revealed that the areas support largely separate migrating populations. Results also revealed that between 13-27\% of the Pacific Flyway Brant Population staged in Boundary Bay and Roberts Bank and Parksville-Qualicum in 1999 and 2000. Conservation of


these two staging areas may be critical for the successful management of the Pacific Flyway Brant Population.

## RÉSUMÉ

On peut définir une population comme un groupe d'organismes qui vivent en un espace et une période donnés. La connaissance des processus qui déterminent les changements démographiques est essentielle pour la conservation et la gestion de nombreuses populations, mais il faut garder à l'esprit que cette connaissance est fonction des échelles spatiale et temporelle à l'intérieur desquelles nous étudions les populations. La population de Bernaches cravants de la voie migratoire du Pacifique de l'Amérique du Nord comprend des populations hivernantes et des populations migratrices qu'on trouve en divers endroits à différentes périodes du cycle annuel. De plus, on compte deux taxons distincts dans la population de cette voie migratoire : la Bernache cravant noire (Branta bernicla nigricans), qui représente plus de $90 \%$ de la population, et la Bernache cravant à ventre gris - l'une des quatre formes de la Bernache cravant à ventre pâle (Branta bernicla hrota) -, qui constitue le reste de la population. Une petite population de Bernaches cravants hivernent dans la région de la baie Boundary et du banc Roberts, dans le delta du Fraser, en Colombie-Britannique (Canada). Cette population s'est rapidement accrue depuis les années 1990, grâce à une bonne productivité annuelle et à l'immigration. Les immigrants sont surtout des Bernaches cravants à ventre gris, forme quia l'habitude d'hiverner dans le nord de l'État de Washington (États-Unis) à 60 km au sud du delta du Fraser. On recommande donc une gestion intégrée accrue des bernaches hivernantes du secteur baie Boundary-banc Roberts et du Washington.
Les Bernaches cravants qui hivernent dans la région de la baie Boundary et du banc Roberts sont exposées à de forts niveaux de perturbation. On a observé des fréquences de perturbation de $2,17 / \mathrm{h}$ à la baie Boundary et de $1,32 / \mathrm{h}$ au banc Roberts. Les bernaches passaient environ $10 \%$ de leur temps à réagir aux perturbations, ce qui, selon nos estimations, requérait une quantité d'énergie correspondant à $45 \%$ de leurs dépenses énergétiques quotidiennes. Il est possible que les bernaches cessent d'hiverner à cet endroit si les niveaux de perturbation demeurent élevés ou s'accroissent. La principale source de perturbation observée étant les Pygargues à tête blanche, la réduction des niveaux de perturbation représentera un défi de taille pour les gestionnaires.

Durant la migration printanière, les bernaches de passage font halte dans deux secteurs clés dans le détroit de Géorgie (Colombie-Britannique), soit celui de la baie-Boundary et du banc Roberts, et celui de Parksville-Qualicum sur la côte est de l'île de Vancouver (Colombie-Britannique). Pour estimer le nombre total de bernaches présentes à ces deux
endroits, on a effectué des dénombrements hebdomadaires et une analyse de données de marquage-recapture au moyen du programme Mark en vue d'élaborer des modèles décrivant les profils de migration printanière. Les modèles obtenus ont montré que ces deux régions sont fréquentées par des populations migratrices largement distinctes. Les résultats ont aussi montré que de 13 à $27 \%$ de la population de Bernaches cravants de la voie migratoire du Pacifique a fait halte dans les secteurs baie Boundary-banc Roberts et Parksville-Qualicum en 1999 et 2000. La conservation de ces deux haltes migratoires pourrait s'avérer essentielle pour assurer le succès de la gestion de la population de Bernaches cravants de la voie migratoire du Pacifique.

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## 1. GENERAL INTRODUCTION

### 1.1 Population Dynamics

A population may be defined as a group of organisms of the same species living in a particular space at a particular time (Krebs 1985). This definition necessarily changes depending on the spatial scale being examined and is also complicated by the fact that a population may be divided into distinct components, which may have behavioural, morphological or geographical differences.

Population size is primarily regulated by four factors: 1) natality, 2) mortality, 3) immigration and 4) emigration (Begon et al. 1990). Ultimately, each of these is a response to limiting factors in the environment (e.g. availability of food and water, or nest sites). Accordingly, an overabundance of a limiting resource leads to an increase in population size. If there are no limiting factors, growth will remain unchecked so that a population experiences exponential growth (Johnson 1994). However, in most natural systems, the growth rate declines once a balance is achieved between the population and its environment (i.e. the carrying capacity).

Understanding this balance is particularly critical when humans influence the environment through habitat loss or degradation.

Within a population, differences in migration patterns and annual life history strategies likely develop from trade-offs made by individuals in order to maximize fitness (Cox 1968, Gauthreaux 1982, Cox 1985). Migration patterns are commonly associated with bird species, particularly Arctic-nesting waterfowl and shorebirds, many of which have evolved highly specialized strategies (Raveling 1979, Gauthreaux 1980, Owen and Black 1990). Differences in migration strategies within a population may complicate the calculation of estimates of that population in temporary, but critical staging areas. Yet, estimates of migratory populations are essential as the importance of spring and fall staging habitats is often measured by the volumes of birds using the sites (Thompson 1993, Routledge et al. 1999, Frederickson et al. 2001).

In this study I investigated the population demography of Arctic-nesting Brant Geese (Branta bernicla), both wintering and migrating, in the Strait of Georgia, British Columbia. In doing so, I identified different segments of the Pacific Flyway Brant Population and described their interactions with the environment.

### 1.2 Study Species: Brant (Branta bernicla)

Brant (Branta bernicla) is a small sea goose that breeds in the North American and Eurasian Arctic. In North America, Brant are classified into two subspecies, the light-bellied Branta
bernicla hrota and the Black Brant, Branta bernicla nigricans (Madsen et al. 1999). B. b. hrota is further subdivided into two distinct stocks, Atlantic Brant and Grey-bellied Brant, which should be considered as separate taxa in terms of conservation and management (Madsen et al. 1999). The taxa are genetically, morphologically, and geographically distinct (Shields 1990). The number of Grey-bellied Brant is estimated between 6000-9000 birds (Subcommittee on Pacific Brant 2001) and is therefore one of the smallest Arctic goose stocks in the world (Brewer et al. 1999).
The Pacific Flyway Brant Population contains Black Brant and Grey-bellied Brant. Together, between 120,000-150,000 Brant migrate along the Pacific Flyway each year (Subcommittee on Pacific Brant 2001). Grey-bellied Brant breed exclusively in the Canadian High Arctic on Melville, Prince Patrick and Eglington Islands and winter in a discrete area in northern Washington State (Padilla, Fidalgo and Samish Bays) (Reed et al. 1989, Sedinger et al. 1994). Black Brant have a broader breeding range, the largest colonies of which are on the YukonKuskokwim (Y-K) Delta in southwestern Alaska. Ninety percent of the Black Brant winter in Mexico (approximately 75\% of the entire Pacific Flyway Brant Population). The remaining birds winter as small units along the Pacific Coast between Alaska and California (Figure 1) (Sedinger et al. 1994). Throughout this thesis, I will refer to Black Brant and Grey-bellied Brant together as Brant unless I make a distinction.

### 1.3 Population Structure

The Pacific Flyway Brant Population is regulated by factors that influence mortality and recruitment. It is widely accepted that harvest mortality is additive to natural mortality and, particularly in years when there has been low productivity combined with high harvest, harvest mortality negatively influences population size (Ebbinge 1991, Sedinger et al. 1994, Madsen et al. 1999). The structure of the Pacific Flyway Brant Population can be divided into three parts: 1) $30 \%$ are successful breeding adults, 2) between $16-28 \%$ are young, and 3 ) the remainder, non-breeders or failed breeders (Sedinger et al. 1994). Changes in the non-breeding portion do not affect the population as severely as changes in successful breeders; however, non-breeders may act as a buffer, which add to the breeding population over time to compensate for periods of low recruitment. Brant develop long-term, pair bonds and both members of the pair are involved in parental care. Family units migrate together during the year, although their family cohesiveness is not as easily detected as in other species of Arctic-nesting geese (Reed 1993).


Figure 1. Important breeding, wintering and spring staging areas of the Pacific Flyway Brant Population.

### 1.4 Habitat

The primary food of Brant is eelgrass (Zostera species) (Cottam et al. 1944, Einarsen 1965, Charman 1977, Ganter 2000), but other foods include algae (Ulva spp and Enteromorpha spp) and roe deposited by Pacific Herring (Clupea harengus) (Nygren 1990). Brant also ingest grit and regularly haul out on sandy beaches and spits (Campbell et al. 1990).

In winter Brant use sheltered coastal areas, typically estuaries, beaches, bays, lagoons and mudflats (Campbell et al. 1990). One of the greatest threats to Brant is the loss of wintering habitat (Sedinger et al. 1994). Development, driven by urbanization in temperate, coastal areas has caused increased disturbance to Brant and the deterioration or complete loss of habitat (Sedinger et al. 1994).
During migration, Brant stage in areas with abundant eelgrass. The most important of these staging areas is Izembek Lagoon on the tip of the Alaskan Peninsula (Figure 1). This is the final staging ground before Brant depart to the breeding grounds and is again important during autumn migration (Dau 1992).
The Y-K Delta contains the largest colonies of nesting Brant. In years when the area is covered in snow when the Brant arrive, they must delay nesting until conditions are suitable. Therefore, females must rely on stored nutrients acquired during migration when waiting to initiate nesting and they utilize these endogenous reserves during incubation. In late years fewer and smaller clutches occur (Sedinger et al. 1994).

### 1.5 Study Area

Research was conducted in the Strait of Georgia, British Columbia (Figure 1). During the winter (November to February), research was focused in Boundary Bay and Roberts Bank on the Fraser River delta (Figure 2). During the spring (March-May) the study area was expanded to include the east coast of Vancouver Island between Campbell River and Parksville. This allowed for observation of migrating Brant staging in two key areas as they moved north to the breeding grounds. Effort on Vancouver Island was concentrated on the main staging area between Parksville and Qualicum.

### 1.6 Study Objectives

I had two major goals in this study. The first was to understand the population dynamics of Brant wintering in Boundary Bay and Roberts Bank. This includes distribution, abundance, population composition, activity budgets, and examination of the current harvest regime. The second goal was to develop a model that will facilitate estimates of the number of Brant staging in Parksville-Qualicum and the Fraser River delta during spring migration.


Figure 2. Brant staging areas along the Strait of Georgia.

## 2. BRANT WINTERING IN BOUNDARY BAY AND ROBERTS BANK, BRITISH COLUMBIA: POPULATION DYNAMICS

### 2.1 Introduction

The global Brant population is comprised of three subspecies, Black Brant (Branta bernicla nigricans), Dark-bellied Brant (B. b. bernicla) and Light-bellied Brant (B. b. hrota). Dark-bellied Brant are found in Europe, Black Brant are found in North America and Light-bellied Brant are a combination of four stocks, which should be considered separately for conservation purposes (Madsen et al. 1999). Two of the Light-bellied stocks winter in Europe (the Eastern Canadian High Arctic Light-bellied Brent and the Svalbard Brent) and two winter in North America (the Atlantic Brant and the Grey-bellied Brant). In North America, the Light-bellied Atlantic Brant breed in the Eastern Canadian High Arctic and winter along the Atlantic Coast of the United States from Massachusetts to South Carolina. The other North American Light-bellied stock is the Grey-bellied Brant which breed on Melville, Prince Patrick and Eglington Islands in the Western Canadian High Arctic and winter along the northern coast of Washington State (Figure 1).

The Pacific Flyway Brant Population is comprised of Black Brant and Grey-bellied Brant (Reed et al. 1989, Shields 1990, Reed 1997). The two taxa have geographical, morphological and genetic differences (Shields 1990). Black Brant occur as an aggregate of smaller wintering populations. The majority ( $\sim 75 \%$ ) winter on the Baja California peninsula and the west coast of Mexico (Sedinger et al. 1994). The rest winter as small units in sheltered bays and estuaries between Alaska and California. Unlike Black Brant, Grey-bellied Brant have a small winter range concentrated on Samish, Fidalgo and Padilla Bays in Washington State (Reed et al. 1989) and comprise less than 10\% of the total Pacific Flyway Brant Population (Brewer et al. 1999).

Historically in British Columbia, a wintering population of Black Brant appeared in late autumn in Boundary Bay on the Fraser River delta. In addition, from approximately March until May, Brant migrating from more southern wintering areas (Washington, California, Oregon and Mexico) formed a transient spring population in Boundary Bay that intermingled with the wintering birds. By late May, all Brant had departed. However, Brant abundance has changed considerably in British Columbia during the last century. Records show that Brant were abundant winter residents in Boundary Bay in the late 1800's (Fannin 1891), but were considered primarily spring transients by the 1940's (Campbell et al. 1990). Ladner Christmas Bird Counts show declines in wintering Brant, e.g., 600 birds in 1960, 83 in 1962, and zero in 1980 (Campbell et al. 1990). The overall decline has been attributed to disturbance and over-hunting (Leach 1979).

Declines occurred elsewhere in the Pacific Flyway Brant Population, reaching historic lows in the 1970's and 1980's (Sedinger et al. 1994). This prompted changes in the British Columbia Brant harvest regulations. In 1977 hunting was restricted to the first 10 days of March on the Fraser River delta, was permanently closed on Vancouver Island in 1979, and also on the Queen Charlotte Islands in 1984 (BC Ministry of Environment, Lands and Parks unpubl. data.). The spring hunting season on the Fraser River delta was modified to concentrate the harvest on transient spring migrants. The impact of the British Columbia hunting season, however, remains a management concern. Variation in the timing of migration influences the number of spring migrants present on the Fraser River delta before and during the hunt. Furthermore, the harvest of spring migrants may adversely affect the recovery of other small wintering units on the Pacific Flyway.
By 2000, the number of Brant wintering in Boundary Bay increased to approximately 1200 birds (unpubl. data). Brant also expanded their winter range to include Roberts Bank, west of

Boundary Bay, an area not historically used by wintering Brant. The greatest changes in abundance and distribution occurred in the late 1990's (Reed 1997, unpubl. data), but the specific mechanisms for these rapid changes are not well understood.

In recent years, large numbers of Grey-bellied Brant (~200 or 18\%) have been observed wintering in Boundary Bay and Roberts Bank. This is likely a new phenomenon as there are no historical references of Grey-bellied Brant wintering in Boundary Bay. According to local hunters, a Grey-bellied Brant shot during the hunting season in the 1980's and early 1990's was a rare event (Garry Grigg pers. comm.). In 1994-95 and 1995-96 estimates of Grey-bellied Brant in the Boundary Bay wintering population were $3.0 \%$ ( $\pm 2.6$ SD) and $5.3 \% ~( \pm 2.5$ SD) respectively (Reed unpubl. data).

Adequate monitoring of Brant and understanding the factors influencing population change are essential for the conservation of the Boundary Bay-Roberts Bank wintering population. In addition, identifying the impacts of the spring hunt could have management implications that extend to other wintering units on the Pacific Flyway. Therefore, this chapter examines population demographics of Brant in Boundary Bay and Roberts Bank. Specific objectives were to estimate the annual numbers of Brant wintering in Boundary Bay and Roberts Bank, identify the population composition in terms of annual productivity, breeding ground origin, and the proportion of Grey-bellied Brant, compare daily use of Boundary Bay and Roberts Bank, and determine the impact of hunting.

### 2.2 Methods

### 2.2.1 Study area

In Boundary Bay and Roberts Bank, Brant forage on eelgrass and haul out on spits. In 1989, $55 \%$ (2839 ha) of Boundary Bay was covered with eelgrass (Zostera marina and Zostera japonica) (Ward et al. 1992) and the eelgrass beds have continued to increase (Tarbotton and Harrison 1996). The sandy spit at Beach Grove is used for loafing, preening, and gritting (Reed 1997). During the time frame of this study, wintering Brant expanded their range to include Roberts Bank (pers. obs.). Between 1959 and 1994 the surface area of Roberts Bank covered with eelgrass increased from 395 to 700 ha (Tarbotton and Harrison 1996). Boundary Bay and Roberts Bank are both areas of high recreational activity and support the only Brant hunt in British Columbia.

### 2.2.2 Data Collection

Researchers examining breeding and migratory patterns of the Pacific Flyway Brant Population have marked Black Brant on the breeding and moulting grounds with coded, coloured leg-bands since 1985 (Bollinger unpubl. data). As a result, 8-10\% of Black Brant in the Pacific Flyway Brant Population are individually marked (Reed 1997). Leg-banded Brant can be observed on winter and migratory staging areas and I capitalized on this during my research. During my research Grey-bellied Brant did not carry leg-bands.

During the winters and springs of 1997-98, 1998-99, and 1999-2000 Brant were surveyed daily in Boundary Bay and Roberts Bank. Each year surveys began when Brant arrived from the breeding grounds (approximately November) and lasted until spring migration (approximately May). Each survey was conducted with a 20-60X spotting scope to record leg-band data and conduct counts. Leg-band data were collected in order to identify winter residents of Boundary Bay and Roberts Bank and their breeding ground affiliations. Leg-band data were also used to identify when Brant from more southern wintering areas arrived in Boundary Bay and Roberts Bank.

Each day, the maximum number of Brant was recorded at each site. Concurrent counts of both sites were conducted 2-3/week to estimate the total number of Brant in the study area. I examined population composition by scanning groups of Brant (usually $\sim 100$ birds) and counting the number of each taxon, the number of banded birds (recording colour and code), and the number of immature Black Brant. Immature Brant were distinguished by the presence of white on the edges of their wing coverts and secondary feathers. Surveys for immature Greybellied Brant were added in the final year of the study.

Grey-bellied Brant were distinguished from Black Brant by belly plumage. This difference was observed when the birds were standing out of the water. The black plumage is continuous from the neck to the lower abdomen on Black Brant, whereas there is a sharp contrast between the black breast and the pale abdomen on Grey-bellied Brant. Because of plumage variability (i.e. some Grey-bellied Brant appear dark and cannot be easily differentiated from Black Brant) the estimates of Grey-bellied Brant in this research are minimum estimates. During subsequent research, I have learned that the necklace is a reliable feature for differentiating Black and Greybellied Brant (see Boyd and Maltby, 1979).

### 2.2.3 Analysis

## Abundance

In this study, the wintering Brant population was defined as those birds present between December 15 and January 31. This definition was likely conservative as the wintering population appeared to have formed by the end of November and spring migrants did not appear to arrive until mid-February (Reed 1997, pers. obs). Fall migrants seem to be absent in Boundary Bay and Roberts Bank (Reed 1997). To estimate the annual abundance of the wintering population, I used maximum and mean counts observed between December 15 and January 31 each year.
I also estimated Brant goose-day use for each winter (where a goose day is defined as the presence of one goose for one day). To calculate cumulative goose-days I used the formula:

$$
\begin{equation*}
\text { Goose-days }=\sum_{i} \frac{\mathrm{~N}_{\mathrm{i}}\left(\mathrm{D}_{\mathrm{i}+1}-\mathrm{D}_{\mathrm{i}-1}\right)}{2} \tag{Equation1}
\end{equation*}
$$

where $\mathrm{N}_{\mathrm{i}}=$ population at $i^{\text {th }}$ survey, $\mathrm{D}_{i}=$ date of $i^{\text {th }}$ survey, $\mathrm{D}_{i+1}=$ next survey date, $\mathrm{D}_{\mathrm{i}-1}=$ previous survey date. By definition, $D_{0}=D_{1}$ and $D_{n+1}=D_{n}$ (Boyd 1995).

## Population Composition and Increase

Like all populations, the wintering Brant population is influenced by four factors: immigration, emigration, annual productivity and mortality (Begon et al. 1990). Accordingly, when population increase cannot be accounted for by annual productivity and mortality, the balance must be from immigration. To determine if annual productivity and/or immigration were responsible for the rapid growth in the population, I examined population composition.
To estimate population composition, the numbers of immature Black Brant, Grey-bellied Brant, and Brant from each breeding ground were converted to proportions of the population using the Jackknife technique (Cochran 1977). The proportions of immature Black Brant in the population (i.e. annual productivity) and Grey-bellied Brant were examined to determine how each group contributed to annual increases in the population. I used the equations:

$$
\begin{align*}
& N_{y+1}=N_{y}-M_{y}+\left(I_{y+1}\right)  \tag{Equation2a}\\
& N_{y+1}=N_{y}-M_{y}+\left(G_{y+1}\right)  \tag{Equation2b}\\
& N_{y+1}=N_{y}-M_{y}+\left(I_{y+1}+G_{y+1}\right) \tag{Equation2c}
\end{align*}
$$

where $\mathrm{N}=$ total number of wintering Brant, $\mathrm{y}=$ year, $\mathrm{M}=$ number of mortalities, $\mathrm{I}=$ number of immature Black Brant and $G=$ the number of Grey-bellied Brant.
I calculated this set of equations for each index of abundance and compared the results with the observed values to determine which equation best predicted population growth. The estimates of immature Black Brant and Grey-bellied Brant were obtained from the survey data. Mortality consisted of two components, natural and hunting mortality. I assumed a 6\% annual natural mortality as observed in other Brant and goose populations (Ebbinge 1991). Hunting mortality is estimated under the harvest section of this chapter.

I also modeled population increase using exponential growth equations. The models were designed to show the level of annual productivity necessary to account for the annual increases in the population. I modeled maximum and mean count data that were available between 199394 and 1999-2000. I compared the results with the observed annual productivity to see if the values generated by the models were supported by the data. An exponential model is based on continuous growth so the instantaneous rate of increase was modified to provide an estimate of annual productivity based on discrete growth (Caughley and Birch 1971) as Brant have one reproductive season each year:

$$
\begin{equation*}
r_{d}=e^{r t}-1 \tag{Equation3}
\end{equation*}
$$

Where $r_{d}=$ estimate of annual productivity based on discrete growth, $r=$ the instantaneous rate of increase and $t=$ time in years.

## Distribution

Brant wintering in Boundary Bay have recently extended their winter range to include Roberts Bank. In terms of management, it may be important to determine the extent that Brant are using Roberts Bank, if the use is increasing, and if Roberts Bank is used by a specific component of the population. Using Z-tests (Zar 1974), the mean daily numbers and population composition data from each site were compared to see if the wintering population was randomly distributed. Leg-band data were used to determine if breeding ground origin influenced distribution. The proportion of Brant from each breeding ground was transformed using the arcsine $\sqrt{ }$ p transformation (Zar 1974) and tested with ANOVAs.

## Harvest

The impact of the spring hunt (March 1-10) is an important management issue for Brant wintering in Boundary Bay and Roberts Bank and to other wintering populations on the Pacific Flyway. To estimate the number of Brant shot from each wintering area I constructed a
population profile of Brant present in Boundary Bay and Roberts Bank. To do this leg-band data were collected during three 10-day observation periods, before, during and after the hunting season. For each period I identified the wintering ground affiliation of observed individual bands using leg-band sighting data collected throughout the Pacific Flyway (Washington, Oregon, California, and Mexico). Therefore, for each 10-day period, I was able estimate the proportion of Brant that had wintered in Boundary Bay and Roberts Bank compared to the proportion of spring migrants from Washington, Oregon, California and Mexico. Using the population profile I was able to identify when spring migrants arrived in Boundary Bay and Roberts Bank. To determine if the timing of arrival of migrants was different each year I conducted a chi-square frequency analysis on the leg-band data (Zar 1974).

Using the population profile I was also able to estimate how many birds from each wintering area were harvested during the British Columbia Brant hunting season. I pooled the data from the 10-day periods before and during the hunt across the three years of study. By pooling the data I estimated the maximum numbers of Brant from each wintering area that were susceptible to hunting. I expressed these numbers as proportions of the average number of Brant harvested during the hunt and as proportions of each of the respective wintering populations. I assumed that $8-10 \%$ of Black Brant in the Pacific Flyway Brant Population was banded and bands were equally represented in each wintering area. This assumption was based on legband observations collected from wintering areas in British Columbia, Oregon, and California. The proportion of banded Black Brant at each wintering area was consistent as was the mix of birds from each of the Arctic breeding and moulting areas (Lee unpubl. data, Morrow unpubl. data, Pitkin unpubl. data, Reed 1997). This mixing is consistent throughout the Flyway; therefore, it is reasonable to assume the band distribution is also consistent. Grey-bellied Brant were not banded at the time of this study and therefore excluded from the analysis (see Appendix A for assumptions and calculations).

### 2.3 Results

### 2.3.1 Abundance

The maximum number of wintering Brant observed in Boundary Bay and Roberts Bank for each winter period was 619 (Dec 23, 1997), 985 (Jan 17, 1999) and 1238 (Dec 17, 1999) (Figure 3). The estimates of cumulative Brant goose-days showed that Boundary Bay and Roberts Bank supported increasing Brant usage each year (Table 1).

Table 1. Indices of Brant abundance and cumulative Brant goose-days for each year (15 December to 31 January).

| Index | Year |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 95\% CI | 1998-99 | 95\% CI | 1999- | $\mathbf{9 5 \%} \mathbf{C l}$ |
| Number of Surveys |  | $*$ | 20 | $*$ | 26 |  |
| Cumulative Goose- |  | $*$ | 37529 | $*$ | 54094 | $*$ |
| days |  |  |  |  |  |  |
| Mean Count | 225.9 | 93.6 | 506.0 | 103.1 | 690.4 | 110.2 |
| Maximum Count | 619 | $*$ | 985 | $*$ | 1238 | $*$ |





Figure 3. Maximum daily number of Brant in Boundary Bay and Roberts Bank. White points between the two bars indicate surveys conducted during the Brant hunt (March 1-10).

### 2.3.2 Population Composition and Increase

Population composition is shown in Table 2. The results from Equations 2a-c are in Table 3 and demonstrate that Equation 2c best predicts population increase.

Table 2. Population composition of Brant wintering in Boundary Bay (BB) and Roberts Bank (RB).

| Year | Site | BLACK |  | IMMATURE BLACK |  | GREY-BELLY |  | IMMATURE GREY-BELLY |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | \% | $\begin{gathered} 95 \% \\ \mathrm{Cl} \end{gathered}$ | \% | $\begin{gathered} 95 \% \\ \mathrm{Cl} \end{gathered}$ | \% | $\begin{gathered} 95 \% \\ \mathrm{Cl} \\ \hline \end{gathered}$ | \% | $\begin{gathered} 95 \% \\ \mathrm{Cl} \end{gathered}$ |
| 1997-98 | BB | 86.6 | 3.6 | 7.3 | 1.4 | 13.4 | 7.2 | * | * |
|  | RB | * | * | * | * | * | * | * | * |
|  | $\begin{gathered} \text { BB\&R } \\ \text { B } \end{gathered}$ | * | * | * | * | * | * | * | * |
| 1998-99 | BB | 85.4 | 3.2 | 9.7 | 2.2 | 14.6 | 6.4 | * | * |
|  | RB | 97.0 | 1.1 | 45.1 | 9.4 | 3.0 | 2.2 | * | * |
|  | $\begin{gathered} \text { BB\&R } \\ \text { B } \end{gathered}$ | 88.7 | 2.2 | 17.2 | 5.2 | 11.3 | 4.6 | * | * |
| 1999-2000 | BB | 80.5 | 3.0 | 16.1 | 1.2 | 19.5 | 3.0 | 27.6 | 13.8 |
|  | RB | 93.3 | 5.4 | 34.7 | 7.2 | 6.7 | 5.4 | $60.0{ }^{1}$ | * |
|  | $\begin{gathered} \text { BB\&R } \\ \text { B } \end{gathered}$ | 81.9 | 3.0 | 19.6 | 3.8 | 18.1 | 3.0 | 27.9 | 12.6 |

${ }^{*}$ indicates no data, ${ }^{1}$ this estimate based on one sample only, others based on minimum $\mathrm{n}=20$.

Table 3. Results from Equations 2a-c using maximum and mean indices of abundance. The observed maximums and means are provided for comparison.

| $\mathrm{y}=$ | 2a Equation $2 c$ |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | 199 | -98 | 1998 |  | 1997-98 |  | 1998-99 |  | 1997-98 |  | 1998-99 |  |
|  | $\mathbf{N}_{(y+1)}$ | $\begin{gathered} 95 \% \\ \text { CI } \end{gathered}$ | $\mathbf{N}_{(\mathrm{y}+1)}$ | $\begin{gathered} 95 \% \\ \text { CI } \end{gathered}$ | $\mathrm{N}_{(\mathrm{y}+1)}$ | $\begin{gathered} 95 \% \\ \text { CI } \end{gathered}$ | $\mathbf{N}_{(y+1)}$ | $\begin{gathered} 95 \% \\ \mathrm{CI} \end{gathered}$ | $\mathbf{N}_{(y+1)}$ | $\begin{gathered} 95 \% \\ \text { CI } \end{gathered}$ | $\mathrm{N}_{(\mathrm{y}+1)}$ | $\begin{gathered} 95 \% \\ \mathrm{CI} \end{gathered}$ |
| Max | 532 | * | 1017 | * | 543 | * | 915 | * | 615 | * | 1128 | * |
| Obs. Max | 985 | * | 1254 | * | 985 | * | 1254 | * | 985 | * | 1254 | * |
| Mean | 176 | 139 | 488 | 166 | 147 | 149 | 478 | 166 | 234 | 152 | 613 | 188 |
| Obs. <br> Mean | 506 | 103 | 690 | 110 | 506 | 103 | 690 | 110 | 506 | 103 | 690 | 110 |

Using maximum counts as the index of abundance, the exponential growth model equation (Equation 4a) was:

$$
\begin{equation*}
N_{t}=284 e^{0.1964 t} \tag{Equation4a}
\end{equation*}
$$

where $\mathrm{N}=$ population size, $284=$ population size at $\mathrm{t}=0, \mathrm{t}=$ time in years, and $0.1964=$ the instantaneous rate of increase (Figure 4). Using Equation 3 to adjust the instantaneous rate of increase to one based on discrete growth, the estimate of annual productivity was $21.7 \%$.
The exponential growth model equation generated with mean counts (Equation 4b) was

$$
\begin{equation*}
N_{t}=114 e^{0.2274 t} \tag{Equation4b}
\end{equation*}
$$

where $N=$ mean population size, $114=$ mean population size at $t=0, t=$ time in years, and 0.2274 = the instantaneous rate of increase (Figure 5). Therefore, the estimate of annual productivity was $25.5 \%$.


Figure 4. Exponential growth curve model from maximum counts of the Boundary Bay-Roberts Bank Brant population between 1993-94 and 1999-2000 (other data sources: 1993-94 Reed unpubl. data, 1994-95 and 1995-96 Reed 1997, 1996-97 Gowans unpubl. data).


Figure 5. Exponential growth curve model from mean counts of the Boundary Bay-Roberts Bank Brant population between 1993-94 and 1999-2000 (other data sources: 1993-94 Reed unpubl. data, 1994-95 and 1995-96 Reed 1997, 1996-97 Gowans unpubl. report).

### 2.3.3 Distribution

In 1997-98, during regular surveys, Brant were observed at Roberts Bank on only three occasions. In 1998-99 there was no difference in the mean daily use between the two sites ( $\mathrm{t}_{38}$ $=2.02, \mathrm{p}=0.207$ ), and in 1999-2000, Brant used Roberts Bank more than Boundary Bay ( $\mathrm{t}_{118}=$ 3.20, p = 0.0017).

In 1998-99 and 1999-2000, larger proportions of immature Brant used Roberts Bank compared to Boundary Bay (1998-99: $Z=6.021, p<0.0001 ; 1999-2000: Z=5.032, p<0.0001$ ). Conversely, larger proportions of Grey-bellied Brant used Boundary Bay compared to Roberts Bank (1998-99: $Z=3.480, p=0.0003 ; 1999-2000: Z=4.146, p<0.0001$ ). Analysis of the leg band data suggests that site selection was not based on breeding ground origin ( $F_{4,24}=0.6831$, $p=0.6119$ ) (Figure 6).

### 2.3.4 Harvest

Counts indicated no increase in the number of Brant in Boundary Bay and Roberts Bank prior to the hunt, suggesting that spring migrants did not arrive before the hunt each year (Figure 3), but leg-band observations suggested otherwise. Birds carrying leg-bands, showing that they had wintered south of BC, were identified in Boundary Bay and Roberts Bank before and during the Brant hunt (Table 4). Over the three years, migrants comprised $21.9 \%$ of the population before, $33.1 \%$ during and $52.9 \%$ after the hunt, strong evidence that migrants were present in the area during the March hunt (Table 4). However, migration appears to have been delayed in the last two years, given that wintering Brant comprised the majority of the birds before and during the hunt ( $\mathrm{X}_{2}^{2}=20.56, \mathrm{p}<0.0001$ ). The proportion of migrants was higher in all three time periods in 1998 suggesting that migrants arrived earlier this year than the other years. The largest numbers of migrants were from Mexico, which is to be expected as $90 \%$ of Black Brant winter there (Sedinger et al. 1994).
During the study, approximately 150-200 Brant were killed during the hunt each year (including crippling mortality) (BC Ministry of Environment Land and Parks unpubl. data and Canadian Wildlife Service unpubl. data). Of these, approximately $70 \%$ were estimated to have wintered in Boundary Bay and Roberts Bank (Table 5), hence 105-140 of these Brant were harvested in each hunting season. This is equivalent to $11-15 \%$ of the Boundary Bay and Roberts Bank wintering population. Less than $1 \%$ of each of the Mexico, California, Oregon and Washington wintering populations were harvested (Appendix A).


Figure 6. Breeding ground origins of Black Brant estimated from leg-band observations.

Table 4. The number of bands and the composition of Brant observed in Boundary Bay and Roberts Bank classified by wintering affiliation during the 10-day periods before, during and after the hunt.

| 10 Day Period | Wintering Ground | 1998 |  | 1999 |  | 2000 |  | 3-Year Mean \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | \# | \% | \# | \% | \# | \% |  |
| Before | Mexico | 11 | 36.7 | 1 | 12.5 | 1 | 3.1 | 17.4 |
|  | California | 4 | 13.3 | 0 | 0.0 | 0 | 0.0 | 4.4 |
|  | Oregon | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 | 0.0 |
|  | Washington | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 | 0.0 |
|  | BC | 15 | 50.0 | 7 | 87.5 | 31 | 96.9 | 78.1 |
|  | Total | 30 | 100.0 | 8 | 100.0 | 32 | 100.0 | 100.0 |
| During | Mexico | 11 | 42.3 | 3 | 23.1 | 2 | 10.5 | 25.3 |
|  | California | 2 | 7.7 | 0.0 | 0.0 | 0 | 0.0 | 2.6 |
|  | Oregon | 0 | 0.0 | 0.0 | 0.0 | 0 | 0.0 | 0.0 |
|  | Washington | 0 | 0.0 | 0.0 | 0.0 | 3 | 15.8 | 5.3 |
|  | BC | 13 | 50.0 | 10 | 76.9 | 14 | 73.7 | 66.9 |
|  | Total | 26 | 100.0 | 13 | 100.0 | 19 | 100.0 | 100.0 |
| After | Mexico | 25 | 58.1 | 2 | 33.3 | 20 | 48.8 | 46.8 |
|  | California | 3 | 7.0 | 0 | 0.0 | 3 | 7.3 | 4.8 |
|  | Oregon | 0 | 0.0 | 0 | 0.0 | 0 | 0.0 | 0.0 |
|  | Washington | 0 | 0.0 | 0 | 0.0 | 1 | 2.4 | 0.8 |
|  | BC | 15 | 34.9 | 4 | 66.7 | 17 | 41.5 | 47.7 |
|  | Total | 43 | 100.0 | 6 | 100.0 | 41 | 100.0 | 100.0 |

Table 5. Estimates of the population composition susceptible to hunting in British Columbia and estimates of the number of Brant harvested from each wintering area.

| $\begin{array}{l}\text { \# Leg-bands Observed From Each Wintering Area } \\ \text { 10 Days }\end{array}$ |  |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
|  | MEX | Cere and 10 Days of the Hunt |  |  |  |  |$]$

MEX = Mexico, CAL = California, OR = Oregon, WA = Washington, BC = British Columbia, T = Total.

### 2.4 Discussion

### 2.4.1 Abundance, Population Composition and Increase

During the course of this study, the number of Brant wintering in Boundary Bay and Roberts Bank increased. I suspect two main factors contributed to the increase: 1) annual productivity, and 2) immigration of Grey-bellied Brant. Another factor may be my own improved ability to survey Brant. However, to reduce measurement error and standardize counts, I worked closely with experienced Brant observers and believe the numbers in this study are valid indicators of population change.

I used exponentials models to estimate the annual rate of population increase between 1993-94 and 1999-2000. The estimated rates were clearly higher than the annual productivity as deduced from juvenile ratios ( $\sim 22 \%$ and $\sim 26 \%$ compared to mean $=14 \%, n=6$ ) (other data sources: Reed 1997, Reed unpubl. data, Gowans unpubl. report).
It was appropriate to use exponential modeling to measure population change because when populations occupy new habitat or rebound from near extinction they are not subjected to limiting conditions and may increase exponentially (Johnson 1994). Exponential increases have been observed in other wintering bird populations recovering from serious declines. Several arctic goose populations have experienced exponential growth after severe population reductions. The Pink-footed Goose (Anser brachyrhynchus) of Iceland and Greenland increased from approximately 20,000 geese in the 1950's to over 200,000 in the 1990's (Fox et al. 1989). A decline in mortality and improved annual productivity ( $\sim 18 \%, n=25$ ), possibly due to the protection of wintering areas, accounted for the increase (Fox et al. 1989). The Svalbard Brant, one of the European stocks of B.b. hrota decreased from 40,000-50,000 Brant in the 1890's (Salomonsen 1958) to 1600-2000 in the 1960-70's (Fog 1972). The cause of the decline was likely a combination of over-hunting, uncontrolled egg-collection, and widespread loss of eelgrass due to eelgrass wasting disease (Madsen et al. 1999). By the 1990's the population was approximately 6000 birds (Clausen et al. 1998) and annual productivity ( $\sim 15 \%, n=16$ ) (Madsen et al. 1989) accounted for the increase.

When I examined population composition to determine if annual productivity and/or immigration were responsible for the rapid growth in Brant in Boundary Bay and Roberts Bank, annual productivity did not account for the increases, even in relatively successful years. Equation 2c, incorporating both annual productivity and immigration of Grey-bellied Brant, best predicted population increase. However, Equations 2a-c are too simplified to detect other processes that may be affecting the population. For example, there may be immigration of Black Brant from
other wintering areas. Grey-bellied Brant are showing changes in their wintering distribution and it is probable that Black Brant are making similar shifts.

Although Arctic-nesting geese show high site-fidelity to winter and spring staging sites (Raveling 1979, Prevett and MacInnes 1980, Hestbeck et al.1991, Percival 1991), Brant have shown the capacity to make dramatic shifts in their wintering sites (Sedinger et al. 1994). Prior to the 1950's, the coast north of Baja, California supported 50-65\% of the Pacific Flyway Brant population. Since then, this proportion has decreased to less than $10 \%$ of the population, coinciding with an increase in Brant numbers in Baja and along the West Coast of Mexico (Sedinger et al. 1994).
Exponential growth cannot occur indefinitely and as a population approaches the carrying capacity of its habitat, the rate of increase will decline (Begon et al. 1990). It is not likely that this population is currently subjected to limiting conditions as historically the Fraser River delta supported thousands of wintering Brant and, in addition, the biomass of eelgrass has increased in recent years (Ward et al. 1992, Tarbotton and Harrison 1996). However, as the wintering population increases, factors such as intraspecific competition and predation pressure are likely to also increase, leading to changes in the population growth rate. The growth observed in this study is likely the first portion of a logistic curve. Continued monitoring of the wintering population will eventually reveal the maximum number of Brant that can be sustained by Boundary Bay and Roberts Bank. Understanding this growth curve is fundamental for Brant management.

### 2.4.2 Distribution

During this study Brant in Boundary Bay and Roberts Bank appeared to change their local winter distribution patterns. Daily use of Roberts Bank did not occur until 1998-99, but by 19992000 the mean daily number of Brant was higher at Roberts Bank than Boundary Bay. Roberts Bank is subjected to fewer disturbances than Boundary Bay (Chapter 3) and combined with the expansion of eelgrass (Tarbotton and Harrison 1996) may provide Brant with new foraging opportunities. Significantly higher proportions of young were observed at Roberts Bank than at Boundary Bay, whereas, there were higher proportions of Grey-bellied Brant observed at Boundary Bay than at Roberts Bank. Leg-band data showed mixing of Black Brant from the different breeding areas at each site; however, anecdotal behaviour observations suggest the Black and Grey-bellied Brant remained separate and agonistic encounters were common when the two groups were in close proximity.

### 2.4.3 Harvest

Band observation data suggest that Brant from more southern wintering areas begin to arrive in Boundary Bay and Roberts Bank before the hunting season each year; however, in the last two years of this study, migration appears to have been delayed and wintering Brant comprised the majority of the population during the hunt. This suggests that a) early migrants were not detected in regular counts and, therefore, counts are not a good indicator of when migrants arrive (e.g. Figure 3) and b) timing of migration in relation to the hunt is important to the conservation of wintering Brant.

There are several possibilities as to why Brant numbers do not appear to increase with the initial arrival of spring migrants to Boundary Bay and Roberts Bank. It is likely that some wintering Brant are leaving at the same time as spring migrants are arriving. Mark-resight analysis (Reed 1997) has shown that Brant wintering in Boundary Bay begin to emigrate from the area in early March. The rate of emigration remains constant throughout spring migration. The research also suggested that spring migrants from other areas arrive in Boundary Bay by the third week in February. Therefore, counts do not detect changes in the number of wintering and migrant birds. Migrants may also use areas outside of Boundary Bay and Roberts Bank that are not observed in regular surveys. With the addition of migrants into the area, local distribution patterns of wintering Brant may change such that wintering Brant are more difficult to detect. The distribution of wintering Brant and spring migrants may be critical when examining the band data to estimate harvest. It is possible that migrants entering the area do not initially stage in areas that are accessible for band reading. If this is the case, the band observations will be biased high for winter birds.
During this study, an estimated 11-15\% of the Boundary Bay-Roberts Bank wintering Brant population was harvested annually. While this estimate is below the overall rate of increase it is close to the estimate of annual productivity ( $\sim 14 \%$ over 6 years). If there were repeated years of low annual productivity and delayed migration, the hunt could heavily impact the wintering population.

The timing of the BC hunt remains controversial as other states along the Pacific Flyway (e.g. Washington, Oregon, and California) are concerned that their respective wintering populations are harvested during the BC spring hunt. Results from this study suggest that less than $1 \%$ of the Brant from each wintering area south of $B C$ are harvested in the $B C$ hunt. However, this estimate does not include the Grey-bellied Brant from Washington State. Furthermore, in this analysis smaller populations are likely to be under-estimated as there is less likelihood of
sighting these bands in Boundary Bay and Roberts Bank. I suspect this may be the case with the estimates of Brant that had wintered in Oregon.

### 2.4.4 Conclusions and Recommendations

The continued increase in the Boundary Bay-Roberts Bank wintering Brant population is not due to annual productivity alone, but immigration from just south of the Canada-USA border. The most apparent contribution is from the Grey-bellied Brant population that traditionally winters in Washington State. The Boundary Bay-Roberts Bank population has expanded its local winter distribution. Should expansion continue, it may lead to a high degree of connectivity between the $B C$ and Washington populations that cannot be ignored when designing management plans for Brant in BC. Moreover, management plans must consider that the wintering population is comprised of Brant from all the Arctic nesting grounds, and during the spring, from the majority of the wintering areas. Continued monitoring of the Boundary Bay-Roberts Bank population is essential for adaptive management and understanding the processes affecting this and other wintering populations along the Pacific Flyway.

## 3. BRANT WINTERING IN BOUNDARY BAY AND ROBERTS BANK, BRITISH COLUMBIA: ACTIVITY BUDGETS AND EFFECTS OF DISTURBANCE

### 3.1 Introduction

Black Brant (Branta bernicla nigricans) have historically wintered in large numbers along the coast of British Columbia (Campbell et al. 1990). However, over the past 100 years the number of wintering Brant has declined drastically, reaching historic lows in the 1980's (Campbell et al. 1990). This was particularly true on the Fraser River delta, near Vancouver, where Christmas Bird Count records show years of few or no Brant during the 1960's to 1980's. Since then, the wintering population increased to approximately 1200 Brant by 2000 (unpubl. data). In Boundary Bay and Roberts Bank Brant forage on eelgrass (Zostera marina and Z. japonica) and haul out on spits to preen and ingest grit (Campbell et al. 1990). Eelgrass is important to Brant as a primary food source (Cottam et al. 1944, Einarsen 1965, Charman 1977, Ganter 2000). Geese have high energy costs and vegetation is generally nutrient-poor so wintering geese need to forage for a large portion of the day to satisfy their energy requirements (Riddington et al. 1996).
Farms surround Boundary Bay and Roberts Bank, but increasing residential and commercial development is also encroaching. The associated increase in disturbance to wintering Brant is a management concern. Disturbance is a discrete event that alters the ability of an individual to
obtain resources (White and Pickett 1995). Disturbance may reduce food intake through interruption of foraging bouts or by displacement from feeding areas (Madsen 1985, Bélanger and Bédard 1989). Disturbance may cause changes in behaviour, which can affect distribution, survival and reproductive output and increase energy expenditure from additional time in flight (Madsen 1994, Riddington et al. 1996). Time and energy spent responding to disturbance reduces the ability of Brant to feed and perform maintenance activities (Percival and Evans 1997). Geese may abandon a wintering site if the amount of energy they expend outweighs the returns gained from available food resources (Bélanger and Bédard 1989, Stock 1993). Brant did not consistently use Roberts Bank as a wintering area until 1998-99 (pers. obs). Here, eelgrass biomass has increased substantially (Ward et al. 1992, Tarbotton and Harrison 1996) and the area is subjected to fewer disturbances than Boundary Bay. One or both of these reasons may be responsible for the current use of Roberts Bank by Brant.
This study examines activity budgets and the effect of disturbance on wintering Brant in Boundary Bay and Roberts Bank. In doing so the following are addressed:

- energy budgets,
- causes and frequency of disturbance, and
- the effect of disturbance on energy budgets.

Results will help managers understand why local shifts in distribution of wintering Brant have occurred and also identify sources of disturbance for management purposes. Further, findings from this study will help determine if wintering Brant in Boundary Bay and Roberts Bank are being subjected to disturbance levels that might lead to eventual abandonment of the area.

### 3.2 Methods

### 3.2.1 Study Area

Boundary Bay and Roberts Bank are 25 km south of Vancouver, British Columbia, on the Fraser River delta (Figure 2). Observations of Brant in Boundary Bay were made from Beach Grove where Brant are frequently found. Two prominent features of Roberts Bank are the Tsawwassen Ferry Terminal and the Deltaport-Westshore Terminal Causeway. Brant haul out on the mudflats between the bases of the two causeways and were observed from shore.

### 3.2.2 Data Collection

In 1998 and 2000, Brant behaviour was monitored during January and the beginning of February in Boundary Bay. Roberts Bank was surveyed only in 2000. Surveys began at sunrise
or with the arrival of Brant and lasted until sunset or until no Brant remained at the site. Surveys were conducted with a 20-60x spotting scope.

## Activity Budgets

Scan-sampling (Altmann 1977) was conducted by scanning flocks of Brant and recording the behaviour of individual birds. Behaviours were classified as swimming, preening, feeding, social, resting and alert (Goudie and Hearne 1994, Stock 1993, Riddington et al. 1996). To ensure random sampling, I began scans by focusing the spotting scope at one edge of a flock, recording the behaviour of the bird at the left side of the field of vision, and then adjusting the scope to a new position. I scanned back and forth across the flock for 15 minutes. Each scan was separated by 1-15 minute intervals, the timing determined by randomly generated numbers. Brant in flight were not included.

## Disturbance

Causes of disturbance were categorized as natural, human-related, or unknown (Goudie and Hearne 1994, Mooij 1992, Stock 1993, Riddington et al. 1996). Specific causes of natural and human-related disturbances were also noted. A disturbance event was defined in terms of a) the length time of the disturbance and b) the effect of the disturbance in terms of distance moved. The length of a disturbance event was defined as the time beginning when $50 \%$ of the flock responded by becoming alert and then flying until the time when $50 \%$ of the flock returned and resumed normal activity. The time between becoming alert and flying was not measurable (less than 1 second). Distances were recorded as $<100 \mathrm{~m},<200 \mathrm{~m},<300 \mathrm{~m},<400 \mathrm{~m},<500,>500 \mathrm{~m}$ (still in view), or out of view. All disturbances were recorded whether or not they occurred within a 15-minute scan.

### 3.2.3 Analysis

Activities were converted into proportions for each scan. The proportional data were transformed using the arcsine $\sqrt{ }$ p transformation (Zar 1974). MANOVAs were used to test for differences between years at Boundary Bay and differences between sites for the 2000 data. The tests were Bonferroni adjusted. Activity budgets were converted to energy budgets using the following equation (Wooley and Owen 1978, Gauthier et al. 1984, Mooij 1992, Riddington et al. 1996):

$$
\text { Energy Expended }=(\text { Activity Cost }) \times(\text { Basal Metabolic Rate }) \times(\% \text { Time of the Activity) }
$$ (Equation 5)

On the basis of previous calculations (Owen et al. 1992, Wooley and Owen 1978, Gauthier et al. 1984, Bélanger and Bédard 1990, Mooij 1992 and Riddington et al. 1996) activity costs were assumed to be as follows: resting $1.1 \times$ BMR, preening 2.3 x , feeding 1.7 x , alert 2.1 x , swimming 2.2 x , social 6.4 x and disturbance 15 x . The cost of disturbance was assumed to be equivalent to the cost of flight as flight was always observed as the response to disturbance. Basal metabolic rate (BMR) was calculated as

$$
\mathrm{BMR}=327.8 \times \mathrm{W}^{723}
$$

## (Equation 6)

where BMR is in $\mathrm{kJ} /$ day and W is the mass in kilograms (Lasiewski and Dawson 1967, Riddington et al. 1996). Therefore, the BMR for a wintering Brant goose of mean weight 1458 g (Reed et al. 1998) was estimated as $430.4 \mathrm{~kJ} /$ day
The daily energy expenditure (DEE) is the amount of energy expended by an individual during average daily activity. DEE is linearly related to BMR; therefore, the larger the bird, the greater the energy expended for daily activities. The relationship between BMR and the theoretical DEE (DEE ${ }_{t}$ ) for an Arctic-nesting goose is (Drent et al. 1978, Bédard and Gauthier 1989, Mooij 1992):

$$
\mathrm{DEE}_{\mathrm{t}}=2.55 \times \mathrm{BMR}
$$

## (Equation 7)

Using Equation 7, the $\mathrm{DEE}_{\mathrm{t}}$ for a wintering Brant in Boundary Bay and Roberts Bank is predicted to be $1097.5 \mathrm{~kJ} /$ day. Sites at which observed DEEs ( $\mathrm{DEE}_{0}$ ) are above this level may be too energetically demanding for Brant. The DEE ${ }_{0}$ for each site and year was calculated by summing activity costs. The amount of DEE specifically due to human disturbance was also calculated. The frequency of disturbance, the mean duration, and the mean distance moved in response to disturbance was calculated for each site. T-tests were used to identify differences in disturbance between years at Boundary Bay and between sites in 2000.

### 3.3 Results

### 3.3.1 Activity Budgets

In 1998, 63 15-minute scans were conducted at Boundary Bay. In 2000, 69 were conducted at Boundary Bay and 45 at Roberts Bank. Swimming was the dominant activity, followed by
feeding and preening. The remaining time was spent responding to disturbance with small amounts of time resting, socializing, and being alert (Table 6).

Table 6. Time-activity budgets (\%) for Brant in Boundary Bay and Roberts Bank.

| Behaviour | $\begin{gathered} \hline \text { Boundary Bay } \\ 1998 \\ \mathrm{n}=63 \\ \hline \end{gathered}$ |  | $\begin{gathered} \hline \text { Boundary Bay } \\ 2000 \\ \mathrm{n}=69 \\ \hline \end{gathered}$ |  | Roberts Bank 2000 $\mathrm{n}=45$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | \% | SE | \% | SE | \% | SE |
| Disturbed | 6.7 | 2.2 | 10.7 | 2.5 | 3.8 | 0.7 |
| Swimming | 44.2 | 2.4 | 29.8 | 2.0 | 44.9 | 2.4 |
| Preening | 16.5 | 2.6 | 26.9 | 0.7 | 16.9 | 1.3 |
| Feeding | 31.5 | 2.4 | 18.9 | 0.5 | 24.4 | 0.5 |
| Social | 0.5 | 0.1 | 2.1 | 0.1 | 0.5 | 0.3 |
| Resting | 0.2 | 0.1 | 0.1 | 0.0 | 1.7 | 0.0 |
| Alert | 0.4 | 0.1 | 11.5 | 0.1 | 7.8 | 0.0 |

Overall, time-activity budgets differed across years in Boundary Bay ( $F_{7,123}=78.75, p<0.0001$ ) and between sites in $2000\left(F_{7,106}=18.36, p<0.0001\right)$ (Table 7), but there were no significant differences in the total amount of disturbance.

Table 7. $F$ - and $p$-values from MANOVAs testing differences in time-activity budgets from Boundary Bay (1998 and 2000) and Boundary Bay and Roberts Bank (2000). The Bonferroni adjustment is $p=0.0073$ for $\alpha=0.05$.

| Activity Tested | Boundary Bay 1998 \& 2000 | Boundary Bay and Roberts Bank |
| :--- | :--- | :---: |
| 2000 |  |  |
| Overall | $\mathrm{F}_{7,123}=78.75, \mathrm{p}<0.0001$ | $\mathrm{~F}_{7,106}=18.36, \mathrm{p}<0.0001$ |
| Disturbed | $\mathrm{F}_{1,129}=2.83, \mathrm{p}=0.0950^{*}$ | $\mathrm{~F}_{1,112}=4.38, \mathrm{p}=0.0337^{*}$ |
| Swimming | $\mathrm{F}_{1,129}=10.86, \mathrm{p}=0.0013$ | $\mathrm{~F}_{1,112}=16.96, \mathrm{p}<0.0001$ |
| Preening | $\mathrm{F}_{1,129}=13.05, \mathrm{p}=0.0004$ | $\mathrm{~F}_{1,112}=12.53, \mathrm{p}=0.0006$ |
| Feeding | $\mathrm{F}_{1,129}=19.70, \mathrm{p}<0.0001$ | $\mathrm{~F}_{1,112}=4.12, \mathrm{p}=0.0446^{*}$ |
| Social | $\mathrm{F}_{1,129}=79.75, \mathrm{p}=0.0001$ | $\mathrm{~F}_{1,112}=65.42, \mathrm{p}<0.0001$ |
| Resting | $\mathrm{F}_{1,129}=1.19, \mathrm{p}=0.2778^{*}$ | $\mathrm{~F}_{1,112}=8.38, \mathrm{p}=0.0045$ |
| Alert | $\mathrm{F}_{1,129}=360.51, \mathrm{p}<0.0001$ | $\mathrm{~F}_{1,112}=9.65, \mathrm{p}=0.0024$ |

* not significant

Table 8 shows the observed DEE in the three samples and the proportion associated with disturbance.

Table 8. Energy costs ( $k J /$ day ) for daily activities, total $D E E_{o}$ and the proportion of the $D E E_{o}$ due to disturbance ( $\mathrm{DEE}_{\mathrm{t}}$ for a Brant is $1097.5 \mathrm{~kJ} /$ day ).

| Activity | Boundary Bay |  | 1998 |  | Boundary Bay <br> 2000 |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| Roberts Bank <br> 2000 |  |  |  |  |  |  |
|  | Cost | SE | Cost | SE | Cost | SE |
| Disturbed | 432.6 | 142.0 | 690.8 | 161.4 | 245.3 | 45.2 |
| Swimming | 418.5 | 22.7 | 282.2 | 18.9 | 425.1 | 22.7 |
| Preening | 163.3 | 25.7 | 266.3 | 6.9 | 167.3 | 12.9 |
| Feeding | 230.4 | 17.5 | 138.3 | 3.7 | 178.5 | 3.7 |
| Social | 13.8 | 2.75 | 57.8 | 2.8 | 13.8 | 8.3 |
| Resting | 0.9 | 0.5 | 0.5 | 0.0 | 8.0 | 0 |
| Alert | 3.6 | 0.9 | 103.9 | 0.9 | 70.5 | 0 |
| Total (DEE $)$ | 1263.2 | 212.2 | 1539.8 | 194.6 | 1108.6 | 92.7 |
| Energy expended <br> if Disturbance <br> removed | 830.7 | 70.1 | 849.0 | 33.1 | 863.3 | 47.5 |
| Energy for <br> Disturbance (\%) | $34.2 \%$ | $11.2 \%$ | $44.9 \%$ | $10.5 \%$ | $22.1 \%$ | $4.1 \%$ |

In 1998, disturbance comprised only $6.7 \%$ of the time-activity budget in Boundary Bay, but accounted for $34 \%$ of the total energy budget (Table 8). In 2000, estimated energy expenditure due to disturbance was $45 \%$ in Boundary Bay and $22 \%$ in Roberts Bank. The DEE ${ }_{0}$ for Boundary Bay was greater than the $\mathrm{DEE}_{\mathrm{t}}$ in both years, but close to the $\mathrm{DEE}_{\mathrm{t}}$ in Roberts Bank in 2000. Subtracting energy costs due to disturbance resulted in totals that were similar in all three cases and below the $\mathrm{DEE}_{\mathrm{t}}$ of $1097.5 \mathrm{~kJ} /$ day.

### 3.3.2 Disturbance

Natural causes were the most important source of disturbance in both years at Boundary Bay. At Roberts Bank, the largest number of causes was unknown (Table 9).

Table 9. Disturbance (\%) at Boundary Bay and Roberts Bank from natural, human and unknown causes.

| Cause | Boundary Bay <br> $\mathbf{1 9 9 8}(\mathbf{n}=\mathbf{9 1})$ |  | Boundary Bay <br> $\mathbf{2 0 0 0}(\mathbf{n}=\mathbf{1 0 6})$ |  | Roberts Bank 2000 <br> $(\mathbf{n}=\mathbf{3 7})$ |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\%$ | SE | $\%$ | SE | $\%$ | SE |
| Human | 13.2 | 2.4 | 8.5 | 2.3 | $11.4^{*}$ | 11.0 |
| Natural | 76.9 | 3.0 | 61.3 | 5.7 | 25.7 | 8.9 |
| Unknown | 9.9 | 3.1 | 30.2 | 6.8 | 62.9 | 9.9 |

*All the human disturbances at Roberts Bank were observed on one day.

Bald eagles (Haliaeetus leucocephalus) were the dominant source of natural disturbance, causing $93 \%$ (1998) and $97 \%$ (2000) of natural disturbances in Boundary Bay and all natural disturbances in Roberts Bank. In both years, the mean frequency of all disturbances was 2.17/hour in Boundary Bay. At Roberts Bank the frequency was significantly less at 1.32 /hour ( $\mathrm{t}_{17}=-2.22, \mathrm{p}=0.0199$ ).
There was no difference in the mean duration of disturbance between years at Boundary Bay ( $t_{129}=1.166, p=0.2456$ ); however responses to disturbance were longer at Boundary Bay than at Roberts Bank in $2000\left(\mathrm{t}_{112}=2.138, \mathrm{p}=0.0340\right)$. There was no difference in the mean distance moved by Brant in response to disturbance either between years at Boundary Bay $\left(\mathrm{t}_{199}\right.$ $=1.130, p=0.2584)$ or between sites in $2000\left(\mathrm{t}_{142}=0.5151, \mathrm{p}=0.6073\right)$.
Using estimates for cause of disturbance (Table 9), the amount of energy specifically expended responding to human disturbance was partitioned (Table 10). I assumed the ratio of known human and known natural disturbances was similar in the unknown category. Removal of human-caused disturbance suggests that Brant at Boundary Bay would likely still have a DEE above 1097.5kJ/day, but that Brant at Roberts Bank could have a DEE ${ }_{0}$ below that level (Table 10).

Table 10. Energy expended ( $\mathrm{kJ} /$ day and $\%$ of $D E E_{0}$ ) due to human disturbances, $\mathrm{DEE}_{\mathrm{o}}$ with human disturbance removed ( $\mathrm{Kj} / \mathrm{day}$ ) and the $\mathrm{DEE}_{\mathrm{t}}$ (for comparison).

|  | Boundary Bay 1998 |  | Boundary Bay 2000 |  | Roberts Bank 2000 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Cost | SE | Cost | SE | Cost | SE |
| Cost of Human Disturbance (kJ/day) | 62.7 | 12.1 | 76.5 | 20.0 | 45.6 | 30.2 |
| Cost of Human Disturbance as a \% of DEE。 | 4.9 | 1.6 | 4.9 | 2.7 | 4.1 | 2.8 |
| DEE ${ }_{0}$ - Cost of Human Disturbance ( $\mathrm{Kj} /$ day) | 1200.5 | 224.3 | 1463.3 | 214.6 | 1063.0 | 122.9 |
| DEE ${ }_{\text {t }}(\mathrm{Kj} /$ day $)$ | 1097.5 | * | 1097.5 | * | 1097.5 | * |

### 3.4 Discussion

The calculation of DEE relies on accurate measurements for activity costs. In this study, these costs were obtained from the literature. There is agreement among most studies on the costs of activities; however, accepted costs for flight range between 8 and $15 \times$ BMR (Tucker 1969, Hart and Berger 1972, Drent et al. 1978, Gauthier et al. 1984, Bélanger and Bédard 1990, Riddington et al. 1996). Long flights are less energy demanding than shorter flights as take-off is the most energy demanding part of flying (Utter and Lefevre 1970, Butler et al. 1977). Therefore, in this study, flight caused by disturbance was estimated as $15 \times$ BMR which is consistent with other energetics studies of Arctic-nesting geese (Gauthier et al. 1984, Bélanger and Bédard 1990, Riddington et al. 1996).
Thermoregulatory costs were assumed to be negligible, as Arctic-nesting geese maintain their body temperature by the heat generated in normal daily activities (West and Norton 1975). This is not the case if the ambient temperature falls below a critical temperature, which was estimated between $-2^{\circ} \mathrm{C}$ and $2^{\circ} \mathrm{C}$ for Brant (West and Norton 1975, Wooley and Owen 1978, Owen and Reinecke 1977). The average temperatures during this study were $5.0^{\circ} \mathrm{C}$ (1998) and $6.4^{\circ} \mathrm{C}(2000)$; however, in both years there were instances when the temperature dropped below $-2^{\circ} \mathrm{C}$. Therefore, the $\mathrm{DEE}_{0}$ estimates in this study are conservative.

Calculating the DEE $_{0}$ based on activity budgets is a good method for comparing sites; however, only diurnal observations were made in this study and activity budget analysis did not consider differences between nocturnal and diurnal behaviour. Several studies show geese will feed nocturnally which may compensate for interruptions in diurnal feeding. Feeding comprised a lower proportion of the activity budget at both sites and in both years than has been reported in other studies of Arctic-nesting geese. Mooij (1992) estimated that Bean Geese (Anser fabalis rossicus) staging in Germany spent $55 \%$ of their time feeding. Other studies show feeding for $47-80 \%$ of the day (Gauthier et al. 1988, Bélanger and Bédard 1990, Black et al. 1991, Riddington et al. 1996). Boundary Bay and Roberts Bank experience extreme low nocturnal tides in the winter. During these tides, eelgrass beds are exposed and may provide Brant with foraging opportunities. However, Bélanger and Bédard (1990) showed that in greater snow geese (Chen caerulescens atlantica) more than two disturbance events per hour caused an energy deficit that could not be compensated by nocturnal feeding. In Boundary Bay, the frequency of disturbance was $2.17 /$ hour.
On the Yukon-Alaska North Slope staging snow geese that experienced a seasonal average frequency of 0.5 disturbances per hour experienced a $20 \%$ decrease in energy reserves (Davis and Wiseley 1974). In Britain, wintering Brent geese doubled their flying time due to disturbance, resulting in a loss of 4.9-11.7\% of their daily feeding time and an estimated energy expenditure of $31 \%$ (White-Robinson 1982). The frequency of disturbance in Boundary Bay translated to an energy loss of approximately $34 \%$ in 1998 and $45 \%$ in 2000. The frequency of disturbance at Roberts Bank translated to an estimated energy expenditure of $22 \%$.

Choice of wintering site is determined, in part, by trade-offs between energy gain (from food intake) and energy expenditure (through foraging, maintenance, and movement to and within sites) (Riddington et al. 1996). Response to disturbance involves flight, the most energetically expensive of daily activities (Tucker 1969, Hart and Berger 1972, Drent et al. 1978, Gauthier et al. 1984, Bélanger and Bédard 1990, Riddington et al. 1996). Therefore, decreased time available for feeding and increased energy expenditure due to disturbance may result in a change of feeding sites. If alternate sites are not available, birds may respond by increasing their feeding time, feeding at night, increasing their intake rate, processing more efficiently, or by losing weight (Riddington et al. 1996).

Preening, alert and social behaviour were comparable with other studies, but in this study, birds spent more time swimming. This is because Pacific Brant are strictly a coastline species; they
do not graze in fields, but only in estuaries and consequently, swimming in shallow waters between feeding sessions is common.

Disturbance also affects the amount of swimming. After a disturbance event, although the flock may have resumed normal activity, Brant tended to swim until they were secure enough to resume feeding again. Alert is another behaviour affected by disturbance. The amount of swimming and alert behaviour that is due to disturbance could not be partitioned in this study, suggesting the estimated impacts of disturbance are conservative (also see Stock 1993). Predator-like birds with slow wing beats (e.g. Great Blue Heron, Ardea herodias) and raptors are the major sources of natural disturbance in a number of other geese-disturbance studies (e.g. Bélanger and Bédard 1989, Ward et al. 1994, Riddington et al. 1996). The largest identified source of disturbance in this study was from natural causes and in almost all cases was due to Bald Eagles. In most other studies of disturbance, human-related causes were greater than natural ones and management recommendations focused on reducing human impacts.
In this study, the energy expended by Brant due to human disturbance was approximately 4-5\% of the $\mathrm{DEE}_{o}$ at each site. At Roberts Bank all the human disturbances were observed in one day. Excluding these, the energy expended at site would likely have been below the $\mathrm{DEE}_{t}$ and demonstrates the importance of managing human disturbance. Levels of disturbance to wintering Brant in Boundary Bay and Roberts Bank are comparable to other studies, but this study showed that disturbance may be above tolerable levels for Brant in Boundary Bay. Because the causes of disturbance at both sites were mainly natural, managing disturbance will be challenging.
Brant have shown the capacity to make large distributional shifts in their wintering areas (Sedinger et al. 1994). The most dramatic example of this was the shift in wintering Brant from California to Mexico in the 1950's, which has been attributed to disturbance (Sedinger et al. 1994). Brant use of Roberts Bank is a recent phenomenon and may be a consequence of relatively high levels of disturbance and energy expenditure at Boundary Bay. In order to maintain a wintering population in Boundary Bay and Roberts Bank, disturbance must be managed. The wintering Bald Eagle population in BC is increasing at an annual rate of 8\% (CWS unpubl. data). It is possible that disturbance from the growing Bald Eagle population will force the Brant population to abandon Boundary Bay and Roberts Bank within the next 10 years (see Appendix B). Recreational use of Boundary Bay and Roberts Bank will also increase
and controlling the current level of human disturbance may be an essential part of the overall management of disturbance.

For management purposes, more information is required on nocturnal feeding and methods of controlling natural disturbance. In the interim, managing human disturbance may reduce the impacts of disturbance enough to ensure Boundary Bay and Roberts Bank are energetically viable for wintering Brant. However, if the Bald Eagle population continues to increase, no amount of control of human disturbance will compensate for the increase in natural disturbance.

## 4. SPRING STAGING PATTERNS OF BRANT IN THE STRAIT OF GEORGIA, BRITISH COLUMBIA

### 4.1 Introduction

Arctic nesting geese are characterized by their sudden arrival to the Arctic to breed and their equally sudden departure in late summer to return to wintering areas (Johnson and Herter 1990). Arctic goose reproductive success is linked with migratory strategies, and spring fattening and nutrient accumulation are essential to their fitness (Ankney and MacInnes 1978, Raveling 1979, Ebbinge 1988, Black et al. 1991). During nesting at northern latitudes there are few opportunities for feeding (Ankney and MacInnes 1978). Therefore, geese must reach their Arctic nesting grounds when their body conditions are optimal to meet the demands of egglaying, incubating and rearing broods (Ankney and MacInnes 1978, Raveling 1979, Ebbinge 1988, Black et al. 1991).
Geese build up their nutrient reserves on staging areas during migration. The importance of a particular winter or spring staging area is often assessed by the volume of birds using that area. Volume can be defined as the total number of birds using a stopover site (Thompson 1993, Frederiksen et al. 2001). For static populations (e.g. wintering birds), volume estimates are readily obtained through counts, but for transient populations (e.g. migrating birds) estimating volume is more complex (Frederiksen et al. 2001) and requires incorporation of estimates of arrival and residence times.

Recently, volume estimates of Arctic geese have been produced through statistical (e.g. Routledge et al. 1999) and modeling approaches (e.g. Frederiksen et al. 2000). The statistical approach is computationally complex and uses the mean length of stay to estimate total volume. Since length of stay varies among individuals (van Eerden et al. 1991), a mean value is not a good estimator and may bias volume estimates and underestimate uncertainty in the calculation of volume. The modeling approach uses capture-mark-recapture (CMR) information that
provides probability estimates of immigration and emigration among multiple staging sites for specific time periods (i.e., weekly intervals), which can be used to calculate lengths of stay per site and per time period. Program MARK (White and Burnham 1999) provides a methodology to calculate such probabilities. I used Program MARK's multi-strata modeling option (described in detail below) that synthesises the dynamics of survival probabilities, resighting probabilities and movement among site probabilities for weekly time intervals to analyze my mark-resighting data. The results, together with counts for time periods, yield robust estimates of length of stay, and volume, and the associated uncertainty of these estimates (Frederiksen et al. 2001). CMR models require repeated observations of marked birds in a population. Black Brant have been marked with individually coded, coloured leg bands on their breeding and moulting grounds since 1985 (Bollinger unpubl. data). Consequently, 8-10\% of the Pacific Flyway Brant Population has been marked with leg bands, which can, and have been, identified by observers at spring staging locations (Reed 1997).
Arctic-nesting geese exhibit high philopatry (i.e. the tendency of individuals to exhibit long term fidelity to a specific area) (Greenwood 1980). One consequence of philopatry is local adaptation to specific sites. Hence, changes to, or loss of, staging areas can be detrimental to goose populations. In the Strait of Georgia, numbers of spring-staging Black Brant (Branta bernicla nigricans) begin to increase in late February and reach a peak by late March or early April. The two main staging areas in the Strait of Georgia are the Fraser River delta on the British Columbia Lower Mainland and the Parksville-Qualicum area on the east coast of Vancouver Island. High philopatry ( $81 \%$ and $87 \%$ respectively) has been observed across years in Brant on these staging grounds (Reed 1997).
A contemporary challenge for conservationists is the management of the Strait of Georgia for the sustainability of Brant as the Strait is undergoing increasing development and urbanization. Consequently, one objective of this study was to estimate the total number of Brant using the Fraser River delta and Parksville-Qualicum as staging areas, and estimate what proportion of the Pacific Flyway Brant Population uses these areas. In doing so, I examined the population composition at each site and the differences in migration and staging patterns between each site.

### 4.2 Methods

### 4.2.1 Study Area

The study was conducted at two sites, the Fraser River delta and the east coast of Vancouver Island. Specific areas in the Fraser River delta included Boundary Bay and Roberts Bank in British Columbia and Point Roberts in Washington State. On Vancouver Island, the study area was concentrated between Parksville and Qualicum (Figure 2).

To facilitate the multi-strata model structure I needed to define two other study sites. Although no data were collected at these sites, they were required in order to model the northern migration through the Strait of Georgia. The first site was defined as all areas south of the Strait of Georgia, where Brant migration begins. The final site was defined as all areas north of the Strait of Georgia, where Brant migration ends. Hence, four sites, or strata (A to D) were defined for the study:
$A=$ South - where Brant begin their migration,
$B=$ Fraser River delta - staging area during migration,
C = Parksville-Qualicum - staging area during migration, and
$D=$ North - where Brant end their migration.

### 4.2.2 Data Collection

Daily surveys were conducted from mid-February until mid-May in 1999 and 2000. Each survey consisted of driving a specific route to check locations where Brant could be observed. A 20$60 x$ spotting scope was used to read leg-bands and conduct counts. The total number of Brant was estimated for each survey day. On Vancouver Island, additional surveys were conducted on a weekly basis between Parksville and Campbell River. These surveys encompassed the Parksville-Qualicum study area and all areas north to Campbell River. The Parksville-Campbell River surveys were only done to determine if Brant staged outside of the Parksville-Qualicum area, and were not included in the MARK analysis. Population composition data were collected by scanning groups of Brant (usually $\sim 100$ birds) and counting the number of immature Brant, coloured leg-bands and Grey-bellied Brant (see Chapter 2 for detailed methodology).

### 4.2.3 Counts

Using the daily abundance estimates, Brant goose-days were calculated for the Fraser River delta, Parksville-Qualicum and Parksville to Campbell River using the formula:

$$
\begin{equation*}
\text { Goose-days }=\sum_{i} \frac{\mathrm{~N}_{\mathrm{i}}\left(\mathrm{D}_{\mathrm{i}+1}-\mathrm{D}_{\mathrm{i}-1}\right)}{2} \tag{Equation1}
\end{equation*}
$$

where $N_{i}=$ population at $i^{\text {th }}$ survey, $D_{i}=$ date of $i^{\text {th }}$ survey, $D_{i+1}=$ next survey date, and $D_{i-1}=$ previous survey date (Boyd 1995). By definition, $D_{0}=D_{1}$ and $D_{n+1}=D_{n}$ (Boyd 1995).

To estimate population composition on the Fraser River delta and Parksville-Qualicum, the samples of immature, Grey-bellied, and leg-banded Brant were converted to ratios, and the overall estimate was calculated for each group at each site using the Jackknife technique (Cochran 1977).

### 4.2.4 Model Data Sets

The CMR data from the leg-band observations on the Fraser River delta and ParksvilleQualicum were analyzed using the multi-strata model option provided by Program MARK (White and Burnham 1999). This model infrastructure facilitated the construction of a model of spring migration as Brant moved from the south, onto the staging areas of the Fraser River delta and Parksville-Qualicum, then to the north. The model provided an opportunity to challenge the CMR data with competing hypotheses concerning migration dynamics. For example, I tested if migration rates change weekly, and if birds moved between the Fraser River delta and Parksville-Qualicum.

Given that Program MARK has a limit to the amount of data it can accommodate, some consolidation of the data was necessary so that Program MARK could test the competing migration hypotheses. My choice was to analyze the data for each year in separate analyses. For both 1999 and 2000, leg-band observations were divided into 7-day periods, a compromise between a loss of resolution in the raw data, and a manageable amount of data and associated parameters for analysis. A total of 15 periods covered the weekly periods from mid-February until the end of May. Weekly periods corresponded across years and strata.
Based on initial data exploration (Appendices $C$ and $D$ ), within a year, leg-band data were divided into two groups and modeled separately. Group one contained band data for Brant observed only once within a year (hereafter referred to as single-sight data), and group two contained bands for birds observed multiple times (hereafter referred to as multiple-sight data). These groups represented two distinct staging patterns that were consistent with previous observations of Brant staging in the Strait of Georgia (e.g. Routledge et al. 1999).

### 4.2.5 Multi-strata Model Notation

Model notation followed Cooch and White (1998) where the strata (s) are defined by letters A to D as previously defined. Sighting probabilities were defined as:
$p_{i}^{s}=\quad$ the probability that a marked bird present in stratum $s$ at time $i$ is sighted at time $i$. Transition probabilities between strata were defined as:
$\Psi_{i}^{r s}=$ the probability that a bird present in stratum $r$ at time $i$ is in stratum $s$ at time $i+1$. Survival between strata was defined as:

$$
\boldsymbol{\Phi}_{i}^{r s}=S_{i}^{r} \boldsymbol{\Psi}_{i}^{r s}
$$

where the probability of individual survival $S_{i}^{r}$ was always assumed to be 1 as mortality during the low-risk staging tenures is considered insignificant (Hestbeck et al. 1991, Reed 1997), therefore $\Phi_{i}^{r s}=\Psi_{i}^{r s}$.

### 4.2.6 Assumptions

Several standard and specific assumptions were made in order to model the data. The standard assumptions for CMR analysis included (Seber 1982, Pollock et al. 1990, Hestbeck et al. 1991):
a) every marked animal present in the population at time $i$ had the same probability of being observed, $p_{i}$,
b) all samples were instantaneous relative to the interval between $i$ and $i+1$,
c) individuals and individual fates were independent,
d) marking of individuals was random,
e) there were no behavioural effects from marking,
and further assumptions specific to these analyses included:
f) leg-band data sets contained no observation errors,
g) mortality between weekly time periods was insignificant (i.e. survival $=1$ ),
h) the probability of sighting on Stratum $A=1$, because an individual was considered in Stratum A until it was observed in either B or C
i) the probability of sighting on Stratum $D=1$, because an individual was considered in Stratum D when it was no longer observed in either B or C,
j) once in Strata B or $C$ individuals could not return to $A$ (i.e. spring migration was consistently north), and
k) once in Stratum $D$, individuals did not return to $B$ or $C$ (i.e. spring migration was consistently north).

When modeling the migration of Arctic-nesting geese, standard CMR assumptions are often violated (Hestbeck et al. 1991, Anderson et al. 1994). Geese exhibit strong bonds between mates and families migrate as cohesive units (Prevett and MacInnes 1980, Reed 1993). Fates of pairs and families are closely linked and violate the assumption of independence. Marking of individuals is not random as families are captured and marked together. Therefore, the true
effective sample size is smaller than the actual number of marked birds and results in an underestimation of variance (Schmutz et al. 1995).

These violations are reflected in my results by high values of $\hat{c}$ (variance inflation factor). A $\hat{c}$ value $=1$ indicates independence and non-violation of standard assumptions as tested by GOF tests, i.e., random binomial error. Values larger than 1 suggest overdispersion of the data, i.e. the clumping of individuals such that a co value of 2 implies that birds are acting and making decisions as a pair. This is a reasonable expectation for Arctic-nesting geese and is observed in many species, which migrate as schools or flocks (Anderson et al. 1994).

Assumptions $h$ and $i$ were necessary in order to create the appropriate encounter histories to be read by Program MARK. However, this meant it was possible for a bird to either arrive on Strata B or C from A before it was actually observed, or to still be present on Strata B or C although it was assumed to have departed for D . To adjust for these possibilities, an a posteriori correction was applied to the results for the multiple-sight data models so that I could more accurately estimate arrival and departure dates to and from Strata B and C, and also length-of-stay.

### 4.2.7 Migration Model Construction and Selection

The multi-strata models are an extension of the generalized CMR model (the Cormack-JollySeber model), and allow for the estimation of transition probabilities between strata (Brownie et al. 1993, Hestbeck et al. 1991). In this study, strata refer to geographic locations along a migration route. A data record for a single bird as it is sighted at specific intervals (e.g. weekly) along its migration route is called an encounter history. A complete encounter history records the stratum where the bird is seen in a particular week, e.g. AAAB0BB0CCCDDDDD, where 0 represents no sighting. Using the Multi-strata Recaptures Only option in Program MARK, I used separate models for each year to estimate transition probabilities for the following movements: Single-sight (1999, 2000):

- $A$ to $B$
- A to C

Multiple-sight (1999, 2000):

- A to B
- A to C
- $(B, C)$ to $(B, C, D)$

I was fortunate that only on four occasions was a bird seen at more than one stratum within a single week. I dealt with that data issue by randomly choosing (coin toss) one of the strata for that period. I challenged the data by statistically fitting hypothesized competing models that described the pattern of the transition probabilities among strata over time. I then chose the best models based on the Quasi Akaike Information Criterion (QAIC $C_{c}$, which takes into account the extra-binomial dispersion of error variance resulting from the non-independent behaviour of individual birds (Anderson et al. 1994, Burnham and Anderson 1998)
Program MARK is not capable at this time of performing a parametric bootstrap goodness-of-fit (GOF) test for multi-strata models. Therefore I conducted non-parametric bootstraps of the observed data to simulate deviances (Schwarz et al. 1993). I compared the $\hat{c}$ for my data to the distribution of $\hat{c}$ values generated by simulations to judge the probability of my data. That is, I was asking the question "could my data have reasonably been generated by the model I selected based on QAIC?" To compute $\hat{c} \mid$ took the observed and expected deviance values for the encounter histories in the data and used the equation (White 2000):

$$
\begin{equation*}
\widehat{c}=2 \sum_{i=1}^{n} O_{i} \ln \left(O_{i} / E_{i}\right) \tag{Equation8}
\end{equation*}
$$

where $O_{i}=$ observed and $E_{i}=$ expected frequencies of the observed encounter histories, subject to the constraint that and $\Sigma O_{i}=\Sigma E_{i}$ over all possible encounter histories.

### 4.2.8 Adjusted Arrival and Departure Weeks

Since birds were never observed on strata A or D, I had to assume that birds arrived on strata B or $C$ the first week they were observed, and arrived on stratum $D$ the week after they were last seen on B or C . Therefore $p_{i}^{\mathrm{A}}$ and $p_{i}^{\mathrm{D}}$ (i.e., the probability of being on strata A or D , respectively in weekly period 1) were fixed at 1 in all competing models. However, it was possible, and even somewhat likely, that a bird arrived on strata B or C from A before it was observed in B or C. Conversely a bird not seen on B or C and necessarily assumed to have left for D, may just not have been observed on B or C . To correct for these a priori assumptions I made a posteriori adjustments of the time period of arrival for the multi-sight data using the sighting probabilities $\left(p_{i}^{r}\right)$ generated from the models. These adjustments were not applied to the single-sight data, since I chose to define birds in the single-sight category as those that were observed to arrive and depart within the one-week time period in which they were observed.

The revised estimate of the week of arrival was estimated using the following equations:

$$
\begin{equation*}
\hat{A}=\sum_{w=1}^{F} w P_{A, w}^{\prime} \tag{Equation9}
\end{equation*}
$$

where $\hat{A}=$ the estimated period (week) of arrival, $F=$ the first period a marked bird was seen, $w$ $=$ period and $P_{A, w}^{\prime}=$ the probability that the marked bird arrived in period $w$ or later if it had not been seen prior to $F$. The term $P_{A, w}^{\prime}$ was calculated using:

$$
P_{A, w}^{\prime}=P_{A, w} \prod_{i=w}^{F}\left(1-p_{i}^{r}\right)
$$

(Equation 10)
where $\mathrm{P}_{\mathrm{A}, \mathrm{w}}=\left(\frac{1}{N-\sum_{i=1}^{w-1} 1}\right)$ i.e. the probability that the marked bird arrived in period $w$, given it had not arrived prior to w , and $p_{i}^{r}=$ the probability of a marked bird being seen in period $\mathrm{w}=$ $2, \ldots . N$, given that it was present. The standard error of Â was estimated as:

$$
\begin{equation*}
\hat{s}_{A}=\sqrt{\sum_{w=1}^{F} w^{2} P_{A, w}^{\prime}-\hat{A}^{2}} \tag{Equation11}
\end{equation*}
$$

The same approach was used to estimate modified departure times:

$$
\begin{equation*}
\hat{D}=\sum_{w=L}^{N} w P_{D, w}^{\prime} \tag{Equation12}
\end{equation*}
$$

where $\hat{D}$ = the estimated period (week) of departure, $\mathrm{L}=$ the last period that the marked bird was observed, and $P_{D, w}^{\prime}=$ the probability that the marked bird departed in period $w$ or earlier if it had not been seen after period $L$ :

$$
\begin{equation*}
P_{D, w}^{\prime}=P_{D, w} \prod_{i=L}^{w}\left(1-p_{i}^{r}\right) \tag{Equation13}
\end{equation*}
$$

where $P_{D, w}=\left(\frac{1}{N-\sum_{i=w+1}^{L} 1}\right)$ i.e., the probability that the marked bird departed in period $w$,
given it had not departed later than period w. The standard error of $\hat{D}$ was estimated as:

$$
\begin{equation*}
\hat{s}_{D}=\sqrt{\sum_{w=L}^{N} w^{2} P_{D, w}^{\prime}-\hat{D}^{2}} \tag{Equation14}
\end{equation*}
$$

### 4.2.9 Length of Stay

The mean length of stay for an individual Brant goose in the multiple-sight category was estimated for each cohort of birds (where a cohort was defined as the group of birds that arrived in a week period) on each stratum using the probabilities of remaining (R) and leaving (L) a stratum where:

$$
\begin{array}{ll}
p\left(R_{i}\right)=1-\left(\Phi_{i}^{r s}+\Phi_{i}^{r t}\right) & \\
p\left(L_{i}\right)=1-p\left(R_{i}\right)\left(1-\sum_{j=1}^{i-1} P\left(L_{j}\right)\right) & \text { for } i<\text { the last week } \\
\left.p\left(L_{i}\right)=\left(1-\sum_{j=1}^{i-1} P\left(L_{j}\right)\right)\right) & \text { for } i=\text { the last week } \tag{Equation17}
\end{array}
$$

and length of stay $(T)$ was calculated as:

$$
T=\sum_{i=1}^{k} p\left(L_{i}\right)
$$

(Equation 18)
where $i=$ period, $j=$ period and $k=$ number of periods.

### 4.2.10 Volume Estimates

Volume estimates were calculated for each stratum by independently estimating, then summing, the number of single-sight and multiple-sight Brant that arrived each period. First, for each period I estimated the abundance from the count data and the proportion of birds that were classified as single-sight or multiple-sight from the band observation data. This led directly to an estimate of the volume of single-sight Brant:

$$
\begin{equation*}
\hat{V}_{S S}=\sum_{i=1}^{k}\left(n_{i}\right)\left(p S S_{i}\right) \tag{Equation19}
\end{equation*}
$$

where $\hat{V}_{S S}=$ volume of single-sight Brant estimate, $\mathrm{n}_{i}=$ estimate of period abundance, and $\mathrm{pSS}_{i}$ $=$ the proportion of single-sight Brant in the period.

Second, for each period an estimate of the number of multiple-sight Brant was derived from:

$$
\begin{equation*}
M_{i}=\left(n_{i}\right)\left(p M_{i}\right) \tag{Equation20}
\end{equation*}
$$

where $\mathrm{M}_{i}=$ the number of multiple-sight Brant, $\mathrm{n}_{i}=$ estimate of period abundance, and $\mathrm{pM}_{i}=$ the proportion of multiple-sight Brant in the period.
The total volume of multiple-sight Brant was estimated using $M_{i}$ and $p\left(R_{i}\right)$ :

$$
\begin{equation*}
\hat{V}_{M}=M_{1,1}+\sum_{i=2}^{k}\left\{M_{i, i}-\sum_{j=1}^{i=1}\left(M_{i-1, j} p\left(R_{i-1, j}\right)\right\}\right. \tag{Equation21}
\end{equation*}
$$

where $i=$ index of periods, $j=$ index of cohorts and $k=$ number of periods. The estimate of volume of Brant arriving over all periods is:

$$
\begin{equation*}
\hat{V}_{T}=\hat{V}_{S S}+\hat{V}_{M S} \tag{Equation22}
\end{equation*}
$$

Error in the volume estimates arises from two sources, 1) counting error in abundance estimates, and 2) statistical error in Program MARK's estimates of the transition probabilities. There was no explicit measure of error for abundance estimates as they were from single counts; therefore, I simulated different degrees of error in the volume estimates by imposing three different levels of counting error (5\%, 10\% and 20\%). Due to model complexity I was forced to avoid the computational complexity of trying to precisely estimate the total variance of the volume estimates due to the individual transition probability variances and covariances. Therefore, I present three extreme scenarios to characterize the uncertainty of my estimates arising from uncertainty in the transition probability estimates: 1) no uncertainty in the parameter estimates (extreme underestimation, only uncertainty is due to counting errors), 2) parameter estimates are uncorrelated (overestimate of uncertainty), and 3) parameter estimates are $100 \%$ positively correlated (intermediate uncertainty). Simulations were computed 1000 times to obtain randomized estimates. In doing so, the range from no uncertainty to extreme uncertainty in volume estimates was estimated.

### 4.3 Results

### 4.3.1 Counts

The peak of migration occurred at similar times each year, but in both years the peak occurred about two weeks later on the Fraser River delta than on Vancouver Island (Figures 7 and 8). The peak abundance and the number of Brant goose-days were lower on the Fraser River delta in 2000 than in 1999 (Table 11). Approximately 5-6\% of the Brant observed on the Fraser River delta and about 8-9\% of the Brant on Vancouver Island wore leg-bands (Table 12). No Greybellied Brant were observed on Vancouver Island.

The Parksville-Campbell River surveys encompassed the Parksville-Qualicum area; however, the surveys were only conducted $1 /$ week. In both years the maximum daily number of Brant observed at Parksville-Qualicum was not during a Parksville-Campbell River survey.
Consequently, the Parksville-Campbell River peak count was less than the one for ParksvilleQualicum, although the survey covered a larger geographical area. Overall, the surveys did show that the larger survey area supports a greater number of goose-days than just ParksvilleQualicum (Table 11).

Table 11. Brant Goose-days and peak abundance during spring migration on the Fraser River delta and the east coast of Vancouver Island.

| Year | Site | Goose-days | Peak <br> Abundance | Date of Peak Abundance |
| :---: | :---: | :---: | :---: | :---: |
| $\mathbf{1 9 9 9}$ | FRD | 154519 | 5347 | April 16 |
|  | PQ | 128949 | 6900 | April 1 |
|  | PQ-CR | 141452 | 5226 | April 8 |
| $\mathbf{2 0 0 0}$ | FRD | 67532 | 2428 | April 18 |
|  | PQ | 150145 | 5415 | March 29 |
|  | PQ-CR | 156243 | 4381 | April 2 |
| FRD = Fraser River delta, PQ = Parksville-Qualicum and PQ-CR = surveys between Parksville |  |  |  |  |
|  |  |  |  |  |
| and Campbell River. |  |  |  |  |

Table 12. Proportions (\%) of leg-banded, immature and Grey-bellied Brant observed on the Fraser River delta (FRD) and Vancouver Island (VI).

|  |  | Leg-banded |  |  |  | $\begin{array}{c}\text { Immature } \\ \text { Black Brant }\end{array}$ |  |  | Grey-bellied |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | \(\left.\begin{array}{c}Immature <br>

Grey-bellied\end{array}\right]\)

* no data


### 4.3.2 Band Observations

In 1999, 1094 discrete bands were observed in Parksville-Qualicum and 316 on the Fraser River delta. In 2000, 900 discrete bands were observed in Parksville-Qualicum and 382 on the Fraser River delta. The most common observations were of single-sight bands, which comprised approximately 40\% of Parksville-Qualicum and 70\% of Fraser River delta observations (Appendix C).

### 4.3.3 Program MARK Multi-strata Model Selection

A model that constrained the logit of transition probabilities to conform to a quadratic function of time period was the most parsimonious model for each year of the single-sight data. The two years of single-sight data were treated as different groups within the same model; therefore, the model characteristics generated by Program MARK are identical. A model that constrained the logit of transition probabilities to conform to quadratic functions of both time period and cohort provided the most parsimonious models for the multiple-sight data for each year (Table 13). The high ĉ values (~3-6) indicate the data are overdispersed and, as anticipated for reasons explained earlier, clearly violate the assumption of independence in the data.


Figure 7. Brant abundance during spring migration on the Fraser River delta. A. 1999. B. 2000.


Figure 8. Brant abundance during spring migration in Parksville-Qualicum. A. 1999. B. 2000.

Table 13. Best models for single-sight and multiple-sight data.

| Model | Model Characteristics |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  | \# Parameters | Deviance | QAICc | $\hat{\text { c }}$ | c SE |
| Single-sight |  |  |  |  |  |
| 1999 from A to B or C | 12 | 221.3 | 6743.6 | 6.33 | 1.97 |
| 2000 from A to B or C | 12 | 221.3 | 6743.6 | 3.10 | 0.96 |
| Multiple-sight |  |  |  |  |  |
| 1999 from A to B or C | 6 | 212.3 | 3421.2 | 2.67 | 1.17 |
| $\begin{aligned} & 2000 \text { from } A \text { to } B \\ & \text { or } C \end{aligned}$ | 6 | 992.3 | 4556.1 | 4.62 | 1.61 |
| $\begin{aligned} & 1999 \text { from } B \text { or } C \\ & \text { to } D \end{aligned}$ | 27 | 556.0 | 2967.1 | 6.48 | 0.92 |
| $\begin{aligned} & 2000 \text { from } B \text { or } C \\ & \text { to } D \end{aligned}$ | 35 | 829.8 | 3580.9 | 6.20 | 0.68 |

## Single-Sight Models

Single-sight models were used to estimate transition probabilities from Stratum A to either Stratum B or C for individuals only observed once. These models predicted that early in migration, the probability of moving from Stratum A to Stratum B was low, but as time progressed the transition probability increased (Figure 9). The probability of moving from Stratum A to $C$ was initially low but as time increased, the probability increased until approximately week 8 (Figure 9). These models showed that Brant migrating from Stratum A after week 9 had a higher probability of moving to Stratum B than to Stratum C (see Appendix E for transition probabilities). The second part of the migration movement is from Strata B and C to Stratum D, but by definition, single-sight birds arrive and leave within the same weekly period.

## Multiple-Sight Models

Separate multiple-sight models were used to estimate the weekly transition probabilities for each cohort of birds in 1999 and 2000. The first part of the migration was the transition from Stratum A to B or C. In 1999 there was a higher probability of migrating from Stratum A to Stratum C than to Stratum B early in migration. Late in migration, all Brant were expected to
migrate to Stratum B. In 2000, the probability of migrating from Stratum A to Stratum C was higher than to Stratum B with the exception of the last period Brant were observed (week 13) (Figure 10). The highest probability of migrating to Stratum C was similar in both years (approximately week 8). On Stratum B, although the pattern was similar across years, the transition probabilities were much lower in 2000 than in 1999 (Figure 10).
The second part of the migration was the transition from Stratum B or C to D after a length of stay of a week or more at Stratum B or C. In general, the probability of a bird migrating to Stratum D increased as period (time) increased. Hence, a bird that arrived to either Stratum B or C in the first week of migration tended have a longer residence time there than one that arrived later (Figures 11 and 12). Differences existed in the movement patterns through Strata B and C. The models in Figure 11 and 12 show steeper curves with flatter tails for Stratum C compared to B. This suggests that early migrating Brant that arrived in Stratum C, staged longer than those on Stratum B. Limited interchange occurred between Strata B and C. In 1999, $0.5 \%$ and in 2000, 3\% of the bands were observed at both sites (see Appendix E for transition probabilities).
The probably of remaining on Stratum B (or C) in a given week $i$, is the complement to this probability 1- $\Phi_{i}^{B D}$ (or 1- $\Phi_{i}^{C D}$ ). The arrival and departure weeks are listed in Appendix F.

### 4.3.4 Length of Stay

Length of stay for multiple-sight birds generally decreased as the week that a cohort arrived increased (Figure 13) (See Appendix F for associated standard errors).

### 4.3.5 Volume Estimates

The estimated volumes were similar between years on Stratum C, but decreased between years on Stratum B (Table 14). In 1999 and 2000, the midwinter indices for the Pacific Flyway Brant Population were 119,993 and 127,129 respectively (Pacific Flyway Committee 2001). Table 15 contains volume estimates and their uncertainty from this study expressed as proportions of the Pacific Flyway Brant Population.



Figure 9. Each curve models the probability of transition from Stratum A to Stratum B or C during each period for single-sight data. The probably of remaining on Stratum $A$ in a given week $i$, is the complement to these probabilities 1- $\Phi_{i}^{A B}+\Phi_{i}^{A C}$.



Figure 10. Each curve models the transition probability from Stratum $A$ to either $B$ or $C$ for the multiple-sight data. The probably of remaining on Stratum A in a given week $i$, is the complement to these probabilities $1-\Phi_{i}^{A B}+\Phi_{i}^{A C}$.


Figure 11. Each graph depicts the transition probability for each cohort of birds, each period, in 1999. A) Transition from Stratum B to D. B) Transition from Stratum C to D. The probably of remaining on Stratum $B$ (or $C$ ) in a given week $i$ is the complement to this probability 1- $\Phi_{i}^{B D}$ (or 1- $\Phi_{i}^{C D}$ ). The arrival and departure weeks are listed in Appendix F.


Figure 12. Each graph depicts the transition probability for each cohort of birds, each period, in 2000. A) Transition from Stratum B to D. B) Transition from Stratum C to D. B) Transition from Stratum C to D. The probably of remaining on Stratum B (or C) in a given week $i$, is the complement to this probability $1-\boldsymbol{\Phi}_{i}^{B D}$ (or $1-\boldsymbol{\Phi}_{i}^{C D}$ ). The arrival and departure weeks are listed in Appendix F.


Figure 13. The mean length of stay, estimated in weeks for each cohort, year and stratum.

Table 14. Volume estimates adjusted for three proposed levels of error in abundance counts (5\%, 10\% and 20\%) and three levels of uncertainty among the standard error estimates for transition parameters.

| Uncertainty in Transition Parameters | Volume Estimate | Stratum B 1999 |  |  | Stratum B 2000 |  |  | Stratum C 1999 |  |  | Stratum C 2000 |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 5\% | 10\% | 20\% | 5\% | 10\% | 20\% | 5\% | 10\% | 20\% | 5\% | 10\% | 20\% |
| Underestimate | Total | 17562 | 17641 | 18040 | 10421 | 10858 | 11119 | 14025 | 14111 | 14356 | 14727 | 14828 | 15072 |
|  | S.D. | 1437 | 1587 | 2069 | 456 | 603 | 982 | 494 | 734 | 1449 | 980 | 1170 | 1731 |
| Intermediate | Total | 16783 | 16876 | 17242 | 10645 | 10685 | 10898 | 13838 | 13921 | 14210 | 14242 | 14304 | 14603 |
|  | S.D. | 315 | 628 | 1416 | 195 | 356 | 832 | 299 | 617 | 1383 | 289 | 619 | 1309 |
| Overestimate | Total | 10987 | 11037 | 11206 | 7460 | 7492 | 7710 | 5388 | 5420 | 10841 | 8135 | 8150 | 7940 |
|  | S.D. | 215 | 429 | 912 | 138 | 261 | 594 | 132 | 231 | 1109 | 167 | 343 | 699 |

Table 15. Volumes estimates as proportions (\%) of the Pacific Flyway Brant Population for each abundance error-transition parameter uncertainty combination.

| Uncertainty in <br> Iransition <br> Parameters | Volume <br> Estimate | Stratum B 1999 | Stratum B 2000 | Stratum C 1999 | Stratum C 2000 |  |  |  |  |  |  |  |  |
| :--- | :---: | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- | :--- |
|  |  | $\mathbf{5 \%}$ | $\mathbf{1 0 \%}$ | $\mathbf{2 0 \%}$ | $\mathbf{5 \%}$ | $\mathbf{1 0 \%}$ | $\mathbf{2 0 \%}$ | $\mathbf{5 \%}$ | $\mathbf{1 0 \%}$ | $\mathbf{2 0 \%}$ | $\mathbf{5 \%}$ | $\mathbf{1 0 \%}$ | $\mathbf{2 0 \%}$ |
| Underestimate | Total | $\mathbf{1 4 . 6}$ | 14.7 | 15.0 | 8.7 | 9.0 | 9.3 | 11.7 | 11.8 | 12.0 | 12.3 | 12.4 | 12.6 |
|  | S.D. | 1.2 | 1.3 | 1.7 | 0.4 | 0.5 | 0.8 | 0.4 | 0.6 | 1.2 | 0.8 | 1.0 | 1.4 |
| Intermediate | Total | 14.0 | 14.1 | 14.4 | 8.9 | 8.9 | 9.1 | 11.5 | 11.6 | 11.8 | 11.9 | 11.9 | 12.2 |
|  | S.D. | 0.3 | 0.5 | 1.2 | 0.2 | 0.3 | 0.7 | 0.2 | 0.5 | 1.2 | 0.2 | 0.5 | 1.1 |
| Overestimate | Total | 9.2 | 9.2 | 9.3 | 6.2 | 6.2 | 6.4 | 4.5 | 4.5 | 9.0 | 6.8 | 6.8 | 6.6 |
|  | S.D. | 0.2 | 0.4 | 0.8 | 0.1 | 0.2 | 0.5 | 0.1 | 0.2 | 0.9 | 0.1 | 0.3 | 0.6 |

### 4.4 Discussion

### 4.4.1 Band Observations, Counts, and Population Composition

During this study, the timing of migration was similar in both years; however, there was a notable difference between years in the number of Brant staging on the Fraser River delta. In 2000, the counts were much lower on the delta. This was not likely a result of observer bias as counts were consistently lower for all researchers and volunteers, all of whom were experienced at surveying Brant.

During spring migration, the proportion of Brant wearing leg-bands was lower on the Fraser River delta than at Parksville-Qualicum. This was attributed the presence of Grey-bellied Brant on the Fraser River delta. Grey-bellied Brant, which are genotypically and phenotypically distinct from Black Brant (Shields 1990), primarily winter in northern Washington (Reed et al. 1989). This group of Brant was not leg-banded at the time of the study and their presence on the Fraser River delta during migration likely caused the band ratio to be low at this site. Greybellied Brant were not observed on Vancouver Island.

The acquisition of leg-band observation data depended on accessibility to sites where Brant could be observed out of the water. The Fraser River delta had fewer sites than ParksvilleQualicum for observing Brant. This may have contributed to the relatively low number of legbands observed on the Fraser River delta and may also have biased the number of observations classified as single-sight. To what extent this may have occurred is unknown.

### 4.4.2 Migration and Staging Patterns

Differences were observed in the migration and staging patterns on the Fraser River delta compared to Parksville-Qualicum. Early in migration, Brant from the south were more likely to migrate to Parksville-Qualicum, but as weekly periods increased, Brant were more likely to migrate to the Fraser River delta. This was consistent with the count data that suggested the peak of migration occurred approximately two weeks later on the Fraser River delta than in Parksville-Qualicum each year.

There was little evidence of interchange between the two sites with only 0.5\% (82-162 Brant) (1999) and 3\% (468-786 Brant) (2000) of the bands observed at both sites. Brant exhibit high philopatry to their spring staging areas (e.g. Raveling 1979, Reed 1997) and the lack of interchange suggests different segments of the Pacific Flyway Brant Population were using specific sites.

### 4.4.3 Length of Stay

Length of stay for multiple-sight Brant, as estimated by the models, ranged from approximately 1 to 9 weeks, but decreased as the time a cohort first arrived on Stratum B or C progressed. Early in migration, Brant may have experienced less pressure to reach the breeding areas. Although the 2000 model estimated lengths of stay of up to 9 weeks, very few birds likely staged for this long since few birds arrived on these strata this early.

### 4.4.4 Volume Estimates

The volume estimates provided in this study are minimum estimates. This is largely due to the confidence in the abundance counts, which likely underestimates the total number of Brant present during a period. The two main reasons are that daily counts are minimum estimates of the number of Brant in the survey area and the lack of accuracy in the single-sight turnover rates. Turnover time for some single-sight Brant is likely greater than 1 day. Also, some unknown proportion of single-sight Brant is never seen. These two factors could affect the volume estimates for each weekly period. This uncertainty likely leads to a significant underestimation of the true number of birds staging in the study areas.
This study estimated that between $13-27 \%$ of the Pacific Flyway Population staged on the Fraser River delta and Parksville-Qualicum in 1999 and 2000. This range represented the minimum and maximum extremes based on the transition probabilities generated in the models. These study results were based on minimum estimates of abundance and exclude other smaller staging areas on the Strait of Georgia, such as Denman Island, the Seal Islets, and the Trent River Estuary on Vancouver Island. Over one thousand Brant can be observed at each of these sites during the peak of migration (Hagmeier unpubl. data).
To further refine the volume estimates in this study, the extent of overlap between minor and main staging areas must be estimated. The counts at smaller sites cannot be added without this information. Although leg-band observation data are limited at smaller sites, preliminary data suggests Denman Island and Texada Island (Martin and Dove unpubl. data) support Brant, which do not use Parksville-Qualicum or the Fraser River delta. Therefore, the numbers observed at these sites could be added to the Parksville-Qualicum and Fraser River delta estimates to achieve a better estimate of the total number of Brant staging in the entire Strait of Georgia.

Some leg-banded Brant observed in Campbell River were also observed in Parksville-Qualicum and the Fraser River delta (Hagmeier unpubl. data). Although not conclusive, this suggests that some Brant travel only short distances between stopovers in the Strait of Georgia. Therefore,
birds observed north of Parksville-Qualicum should not necessarily be considered as different from ones observed at Parksville-Qualicum.

The volume estimates from this study are comparable to other studies. Routledge et al. (1999) estimated approximately 12,100 Brant staged in Parksville-Qualicum during the spring of 1990 and Nygren (1991) estimated that 20\% of the Pacific Flyway Brant Population staged in the entire Strait of Georgia, approximately 14,100 of which staged in Parksville-Qualicum. Both of these studies used a mean length of stay to calculate volumes.

Relatively little attention has been given to the Fraser River delta as a spring staging area for Brant, yet the volume estimates in this study suggest that it is just as important as the east coast of Vancouver Island. Moreover, separate groups of Brant use these two major sites in the Strait of Georgia as evidenced by the lack of interchange between the two sites and the absence of Grey-bellied Brant on Vancouver Island.

This method of calculating volume and identifying differences in staging patterns at the two sites has limitations; however, Program MARK and population modeling methodologies are improving. Limitations in this analysis may not be problematic in the future. Currently however, it is important for managers and researchers to recognize that even based on minimum estimates, approximately one quarter $(30,000-35,000)$ of the Pacific Flyway Brant Population may stage in the Strait of Georgia.

## 5. SUMMARY AND CONCLUSIONS

### 5.1 Introduction

The Strait of Georgia in British Columbia is an important over-wintering and spring staging area for Brant. Historically, Brant wintered throughout the strait, but they are now largely confined to the Fraser River delta, specifically Boundary Bay and Roberts Bank. The Boundary Bay-Roberts Bank population decreased dramatically over the last century, but rebounded in the last 10 years. In spring, the primary staging areas for spring migrants are the Fraser River delta and Parksville-Qualicum on the east coast of Vancouver Island.
The Strait of Georgia is experiencing tremendous growth in development and urbanization.
Because of this, the sustainability of the strait to support wintering and spring staging Brant is an important management concern. With this in mind, I examined the population demographics of the Boundary Bay-Roberts Bank wintering population to understand the factors that are important to its growth and maintenance, recent changes in the local distribution, and the impact
of the Brant hunt in British Columbia on the different wintering segments of the Pacific Flyway Brant Population (Chapter 2). I also studied two aspects of Brant ecology in the Strait of Georgia: 1) time-activity budgets and the impacts of disturbance on Brant wintering on the Fraser River delta (Chapter 3) and 2) the migration patterns and the volume of Brant using the Fraser River delta and Parksville-Qualicum as spring staging areas in the Strait of Georgia (Chapter 4).

### 5.2 Population Composition and Increase and Distribution of Brant Wintering in Boundary Bay and Roberts Bank

The Boundary Bay-Roberts Bank wintering population grew dramatically in the 1990s, after having been nearly decimated in previous decades. In recent decades, several European Arctic goose populations have experienced similar trends in response to decreased hunting pressure and increases in wintering refuges (e.g. Svalbard Brant Geese) (Madsen et al. 1999). Exponential growth models suggested that the Boundary Bay-Roberts Bank population grew at an annual rate of approximately 22-26\% between 1993-94 and 1999-2000. This is well above growth rates observed for other recovering goose populations, but mean annual productivity over the same time period was approximately $14 \%$, a rate consistent with other populations. Therefore, unlike other goose populations, productivity alone did not account for the increase. Concurrent with the population growth is the increased presence of Grey-bellied Brant in the wintering population (15-20\%). Grey-bellied Brant are geographically, morphologically and genetically distinct from Black Brant (Shields 1990). The traditional wintering area of GreyBellied Brant is in northern Washington State, 60km south of the Fraser River delta. Greybellied have expanded their northern range to include the Fraser River delta.
During this study, Brant wintering on the Fraser River delta expanded their local winter range to include Roberts Bank. Distinct segments of the population seemed to favour one site over the other. Significantly higher proportions of Grey-bellied Brant were observed at Boundary Bay compared to Roberts Bank and significantly higher proportions of juveniles were observed at Roberts Bank compared to Boundary Bay. Leg-band observation data of Black Brant showed there was similar mixing of birds from different breeding grounds at each site, consistent with previous research conducted in this area (Reed 1997).

### 5.3 Harvest

Evidence from this study suggests that during the British Columbia Brant hunt approximately $70 \%$ (105-140) of the harvested Brant are from the Boundary Bay-Roberts Bank wintering population. Less than $1 \%$ of Brant from each of the wintering areas south of the Fraser River delta are harvested in the British Columbia season. This, of course, depends on the timing of arrival of migrants from more southern wintering areas into Boundary Bay and Roberts Bank and the number of Brant harvested each year.

### 5.4 Activity Budgets and Effects of Disturbance

During the day, Brant spent up to $10 \%$ of their time responding to disturbance. In Boundary Bay (2000) this was equivalent to almost half of their daily energy expenditure. There were fewer disturbances at Roberts Bank than Boundary Bay, which may be contributing to the increasing use of Roberts Bank as documented in Chapter 2.

During this study, disturbance caused by humans resulted in only $5 \%$ of the daily energy expenditure of Brant. Most disturbance events were from natural sources, particularly Bald Eagles. The frequencies of disturbance were $2.17 / \mathrm{h}$ and $1.32 / \mathrm{h}$ at Boundary Bay and Roberts Bank respectively. High levels of disturbance ( $>2 / \mathrm{h}$ ) have been shown to lead to abandonment of wintering areas by Arctic-nesting geese even if they engaged in compensatory nocturnal feeding (Bélanger and Bédard 1990). The increase in the Bald Eagle numbers and the effect on the Brant DEE will ultimately negatively impact Brant remaining in the area.

### 5.5 Spring Migration Patterns and Volume Estimates

This study identified differences in migration patterns of Brant using the Fraser River delta and the east coast of Vancouver Island. Each site is an important staging area and is used by a separate component of the Pacific Flyway Brant Population. This is to be expected as Arcticnesting geese exhibit high philopatry to staging and nesting areas and high philopatry has been observed in spring staging Brant at Parksville-Qualicum (Reed 1997).
This study estimated that between $13-27 \%$ of the Pacific Flyway Population stages on the Fraser River delta and Parksville-Qualicum. This is a minimum estimate and suggests that these are critical staging areas for the Pacific Flyway Brant Population.

### 5.6 Research and Management Implications

The results of this research have considerable management implications, a primary one being that the Boundary Bay-Roberts Bank wintering population cannot be effectively managed in
isolation. Although the wintering population may be closed within a year (Reed 1997), across years, immigration of Brant from Washington State affects the size of the population. Furthermore, immigration likely contains both Black and Grey-bellied Brant, but it is only the morphologically distinct Grey-bellied Brant that can be identified as a specifically increasing component of the wintering population. Continued and increased effort with Washington State for management of wintering Brant as a trans-boundary unit will undoubtedly yield the most successful management results.

Because the causes of disturbance in most studies of Arctic-nesting geese are human-related, management problems in those situations can be fairly easily resolved (e.g. changing aircraft flight paths or hunting boundaries); however in this study the causes of disturbance were primarily from Bald Eagles. Therefore, management of disturbance will be challenging. I suggest that it may be critical to maintain or decrease the current levels human disturbance, should the Brant be approaching a threshold that would force them to abandon this wintering area. The information presented in this thesis is important for current management planning; however also provides a baseline for future research which may answer some critical questions about the Pacific Flyway Brant Population, for example,

- Correlating the probability of arrival to staging areas and lengths of stay with respect to resource availability (e.g. timing and abundance of herring spawn),
- Determining which components of the Pacific Flyway Brant Population are staging on the Fraser River delta and Vancouver Island. Currently, the amount of data on Brant migrating through the Strait of Georgia and their reproductive fates is limited, but as this data set increases, the importance of the Strait of Georgia in terms of nutrient accumulation of breeding Brant could be quantified and used as a tool in the conservation and management,
- Understanding the population dynamics of Grey-bellied Brant.

It seems essential that the Pacific Flyway Brant Population be researched on a flyway-wide scale and that the connectivity between different projects be increased. Therefore, decisions made by regional biologists and managers will consider the impacts all along the Flyway, and through adaptive management, our capability to manage and conserve this population will continue to improve.

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## Appendix A

## Calculations for Harvest Estimates using Leg-band Data

1. Assumed $8-10 \%$ of the Pacific Flyway Brant Population (PFBP) is banded and equally distributed along the Flyway

- Based on observations from Boundary Bay, Roberts Bank, Vancouver Island, Oregon and California.

2. Removed Grey-bellied Brant from analysis

- Grey-bellied Brant were not banded at the time of this study

3. Estimated 3-year population means for each wintering area (Subcommittee on Pacific Brant 2001).

Table 16. 3-year population means for each wintering area on the Pacific Flyway

| Year |  |  |  |  |  |
| :--- | :---: | :---: | :---: | :---: | :---: |
| Site | $\mathbf{1 9 9 7 - 9 8}$ | $\mathbf{1 9 9 8 - 9 9}$ | $\mathbf{1 9 9 9 - 0 0}$ | Mean | SD |
| BC | 619 | 985 | 1238 | 947 | 311 |
| WA | 4541 | 6037 | 5944 | 5507 | 838 |
| OR | 580 | 645 | 523 | 583 | 61 |
| CA | 6091 | 4296 | 3389 | 4592 | 1375 |
| MEX | 112105 | 100760 | 108440 | 107102 | 5790 |

$B C=$ British Columbia, WA $=$ Washington, $O R=$ Oregon, $C A=$ California, $M E X=$ Mexico.
4. Pooled the band data from 10 days before and 10 days of the hunt to establish which segments of the PFBP were vulnerable to hunting and estimated the proportion of birds from each wintering area present in Boundary Bay and Roberts Bank before and during the hunt using the leg-band data.

Table 17. The number of leg bands observed in Boundary Bay and Roberts Bank from each of the wintering areas.

| Year |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Site | 1997-98 | 1998-99 | 1999-00 | Total | Proportion of Population (\%) |
| 10 days before |  |  |  |  |  |
| BC | 15 | 7 | 31 | 53 | * |
| WA | 0 | 0 | 0 | 0 | * |
| OR | 0 | 0 | 0 | 0 | * |
| CA | 4 | 0 | 0 | 4 | * |
| MEX | 11 | 1 | 1 | 13 | * |
| 10 days of hunt |  |  |  |  |  |
| BC | 13 | 10 | 14 | 37 | * |
| WA | 0 | 0 | 3 | 3 | * |
| OR | 0 | 0 | 0 | 0 | * |
| CA | 2 | 0 | 0 | 2 | * |
| MEX | 11 | 3 | 2 | 16 | * |
| Total |  |  |  |  |  |
| BC | 28 | 17 | 45 | 90 | 70.3 |
| WA | 0 | 0 | 3 | 3 | 2.3 |
| OR | 0 | 0 | 0 | 0 | 0.0 |
| CA | 6 | 0 | 0 | 6 | 4.7 |
| MEX | 22 | 4 | 3 | 29 | 22.7 |

5. Estimated the number of birds harvested from each wintering ground based on a total harvest of 150-200 Brant.

Table 18. Estimated number of Brant shot in Boundary Bay-Roberts Bank.

| Wintering <br> area | Estimated \# <br> Brant shot |
| :--- | :---: |
| BC | $105-140$ |
| WA | $3-5$ |
| OR | 0 |
| CA | $7-9$ |
| MEX | $34-45$ |

6. Estimated the proportion of birds harvested from each wintering population by dividing the estimate of birds harvested from each wintering area (Step 5) by the mean wintering population size of each area (Step 3). The harvested proportions from each wintering population were $B C=11-15 \%$, Washington $<1 \%$, Oregon $<1 \%$, California $<1 \%$, and Mexico < 1\%.

## Appendix B

## Effects of Bald Eagle and Human Population Growth on DEE of Brant in Boundary Bay.



Figure 14. The predicted sizes of the BC Coastal Bald Eagle population and the human population of Delta. In 1994 there were 30000 Bald Eagles, increasing at an annual rate of 8\% (CWS unpubl. data). The human population of Delta is increasing by $0.4 \%$ each year (Corportation of Delta, Department of Community Planning and Development).


Figure 15. The relationship between Bald Eagles and DEE and between people and DEE using the two years of data from Boundary Bay.


Figure 16. The predicted growth of the Bald Eagle population in BC and the associated increase in Brant DEE due to Bald Eagle disturbance (DEEe).


Figure 17. The predicted growth of the human population in Delta, BC and the associated increase in DEE due to human disturbance (DEEp).


Figure 18. The predicted DEE caused by human and Bald Eagle disturbances. Using 1097.5 $\mathrm{Kj} /$ day as the DEEt guideline, by 2007 all energy would be expended responding to disturbance.

## Appendix C.

## Frequency Distribution of Leg-Band Observations



Figure 19. Frequency distribution of length of stay (first and last days a band was observed) on A) Fraser River Delta 1999 and B) Fraser River Delta 2000.
A. Frequency Histogram Showing the Number of Individuals Vs. Length of Stay (Parksville-Qualicum, 1999)

B. Frequency Histogram Showing the Number of Individuals Vs. Length of Stay
(Parksville-Qualicum, 2000)


Figure 20. Frequency distribution of length of stay (first and last days a band was observed in A) Parksville-Qualicum 1999 and B) Parksville-Qualicum 2000.

## Appendix D.

## Lengths of Stay for Individually Leg-Banded Brant Observed in 1999 and 2000.

Data exploration suggested that individual Brant staged for similar lengths of time across years (Figure 21).



Figure 21. The relationship between lengths of stay for individually banded Brant observed in consecutive years at A) Parksville-Qualicum ( $r=0.63, p<0.001, n=1095$ ) and $B$ ) the Fraser River delta $(r=0.51, p=0.05, n=318)$. Each axis represents the number of days an individual Brant staged at a site, each year, based on the first and the last days a band was observed.

## Appendix E.

## Transition Probabilities

Table 19. Single Sight Transitions Probabilities from A to B and A to C. Parameters 4-30 are 1999 and 31-59 are 2000.

| Parameter | Estimate | Standard Error | $\begin{aligned} & \hline 95 \% \text { Confidence } \\ & \text { Interval } \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower | Upper |
| 1:S A:South | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 2:p A:South | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 3:Psi B to A | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 4:Psi A to B | 0.0076753 | 0.0023437 | 0.0042137 | 0.0139406 |
| 5:Psi A to B | 0.0087745 | 0.0018933 | 0.0057445 | 0.0133812 |
| 6:Psi A to B | 0.0109857 | 0.0016972 | 0.0081119 | 0.0148624 |
| 7:Psi A to B | 0.0150517 | 0.0018440 | 0.0118338 | 0.0191276 |
| 8:Psi A to B | 0.0225313 | 0.0025108 | 0.0181012 | 0.0280148 |
| 9:Psi A to B | 0.0367211 | 0.0038828 | 0.0298247 | 0.0451378 |
| 10:Psi A to B | 0.0646522 | 0.0061565 | 0.0535837 | 0.0778191 |
| 11:Psi A to B | 0.1208111 | 0.0097065 | 0.1030407 | 0.1411639 |
| 12:Psi A to B | 0.2304773 | 0.0169049 | 0.1990190 | 0.2652613 |
| 13:Psi A to B | 0.4171616 | 0.0329011 | 0.3544211 | 0.4827041 |
| 14:Psi A to B | 0.6522114 | 0.0478199 | 0.5536855 | 0.7392308 |
| 15:Psi A to B | 0.8434351 | 0.0414580 | 0.7443435 | 0.9088238 |
| 16:Psi A to B | 0.9443429 | 0.0231755 | 0.8772925 | 0.9757673 |
| 17:Psi A to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 18:Psi A to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 19:Psi A to B | 0.0063544 | 0.0020156 | 0.0034089 | 0.0118146 |
| 20:Psi A to B | 0.0114731 | 0.0023774 | 0.0076372 | 0.0172024 |
| 21:Psi A to B | 0.0205537 | 0.0027782 | 0.0157593 | 0.0267670 |
| 22:Psi A to B | 0.0364231 | 0.0037485 | 0.0297481 | 0.0445270 |
| 23:Psi A to B | 0.0635233 | 0.0060104 | 0.0527126 | 0.0763724 |
| 24:Psi A to B | 0.1081527 | 0.0094667 | 0.0909479 | 0.1281533 |
| 25:Psi A to B | 0.1776209 | 0.0137555 | 0.1522422 | 0.2062011 |
| 26:Psi A to B | 0.2770556 | 0.0220363 | 0.2359972 | 0.3222443 |
| 27:Psi A to B | 0.4038512 | 0.0406682 | 0.3272772 | 0.4854136 |
| 28:Psi A to B | 0.5440011 | 0.0667247 | 0.4132020 | 0.6690000 |
| 29:Psi A to B | 0.6766897 | 0.0882623 | 0.4869732 | 0.8219068 |
| 30:Psi A to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 31:Psi A to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 32:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 33:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 34:Psi A to C | 0.0088932 | 0.0021410 | 0.0055432 | 0.0142390 |
| 35:Psi A to C | 0.0338747 | 0.0045555 | 0.0259967 | 0.0440322 |
| 36:Psi A to C | 0.0925354 | 0.0067953 | 0.0800497 | 0.1067427 |
| 37:Psi A to C | 0.1808132 | 0.0101094 | 0.1618373 | 0.2014793 |
| 38:Psi A to C | 0.2623136 | 0.0134250 | 0.2368625 | 0.2894622 |
| 39:Psi A to C | 0.2989183 | 0.0149620 | 0.2704405 | 0.3290424 |
| 40:Psi A to C | 0.2756233 | 0.0205254 | 0.2372579 | 0.3176093 |
| 41:Psi A to C | 0.2017416 | 0.0293701 | 0.1502170 | 0.2654193 |
| 42:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 43:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 44:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |

## Table 19 Continued.

| Parameter | Estimate | Standard <br> Error |  | $95 \%$ Confidence Interval |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | 0.0000000 | 0.0000000 | 0.0000000 |  |
| Upper |  |  |  |  |  |
| 45:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.00000000 |  |
| 46:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |  |
| 47:Psi A to C | 0.0295051 | 0.0049265 | 0.0212388 | 0.0408544 |  |
| 48:Psi A to C | 0.0667803 | 0.0061635 | 0.0556678 | 0.0799235 |  |
| 49:Psi A to C | 0.1249149 | 0.0077036 | 0.1105812 | 0.1408124 |  |
| 50:Psi A to C | 0.1944088 | 0.0111445 | 0.1734911 | 0.2171859 |  |
| 51:Psi A to C | 0.2569263 | 0.0138452 | 0.2307386 | 0.2849850 |  |
| 52:Psi A to C | 0.2956983 | 0.0161588 | 0.2650386 | 0.3283203 |  |
| 53:Psi A to C | 0.3017040 | 0.0241071 | 0.2566466 | 0.3509373 |  |
| 54:Psi A to C | 0.2736873 | 0.0384968 | 0.2049606 | 0.3551643 |  |
| 55:Psi A to C | 0.2178465 | 0.0526729 | 0.1319117 | 0.3379685 |  |
| 56:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |  |
| 57:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |  |
| 58:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |  |
| 59:Psi A to C |  |  |  |  |  |

Table 20. Multiple-Sight Transitions Probabilities from A to B and A to C, 1999.

|  |  | $95 \%$ Confidence Interval |  |  |
| :---: | :---: | :--- | :---: | :---: |
| Parameter | Estimate | Standard Error | Lower | Upper |
|  |  |  |  |  |
| 1:S A:South | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 2:p A:South | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 3:Psi B to A | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 4:Psi A to B | 0.0056686 | 0.0025070 | 0.0023785 | 0.0134484 |
| 5:Psi A to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 6:Psi A to B | 0.0048935 | 0.0011902 | 0.0030365 | 0.0078772 |
| 7:Psi A to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 8:Psi A to B | 0.0094019 | 0.0018472 | 0.0063931 | 0.0138072 |
| 9:Psi A to B | 0.0175222 | 0.0031606 | 0.0122919 | 0.0249221 |
| 10:Psi A to B | 0.0393741 | 0.0060163 | 0.0291352 | 0.0530147 |
| 11:Psi A to B | 0.1033071 | 0.0128682 | 0.0806701 | 0.1313885 |
| 12:Psi A to B | 0.2836972 | 0.0343119 | 0.2214674 | 0.3554299 |
| 13:Psi A to B | 0.6247965 | 0.0635073 | 0.4947458 | 0.7390299 |
| 14:Psi A to B | 0.8954304 | 0.0399489 | 0.7877181 | 0.9518313 |
| 15:Psi A to B | 0.9817700 | 0.0111792 | 0.9405896 | 0.9945709 |
| 16:Psi A to B | 0.9975918 | 0.0020682 | 0.9871204 | 0.9995536 |
| 17:Psi A to B | 0.9997435 | $0.2908 \mathrm{E}-03$ | 0.9976368 | 0.9999722 |
| 18:Psi A to C | $0.5860 \mathrm{E}-03$ | $0.2382 \mathrm{E}-03$ | $0.2641 \mathrm{E}-03$ | 0.0012997 |
| 19:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 20:Psi A to C | 0.0286803 | 0.0041660 | 0.0215508 | 0.0380766 |
| 21:Psi A to C | 0.1069731 | 0.0075425 | 0.0930635 | 0.1226805 |
| 22:Psi A to C | 0.2507931 | 0.0107437 | 0.2303285 | 0.2724323 |
| 23:Psi A to C | 0.3918625 | 0.0145869 | 0.3636716 | 0.4207935 |
| 24:Psi A to C | 0.4607439 | 0.0165285 | 0.4285584 | 0.4932602 |
| 25:Psi A to C | 0.4383255 | 0.0227628 | 0.3943225 | 0.4833204 |
| 26:Psi A to C | 0.3293054 | 0.0355050 | 0.2637808 | 0.4022143 |
| 27:Psi A to C | 0.1754548 | 0.0389583 | 0.1115248 | 0.2650971 |
| 28:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 29:Psi A to C | 0.0128947 | 0.0075439 | 0.0040717 | 0.0400669 |
| 30:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 31:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |

Table 21. Multiple-Sight Transitions Probabilities from B to D and C to D, 1999.

| Parameter | Estimate | Standard Error | 95\% Confidence Interval |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower | Upper |
| 1:S B:BB | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 2:p D:North | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 3:Psi D to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 4:p B:BB | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 5:p B:BB | 0.6666658 | 0.2721615 | 0.1535119 | 0.9566281 |
| 6:p B:BB | 0.4999970 | 0.3535508 | 0.0588644 | 0.9411343 |
| 7:p B:BB | 0.4000008 | 0.2190885 | 0.1002289 | 0.7995942 |
| 8:p B:BB | 0.4920114 | 0.2042714 | 0.1633032 | 0.8277750 |
| 9:p B:BB | 0.0901113 | 0.0859984 | 0.0125171 | 0.4362278 |
| 10:p B:BB | 0.7807219 | 0.1197499 | 0.4747393 | 0.9334470 |
| 11:p B:BB | 0.0454544 | 0.0444092 | 0.0063644 | 0.2614575 |
| 12:p B:BB | 0.8461535 | 0.0707588 | 0.6546128 | 0.9410393 |
| 13:p B:BB | 0.6842096 | 0.1066385 | 0.4516117 | 0.8507546 |
| 14:p B:BB | 0.9642854 | 0.0350708 | 0.7858133 | 0.9949924 |
| 15:p B:BB | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 16:p B:BB | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 17:p B:BB | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 18:p C:PQ | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 19:p C:PQ | 0.6666678 | 0.2721617 | 0.1535119 | 0.9566289 |
| 20:p C:PQ | 0.3333346 | 0.2721635 | 0.0433712 | 0.8464905 |
| 21:p C:PQ | 0.6363638 | 0.1450404 | 0.3387485 | 0.8566953 |
| 22:p C:PQ | 0.6152076 | 0.0497300 | 0.5143743 | 0.7070292 |
| 23:p C:PQ | 0.8028497 | 0.0258188 | 0.7473423 | 0.8486324 |
| 24:p C:PQ | 0.9003335 | 0.0153500 | 0.8659546 | 0.9266422 |
| 25:p C:PQ | 0.9135338 | 0.0172324 | 0.8732467 | 0.9418686 |
| 26:p C:PQ | 0.9083334 | 0.0263413 | 0.8420271 | 0.9485107 |
| 27:p C:PQ | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 28:p C:PQ | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 29:p C:PQ | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 30:p C:PQ | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 31:p C:PQ | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 32:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 33:Psi B to C | 0.2211473 | 0.1934258 | 0.0304702 | 0.7195187 |
| 34:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 35:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 36:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 37:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 38:Psi B to C | 0.0744050 | 0.0706254 | 0.0106562 | 0.3749773 |
| 39:Psi B to C | 0.0928602 | 0.0544454 | 0.0280308 | 0.2665133 |
| 40:Psi B to C | 0.0263191 | 0.0258673 | 0.0037240 | 0.1635068 |
| 41:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 42:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 43:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 44:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 45:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 46:Psi B to D | 0.1516422 | 0.1764692 | 0.0120047 | 0.7244850 |
| 47:Psi B to D | 0.1544299 | 0.1357127 | 0.0232644 | 0.5834009 |
| 48:Psi B to D | 0.1700413 | 0.1190211 | 0.0377488 | 0.5169065 |
| 49:Psi B to D | 0.2014959 | 0.1285200 | 0.0500883 | 0.5470203 |
| 50:Psi B to D | 0.2544193 | 0.1638867 | 0.0590468 | 0.6498098 |
|  |  |  | 79 |  |

## Table 21 Continued.

| Parameter | Estimate | Standard Error | 95\% Confidence Interval |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower | Upper |
| 51:Psi B to D | 0.3362700 | 0.2169983 | 0.0700732 | 0.7730552 |
| 52:Psi B to D | 0.4523167 | 0.2675238 | 0.0904665 | 0.8727308 |
| 53:Psi B to D | 0.5964655 | 0.2810995 | 0.1303052 | 0.9358230 |
| 54:Psi B to D | 0.7438853 | 0.2345681 | 0.2063793 | 0.9700963 |
| 55:Psi B to D | 0.8623789 | 0.1511694 | 0.3404387 | 0.9870256 |
| 56:Psi B to D | 0.9368818 | 0.0772576 | 0.5341599 | 0.9948225 |
| 57:Psi B to D | 0.9747494 | 0.0331630 | 0.7335031 | 0.9981564 |
| 58:Psi B to D | 0.9910094 | 0.0126541 | 0.8720097 | 0.9994396 |
| 59:Psi B to D | 0.9971145 | 0.0044609 | 0.9430148 | 0.9998614 |
| 60:Psi B to D | 0.1154082 | 0.1083837 | 0.0160242 | 0.5110474 |
| 61:Psi B to D | 0.1276704 | 0.0854586 | 0.0315021 | 0.3970579 |
| 62:Psi B to D | 0.1527295 | 0.0786662 | 0.0519135 | 0.3724228 |
| 63:Psi B to D | 0.1959880 | 0.0939298 | 0.0704436 | 0.4394915 |
| 64:Psi B to D | 0.2657406 | 0.1316037 | 0.0880005 | 0.5758142 |
| 65:Psi B to D | 0.3710546 | 0.1807737 | 0.1144637 | 0.7291956 |
| 66:Psi B to D | 0.5135914 | 0.2156651 | 0.1627802 | 0.8515037 |
| 67:Psi B to D | 0.6747791 | 0.2044773 | 0.2504201 | 0.9279845 |
| 68:Psi B to D | 0.8173969 | 0.1469353 | 0.3939656 | 0.9685772 |
| 69:Psi B to D | 0.9138178 | 0.0815153 | 0.5823592 | 0.9877496 |
| 70:Psi B to D | 0.9650058 | 0.0373156 | 0.7597189 | 0.9958594 |
| 71:Psi B to D | 0.9874594 | 0.0151020 | 0.8782387 | 0.9988380 |
| 72:Psi B to D | 0.9959654 | 0.0056286 | 0.9406643 | 0.9997399 |
| 73:Psi B to D | 0.0957584 | 0.0699236 | 0.0212891 | 0.3401793 |
| 74:Psi B to D | 0.1153823 | 0.0566413 | 0.0421052 | 0.2790374 |
| 75:Psi B to D | 0.1499347 | 0.0552805 | 0.0700958 | 0.2921414 |
| 76:Psi B to D | 0.2075275 | 0.0738157 | 0.0980028 | 0.3869447 |
| 77:Psi B to D | 0.2991712 | 0.1108794 | 0.1314999 | 0.5461846 |
| 78:Psi B to D | 0.4331100 | 0.1508247 | 0.1864586 | 0.7180563 |
| 79:Psi B to D | 0.6002071 | 0.1648610 | 0.2808460 | 0.8523214 |
| 80:Psi B to D | 0.7640940 | 0.1353776 | 0.4263462 | 0.9338437 |
| 81:Psi B to D | 0.8846902 | 0.0839203 | 0.6047368 | 0.9746670 |
| 82:Psi B to D | 0.9522750 | 0.0420537 | 0.7649037 | 0.9918943 |
| 83:Psi B to D | 0.9827512 | 0.0183483 | 0.8722595 | 0.9979009 |
| 84:Psi B to D | 0.9944326 | 0.0072500 | 0.9320490 | 0.9995703 |
| 85:Psi B to D | 0.0872545 | 0.0486537 | 0.0280582 | 0.2404451 |
| 86:Psi B to D | 0.1144737 | 0.0411571 | 0.0551162 | 0.2226905 |
| 87:Psi B to D | 0.1610254 | 0.0437593 | 0.0923322 | 0.2658559 |
| 88:Psi B to D | 0.2383092 | 0.0642145 | 0.1352461 | 0.3849489 |
| 89:Psi B to D | 0.3589561 | 0.0976637 | 0.1959540 | 0.5626640 |
| 90:Psi B to D | 0.5238828 | 0.1240612 | 0.2933311 | 0.7446859 |
| 91:Psi B to D | 0.7036070 | 0.1193048 | 0.4361626 | 0.8793000 |
| 92:Psi B to D | 0.8490142 | 0.0853300 | 0.6040101 | 0.9539805 |
| 93:Psi B to D | 0.9359966 | 0.0479637 | 0.7527677 | 0.9859630 |
| 94:Psi B to D | 0.9766125 | 0.0227272 | 0.8558960 | 0.9966054 |
| 95:Psi B to D | 0.9924192 | 0.0094636 | 0.9175150 | 0.9993514 |
| 96:Psi B to D | 0.0875652 | 0.0376574 | 0.0367023 | 0.1946705 |
| 97:Psi B to D | 0.1247155 | 0.0348539 | 0.0708096 | 0.2103678 |
| 98:Psi B to D | 0.1884876 | 0.0413620 | 0.1202703 | 0.2829526 |
| 99:Psi B to D | 0.2936352 | 0.0623210 | 0.1874433 | 0.4282786 |
| 100:Psi B to D | 0.4495988 | 0.0895370 | 0.2866995 | 0.6240758 |
|  |  |  | 80 |  |

Table 21 Continued.

| Parameter | Estimate | Standard Error | 95\% Confidence Interval |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower | Upper |
| 101:Psi B to D | 0.6379869 | 0.1020988 | 0.4256046 | 0.8073811 |
| 102:Psi B to D | 0.8067446 | 0.0863946 | 0.5848903 | 0.9251942 |
| 103:Psi B to D | 0.9156593 | 0.0552616 | 0.7275568 | 0.9778452 |
| 104:Psi B to D | 0.9687501 | 0.0283891 | 0.8314560 | 0.9948929 |
| 105:Psi B to D | 0.9898153 | 0.0123511 | 0.8980058 | 0.9990687 |
| 106:Psi B to D | 0.0967751 | 0.0333784 | 0.0482443 | 0.1846534 |
| 107:Psi B to D | 0.1486871 | 0.0350950 | 0.0920917 | 0.2312064 |
| 108:Psi B to D | 0.2381478 | 0.0448353 | 0.1614863 | 0.3365934 |
| 109:Psi B to D | 0.3805158 | 0.0649423 | 0.2635994 | 0.5131531 |
| 110:Psi B to D | 0.5699311 | 0.0855302 | 0.4007440 | 0.7242228 |
| 111:Psi B to D | 0.7583999 | 0.0863194 | 0.5549204 | 0.8876823 |
| 112:Psi B to D | 0.8908751 | 0.0631973 | 0.6954197 | 0.9668770 |
| 113:Psi B to D | 0.9588663 | 0.0351183 | 0.8027816 | 0.9925649 |
| 114:Psi B to D | 0.9865012 | 0.0158936 | 0.8756941 | 0.9986827 |
| 115:Psi B to D | 0.1174114 | 0.0337896 | 0.0656031 | 0.2013189 |
| 116:Psi B to D | 0.1923052 | 0.0388075 | 0.1273268 | 0.2798114 |
| 117:Psi B to D | 0.3187333 | 0.0496427 | 0.2301099 | 0.4227468 |
| 118:Psi B to D | 0.5023326 | 0.0690265 | 0.3700857 | 0.6342538 |
| 119:Psi B to D | 0.7050967 | 0.0825531 | 0.5233683 | 0.8388685 |
| 120:Psi B to D | 0.8614603 | 0.0698543 | 0.6637992 | 0.9514167 |
| 121:Psi B to D | 0.9466818 | 0.0422367 | 0.7749720 | 0.9891938 |
| 122:Psi B to D | 0.9823520 | 0.0199364 | 0.8538848 | 0.9981175 |
| 123:Psi B to D | 0.1551834 | 0.0372938 | 0.0951716 | 0.2428789 |
| 124:Psi B to D | 0.2652202 | 0.0420588 | 0.1912330 | 0.3552589 |
| 125:Psi B to D | 0.4378036 | 0.0524489 | 0.3390023 | 0.5417976 |
| 126:Psi B to D | 0.6484599 | 0.0723359 | 0.4975828 | 0.7745566 |
| 127:Psi B to D | 0.8275070 | 0.0725866 | 0.6390745 | 0.9285597 |
| 128:Psi B to D | 0.9319650 | 0.0486287 | 0.7528875 | 0.9840226 |
| 129:Psi B to D | 0.9772441 | 0.0241714 | 0.8361026 | 0.9972415 |
| 130:Psi B to D | 0.2200126 | 0.0427454 | 0.1475676 | 0.3148850 |
| 131:Psi B to D | 0.3783250 | 0.0423897 | 0.2994499 | 0.4642091 |
| 132:Psi B to D | 0.5904179 | 0.0563806 | 0.4771973 | 0.6948008 |
| 133:Psi B to D | 0.7894275 | 0.0693086 | 0.6234636 | 0.8946062 |
| 134:Psi B to D | 0.9145648 | 0.0529940 | 0.7391132 | 0.9758734 |
| 135:Psi B to D | 0.9710647 | 0.0281651 | 0.8247196 | 0.9958397 |
| 136:Psi B to D | 0.3251029 | 0.0525258 | 0.2315407 | 0.4350675 |
| 137:Psi B to D | 0.5329369 | 0.0488182 | 0.4372170 | 0.6262919 |
| 138:Psi B to D | 0.7479527 | 0.0619970 | 0.6090099 | 0.8497056 |
| 139:Psi B to D | 0.8944416 | 0.0546823 | 0.7314069 | 0.9634591 |
| 140:Psi B to D | 0.9637214 | 0.0315101 | 0.8195217 | 0.9936064 |
| 141:Psi B to D | 0.4777677 | 0.0746437 | 0.3372990 | 0.6218430 |
| 142:Psi B to D | 0.7040787 | 0.0652317 | 0.5629521 | 0.8146393 |
| 143:Psi B to D | 0.8716927 | 0.0560342 | 0.7178956 | 0.9477461 |
| 144:Psi B to D | 0.9551545 | 0.0342782 | 0.8161025 | 0.9903120 |
| 145:Psi B to D | 0.6589679 | 0.0996388 | 0.4476148 | 0.8216709 |
| 146:Psi B to D | 0.8465651 | 0.0662109 | 0.6701425 | 0.9374380 |
| 147:Psi B to D | 0.9453471 | 0.0380473 | 0.8033217 | 0.9865325 |
| 148:Psi B to D | 0.8194522 | 0.0972072 | 0.5559816 | 0.9426981 |
| 149:Psi B to D | 0.9343357 | 0.0470354 | 0.7599968 | 0.9846005 |
| 150:Psi B to D | 0.9222157 | 0.0666441 | 0.6574409 | 0.9865305 |
|  |  |  | 81 |  |

Table 21 Continued.

| Parameter | Estimate | Standard Error | 95\% Confidence Interval |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower | Upper |
| 151:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 152:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 153:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 154:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 155:Psi C to B | 0.0110883 | 0.0110602 | 0.0015504 | 0.0748997 |
| 156:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 157:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 158:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 159:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 160:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 161:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 162:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 163:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 164:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 165:Psi C to D | $0.926 \mathrm{E}-04$ | $0.175 \mathrm{E}-03$ | 0.226E-05 | 0.0037787 |
| 166:Psi C to D | $0.139 \mathrm{E}-03$ | $0.197 \mathrm{E}-03$ | 0.862E-05 | 0.0022462 |
| 167:Psi C to D | $0.251 \mathrm{E}-03$ | $0.270 \mathrm{E}-03$ | 0.306E-04 | 0.0020672 |
| 168:Psi C to D | $0.546 \mathrm{E}-03$ | $0.478 \mathrm{E}-03$ | 0.981E-04 | 0.0030386 |
| 169:Psi C to D | 0.0014256 | 0.0011533 | $0.291 \mathrm{E}-03$ | 0.0069373 |
| 170:Psi C to D | 0.0044566 | 0.0036345 | $0.897 \mathrm{E}-03$ | 0.0218105 |
| 171:Psi C to D | 0.0165866 | 0.0137286 | 0.0032299 | 0.0807061 |
| 172:Psi C to D | 0.0709399 | 0.0555487 | 0.0144247 | 0.2848773 |
| 173:Psi C to D | 0.2934636 | 0.1690703 | 0.0774982 | 0.6725172 |
| 174:Psi C to D | 0.7308077 | 0.1531164 | 0.3712694 | 0.9258227 |
| 175:Psi C to D | 0.9551993 | 0.0334944 | 0.8213630 | 0.9899868 |
| 176:Psi C to D | 0.9950544 | 0.0044315 | 0.9717857 | 0.9991499 |
| 177:Psi C to D | 0.9995619 | $0.512 \mathrm{E}-03$ | 0.9956651 | 0.9999559 |
| 178:Psi C to D | 0.9999678 | $0.510 \mathrm{E}-04$ | 0.9992812 | 0.9999986 |
| 179:Psi C to D | $0.443 \mathrm{E}-03$ | $0.620 \mathrm{E}-03$ | $0.2852 \mathrm{E}-04$ | 0.0068539 |
| 180:Psi C to D | $0.801 \mathrm{E}-03$ | $0.797 \mathrm{E}-03$ | $0.1137 \mathrm{E}-03$ | 0.0056224 |
| 181:Psi C to D | 0.0017389 | 0.0012505 | $0.4242 \mathrm{E}-03$ | 0.0070982 |
| 182:Psi C to D | 0.0045252 | 0.0026388 | 0.0014399 | 0.0141276 |
| 183:Psi C to D | 0.0140536 | 0.0077861 | 0.0047158 | 0.0411172 |
| 184:Psi C to D | 0.0509675 | 0.0279118 | 0.0170355 | 0.1426764 |
| 185:Psi C to D | 0.1955784 | 0.0913157 | 0.0723069 | 0.4313016 |
| 186:Psi C to D | 0.5694378 | 0.1376870 | 0.3055224 | 0.7990315 |
| 187:Psi C to D | 0.8963121 | 0.0510913 | 0.7463820 | 0.9621084 |
| 188:Psi C to D | 0.9854839 | 0.0087842 | 0.9532179 | 0.9955985 |
| 189:Psi C to D | 0.9984415 | 0.0012693 | 0.9923373 | 0.9996846 |
| 190:Psi C to D | 0.9998624 | $0.159 \mathrm{E}-03$ | 0.9986725 | 0.9999857 |
| 191:Psi C to D | 0.9999899 | $0.163 \mathrm{E}-04$ | 0.9997598 | 0.9999996 |
| 192:Psi C to D | 0.0018461 | 0.0018082 | $0.270 \mathrm{E}-03$ | 0.0124992 |
| 193:Psi C to D | 0.0040008 | 0.0025851 | 0.0011250 | 0.0141234 |
| 194:Psi C to D | 0.0103738 | 0.0045232 | 0.0044006 | 0.0242570 |
| 195:Psi C to D | 0.0318236 | 0.0112357 | 0.0158290 | 0.0629464 |
| 196:Psi C to D | 0.1101962 | 0.0358493 | 0.0570364 | 0.2022744 |
| 197:Psi C to D | 0.3592445 | 0.0855858 | 0.2129240 | 0.5374546 |
| 198:Psi C to D | 0.7530740 | 0.0682237 | 0.5977212 | 0.8622578 |
| 199:Psi C to D | 0.9522305 | 0.0179969 | 0.9017620 | 0.9774207 |
| 200:Psi C to D | 0.9936529 | 0.0033590 | 0.9821808 | 0.9977561 |

Table 21 Continued

|  |  |  | 95\% Confidence Interval |  |
| :--- | :---: | :--- | :--- | :--- |
| Parameter | Estimate | Standard Error | Lower | Upper |
|  |  |  |  |  |
| 201:Psi C to D | 0.9993236 | $0.541 \mathrm{E}-03$ | 0.9967576 | 0.9998592 |
| 202:Psi C to D | 0.9999403 | $0.706 \mathrm{E}-04$ | 0.9993928 | 0.9999941 |
| 203:Psi C to D | 0.9999956 | $0.730 \mathrm{E}-05$ | 0.9998849 | 0.9999998 |
| 204:Psi C to D | 0.0066636 | 0.0042224 | 0.0019178 | 0.0228847 |
| 205:Psi C to D | 0.0172051 | 0.0063810 | 0.0082862 | 0.0353815 |
| 206:Psi C to D | 0.0520375 | 0.0116657 | 0.0333796 | 0.0802584 |
| 207:Psi C to D | 0.1713793 | 0.0296699 | 0.1207457 | 0.2375101 |
| 208:Psi C to D | 0.4835585 | 0.0545107 | 0.3790420 | 0.5895334 |

Table 22. Multiple-Sight Transitions Probabilities from $A$ to $B$ and $A$ to $C$, 2000.

|  |  |  | $95 \%$ Confidence Interval |  |
| :--- | :--- | :--- | :--- | :--- |
| Parameter | Estimate | Standard Error | Lower | Upper |
|  |  |  |  |  |
| 1:S A South | 1.0000000 | 0.0000000 | 1.0000000 | 1.00000000 |
| 2:p A:South | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 3:Psi B to A | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 4:Psi A to B | 0.0035139 | 0.0013631 | 0.0016415 | 0.0075061 |
| 5:Psi A to B | 0.0073857 | 0.0019004 | 0.0044567 | 0.0122160 |
| 6:Psi A to B | 0.0144667 | 0.0024427 | 0.0103831 | 0.0201237 |
| 7:Psi A to B | 0.0263622 | 0.0033676 | 0.0205070 | 0.0338313 |
| 8:Psi A to B | 0.0446116 | 0.0053898 | 0.0351625 | 0.0564515 |
| 9:Psi A to B | 0.0700205 | 0.0084683 | 0.0551346 | 0.0885488 |
| 10:Psi A to B | 0.1019389 | 0.0120107 | 0.0806899 | 0.1280044 |
| 11:Psi A to B | 0.1379297 | 0.0168755 | 0.1080550 | 0.1744485 |
| 12:Psi A to B | 0.1741409 | 0.0268909 | 0.1275226 | 0.2332456 |
| 13:Psi A to B | 0.2062362 | 0.0454092 | 0.1310799 | 0.3091527 |
| 14:Psi A to B | 0.2303730 | 0.0725359 | 0.1183589 | 0.4002679 |
| 15:Psi A to B | 0.2438130 | 0.1064781 | 0.0941619 | 0.5000177 |
| 16:Psi A to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 17:Psi A to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 18:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 19:Psi A to C | 0.0701546 | 0.0076460 | 0.0565697 | 0.0867021 |
| 20:Psi A to C | 0.1290086 | 0.0079990 | 0.1141251 | 0.1455143 |
| 21:Psi A to C | 0.2042240 | 0.0093453 | 0.1865163 | 0.2231517 |
| 22:Psi A to C | 0.2818389 | 0.0127977 | 0.2574462 | 0.3075855 |
| 23:Psi A to C | 0.3462584 | 0.0158940 | 0.3157987 | 0.3780327 |
| 24:Psi A to C | 0.3868420 | 0.0185851 | 0.3511064 | 0.4238390 |
| 25:Psi A to C | 0.398690 | 0.0238645 | 0.3530427 | 0.4463319 |
| 26:Psi A to C | 0.3809547 | 0.0339828 | 0.3169263 | 0.4494080 |
| 27:Psi A to C | 0.3350771 | 0.0475175 | 0.2491170 | 0.4335717 |
| 28:Psi A to C | 0.2669734 | 0.0597100 | 0.1668592 | 0.3984293 |
| 29:Psi A to C | 0.1885183 | 0.0639695 | 0.0928555 | 0.3452292 |
| 3:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.000000 |
| 31:Psi A to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |

Table 23. Multiple-Sight Transitions Probabilities from B to D and C to D, 2000.

| Parameter | Estimate | Standard Error | 95\% Confidence Interval |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower | Upper |
| 1:S B:BB | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 2:p D:North | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 3:Psi D to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 4:p B:BB | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 5:p B:BB | 0.1111105 | 0.1047560 | 0.0153928 | 0.4998616 |
| 6:p B:BB | 0.2727272 | 0.1342815 | 0.0904828 | 0.5856704 |
| 7:p B:BB | 0.2142857 | 0.1096643 | 0.0707051 | 0.4943358 |
| 8:p B:BB | 0.5066097 | 0.1226048 | 0.2819218 | 0.7286591 |
| 9:p B:BB | 0.4826051 | 0.1000343 | 0.2984212 | 0.6716400 |
| 10:p B:BB | 0.7244658 | 0.0718127 | 0.5650299 | 0.8418216 |
| 11:p B:BB | 0.6900445 | 0.0734593 | 0.5317458 | 0.8135888 |
| 12:p B:BB | 0.7333332 | 0.0807371 | 0.5504265 | 0.8606628 |
| 13:p B:BB | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 14:p B:BB | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 15:p B:BB | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 16:p B:BB | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 17:p B:BB | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 18:p C:PQ | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 19:p C:PQ | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 20:p C:PQ | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 21:p C:PQ | 0.8972603 | 0.0251277 | 0.8365650 | 0.9371095 |
| 22:p C:PQ | 0.7908068 | 0.0279441 | 0.7308067 | 0.8403550 |
| 23:p C:PQ | 0.8462427 | 0.0225893 | 0.7965967 | 0.8855128 |
| 24:p C:PQ | 0.8299265 | 0.0242125 | 0.7771076 | 0.8722859 |
| 25:p C:PQ | 0.9206005 | 0.0203909 | 0.8703173 | 0.9524521 |
| 26:p C:PQ | 0.9056605 | 0.0401505 | 0.7926057 | 0.9601828 |
| 27:p C:PQ | 0.5384619 | 0.1382623 | 0.2816530 | 0.7763604 |
| 28:p C:PQ | 0.9333334 | 0.0644058 | 0.6480034 | 0.9906949 |
| 29:p C:PQ | 1.0000000 | 0.0000000 | 1.0000000 | 1.0000000 |
| 30:p C:PQ | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 31:p C:PQ | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 32:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 33:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 34:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| $35:$ Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 36:Psi B to C | 0.1878565 | 0.0821892 | 0.0744767 | 0.3993618 |
| 37:Psi B to C | 0.1427766 | 0.0650123 | 0.0555386 | 0.3205386 |
| 38:Psi B to C | 0.0248283 | 0.0244161 | 0.0035151 | 0.1552383 |
| 39:Psi B to C | 0.1426296 | 0.0417642 | 0.0784921 | 0.2452285 |
| 40:Psi B to C | 0.0836084 | 0.0388863 | 0.0326389 | 0.1978904 |
| 41:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 42:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 43:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 44:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 45:Psi B to C | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 46:Psi B to D | 0.0638400 | 0.0646908 | 0.0081064 | 0.3626580 |
| 47:Psi B to D | 0.1004327 | 0.0685717 | 0.0246013 | 0.3307475 |
| 48:Psi B to D | 0.1615647 | 0.0779274 | 0.0587362 | 0.3730634 |
| 49:Psi B to D | 0.2596049 | 0.1063844 | 0.1059454 | 0.5091981 |

Table 23 Continued.

| Parameter | Estimate | Standard Error | 95\% Confidence Interval |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower | Upper |
| 50:Psi B to D | 0.4021294 | 0.1518134 | 0.1632508 | 0.6986834 |
| 51:Psi B to D | 0.5763084 | 0.1779221 | 0.2459118 | 0.8501542 |
| 52:Psi B to D | 0.7435856 | 0.1539844 | 0.3732655 | 0.9338636 |
| 53:Psi B to D | 0.8669845 | 0.0990857 | 0.5474841 | 0.9723099 |
| 54:Psi B to D | 0.9391884 | 0.0507943 | 0.7298962 | 0.9887978 |
| 55:Psi B to D | 0.9747346 | 0.0225551 | 0.8650209 | 0.9957128 |
| 56:Psi B to D | 0.9902534 | 0.0093407 | 0.9384390 | 0.9985254 |
| 57:Psi B to D | 0.9964673 | 0.0038004 | 0.9714231 | 0.9995729 |
| 58:Psi B to D | 0.9987902 | 0.0015442 | 0.9853899 | 0.9999011 |
| 59:Psi B to D | 0.9996076 | 0.616E-03 | 0.9915098 | 0.9999820 |
| 60:Psi B to D | 0.0612414 | 0.0462856 | 0.0132853 | 0.2401695 |
| 61:Psi B to D | 0.1012019 | 0.0479155 | 0.0385525 | 0.2402221 |
| 62:Psi B to D | 0.1700416 | 0.0544732 | 0.0877146 | 0.3038988 |
| 63:Psi B to D | 0.2821323 | 0.0807890 | 0.1524210 | 0.4620528 |
| 64:Psi B to D | 0.4428331 | 0.1185161 | 0.2366481 | 0.6707990 |
| 65:Psi B to D | 0.6288718 | 0.1305889 | 0.3614005 | 0.8353531 |
| 66:Psi B to D | 0.7920367 | 0.1017722 | 0.5315135 | 0.9274573 |
| 67:Psi B to D | 0.9002430 | 0.0594912 | 0.7112653 | 0.9706398 |
| 68:Psi B to D | 0.9575244 | 0.0291216 | 0.8470982 | 0.9892157 |
| 69:Psi B to D | 0.9834345 | 0.0132161 | 0.9236976 | 0.9965769 |
| 70:Psi B to D | 0.9939694 | 0.0058620 | 0.9603845 | 0.9991084 |
| 71:Psi B to D | 0.9979314 | 0.0025409 | 0.9773875 | 0.9998143 |
| 72:Psi B to D | 0.9993286 | 0.0010539 | 0.9856069 | 0.9999691 |
| 73:Psi B to D | 0.0636864 | 0.0348026 | 0.0212086 | 0.1759476 |
| 74:Psi B to D | 0.1101343 | 0.0349223 | 0.0579880 | 0.1992548 |
| 75:Psi B to D | 0.1918634 | 0.0406875 | 0.1243055 | 0.2842210 |
| 76:Psi B to D | 0.3243812 | 0.0654339 | 0.2110004 | 0.4629397 |
| 77:Psi B to D | 0.5058353 | 0.0925425 | 0.3313048 | 0.6789551 |
| 78:Psi B to D | 0.6970329 | 0.0916250 | 0.4957032 | 0.8433813 |
| 79:Psi B to D | 0.8449971 | 0.0647931 | 0.6739918 | 0.9349589 |
| 80:Psi B to D | 0.9315907 | 0.0368799 | 0.8141409 | 0.9769242 |
| 81:Psi B to D | 0.9728722 | 0.0189131 | 0.8979918 | 0.9932018 |
| 82:Psi B to D | 0.9900562 | 0.0091399 | 0.9416499 | 0.9983747 |
| 83:Psi B to D | 0.9965802 | 0.0041566 | 0.9638845 | 0.9996858 |
| 84:Psi B to D | 0.9988891 | 0.0017630 | 0.9755741 | 0.9999506 |
| 85:Psi B to D | 0.0717484 | 0.0279317 | 0.0328582 | 0.1495507 |
| 86:Psi B to D | 0.1291254 | 0.0283477 | 0.0829591 | 0.1955057 |
| 87:Psi B to D | 0.2306798 | 0.0362717 | 0.1672738 | 0.3091962 |
| 88:Psi B to D | 0.3899733 | 0.0581779 | 0.2835845 | 0.5079746 |
| 89:Psi B to D | 0.5896314 | 0.0738994 | 0.4412315 | 0.7233318 |
| $90:$ Psi B to D | 0.7729637 | 0.0666411 | 0.6179438 | 0.8775483 |
| 91:Psi B to D | 0.8947888 | 0.0462157 | 0.7646641 | 0.9570086 |
| 92:Psi B to D | 0.9572595 | 0.0270917 | 0.8594928 | 0.9879524 |
| 93:Psi B to D | 0.9841725 | 0.0140865 | 0.9135357 | 0.9972749 |
| 94:Psi B to D | 0.9945354 | 0.0066262 | 0.9434421 | 0.9994966 |
| 95:Psi B to D | 0.9982225 | 0.0028494 | 0.9602003 | 0.9999235 |
| 96:Psi B to D | 0.0873675 | 0.0248221 | 0.0494360 | 0.1498164 |
| 97:Psi B to D | 0.1621969 | 0.0279667 | 0.1145230 | 0.2246809 |
| 98:Psi B to D | 0.2921593 | 0.0387092 | 0.2223900 | 0.3733088 |
| 99:Psi B to D | 0.4812454 | 0.0561485 | 0.3738165 | 0.5904367 |

Table 23 Continued.

| Parameter | Estimate | Standard Error | 95\% Confidence Interval |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower | Upper |
| 100:Psi B to D | 0.6873216 | 0.0648381 | 0.5489178 | 0.7988231 |
| 101:Psi B to D | 0.8459420 | 0.0560178 | 0.7027956 | 0.9272772 |
| 103:Psi B to D | 0.9756971 | 0.0209116 | 0.8769714 | 0.9955970 |
| 104:Psi B to D | 0.9915616 | 0.0101299 | 0.9163360 | 0.9992074 |
| 105:Psi B to D | 0.9972496 | 0.0044096 | 0.9394728 | 0.9998819 |
| 106:Psi B to D | 0.1144373 | 0.0253914 | 0.0732857 | 0.1743491 |
| 107:Psi B to D | 0.2159977 | 0.0315170 | 0.1605788 | 0.2840701 |
| 108:Psi B to D | 0.3824226 | 0.0422875 | 0.3035988 | 596 |
| 109:Psi B to D | 0.5946935 | 0.0575762 | 0.478815 | 0.7009012 |
| 110:Psi B to D | 0.7856494 | 0.0625712 | 0.6389140 | 0.8836174 |
| 111:Psi B to D | 0.9061245 | 0.0488527 | 0.7579649 | 0.9674806 |
| 112:Psi B to D | 0.9640263 | 0.0292956 | 0.8365284 | 0.9929247 |
| 113:Psi B to D | 0.9874111 | 0.0146918 | 0.8855129 | 0.9987443 |
| 114:Psi B to D | 0.9958851 | 0.0064973 | 0.9154012 | 0.9998153 |
| 115:Psi B to D | 0.1597489 | 0.0296065 | 0.1098369 | 0.2265690 |
| 116:Psi B to D | 0.2993844 | 0.0349332 | 0.2356537 | 0.3719630 |
| 117:Psi B to D | 0.5031123 | 0.0449674 | 0.4157832 | 0.5902520 |
| 118:Psi B to D | 0.7166581 | 0.0612270 | 0.5834517 | . 8203808 |
| 119:Psi B to D | 0.8694670 | 0.0576725 | 0.7110055 | 0.9474613 |
| 120:Psi B to D | 0.9486988 | 0.0381817 | 0.7989482 | . 9885134 |
| 121:Psi B to D | 0.9818598 | 0.0200739 | 0.8559831 | 0.9979753 |
| 122:Psi B to D | 0.9940481 | 0.0090823 | 0.8918053 | 0.9997046 |
| 123:Psi B to D | 0.2336324 | 0.0383198 | 0.1669512 | 0.3168177 |
| 124:Psi B to D | 0.4194011 | 0.0385740 | 0.3462146 | 0.4963140 |
| 125:Psi B to D | 0.6434254 | 0.0520851 | 0.5362620 | 0.7379271 |
| 126:Psi B to D | 0.8261484 | 0.0609252 | 0.6741818 | 0.9160603 |
| 127:Psi B to D | 0.9295434 | 0.0459140 | 0.7695234 | 0.9811788 |
| 128:Psi B to D | 0.9747569 | 0.0257336 | 0.8325503 | 0.9966767 |
| 129:Psi B to D | 0.9916771 | 0.0120195 | 0.8728160 | . 9995168 |
| 130:Psi B to D | 0.3475915 | 0.0562896 | 0.2467248 | . 4642786 |
| 131:Psi B to D | 0.5709803 | 0.0526837 | 0.4661312 | 0.6698235 |
| Psi B to D | 0.7780194 | 0.0601555 | 0.6391019 | 0.8740062 |
| 133:Psi B to D | 0.9068094 | 0.0512645 | 0.7476803 | 0.9696545 |
| 134:Psi B to D | 0.9660794 | 0.0309827 | 0.8169955 | 0.9945264 |
| 135:Psi B to D | 0.9887489 | 0.0150611 | 0.8608537 | 0.9991996 |
| 136:Psi B to D | 0.5036781 | 0.0864362 | 0.3400696 | 0.6665026 |
| 137:Psi B to D | 0.7277074 | 0.0705763 | 0.5707507 | 0.8430540 |
| 138:Psi B to D | 0.8812317 | 0.0563160 | 0.7210194 | 0.9551596 |
| 139:Psi B to D | 0.9559796 | 0.0355869 | 0.8054358 | 0.9912987 |
| 140:Psi B to D | 0.9852962 | 0.0179647 | 0.8550074 | 0.9986885 |
| 141:Psi B to D | 0.6781218 | 0.1111252 | 0.4371575 | 0.8510694 |
| 142:Psi B to D | 0.8539941 | 0.0703868 | 0.6592273 | 0.9464804 |
| 143:Psi B to D | 0.9448107 | 0.0410596 | 0.7853030 | 0.9876734 |
| 144:Psi B to D | 0.9814210 | 0.0207685 | 0.8499882 | 0.9979735 |
| 145:Psi B to D | 0.8265943 | 0.1049606 | 0.5315662 | 0.9524351 |
| 146:Psi B to D | 0.9331191 | 0.0518212 | 0.7326493 | 0.9861173 |
| 147:Psi B to D | 0.9772987 | 0.0242318 | 0.8350192 | 0.9972765 |
| 148:Psi B to D | 0.9216004 | 0.0730578 | 0.6183295 | 0.9884119 |
| 149:Psi B to D | 0.9731702 | 0.0300359 | 0.7918868 | 0.9971162 |
| 150:Psi B to D | 0.9693202 | 0.0399606 | 0.6940829 | 0.9977323 |

Table 23 Continued.

| Parameter | Estimate | Standard Error | 95\% Confidence Interval |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower | Upper |
| 151:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 152:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 153:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 154:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 155:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 156:Psi C to B | 0.0095344 | 0.0071609 | 0.0021729 | 0.0408158 |
| 157:Psi C to B | 0.0150341 | 0.0073809 | 0.0057133 | 0.0389649 |
| 158:Psi C to B | 0.0039002 | 0.0038357 | $0.565 \mathrm{E}-03$ | 0.0263991 |
| 159:Psi C to B | 0.0482131 | 0.0190365 | 0.0219713 | 0.1025120 |
| 160:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 161:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 162:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 163:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 164:Psi C to B | 0.0000000 | 0.0000000 | 0.0000000 | 0.0000000 |
| 165:Psi C to D | 0.0015584 | 0.0011847 | $0.350 \mathrm{E}-03$ | 0.0068940 |
| 166:Psi C to D | 0.0046242 | 0.0025555 | 0.0015623 | 0.0136050 |
| 167:Psi C to D | 0.0133286 | 0.0057103 | 0.0057346 | 0.0306687 |
| 168:Psi C to D | 0.0369573 | 0.0143048 | 0.0171562 | 0.0778032 |
| 169:Psi C to D | 0.0962539 | 0.0369693 | 0.0442545 | 0.1967735 |
| 170:Psi C to D | 0.2240714 | 0.0799714 | 0.1049279 | 0.4156732 |
| 171:Psi C to D | 0.4334114 | 0.1184391 | 0.2291220 | 0.6631543 |
| 172:Psi C to D | 0.6643835 | 0.1079596 | 0.4338684 | 0.8364251 |
| 173:Psi C to D | 0.8334689 | 0.0651072 | 0.6661944 | 0.9262050 |
| 174:Psi C to D | 0.9251607 | 0.0313453 | 0.8357996 | 0.9677654 |
| 175:Psi C to D | 0.9675643 | 0.0146227 | 0.9228910 | 0.9867282 |
| 176:Psi C to D | 0.9859794 | 0.0075313 | 0.9602775 | 0.9951355 |
| 177:Psi C to D | 0.9938637 | 0.0042988 | 0.9760098 | 0.9984515 |
| 178:Psi C to D | 0.9972635 | 0.0025635 | 0.9829963 | 0.9995649 |
| 179:Psi C to D | 0.0047127 | 0.0026406 | 0.0015683 | 0.0140724 |
| 180:Psi C to D | 0.0135816 | 0.0051751 | 0.0064162 | 0.0285192 |
| 181:Psi C to D | 0.0376415 | 0.0105162 | 0.0216624 | 0.0646289 |
| 182:Psi C to D | 0.0979243 | 0.0242678 | 0.0595824 | 0.1568250 |
| 183:Psi C to D | 0.2274017 | 0.0522967 | 0.1410662 | 0.3453327 |
| 184:Psi C to D | 0.4380964 | 0.0785771 | 0.2943134 | 0.5930874 |
| 185:Psi C to D | 0.6686190 | 0.0721287 | 0.5159673 | 0.7924899 |
| 186:Psi C to D | 0.8360969 | 0.0441787 | 0.7305869 | 0.9056239 |
| 187:Psi C to D | 0.9264694 | 0.0227218 | 0.8676051 | 0.9603578 |
| 188:Psi C to D | 0.9681570 | 0.0121704 | 0.9334432 | 0.9850551 |
| 189:Psi C to D | 0.9862404 | 0.0071559 | 0.9622619 | 0.9950613 |
| 190:Psi C to D | 0.9939788 | 0.0043599 | 0.9753646 | 0.9985493 |
| 191:Psi C to D | 0.9973150 | 0.0026493 | 0.9816250 | 0.9996129 |
| 192:Psi C to D | 0.0128081 | 0.0050076 | 0.0059345 | 0.0274235 |
| 193:Psi C to D | 0.0355473 | 0.0085638 | 0.0220901 | 0.0567271 |
| 194:Psi C to D | 0.0927997 | 0.0149914 | 0.0672990 | 0.1266510 |
| 195:Psi C to D | 0.2171322 | 0.0297276 | 0.1644852 | 0.2809635 |
| 196:Psi C to D | 0.4235279 | 0.0471160 | 0.3348021 | 0.5174752 |
| 197:Psi C to D | 0.6553249 | 0.0464688 | 0.5595414 | 0.7399593 |
| 198:Psi C to D | 0.8277911 | 0.0315246 | 0.7570571 | 0.8811633 |
| 199:Psi C to D | 0.9223178 | 0.0192883 | 0.8750767 | 0.9526606 |
| 200:Psi C to D | 0.9662734 | 0.0122351 | 0.9320939 | 0.9835528 |

Table 23 Continued.

| Parameter | Estimate | Standard Error | 95\% Confidence Interval |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower | Upper |
| 201:Psi C to D | 0.9854103 | 0.0078043 | 0.9588604 | 0.9949168 |
| 202:Psi C to D | 0.9936126 | 0.0048601 | 0.9719710 | 0.9985690 |
| 203:Psi C to D | 0.9971511 | 0.0029541 | 0.9785383 | 0.9996280 |
| 204:Psi C to D | 0.0311111 | 0.0078936 | 0.0188568 | 0.0509159 |
| 205:Psi C to D | 0.0818248 | 0.0111441 | 0.0624714 | 0.1064928 |
| 206:Psi C to D | 0.1946075 | 0.0158213 | 0.1654542 | 0.2274973 |
| 207:Psi C to D | 0.3902657 | 0.0253590 | 0.3418477 | 409469 |
| 208:Psi C to D | 0.6235494 | 0.0298308 | 0.5635443 | 94 |
| 209:Psi C to D | 0.8072384 | 0.0258686 | 0.7514429 | 0.8529598 |
| 210:Psi C to D | 0.9118451 | 0.0199042 | 0.864250 | 0.9438374 |
| 211:Psi C to D | 0.9614791 | 0.0141079 | 0.9220607 | 0.9813643 |
| 212:Psi C to D | 0.9832894 | 0.0092852 | 0.9510740 | 0.9944170 |
| 213:Psi C to D | 0.9926751 | 0.0058023 | 0.9659437 | 0.9984581 |
| 214:Psi C to D | 0.9967313 | 0.0035130 | 0.9735765 | 0.9996039 |
| 215:Psi C to D | 0.0669716 | 0.0102514 | 0.0494668 | 0.0900836 |
| 216:Psi C to D | 0.1629140 | 0.0121847 | 0.1404175 | 0.1882255 |
| 217: Psi C to D | 0.3401659 | 0.0155528 | 0.3103749 | 0.3712769 |
| 218:Psi C to D | 0.5715754 | 0.0219226 | 0.5281733 | 0.6139036 |
| 219:Psi C to D | 0.7713245 | 0.0253725 | 0.7178581 | 0.8172385 |
| 220:Psi C to D | 0.8928335 | 0.0231083 | 0.8384381 | 343 |
| 221:Psi C to D | 0.9526154 | 0.017414 | 0.904196 | . 9771809 |
| 222:Psi C to D | 0.9793363 | 0.0116702 | 0.9386698 | 0.9932323 |
| 223:Psi C to D | 0.9909219 | 0.0073208 | 0.9567981 | 0.9981447 |
| 224:Psi C to D | 0.9959449 | 0.0044299 | 0.9662298 | 0.9995259 |
| 225:Psi C to D | 0.1265807 | 0.0123160 | 0.1043427 | 0.1527499 |
| 226:Psi C to D | 0.2774019 | 0.0155507 | 0.2479753 | 0.3088863 |
| 227:Psi C to D | 0.4983628 | 0.0199424 | 0.4593651 | 0.5373804 |
| 228:Psi C to D | 0.7152407 | 0.0261727 | 0.6613029 | 0.7636586 |
| 229:Psi C to D | 0.8611871 | 0.0272503 | 0.7987337 | 0.9065290 |
| 230:Psi C to D | 0.9373846 | 0.0220107 | 0.8777266 | 0.9689643 |
| 231:Psi C to D | 0.9724460 | 0.0152248 | 0.9205611 | . 9907820 |
| 232:Psi C to D | 0.9878468 | 0.0096895 | 0.9435362 | 0.9974772 |
| 233:Psi C to D | 0.9945620 | 0.0059035 | 0.95561 | 0.9993567 |
| 234:Psi C to D | 0.2090456 | 0.0176040 | 0.1766275 | 0.2456386 |
| 235:Psi C to D | 0.4061628 | 0.0209198 | 0.3658992 | 0.4477286 |
| 236:Psi C to D | 0.6335954 | 0.0248170 | 0.5837408 | 0.6807447 |
| 237:Psi C to D | 0.8102881 | 0.0301793 | 0.7440426 | 0.8625555 |
| 238:Psi C to D | 0.9115558 | 0.0275519 | 0.8406357 | 0.9526912 |
| 239:Psi C to D | 0.9604701 | 0.0203224 | 0.8948489 | 0.9857897 |
| 240:Psi C to D | 0.9824438 | 0.0133653 | 0.9245510 | 0.9961021 |
| 241:Psi C to D | 0.9921206 | 0.0082872 | 0.9403512 | 0.9990066 |
| 242:Psi C to D | 0.3033016 | 0.0264793 | 0.2540300 | 0.3575494 |
| 243:Psi C to D | 0.5239564 | 0.0244952 | 0.4758724 | 0.5716002 |
| 244:Psi C to D | 0.7310801 | 0.0293462 | 0.6698588 | 0.7845992 |
| 245:Psi C to D | 0.8677267 | 0.0327726 | 0.7894084 | 0.9198751 |
| 246:Psi C to D | 0.9392658 | 0.0272228 | 0.8585449 | 0.9752514 |
| 247:Psi C to D | 0.9726913 | 0.0190375 | 0.8973516 | 0.9931565 |
| 248:Psi C to D | 0.9876762 | 0.0121992 | 0.9182969 | 0.9982532 |
| 249:Psi C to D | 0.3930890 | 0.0363815 | 0.3244835 | 0.4661894 |
| 250 Psi C to D | 0.6153517 | 0.0292448 | 0.5566794 | 0.6708502 |

Table 23 Continued.

| Parameter | Estimate | Standard Error | 95\% Confidence Interval |  |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  | Lower | Upper |
| 251:Psi C to D | 0.7942542 | 0.0354059 | 0.7162839 | 0.8551306 |
| 252:Psi C to D | 0.9009965 | 0.0355275 | 0.8065449 | 0.9520737 |
| 253:Psi C to D | 0.9544627 | 0.0275947 | 0.8579330 | 0.9864404 |
| 254:Psi C to D | 0.9792365 | 0.0186960 | 0.8860808 | 0.9965151 |
| 255:Psi C to D | 0.4653427 | 0.0481625 | 0.3732571 | 0.5598529 |
| 256:Psi C to D | 0.6774430 | 0.0396973 | 0.5953131 | 0.7499072 |
| 257:Psi C to D | 0.8319667 | 0.0440240 | 0.7275851 | 0.9017529 |
| 258:Psi C to D | 0.9193762 | 0.0398296 | 0.7991056 | 0.9703182 |
| 259:Psi C to D | 0.9624880 | 0.0294845 | 0.8381132 | 0.9921974 |
| 260:Psi C to D | 0.5137110 | 0.0649356 | 0.3882627 | 0.6374553 |
| 261:Psi C to D | 0.7134996 | 0.0568004 | 0.5909341 | 0.8110820 |
| 262:Psi C to D | 0.8515378 | 0.0563174 | 0.7054909 | 0.9321278 |
| 263:Psi C to D | 0.9280872 | 0.0470266 | 0.7643426 | 0.9808986 |
| 264:Psi C to D | 0.5366300 | 0.0891315 | 0.3645248 | 0.7004295 |
| 265:Psi C to D | 0.7273174 | 0.0812814 | 0.5443275 | 0.8562309 |
| 266:Psi C to D | 0.8571739 | 0.0744694 | 0.6456081 | 0.9518569 |
| 267:Psi C to D | 0.5341775 | 0.1221329 | 0.3046836 | 0.7500615 |
| 268:Psi C to D | 0.7206890 | 0.1158272 | 0.4551424 | 0.8885166 |
| $\underline{269: P s i ~ C ~ t o ~ D . ~}$ | 0.5063196 | 0.1641481. | 0.2206561 | 0.7879156 |

## Appendix F.

## Adjusted Arrivals and Departures and Mean Lengths of Stay

Adjusted arrival and departure dates for multiple-sight data correct for the assumption $p_{i}{ }^{A}=1$ and $p_{i}^{D}=1$ (Tables 24 and 25).

Table 24. Estimated arrival times if first observed in period $i$.

| Period $i$ | 1999 |  |  |  | 2000 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | B |  | C |  | B |  | C |  |
|  | Estimate | SE | Estimate | SE | Estimate | SE | Estimate | SE |
| 1 | 1.0 | 0.0 | 1.0 | 0.0 | 1.0 | 0.0 | 1.0 | 0.0 |
| 2 | 1.5 | 0.5 | 1.5 | 0.5 | 2.0 | 0.0 | 2.0 | 0.0 |
| 3 | 2.0 | 0.8 | 2.0 | 0.8 | 2.5 | 0.5 | 2.5 | 0.5 |
| 4 | 3.1 | 0.9 | 3.1 | 0.9 | 3.1 | 0.8 | 4.0 | 0.0 |
| 5 | 4.0 | 1.0 | 3.9 | 1.1 | 3.9 | 1.1 | 5.0 | 0.0 |
| 6 | 4.9 | 1.1 | 5.1 | 1.1 | 4.7 | 1.3 | 5.6 | 0.2 |
| 7 | 5.9 | 1.2 | 6.1 | 1.0 | 6.0 | 1.4 | 6.9 | 0.3 |
| 8 | 6.5 | 1.4 | 7.3 | 0.8 | 7.2 | 1.4 | 7.9 | 0.3 |
| 9 | 8.1 | 1.1 | 8.4 | 0.6 | 8.6 | 1.0 | 8.9 | 0.3 |
| 10 | 8.7 | 1.3 | 9.4 | 0.6 | 9.7 | 0.8 | 9.9 | 0.2 |
| 11 | 10.3 | 0.9 | 10.4 | 0.6 | 10.8 | 0.6 | 10.9 | 0.2 |
| 12 | 11.2 | 0.8 | 11.5 | 0.5 | 12.0 | 0.0 | 11.7 | 0.5 |
| 13 | 12.5 | 0.5 | 12.0 | 0.8 | 13.0 | 0.0 | 12.9 | 0.3 |
| 14 | 14.0 | 0.0 | 12.5 | 1.1 | 13.5 | 0.5 | 14.0 | 0.0 |
| 15 | 15.0 | 0.0 | 13.0 | 1.4 | 14.0 | 0.8 | 14.5 | 0.5 |

Table 25. Estimated departure times if last observed in period $i$.

| Period i | 1999 |  |  |  | 2000 |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | B | C |  |  | B |  | C |  |
|  | Estimate | SE | Estimate | SE | Estimate | SE | Estimate | SE |
| 1 | 1.9 | 1.1 | 1.9 | 1.0 | 2.8 | 1.6 | 1.5 | 0.5 |
| 2 | 2.4 | 0.9 | 2.4 | 0.8 | 3.3 | 1.4 | 2.0 | 0.0 |
| 3 | 3.7 | 1.1 | 3.5 | 0.8 | 4.0 | 1.2 | 3.0 | 0.0 |
| 4 | 4.7 | 1.1 | 4.3 | 0.6 | 4.8 | 0.9 | 4.1 | 0.3 |
| 5 | 5.7 | 1.0 | 5.2 | 0.5 | 5.5 | 0.8 | 5.1 | 0.4 |
| 6 | 6.7 | 0.8 | 6.1 | 0.3 | 6.4 | 0.7 | 6.1 | 0.3 |
| 7 | 7.3 | 0.6 | 7.1 | 0.2 | 7.2 | 0.5 | 7.1 | 0.3 |
| 8 | 8.6 | 0.6 | 8.1 | 0.2 | 8.2 | 0.5 | 8.0 | 0.2 |
| 9 | 9.1 | 0.4 | 9.1 | 0.2 | 9.1 | 0.3 | 9.1 | 0.3 |
| 10 | 10.2 | 0.4 | 10.0 | 0.0 | 10.0 | 0.0 | 102 | 0.5 |
| 11 | 11.0 | 0.1 | 11.0 | 0.0 | 11.0 | 0.0 | 11.0 | 0.2 |
| 12 | 12.0 | 0.0 | 12.0 | 0.0 | 12.0 | 0.0 | 12.0 | 0.0 |
| 13 | 13.0 | 0.0 | 13.0 | 0.0 | 13.0 | 0.0 | 13.0 | 0.0 |
| 14 | 14.0 | 0.0 | 14.0 | 0.0 | 14.0 | 0.0 | 14.0 | 0.0 |
| 15 | 15.0 | 0.0 | 15.0 | 0.0 | 15.0 | 0.0 | 15.0 | 0.0 |

Table 26. Mean length of stay (T) and associated standard error (SE) estimated in weeks for each cohort, year and stratum (S).

|  |  | Cohort |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Year | S |  | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 | 13 | 14 |
| 1999 | 1 | T | 3.7 | 3.6 | 4.4 | 4.2 | 3.9 | 3.5 | 3.0 | 2.7 | 2.4 | 2.0 | 1.6 | 1.3 | 1.3 | 1.1 |
|  | B |  | 2 | 6 | 9 | 7 | 3 | 0 | 2 | 1 | 6 | 8 | 9 | 9 | 9 | 9 |
|  |  | S | 2.3 | 2.4 | 2.0 | 1.8 | 1.6 | 1.5 | 1.5 | 1.3 | 1.1 | 0.9 | 0.7 | 0.6 | 0.5 | 0.4 |
|  |  | E | 5 | 6 | 2 | 3 | 8 | 7 | 1 | 6 | 4 | 6 | 8 | 0 | 9 | 2 |
|  |  | T | 4.2 | 4.0 | 3.7 | 3.2 | 2.7 | 2.6 | 2.5 | 2.0 | 1.8 | 1.6 | 1.3 | 1.1 | 1.0 | 1.0 |
|  | C |  | 9 | 6 | 1 | 6 | 3 | 8 | 2 | 6 | 7 | 5 | 7 | 8 | 8 | 3 |
|  |  | S | 1.6 | 1.4 | 1.4 | 1.4 | 1.4 | 1.3 | 1.1 | 1.0 | 0.9 | 0.7 | 0.5 | 0.4 | 0.2 | 0.1 |
|  |  | E | 0 | 8 | 1 | 1 | 7 | 2 | 2 | 6 | 2 | 5 | 8 | 1 | 8 | 7 |
| 2000 |  | T | 9.6 | 7.9 | 6.4 | 5.0 | 3.9 | 3.1 | 2.5 | 2.1 | 1.8 | 1.6 | 1.5 | 1.4 | 1.4 | 1.4 |
|  | B |  | 4 | 5 | 2 | 9 | 9 | 7 | 4 | 0 | 1 | 2 | 1 | 5 | 4 | 4 |
|  |  | S | 1.0 | 1.1 | 1.1 | 1.1 | 1.0 | 0.9 | 0.8 | 0.7 | 0.6 | 0.6 | 0.5 | 0.5 | 0.5 | 0.5 |
|  |  | E | 8 | 2 | 5 | 4 | 9 | 7 | 7 | 7 | 9 | 2 | 8 | 5 | 3 | 0 |
|  |  | T | 6.9 | 5.9 | 4.9 | 4.1 | 3.4 | 2.8 | 2.4 | 2.0 | 1.8 | 1.7 | 1.6 | 1.6 | 1.5 | 1.4 |
|  | C |  | 2 | 1 | 9 | 7 | 6 | 6 | 1 | 8 | 2 | 4 | 5 | 1 | 9 | 9 |
|  |  | S | 1.4 | 1.4 | 1.3 | 1.3 | 1.2 | 1.1 | 1.0 | 0.9 | 0.8 | 0.8 | 0.7 | 0.7 | 0.7 | 0.5 |
|  |  | E | 4 | 1 | 8 | 3 | 5 | 6 | 5 | 5 | 8 | 2 | 7 | 5 | 1 | 0 |

