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MONITORING NUTRIENTS AND PLANKTON IN THE
VICINITY OF THE IONA ISLAND SEWAGE OUTFALL

MS-85-04

By

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INTRODUCTION

Previous to 1963 there were no sewage treatment plants for the Vancouver area (Waldichuk 1983, 1984). As a result of this situation, sewage discharge was mostly in the raw state and via shallow outfalls. By the 1940's, the Vancouver and District Joint Sewerage and Drainage Board developed a new plan for sewage discharge in the Vancouver area. The first phase of this plan (called the Rawn Report; Rawn et al. 1953), called for the construction of a sewage treatment plant on Iona Island. This meant that the raw sewage outfalls into Burrard Inlet and the Fraser River could be eliminated with the hope of improving beaches in English Bay and improving shellfish harvesting in the Fraser River estuary.

Primary treated effluent from the Iona plant is presently discharged through an open excavated channel on Sturgeon Bank tidal flats (see Figure 1). For some time it has shown that the area near the effluent channel has shown a marked deterioration (Figure 1) due to the deposition of organic matter and the accumulation of high metal concentrations in the sediment (McGreer 1982). To correct this situation the greater Vancouver Regional District is considering the installation of a deep submarine outfall of Sturgeon Bank (S & S Consultants 1983). Therefore, the object of this report is to discuss monitoring of nutrients and plankton in the vicinity of the Iona Island sewage outfall, before and after construction of the deep submarine outfall.

1 OTHER MONITORING PROGRAMS

Several deep sea sewage outfalls are already in operation in the Pacific west coast. Four of these sites and the results from their monitoring programs are briefly reviewed here. Some of their recommendations are probably applicable to the proposed monitoring scheme for the deep submarine Iona Island sewage outfall.

1.1 French Creek

French Creek is located on the east coast of Vancouver Island and it is the site of a deep sea domestic sewage outfall. Effluent is discharged via a 2 km outfall and ends in a diffuser 61 m below low water. This new outfall was monitored for three years (Pomeroy 1982). Monitoring included salinity, temperature, dissolved oxygen, nutrients, sediment particle size, organic content of the sediment, heavy metal content of the sediment, and benthic fauna. A submersible was used to examine the diffuser, the sediments around the diffuser and the response of organisms to the presence of the discharge.

No marked changes were observed in the water column but changes in the benthic environment occurred. Organic content of the sediments increased and there were increases in metal concentrations (e.g. copper, manganese, iron, and nickel).

1.2 Macaulay Point

Macaulay Point is located near Victoria, B.C. Monitoring was conducted to assess the influence of discharge through a deep sea outfall to 60 m (Ellis et al. 1971). Monitoring included, temperature, salinity, water stability, dissolved oxygen, turbidity, nutrients, chlorophyll a and coliform bacteria.

Monitoring recommendations were that ammonium and nitrate concentrations are good indicators of sewage. Lower salinities, decreased dissolved oxygen, increased turbidity and coliform bacterial counts can be

used to detect the presence of the discharge. Ellis et al. recommended that total nitrogen and phosphorus budgets be included. This would include all forms of dissolved inorganic and organic nitrogen, particulate nitrogen, and the nitrogen content of planktonic animals, and nitrogen measurements of the sediments and benthic organisms. Sampling should be less frequent during the winter because biological activity is low and mixing by high winds is extreme. It is important to record the phase of the tidal cycle and the direction and speed of tidal currents during sampling.

1.3 Los Angeles

Two large outfalls, Hyperion (Los Angeles city) and Whites Point (serving Los Angeles county) are deep sea outfalls. Monitoring studies by Thomas and Carsola (1980) indicate that the subsurface discharge (about 40 m) could be traced by its high ammonium concentration. The effluent leaving the diffuser contains 2200 μM of ammonium. Maximum ammonium concentrations of 155 μM at 27 m and 2 km from the diffuser were recorded.

Elevated surface concentrations (3 to 10 μM) were found over the outfalls. In stratified water, high values were found below 15 m, but in well-mixed water, high concentrations were detected at the surface. Subsurface high concentrations were associated with turbid layers, coliform bacteria and reduced oxygen levels. Since the distribution of ammonium correlated well with measured subsurface currents, Thomas and Carsola (1980) have suggested that ammonium may be a useful tracer of the discharge of sewage in seawater.

Since elevated ammonium concentrations occur in the surface water, phytoplankton uptake of nitrogen is less than the discharge rate. This reduction in uptake of ammonium appears to be due to inhibitory levels of heavy metals or organics (MacIsaac et al. 1979). Ammonium must exceed 100 μM before it will exert toxic effects on phytoplankton growth (Thomas et al. 1980). Thomas et al. (1974) have also shown that when nitrogen is present in excess, other nutrients such as phosphate or silicate may become limiting.

1.4 Kaneohe Bay, Hawaii

Primary treated and later, secondary treated domestic sewage was discharged into this shallow bay for several decades. The main effect of this discharge was a severe decline in the coral reef ecosystem and substantial phytoplankton eutrophication (Caperon et al. 1971; Laws and Redalje 1978, 1982).

A study of the phytoplankton community and water column chemistry in Kaneohe Bay, before and after the diversion of secondary treated sewage from the bay has shown that changes in total nutrient concentrations in the water column cannot be accurately predicted without taking into account water column-benthos interactions. During the first year after sewage diversion, the decomposition of about 400 tons of benthic organisms which were primarily filter feeders, resulted in water column dissolved organic nitrogen and phosphorus concentrations roughly an order of magnitude higher than those expected in the absence of such interactions. In a shallow water community, close coupling between the water column and benthos cannot be ignored in predicting and assessing the impact of ecological manipulations.

2 IONA ISLAND SEWAGE TREATMENT PLANT

A more comprehensive description of the sewage treatment plant can be found in the Rawn Report (Rawn et al. 1953) and more recently in the report by S & S Consultants (1983). The principle aspects are summarized here.

The plant was built on Iona Island in 1963 and has undergone three expansions. It receives both sanitary sewage and storm runoff from the city of Vancouver, and portions of Richmond and Burnaby. The flow varies from 3.8 to 17.7 m³ s⁻¹. The plant provides primary treatment (i.e., screening, grit removal, scum and grease removal and sedimentation) and seasonal (May 1 to September 30) chlorination to reduce bacterial coliform counts during the summer.

The primary treated effluent is discharged into a 7.3 km long and 40 m wide outfall channel that runs across Sturgeon Bank. At low tide the effluent is confined to the channel but at high tide the channel is submerged and the effluent is diluted and flows over and onto the adjacent mudflats. The sand-filled rock jetty extends along the north side of the outfall channel for about 5.1 km to prevent the effluent from flowing around Point Grey and into nearby English Bay for at least two tidal cycles. This delay allows time for bacterial dilution and inactivation due to seawater exposure.

3 STURGEON BANK STUDY SITE AND SURROUNDING AREA

3.1 Physical Characteristics

The tides in the Strait of Georgia are the mixed semi-diurnal type with two high and two low tides each day. The tide heights are usually asymmetrical with highs ranging from 4 to 5 m.

Tidally induced surface currents result in flow north or south from the discharge pipe. It is common to have a large counter clockwise gyre in the middle of the strait and a smaller clockwise gyre immediately west of the Iona jetty, especially during times of high discharge from the Fraser River. Drogue studies have been conducted (Tabata et al. 1971; Crean and Ages 1972) and they demonstrate that frequently there is a northward flow from Iona Island into Burrard Inlet, especially when the discharge from the Fraser River is high (June-July) and the drogues are released at low tide (Figure 2). Current speeds are approximately 10 to 20 cm.s⁻¹.

Winter winds are predominantly from the east or south-east and the summer winds from the northwest. The Iona Island area is somewhat protected from the southeast winds by land. It is the northwest wind which causes the build-up of large waves in the area because they are the strongest and have the largest fetch.

The Iona study area is influenced by Fraser River runoff and consequently the brackish water upper layer (10 m) varies from 16% in the winter (low river discharge) to 5% in the summer (river discharge is 10 times winter discharge). Heating of surface waters in the summer, in combination with decreased salinities results in a very strong pycnocline at 10 m and a weaker one of about the same depth in winter.

Seasonal changes in irradiance and light penetration into the water (Secchi depth) are shown in Figure 3.

3.2 Chemical Characteristics

The chemical characteristics of the study area include measurements of dissolved oxygen, nitrogen and phosphorus. Dissolved oxygen within

the first 300 m of the outfall discharge are $\sim 2 \text{ mg}\cdot\text{L}^{-1}$ and these values are low enough to kill some species of fish and stress other aquatic organisms (Birtwell et al. 1983). Beyond about 1500 m, the oxygen concentration values return to saturation levels of about $9 \text{ mg}\cdot\text{L}^{-1}$.

Sewage contains high concentrations of nitrogen, phosphorus, urea, ammonium, nitrate and other dissolved organic nitrogen constituents. Total phosphorus levels in the effluent channel close to the outfall are approximately $0.5 \text{ mg}\cdot\text{L}^{-1}$ ($16 \mu\text{M}$) and ammonium is about $9 \text{ mg}\cdot\text{L}^{-1}$ ($650 \mu\text{M}$). These nutrients decrease substantially (by about two orders of magnitude) about 2 km west beyond the jetty (Table 1). Increases in phytoplankton have been reported near the jetty (Harrison 1981). If the increased phytoplankton biomass settles out onto the sediment (rather than being grazed or transported away), the decomposition of this material would aggravate the low O_2 problem. Typical winter and summer values for O_2 and nutrients for a normal area well beyond the jetty are given in Table 1 and shown in Figure 3. There are no data sets which include nitrate, ammonium, urea, and dissolved organic nitrogen and phosphorus concentrations along a transect away from the jetty and over different seasons.

3.3 Biological Characteristics

3.3.1 Phytoplankton. There appear to be no data on phytoplankton biomass or species composition near the outfall. This is important information that should be obtained in the future. The phytoplankton well beyond the outfall (control station) in the Strait of Georgia is typical of cold, temperate coastal waters with some estuarine influence. Diatoms are the most abundant group (Legare, 1957; Stockner et al. 1979, Shim 1977a; 1977b; Harrison et al. 1983). The most common species (frequency and biomass) is Skeletonema costatum, a chain-forming centric diatom. Other diatoms that are also abundant are Thalassiosira spp., Chaetoceros spp. (especially Chaetoceros debilis), Leptocylindrus danicus, Nitzschia spp. and a few very large diatoms such as Coscinodiscus spp., Cerataulina and

Eucampia. There is little indication that large salinity changes caused by the Fraser River discharge influences the composition of the diatom community (Shim 1977b). Buchanan (1966) studied flagellates in Indian Arm and found that cryptomonads, especially Rhodomonas minuta were major primary producers. Some species such as the chrysomonad Heterosigma okashiwa (Olisthodiscus luteus) and the dinoflagellate Protogonyaulax tamarensis both prefer estuarine waters of less than 15 or 20‰ and form blooms primarily on the east side of the strait on either side of the Fraser River plume. Both of these species may form dense blooms which may discolor the water and hence, they are often referred to as "red-tide" species. They are both very common in English Bay in the summer. However, blooms of these species have been observed in areas where there is no influence of sewage and therefore their presence in English Bay cannot be attributed to sewage effects (Harrison et al. 1983).

The standing stock of phytoplankton is generally assessed in terms of chlorophyll a. These values for the strait range from $1 \text{ mg} \cdot \text{m}^{-3}$ in the winter to $15 \text{ mg} \cdot \text{m}^{-3}$ during spring or early summer blooms (Stockner et al. 1979; Parsons et al. 1981; Harrison et al. 1983; see Figure 3). There are no good estimates of phytoplankton standing stocks along a transect away from the jetty.

Primary productivity for the strait ranges from a maximum of $1.2 \text{ g C} \cdot \text{m}^{-2} \cdot \text{d}^{-1}$ in May to low summer values of $0.2 \text{ g C} \cdot \text{m}^{-2} \cdot \text{d}^{-1}$ (Figure 3) (Parsons et al. 1970; Parsons 1979). However recent estimates have been higher, up to $8 \text{ g C} \cdot \text{m}^{-2} \cdot \text{d}^{-1}$ (Stockner et al. 1979). The reasons for the differences between these estimates has been discussed by Harrison et al. (1983). Phytoplankton standing stock and primary productivity are depressed near the mouth of the Fraser River due to the turbidity of the river water and the increased extinction coefficient (Figure 4). The increased nitrogen input into the Fraser River is thought to have only a minor contribution to increased primary productivity. Estimates are that the diversion of sewage from Vancouver Harbour to the Fraser River has only increased the nutrient loading in the river from 5% in 1967 to 12% in 1977

(Clark and Drinnan 1980). In addition, the contribution of sewage to the total nitrogen concentration of the river is, to a minor extent, controlled by the discharge volume of the river for that particular year. Very high primary productivity has been observed in Burrard Inlet, especially at Port Moody Arm ($532 \text{ g C} \cdot \text{m}^{-2} \cdot \text{yr}^{-1}$) (Stockner and Cliff 1979).

3.3.2 Zooplankton. The zooplankton community structure and changes in biomass for the central strait are shown in Figure 5. The winter community is composed chiefly of small copepods such as Pseudocalanus spp. Other less dominate species are Paracalanus parvus, Oithona helgolandica and Corycaeus sp. In the spring there is a massive recruitment of nauplii and copepodite stages of Neocalanus plumchrus from deep water to the surface and some migration of adult Calanus marshallae, C. pacificus and Metridia pacifica. Components of the estuarine plankton community include the genera Acartia, Pseudocalanus, Centropages and Epilabidocera. Decapod zoea and juveniles and barnacle nauplii form the bulk of the meroplankton. During extended periods of calm weather the larvacean Oikopleura may produce extensive blooms (Sibert et al. 1976). Zooplankton biomass and community structure has not been well studied in the vicinity of the outfall.

3.3.3 Bacteria. The distribution of fecal coliform bacteria in the outfall area has been frequently studied (Churchland 1979). Fecal coliform concentrations in the effluent are extremely high, particularly during the non-chlorination period from October to May (23×10^6 fecal coliforms/100 ml). Typical counts close to the outfall are 16×10^5 fecal coliforms/100 ml, with a decrease to 8×10^3 /100 ml at 2 km from the outfall. Control samples from the north side of the jetty had counts of 2×10^3 /100 ml, which is well above the standard of 200/100 ml used as a limit for recreational waters.

The distribution of bacteria in the Strait of Georgia is not well known. Recent investigations by Albright and co-workers have examined the

influence of the Fraser River on bacterial distribution in the plume area (Valdes and Albright 1981; Albright 1983; Bell and Albright 1981). The freshwater bacteria are predominantly rod-shaped and larger in volume ($0.3 \mu\text{m}^3$) than their marine counterparts ($0.07 \mu\text{m}^3$) and are generally attached to particles, in contrast with the predominance of free-floating marine forms. Bacterial numbers in surface sediments were much greater (by several orders of magnitude) than in overlying waters. The Fraser River contains about 1.2×10^7 total bacteria per ml of river water, which is twice as high as the strait (Bell and Albright 1981; Albright 1983). Most of the Fraser River bacteria are killed when the salinity rises above 16 ‰, but a few appear to remain viable but not very active.

3.3.4 Benthic Algae. The benthic algae in the vicinity of the outfall have not been well studied. Reports of Ulva lactuca and Enteromorpha sp. abundance suggests that it has responded to increased nutrient concentrations (Hoos and Packman 1974). Benthic microalgae (primarily diatoms) on top of the sediment or in the first one centimeter suggests that benthic primary productivity may be considerable based on estimates of chlorophyll a concentrations of 0.2 to $0.3 \text{ g} \cdot \text{m}^{-2}$ from July to October for the area north of the jetty (Harrison 1981).

3.3.5 Benthic Invertebrates and Fish. These two components of the food web have been studied the most extensively in the outfall area. Prawns, shrimp, and crabs of commercial importance inhabit Sturgeon Bank (Hoos and Packman 1974; Levins and Constalin 1975; Levings et al. 1983). Other abundant sediment-dwelling invertebrates include polychaetes (e.g. Lumbrineridae) and nematodes (Harrison 1981; Levings et al. 1983). Harpacticoid copepods are an important food source for bottom feeding fish such as flatfish and juvenile salmon (Harrison 1981; Levings et al. 1983). The harpacticoids in the area north of the jetty has been well studied (Harrison 1981). Mud communities were distinct from sand communities and both of these communities showed strong seasonal changes. Therefore, any

comparisons of the outfall area with a control area must be careful to consider sediment type (spatial variability) and seasonal changes (temporal variability). The extent of this temporal and seasonal variability is shown for the area north of the jetty for 1978 (Figure 6). The clam, Macoma balthica has been extensively studied in the area receiving sewage effluent (McGreer 1979, 1982). He found that the distribution was seriously affected by the sewage effluent, probably due to a combination of factors such as, increased heavy metal content of the sediment, and lower O₂ concentrations in the sediment. Sublethal effects such as burrowing behaviour was also affected by the heavy metal contaminated sediments (McGreer 1979). Twenty-seven species of fish have been observed in the sewage outfall area and tumors in some fish have been reported (Birtwell et al. 1983).

3.3.6 Pelagic Food Webs. The pelagic food web in the outfall area is relatively unknown. The food webs for the main part of the Strait of Georgia changes with the seasons and these changes are summarized in Figure 7. The main food web is the diatoms/dinoflagellates — macrozooplankton — small fish, chaetognaths, ctenophores or hydromedusae.

4 MONITORING PLAN FOR PLANKTON

4.1 Physical and Chemical Indices of Water Quality

Turbidity and water colour may be a useful indication of how much the sewage effluent is getting diluted by the surrounding water and the direction the sewage plume is moving. One quick way to measure turbidity and light extinction is with a secchi disc. The measurement is the depth at which a standard 30 cm white disc just disappears from view. When this measurement is used in conjunction with chlorophyll a measurements, it is possible to separate the extinction of light due to the presence of phytoplankton from that due to particles such as silt or debris. Turbidity and light extinction can also be measured with a submersible light meter or with a transmissometer which may even be towed under water.

Temperature and salinity are not indicators of pollution, but they should be measured because they are useful in assessing water movement, stability, and water density.

Large decreases in dissolved oxygen concentrations are generally a result of biological processes such as respiration. Decomposition of high concentrations of dissolved or particulate organic carbon by bacteria consumes oxygen and may result in very low oxygen concentrations ($2 \text{ mg} \cdot \text{L}^{-1}$) either in the water just above the sediments or in the sediments themselves. This utilization of oxygen by these processes is generally referred to as biological oxygen demand (BOD). Low oxygen or anoxic conditions can greatly reduce benthic invertebrate or fish biomass in the area and alter the species diversity. Samples can be taken for oxygen and measured by Winkler titration or an oxygen electrode (Parsons et al. 1984). Oxygen value in the range of $6\text{--}9 \text{ mg} \cdot \text{L}^{-1}$ are normal, while values $2\text{--}5 \text{ mg} \cdot \text{L}^{-1}$ indicate oxygen depletion stress of the environment depending on the species and other physical and chemical characteristics of the water.

Water samples can be collected for nutrient analyses. Nitrate, nitrite, ammonium, urea, phosphate, dissolved organic phosphorus, and nitrogen should be measured. These methods of collection and analysis are

described in Parsons et al. 1984. The normal range of concentrations found during winter and summer are given in Table 1. Increases in dissolved components above normal concentrations indicates potential eutrophication, provided light, temperature, and water stability are appropriate for phytoplankton growth. Of all the nutrients listed above, ammonium is the best one to monitor because normal concentrations are low and concentrations in sewage effluent are very high. In addition, since marine waters are generally nitrogen limited, nitrogen compounds are important to monitor.

4.2 Biological Indices of Water Quality

4.2.1 Phytoplankton. The effects of sewage on indigenous aquatic organisms can be assessed at four different levels. These are: 1) abundance (numbers and biomass), 2) taxonomic composition (species diversity and composition), 3) growth (productivity, respiration, and metabolic rates) and, 4) bioaccumulation of toxic substances (Figure 8; Worf 1980). A list of parameters used to assess these four levels is given in Table 2.

The biomass of phytoplankton in the outfall area has not been carefully studied. Biomass estimates in terms of chlorophyll a or cell numbers are known for the area well beyond the outfall; they range from 1 to 10 mg chl a·m⁻³ and from 10⁴ to 5 x 10⁶ cells·L⁻¹, respectively (Figure 3). Chlorophyll a measurements are the easiest way to monitor phytoplankton biomass and this can be done continuously by using in vivo fluorescence from a small boat using a battery operated fluorometer, strip chart recorder and pump. Using a grid cruise track, horizontal variations in biomass can be measured and the measurements can be used to construct a surface contour map. Similarly, vertical profiles of biomass can be obtained to determine if biomass is patchy with respect to depth. Determining biomass in terms of cell numbers is much more time consuming since this is usually done by counting a sample with a microscope, although estimates of particles (a chain of diatoms is counted as one particle) can be achieved by using a particle counter.

Taxonomic composition of phytoplankton in the sewage outfall area has not been specifically studied. Dominant species in the area well beyond the outfall has been discussed in Section 3.3.1. The only way to determine species composition is by using a microscope. Samples can be taken and preserved and counted with an inverted microscope some time later. Even though microscope counts are very time consuming they can be valuable since species composition does change with sewage stress and generally species diversity decreases with only a few tolerant species remaining under extreme conditions.

Metabolic rates are usually measured in terms of growth or photosynthetic rate. The uptake of ^{14}C -labelled bicarbonate is used to access short term photosynthetic rates (Strickland and Parsons 1972). There are no known measurements made in the outfall area. Values exceeding $5 \text{ mg C m}^{-3} \text{ d}^{-1}$ would indicate eutrophic conditions (Laws 1981).

Phytoplankton may accumulate high levels of metals if concentrations in the surrounding waters are high. Monitoring intracellular levels of copper and mercury, for example, would give some indication of the degree of metal stress in the area.

4.2.2 Zooplankton. The same four levels of assessment can be applied to zooplankton as well as phytoplankton. Zooplankton biomass must be assessed by sampling with nets. Two sizes of mesh are used, $> 200 \mu\text{m}$ to $> 330 \mu\text{m}$ for the macrozooplankton and > 60 to $200 \mu\text{m}$ for the microzooplankton. Vertical hauls are best for assessing the total biomass in the water column. Harpacticoid copepods and amphipods which lie in the water just above the sediment are more difficult to sample quantitatively.

Assessing taxonomic composition is very time consuming since a microscope must be used. Growth of zooplankton is difficult to assess and the easiest measurement of metabolic rate is respiration. This can be done using an O_2 electrode and rates would be expected to increase with pollution stress.

4.2.3 Bacteria. In addition to the standard coliform bacterial counts total bacterial numbers should be assessed using epifluorescence microscopy (Daley and Hobbie 1975). Total counts would aid in assessing the extent of the bacterial activity associated with organic matter breakdown and its relation to the dissolved oxygen concentration in the water.

Bacterial productivity can be assessed by the rate of uptake of a radioactively labelled organic compound such as glucose or by the incorporation of ^3H -labelled thymidine.

4.2.4 Benthic Organisms. Benthic organisms generally show a clearer response to a point source pollutant than planktonic organisms. Since benthic organisms are relatively fixed, spatially, sampling them at different distances from the outfall frequently reveals the gradient effect of the pollutant. Benthic diatoms and benthic macrophytes (seaweeds) should be assessed in terms of biomass, chemical composition and species composition along a transect away from the outfall.

Considerable attention has been given to benthic invertebrates in terms of species composition (Levings and Constalin 1975; Levings et al. 1983; S & S Consultants 1983) and in terms of bioaccumulation of metals (McGreer 1979, 1982). Bottom feeding fish have also been investigated and they are also good indicators of sewage stress (Birtwell et al. 1983; S & S Consultants 1983).

Other benthic characteristics that are useful to measure and that aid in the interpretation of the benthic organisms responses, are particle size, organic content and heavy metal concentrations in the sediment. The use of a submersible (e.g. Pisces IV) is recommended for direct observations on the benthic environment that might otherwise be missed or confused because of the selected sampling schemes (Pomeroy 1982).

4.2.5 Other Indices. In addition to the long term field monitoring approach described above, short term testing (bioassays) can be conducted. Different types of bioassays are summarized in Table 3.

There are two main types of bioassays. First, is the in situ tests which are conducted at the field test site with natural assemblages of organisms. In order to carry a time series of sampling on the same assemblage and avoid the movement of plankton organisms by advection, in situ enclosures are used (Schelske 1984; Shubert 1984). The second type of bioassay is conducted in the laboratory to assess plant effluents or water from the field test site. Usually a "calibrated" bioassay organism is used, rather than a natural assemblage and its response to different concentrations is determined. The parameters monitored during these bioassay tests may include bioaccumulation of metals or organics, acute toxicity (LC₅₀) or low level responses such as changes in behaviour or growth rates (Table 3).

4.3 Summary

A biological and water quality parameter list, sampling frequencies and priority for proposed biomonitoring programs in the USA are given in Table 4 (Worf 1980). This list can be simplified when it is applied to the Iona Island sewage outfall and limited to assessing water quality and plankton (Table 5). Counts of coliform bacteria must be made to assess water quality. Two key parameters for monitoring potential phytoplankton eutrophication in the lowest possible cost monitoring program are ammonium concentrations in the water and chlorophyll a concentration (as estimate of phytoplankton biomass). Because these parameters are good indicators of eutrophication they should be monitored bi-monthly from April to October (Table 1). Parameters of secondary importance are dissolved oxygen, other nutrients (especially the inorganic and organic nitrogen components), phytoplankton and zooplankton species composition and zooplankton biomass. All these parameters should be monitored along a transect away from the outfall and also in English Bay since there is considerable northward flow. Utilization of the high nutrient concentrations by phytoplankton may not take place until the water stratifies in more quiescent areas such as English Bay.

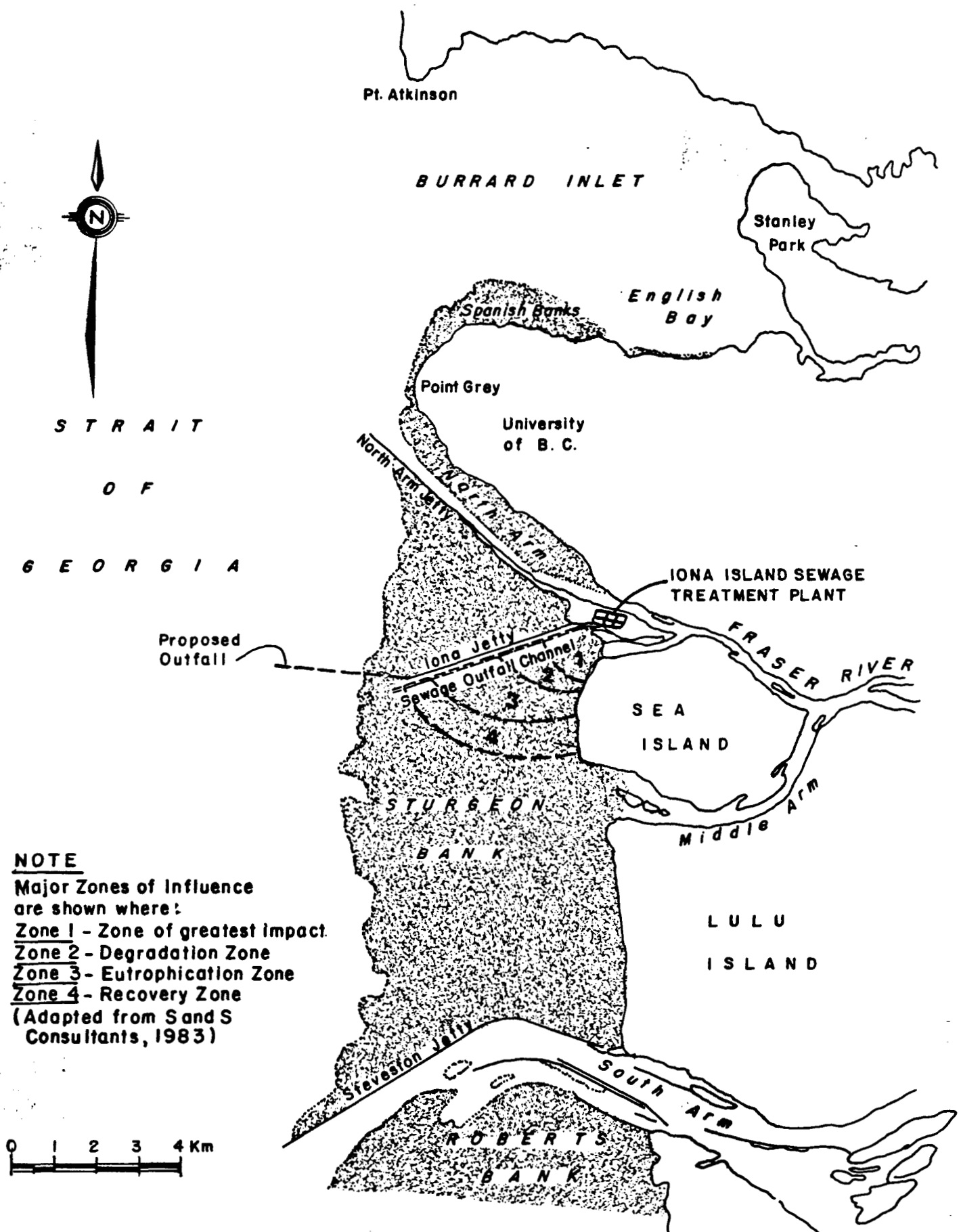


FIGURE 1 IONA ISLAND SEWAGE OUTFALL AND THE SURROUNDING AREA

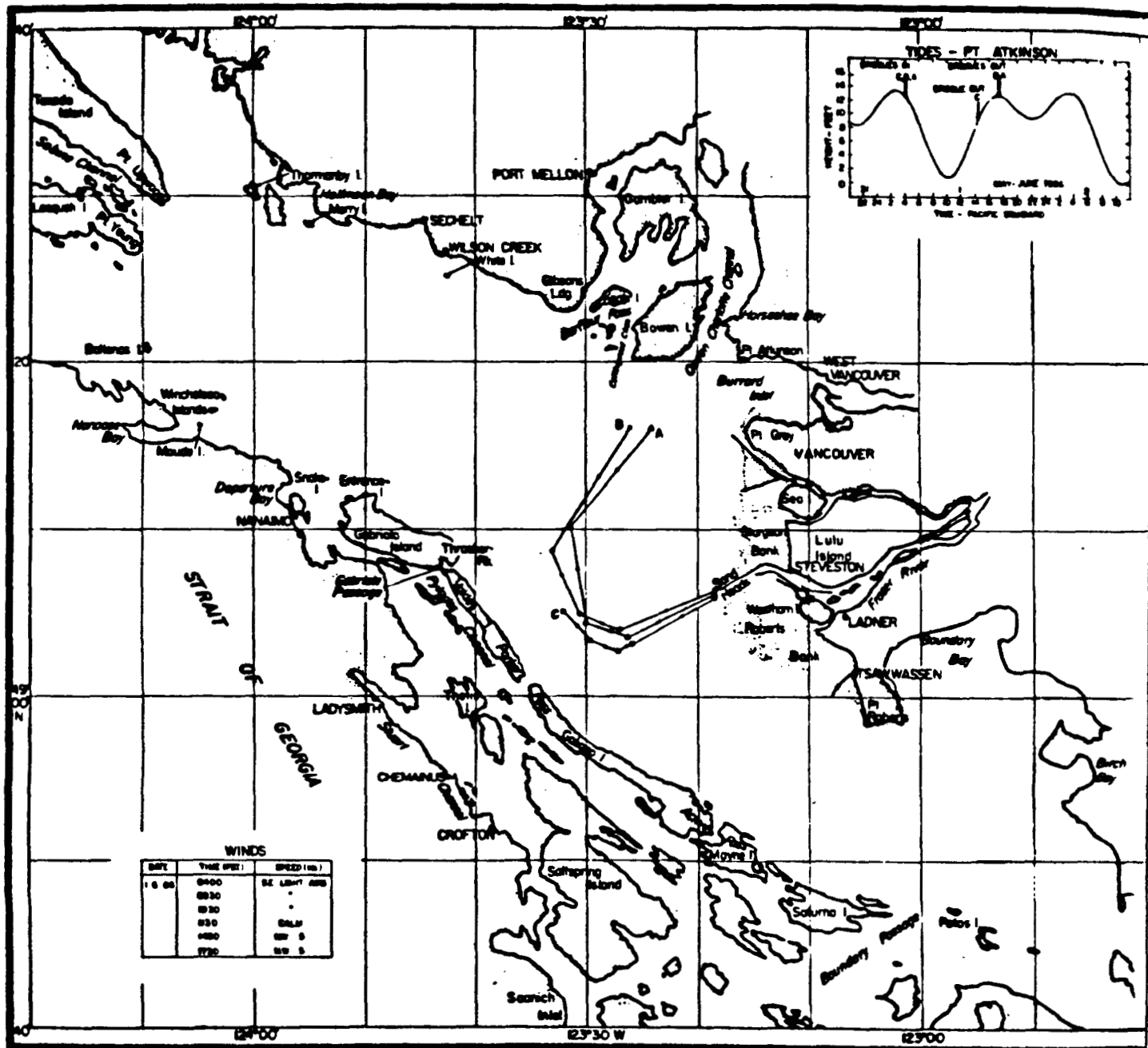


Fig. 2 Movement of surface drogues in the vicinity of Iona jetty during a tidal cycle (Crean and Ages 1972).

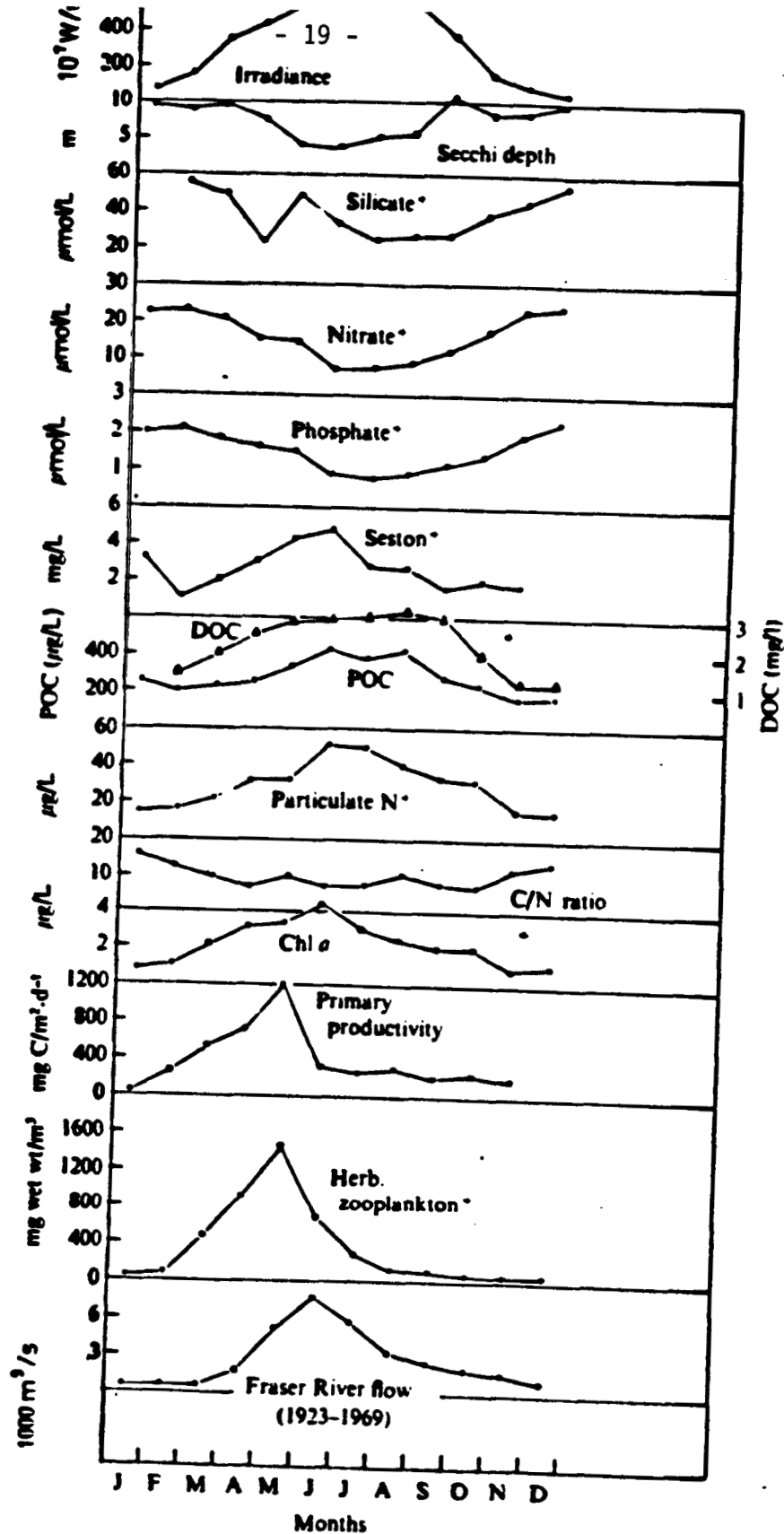


Fig. 3

Seasonal cycle in physical and chemical parameters, standing stock, and productivity expressed as monthly averages for the period 1965-68 for the Strait of Georgia (+ denotes average value for 0-20 m in the water column). DOC and POC represent dissolved and particulate organic carbon, respectively (from Parsons 1979). Recent studies (Stockner et al. 1979) indicate that the values for primary productivity could be twice as high as represented in this figure.

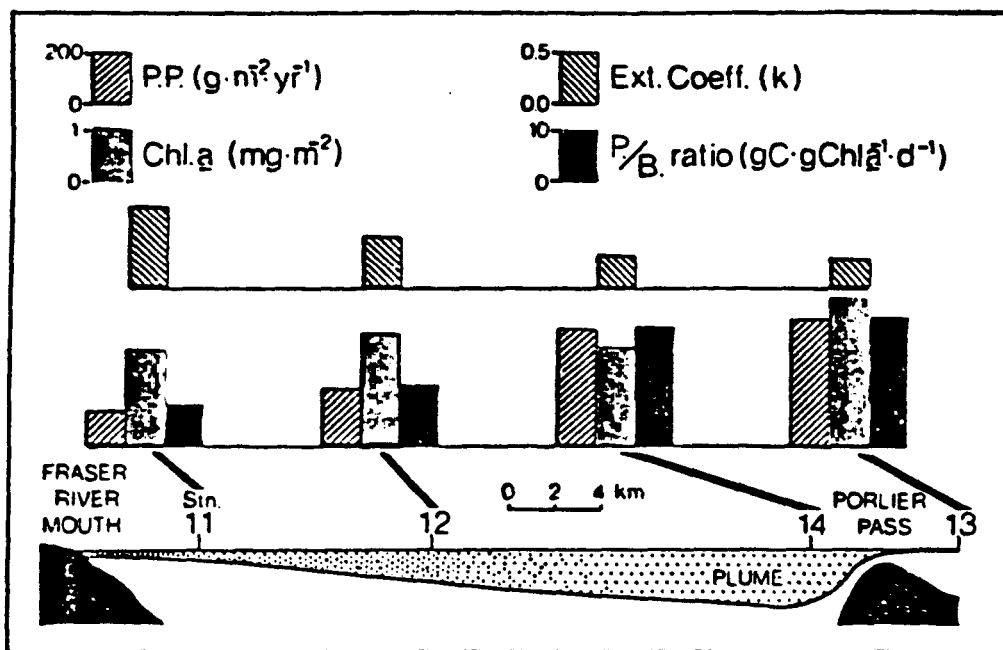


Fig.4 Schematic illustration of the impact of the Fraser River discharge on light attenuation and primary production. Histograms are mean 1976 values for stations on the east-west transect from the river mouth to inside Porlier Pass (from Stockner et al.1979).

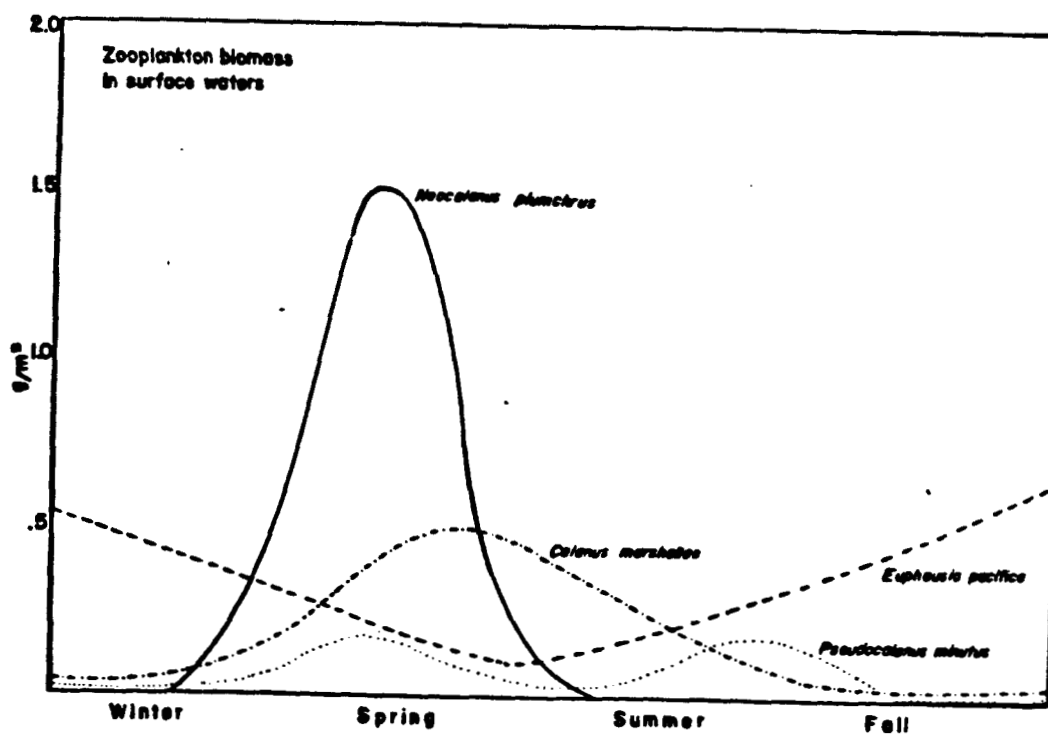
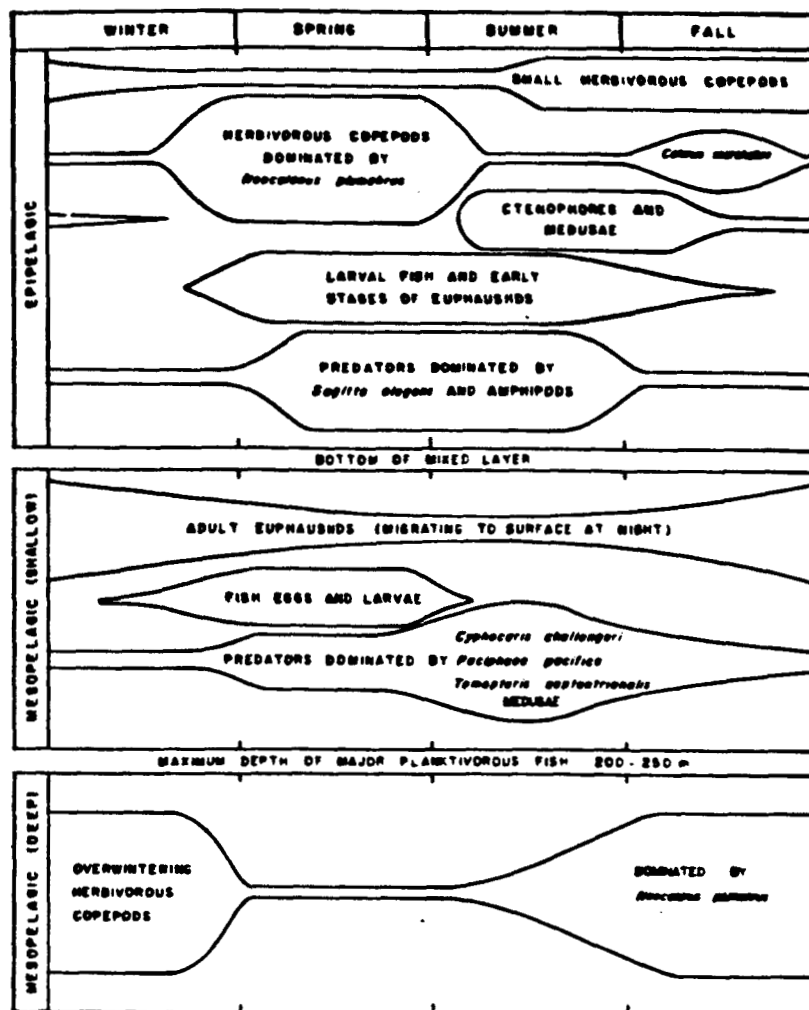


Fig.5

Generalized schematic representation of the zooplankton community structure for the strait. (A) Changes in relative abundance with seasons and depth. (B) Seasonal changes in biomass (g wet wt · m⁻³) of dominant species inhabiting the upper 20-50 m of the water column (from Harrison *et al* 1981)

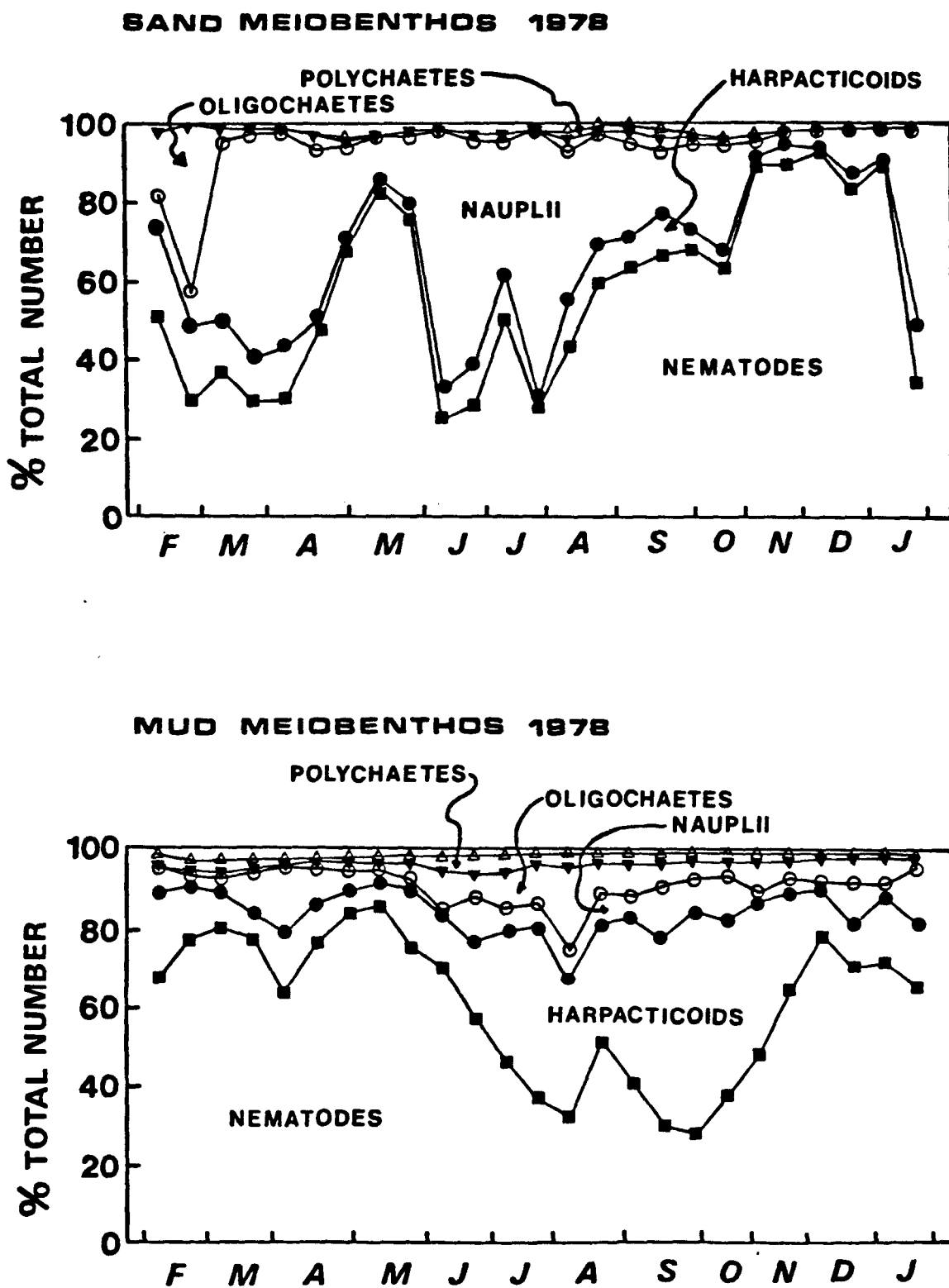


Figure 6 Seasonal variation in percentage contribution of most abundant taxa to the total number of meiobenthic organisms collected from each site through 1978 (from Harrison 1981).

GENERALIZED DEEP WATER FOOD WEBS

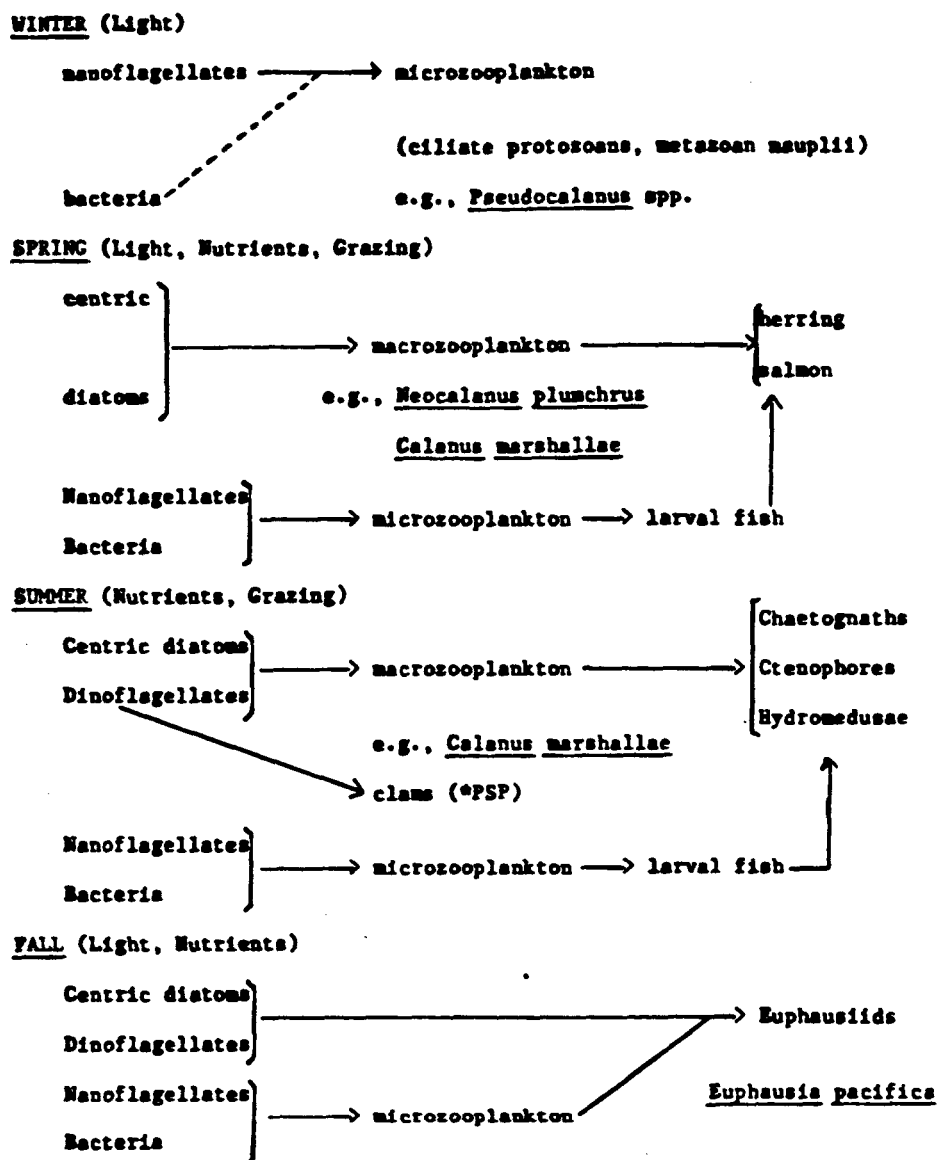


Fig. 7 Generalized schematic representation of food webs during different seasons in the Strait of Georgia. Important factors influencing primary productivity are shown in parentheses for each season. Broken lines indicate that the degree of interaction between the two trophic levels is unknown (*PSP = paralytic shellfish poisoning) (from Harrison et al. 1983)

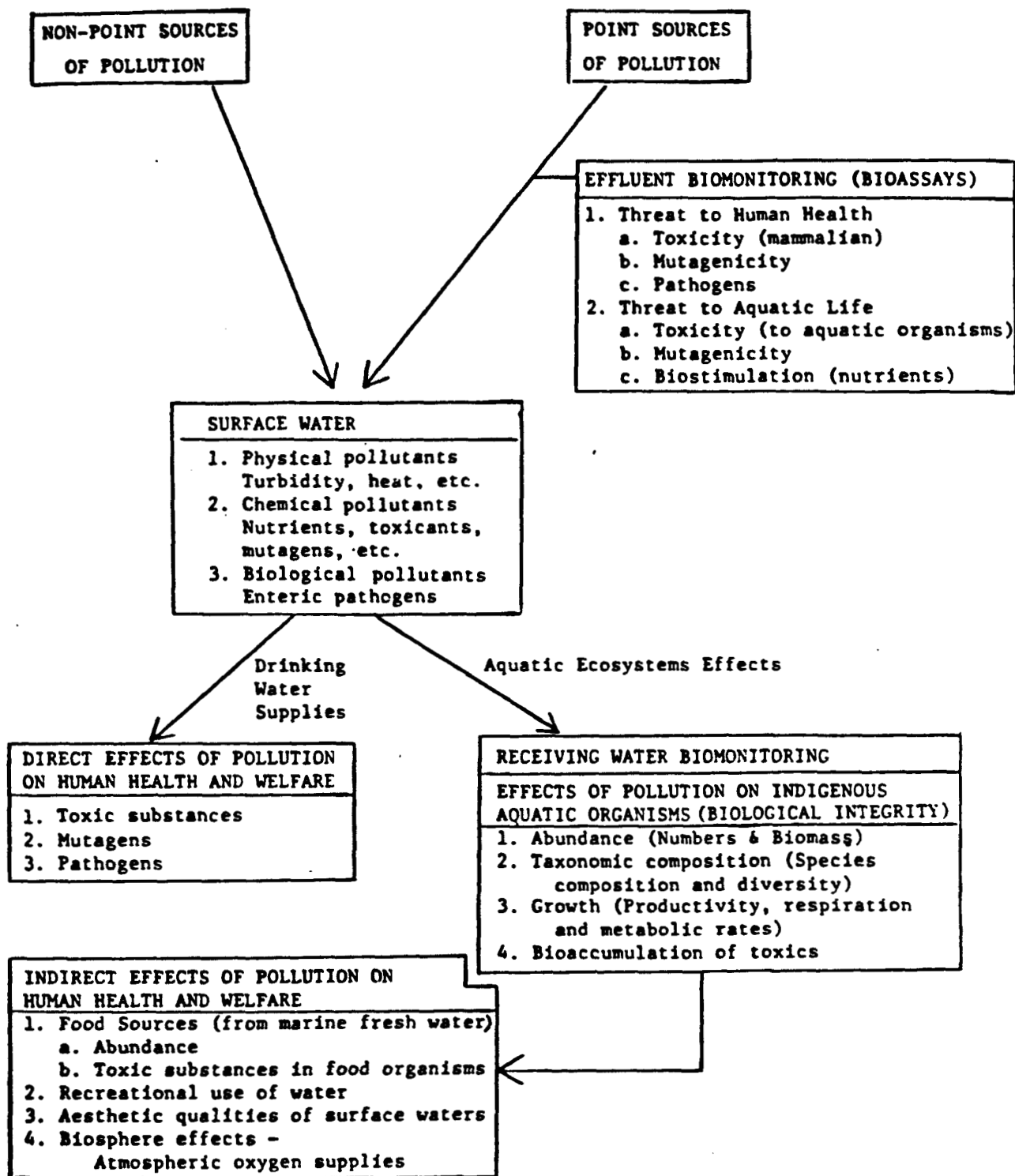


Figure 8 Biological Effects of Pollution on Human Health and Aquatic Life
(After Worf 1980).

TABLE 1 Concentrations of dissolved oxygen and nutrients for central Strait of Georgia, near the outfall and at the mouth of the Fraser River. (DON = dissolved organic nitrogen)

SITE	Nutrient conc'n (mg·l ⁻¹)					
	NC ₃	NH ₄	Urea	DON	PO ₄	O ₂
Central Strait- Winter	0.4	0.03	0.04	0.05	0.06	9
Central Strait-Summer	0	0.01	0.01	0.01	0.01	9
Outfall	-	9.0	-	-	0.5	1.5
Fraser River-Winter	0.2	0.06	-	-	0.009	9
Fraser River-Summer	0.02	0.01	-	-	0.003	9

Table 2

Properties of Indigenous Communities of Aquatic Organisms Used in Determining the Biological Integrity of Surface Waters (from Worf 1980).

Parameters	Community				
	Phyto- plankton	Zoo- plankton	Peri- phyton	Macro- phyton	Macro- invertebrates
Standing Crop					
1. Count	X	X	X	X	X
2. Volume	X	X	X	X	X
3. Wet weight	X	X	X	X	X
4. Dry weight	X	X	X	X	X
5. Ashfree weight	X	X	X	X	X
6. DNA content	X	X	X	X	X
7. ATP content	X	X	X	X	X
8. Chlorophyll <i>a</i> content	X	X	X	X	X
Taxonomic Composition					
1. Species Identification	X	X	X	X	X
Indicator species	X	X	X	X	X
Number of individual species	X	X	X	X	X
Total number of species	X	X	X	X	X
Diversity index	X	X	X	X	X
2. Pigment composition	X	X	X	X	X
Biomass/Chlorophyll <i>a</i>	X	X	X	X	X
Chlorophyll <i>a</i> /Chlorophyll <i>b</i>	X	X	X	X	X
Chlorophyll <i>a</i> /Chlorophyll <i>c</i>	X	X	X	X	X
Phaeophytin content	X	X	X	X	X
3. Nitrogen (N_2) fixation	X	X	X	X	X
Metabolic Activity or Condition					
1. Primary productivity	X	X	X	X	X
Carbon-14 uptake	X	X	X	X	X
Oxygen evolution	X	X	X	X	X
2. Respiration rate	X	X	X	X	X
Plankton dark-bottle O_2 uptake	X	X	X	X	X
Electron transport	X	X	X	X	X
Benthic respirometer O_2 uptake	X	X	X	X	X
3. Nitrogen (N_2) fixation	X	X	X	X	X
4. Chemical composition	X	X	X	X	X
Macronutrient content	X	X	X	X	X
Enzyme content	X	X	X	X	X
Acetyl choline esterase	X	X	X	X	X
Phosphatase	X	X	X	X	X
Nitrate reductase	X	X	X	X	X
Toxic organics and metals content	X	X	X	X	X
5. <i>Flesh tainting</i>	X	X	X	X	X
6. <i>Histopathology</i>	X	X	X	X	X
7. <i>Condition factor</i>	X	X	X	X	X

Table 3
Use of Captive Organisms in Biomonitoring and Toxicity Tests (from Worf 1980).

Type of Test	Organism				
	Phyto- plankton	Zoo- plankton	Peri- phyton	Macro- phyton	Macro- invertebrates
In Situ Tests					
1. Bioaccumulation	X	X	X	X	X
Toxic metals	X	X	X	X	X
Pesticides (organics)	X	X	X	X	X
Flesh tainting	X	X	X	X	X
2. Toxicity tests	X	X	X	X	X
Acute toxicity	X	X	X	X	X
Histopathology	X	X	X	X	X
Histochemistry	X	X	X	X	X
Choline esterase	X	X	X	X	X
In-plant Tests (Effluents)					
1. Bioaccumulation	X	X	X	X	X
Toxic metals	X	X	X	X	X
Pesticides (organics)	X	X	X	X	X
Flesh tainting	X	X	X	X	X
2. Toxicity tests	X	X	X	X	X
Acute toxicity	X	X	X	X	X
Low-level responses (behavioral responses)	X	X	X	X	X
Histopathology	X	X	X	X	X
3. Biostimulatory tests	X	X	X	X	X
Algal growth response (AGP)	X	X	X	X	X
Laboratory Tests					
1. Bioaccumulation	X	X	X	X	X
Toxic metals	X	X	X	X	X
Pesticides (organics)	X	X	X	X	X
Flesh tainting	X	X	X	X	X
2. Toxicity tests	X	X	X	X	X
Acute toxicity	X	X	X	X	X
Chronic toxicity	X	X	X	X	X
Histopathology	X	X	X	X	X
Low-level responses (behavioral responses)	X	X	X	X	X
3. Biostimulatory tests	X	X	X	X	X
Algal growth response (AGP)	X	X	X	X	X

Biological Parameter List, Sampling Frequencies, and Priority for Proposed Biomonitoring Programs (from Worf 1980).

(a) Model State Water Monitoring Program

Community	Parameter	Priority ^a	Collection and Analysis ^b		Sampling Frequency ^c
			Mathematic ^b	Grab samples	
Plankton	Counts and Identification	1		Grab samples	Once each—in spring, summer, and fall
	Chlorophyll <i>a</i>	2			
	Biomass as ashfree weight				
	Counts and Identification	1		Artificial substrates	Minimally once annually during periods of peak periphyton population
Periphyton	Chlorophyll <i>a</i>	2			
	Biomass as ashfree weight	2			
	Areal coverage	2		As circumstances prescribe	Minimally once annually during period of peak macrophyton population density and/or diversity
	Identification	2			
Macrophyton	Biomass as ashfree weight	2			
	Counts and Identification	1		Artificial and natural substrates	Once annually during periods of peak macroinvertebrate population density and/or diversity
	Flesh tainting	2			
	Toxic substances in tissue ^d	2			
Fish	Counts and Identification	1		Electrofishing or netting	Once annually during spawning runs or other times of peak fish population density and/or diversity
	Biomass as wet weight	2			
	Condition factor	2			
	Flesh tainting	2			
	Age and growth	2			

(b) Basic Water Monitoring Program

Parameters	Community of Aquatic Organisms			
	Plankton	Periphyton	Macroinvertebrates	Fish/Shellfish
Counts			X	
Species Identification			X	X
Biomass (ashfree weight)		X		
Chlorophyll <i>a</i>	X	X		
Toxic substances				X ^e
Habitat Types				
Rivers		X	X	X
Lakes	X	X		X
Estuaries	X			

Source: Model State Water Monitoring Program (Washington: Environmental Protection Agency 1975).

Table 4.
Biological Parameter Units

Community	Parameter	Units
Plankton	Counts	Numbers/ml by genus and/or species
	Chlorophyll <i>a</i>	mg/m ³
	Biomass (ashfree dry weight)	mg/m ³
	Counts	Numbers/mm ³
Periphyton	Chlorophyll <i>a</i>	mg/m ²
	Biomass (ashfree weight)	mg/m ²
	Autotrophic index	Ashfree weight (mg/m ²)
		Chlorophyll <i>a</i> (mg/m ²)
Macrophyton	Areal coverage	Maps by species and species associations
	Biomass (ashfree weight)	g/m ²
	Counts	Grab—number/m ²
		Substrate—number/sampler
Fish	Biomass	g/m ³
	Toxic substances	mg/kg
	Toxic substances	mg/kg
	Counts	Number/unit of effort, expressed as per shocker hour or per 100 ft of a 24-hour net set
	Biomass (wet weight)	Same as counts
	Condition	$K(TL) = \frac{10^6 \times \text{weight in grams}}{L^3 \text{ (length in mm)}}$

TABLE 5 PHYSICAL (I), CHEMICAL (II), AND BIOLOGICAL (III) PARAMETERS THAT CAN BE MEASURED IN ORDER TO MONITOR THE EFFECTS OF SEWAGE EFFLUENT DISCHARGE. The methods are listed and the ease of measurement rated as + = easy, ++ = moderate, and +++ = difficult or time consuming.

PARAMETER	METHOD	EASE OF MEASUREMENT	FREQUENCY OF MEASUREMENT
I Light extinction (turbidity) Temperature and salinity	Secchi disc or transmissometer temperature and salinity probe	+ +	bimonthly "
II Dissolved O ₂ Inorganic nutrients (NO ₃ , NH ₄ , urea, PO ₄ , SiO ₄) Dissolved organic N and P Particulate N and P	Winkler titration Autoanalyzer Digestion and autoanalyzer Digestion and autoanalyzer	++ ++ ++ ++	bimonthly " monthly "
III Chlorophyll <u>a</u> - phytoplankton Cell counts - coliform bacteria - phytoplankton Species composition - phytoplankton, zooplankton, & bacteria Metabolic rates - phytoplankton Bioaccumulation - phytoplankton & bacteria (e.g. metals) Zooplankton biomass	<u>in vivo</u> fluorescence microscope particle counter & microscope microscope 14C uptake atomic absorption spectrometry net hauls	+ +++ ++ +++ +++ +++ ++	monthly " " " " " "

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