# DEPARTMENT OF ENVIRONMENT ENVIRONMENTAL PROTECTION PACIFIC REGION YUKON BRANCH 

## ENVIRONMENTAL QUALITY OF RECEIVING WATERS AT UNITED KENO HILL MINES LTD. ELSA, YUKON

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## ABSTRACT

During summer 1985, a receiving environment monitoring study was undertaken by Environmental Protection in the streams potentially influenced by the mining and milling operations of United Keno Hill Mines in the Elsa/Keno area of Yukon.

Some parameters exceeded the water quality guidelines for drinking water and aquatic life at specific station locations. Zinc and manganese levels were noted to be excessively high relative to the guidelines at several station locations. The inputs of metals from mine adit water flow and tailings pond effluent can be detected in the South McQuesten River, Stations 4 and 9.

Sediment chemistry data at specific stations reflect metals input from tailings pond input and a tailings dam break in 1978. As well, elevated metals levels were detected at locations (Stations 2,7) which suggest an influence from mine adit water flows or from undocumented historical tailings releases in previously mined areas. Elevated sediment metal levels were detected in the South McQuesten River at Stations 4 and 9.

Benthic invertebrates were sampled and identified. Generally the populations show good abundance and diversity of species although one location (Station 9) showed a major dominance by one group (Simulium sp.).

## RÉSUMÉ

La Direction générale de la protection de l'environnement a entrepris, au cours de l'été 1985, une étude sur le contrôle des milieux récepteurs dans les cours d'eau susceptibles d'être influencés par les opérations d'extraction et de préparation de la United Keno Hill Mines dans la région de Elsa/Keno au Yukon.

A certains endroits bien précis, certains paramètres étaient plus élevés que les limites prévues dans les lignes directrices pour l'eau potable et la vie aquatique. Les concentrations de zinc et de manganèse étaient excessivement élevées à plusieurs endroits, comparativement aux limites prévues dans les lignes directrices. Les métaux provenant des eaux s'écoulant des galeries à flanc de coteau et des effluents des bassins à résidus peuvent être décelés aux stations 4 et 9 sur la rivière South McQuesten.

Les données sur la chimie des sédiments à différentes stations reflètent les rejets métalliques provenant des bassins à résidus et de la rupture, en 1978, d'un barrage retenant les eaux d'un bassin à résidus. De plus, on a décelé des concentrations élevées d'espèces métalliques à divers endroits (stations 2 et 7), ce qui traduit peut-être l'effet des eaux s'écoulant des galeries à flanc de coteau ou l'effet d'anciens rejets non documentés à partir de zones exlpoitées antérieurement. On a décalé des concentrations élevées de métaux dans les sédiments aux stations 4 et 9 sur la rivière South McQuesten.

On a prélevé et identifié des échantillons d'invertébrés benthiques. En général, l'abondance de ces populations sinsi que la diversité des espèces étaient bonnes, mais un groupe (Simulium sp.) était nettement dominant à un endroit (station 9).

## ADDENDUM

On July 17, 1986 a follow-up water quality survey was conducted at selected locations in the study area to determine the origin of elevated metals found in Flat Creek in 1985 which did not appear to originate from the tailings pond decent (Station 5). The sites sampled included Station 5, Station 6, Station 7 and several small drainages found between these stations.

Among those locations sampled a diversion channel which diverts ground water around the tailings dam from the base of the valley wall on which the mill and town are located, was found to have higher levels of certain metals than the stations sampled. This seepage was not sampled in 1985.

The following table displays levels (mg/L ext.) of selected metals found at Station 5, Station 6, Station 7, the seepage channel and Galena Creek during the follow-up survey.

| STATION | Ag | Cd | Fe | Zn |
| :---: | ---: | ---: | ---: | :---: |
| 5 | 0.0009 | $<0.003$ | 0.507 | 0.153 |
| 6 | $<0.0005$ | $<0.002$ | 0.080 | 0.004 |
| 7 | $<0.0005$ | 0.003 | 0.208 | 0.228 |
| seepage | 0.0020 | 0.006 | 1.600 | 0.476 |
| Galena Cr. | $<0.0005$ | $<0.002$ | 0.840 | 0.122 |

Of the metals shown, Zn and Cd were the only metals at Station 7 to be in excess of the decant although the seepage showed the highest values overall for Ag and Fe .

The data clearly shows the seepage, and to a lesser degree, Galena Creek, contribute to the elevated Zn at Station 7. Other minor ground water seepages in the area were sampled but none showed elevated levels of the above metals.

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### 1.0 INTRODUCTION

An investigation of water quality, stream sediments and aquatic invertebrate populations was carried out by the Environmental Protection Service July $9-10$ and August $21-22$, 1985 in the South McQuesten River watershed in the vicinity of United Keno Hill Mine at Elsa, Yukon.

The purpose of the investigation was to determine if any significant impact from the mining and milling operations was detectable in the receiving waters, namely Flat Creek, Christal Creek and the South McQuesten River.

The information collected by this survey is compared with information from two previous surveys conducted by the Environmental Protection Service in 1974-1975 (Regional Program Report 78-14) and in 1980 (Regional Program Report 81-23). Further comparisons are made with water quality information available through the Northern Affairs Program, (1985), in Whitehorse, Yukon.

The parameters found, by the present study, to exceed recommended levels for drinking water or which are known to be toxic to freshwater invertebrates and fish are identified and discussed. Pollution trends which can be attributed to mine activities are also identified.

### 2.0 STUDY AREA

United Keno Hill Mine is located at the town of Elsa, Yukon (63 ${ }^{\circ}$ $55^{\prime} \mathrm{N}, 135^{\circ} 30^{\prime} \mathrm{W}$ ) approximately 450 kilometres by road north of Whitehorse via the Klondike and the Silver Trail Highways (see Figure 1).

Prospectors first took interest in this area in 1906 when silver, lead and zinc ore deposits were discovered on Galena Hill where the town of Elsa is now situated. A "stampede" resulted when major silver deposits were discovered in 1919. Ore deposits were high-graded for more than 20 years until 1942 when World War II brought most mining activity in the area to a halt (Sinclair et al, 1976).

In 1946 the United Keno Hill Mine Ltd. began mining and milling operations and have operated almost continuously for the past 40 years. Ore from several nearby adits is transported to a crusher and flotation/ recovery mill located at Elsa where lead, zinc and silver concentrates are extracted. The mill is currently operating at 300 tons per day with mineral recovery per ton estimated at 20 oz. silver, $1.4 \%$ zinc and $3 \%$ lead. Ore concentrates are eventually shipped to smelters in southern Canada for complete recovery (Northern Affairs Program, 1985).

Mill tailings and mine water are discharged and contained within a series of three dam structures immediately below the Elsa townsite. The mill discharges approximately 280 tons of water and waste ore (2 to 1 ratio) per day which is then treated with hydrated lime to precipitate heavy metals. In 1984, 39,200 kilograms of lime were added to the tailings (Northern Affairs Program, 1985). Treated effluent is eventually released into Flat Creek which in turn joins the South McQuesten River approximately 10 kilometres downstream from the tailings pond decant. Christal Creek, which originates at Christal Lake 10 kilometres east of Elsa, flows into the South McQuesten River approximately 12 kilometres upstream of the Flat Creek confluence. Although this tributary is not directly associated with the receiving waters it is affected by drainage from several mine adits on the north slope of Galena Hill and the south slope of Keno Hill.

3.0 METHODS

A total of 11 sampling stations were established in the study area, some of which coincide with those established in the two previous investigations, (Environmental Protection Service 1978, Bethel and Soroka, 1981). Table 1 provides station descriptions and Figure 2 identifies station locations. All stations were accessed by road except for Stations 2,3 and 4 which were accessed by helicopter.

## TABLE 1 STATION DESCRIPTION

STATION
DESCRIPTION

1 Christal Creek, 5 meters $d / s$ of culvert at Keno City Road crossing.
2 Christal Creek, 15 meters $u / s$ of confluence with South McQuesten River.
3 South McQuesten River, 50 meters $u / s$ of Christal Creek confluence.
4 South McQuesten River, 50 meters $\mathrm{d} / \mathrm{s}$ of Christal Creek confluence.
5 Tailings pond decant.
6 Flat Creek, u/s of Mayo/Elsa Highway.
7 Flat Creek, 600 meters $\mathrm{u} / \mathrm{s}$ of confluence with South McQuesten River.
8 South McQuesten River, 50 meters $u / s$ of Flat Creek confluence.
9 South McQuesten River, 50 meters $\mathrm{d} / \mathrm{s}$ of Flat Creek confluence.
10 South McQuesten River, approximately 6 kilometers d/s of Flat Creek.
11 South McQuesten River at Bridge downstream of Haggart Creek.


During the July 9-10 sampling period, flow was not observed at the traditional tailings pond decant but water samples were collected from a pond immediately below the decant culvert. It was later determined that the tailing pond discharge was diverted to a second decant location, unknown to the field staff at the time of sampling.

During the August 21-22 sample period water samples were collected at the second decant location.

### 3.1 Water Quality and Quantity

In situ water quality measurements included temperature, conductivity, pH and dissolved oxygen. Temperature and conductivity were measured with a YSI Model 33 Temperature-Conductivity-Salinity Field Meter, pH was measured using a Fisher Scientific Model 640 Field Meter or Horiba Water Quality Checker and dissolved oxygen was measured with a YSI Model 57 Dissolved 0xygen Field Meter. The latter was calibrated using the water saturated air method as described in the YSI Manual. Readings were corrected for temperature, elevation and salinity. Percent saturation was calculated from oxygen saturation tables derived from APHA et al (1981). A full description of field equipment and measurements is given in Appendix $I$, Table 1.

Water quality samples, collected in triplicate at each station, included a 2 litre sample for nutrients analysis and a 100 ml sample for extractable metals analysis. Sample collection, preservation and analysis methods are shown in Appendix I, Table 1. The mean and standard deviation were calculated for each set of three samples collected. During sample collection in August, an attempt was made to characterize the channel cross section at stations on the South McQuesten River. One of each triplicate set was collected along the left bank, at mid stream and along the right bank of the river. Water quality data for each triplicate set in Appendix II, Table 2 is shown in this sequence (left bank, mid stream, right bank).

The parameters analysed in each nutrient sample are as follows:

| pH | total phosphates |
| :--- | :--- |
| conductivity | nitrites |
| colour | nitrates |
| turbidity | ammonia |
| nonfilterable residue | sulfate |
| total alkalinity | chloride |
| total hardness |  |

The following parameters were analysed in each extractable metals sample:

```
aluminum (Al)
antimony (Sb)
arsenic (As)
boron (B)
barium (Ba)
beryllium (Be)
cadmium (Cd)
calcium (Ca)
chromium (Cr)
\begin{tabular}{ll} 
cobalt (Co) & silicon (Si) \\
copper (Cu) & silver ( Ag ) \\
iron ( Fe ) & sodium ( Na ) \\
lead ( Pb ) & strontium ( Sr\()\) \\
magnesium (Mg) & tin ( Sn ) \\
manganese (Mn) & titanium (Ti) \\
molybdenum (Mo) & vanadium (V) \\
nickel (Ni) & zinc ( Zn\()\) \\
selenium (Se) &
\end{tabular}
```

The analyses were completed at the Environmental Protection Service Laboratory, 4195 Marine Drive, West Vancouver, B.C.

Stream flow was measured at selected stations using a Marsh McBirney Electromagnetic Flow Meter. Ten velocity readings, in centimeters per second, were taken across the width of each South McQuesten River Station. On the narrower Christal and Flat Creek Stations usually four readings were taken. Stream flows were calculated by dividing the width of the stream into equal blocks, according to the number of readings taken, then the area of each block was determined (water depth $X$ block width). This area was then multiplied by the stream velocity for each block giving a cubic meter per second value ( $\mathrm{m} 3 / \mathrm{sec}$ ). All block flows were added together to arrive at a measured stream flow.

In some cases where excessive stream depth and velocity made it hazardous for field staff to effectively measure stream flow, discharge was calculated as described below:

July 9-10, 1986

```
Station 4 = Station 2 + Station 3
Station 8 = Station 9 - Station 7
```

August 21-22, 1986

```
Station 3 = Station 4 - Station 2
Station 9 = Station 7 + Station 8
```

Discharge could not be accurately determined at Stations 5 and 6 in July and at Stations 1,5 and 6 in August because stream velocities encountered were less than the minimum velocity required by the instrumentation used.

### 3.2 Sediments

Sediment samples were collected in triplicate at each station, except for Stations 5 and 6, on both visits to the study area. A stainless steel sediment corer device was used to reduce the loss of very fine sediments from samples collected in fast flowing water. In calm or slow moving waters an aluminum scoop shovel was used to collect a sample. The samples were placed in paper geochemical sampling bags, packaged in plastic bags and then frozen within 48 hours of collection. A description of the corer sampler, sediment collection, preparation and analysis methods is given in Appendix I, Table 2.

Each sample was analysed for particle size composition and the following leachable metals:

| aluminum (Al) | iron (Fe) | silicon ( Si ) |
| :--- | :--- | :--- |
| arsenic (As) | lead ( Pb ) | silver ( Ag ) |
| barium ( Ba ) | magnesium ( Mg ) | sodium ( Na ) |
| beryllium (Be) | manganese (Mn) | strontium ( Sr ) |
| cadmium (Cd) | mercury (Hg) | tin (Sn) |
| calcium (Ca) | molybdenum (Mo) | titanium (Ti) |
| chromium (Cr) | nickel (Ni) | vanadium (V) |
| cobalt (Co) | potassium (K) | zinc (Zn) |
| copper (Cu) | selenium (Se) |  |

Particle size analysis was carried out only on samples from Stations 7-11 inclusive. The sediment samples were analysed at the Environmental Protection Service Laboratory, 4195 Marine Drive, West Vancouver, B.C.

### 3.3 Bottom Fauna

Benthic invertebrate samples were collected at all stations except Stations 5 and 6. At each station sampled, 3 artificial substrate samplers were placed on the stream bed on July 9-10, 1985. The samplers used were cylindrical wire baskets (maximum volume $=0.0057$ cubic meters) filled with local hand cleaned substrate material ranging from 2 cm to 6 cm in size. The samplers were placed in the stream where in situ measurements, water and sediment samples were collected. The samplers were left to be colonized for a period of 43 days. On August $21-22$, 1985 the baskets were retrieved and immediately placed into a Wildco wash bucket with 0.5 mm mesh bottom. The bucket was held downstream during retrieval of the sampler in order to capture any escaping organisms. Large rock and wood debris was hand scrubbed in the wash bucket to remove invertebrates and then discarded. Invertebrates and fine debris from each basket were combined into a composite sample for each station. A $10 \%$ formalin solution was used to preserve the samples until sorting could be carried out.

The invertebrate samplers placed at Station 10 were vandalized during the sample period and so the data presented for this station is a result of three samples collected August $21-22$ using a Surber Sampler ( 0.09 m2).

Invertebrate identification and enumeration was carried out by Dr. C. Low, a consulting Invertebrate Biologist in Nanaimo, B.C.

Sorted invertebrate samples were later preserved with methanol and placed in storage at the Environmental Protection Service warehouse facility in Whitehorse.

### 3.4 Laboratory Quality Control

Systematic error and sample contamination during analysis at the EPS Laboratory are minimized through duplicate analysis, procedural blanks and the use of standard reference materials. Internal lab quality control is carried out routinely in all water and sediment analysis before results are released.

### 4.0 RESULTS AND DISCUSSION

### 4.1 Water Quality - Physical and Chemical Parameters

In situ measurement, nutrients and extractable metals results for both sample periods are presented in Appendix 11, Tables 1 and 2. Criteria recommended for drinking water and aquatic life are presented in Appendix 1 , Table 4.
4.1.1 Temperature. In situ temperatures reflect seasonal changes. The South McQuesten River averaged $15^{\circ} \mathrm{C}$ on July $9-10$ and $9^{\circ} \mathrm{C}$ on August 21-22. The tributaries surveyed were slightly cooler ranging from $7.5^{\circ} \mathrm{C}$ to $12.5^{\circ} \mathrm{C}$ in July and $4.0^{\circ} \mathrm{C}$ to $7.5^{\circ} \mathrm{C}$ in August.
4.1.2 Flow. Flow measurements were taken when possible during each visit to the study area except at Stations 4, 5, 6 and 8 on July 9-10 and Stations 3, 5, 6, 9 and 10 on August 21-22.

No discharge data was obtained for Station 5 on July 4-10 sampling period.

On August 21-22 a steady discharge from the tailings pond decant culvert was observed and was estimated by field staff at approximately 0.02 $\mathrm{m}^{3} / \mathrm{sec}$. Accurate measurements could not be obtained because of the shallow nature and slow velocity of the decant.

No flows were obtained at Station 6, Flat Creek, because the irregular stream bed and low level of water made it impossible to accurately measure with the instrumentation available.

Flows measured at stations on the South McQuesten River on July 9-10 were considerably higher than flows measured during the August 21-22 sample period. In July they ranged from $5.9 \mathrm{~m}^{3} / \mathrm{sec}$ at Station 3 , the furthest upstream station, to $9.7 \mathrm{~m}^{3} / \mathrm{sec}$. at Station 11 which is located furthest
downstream. A similar degree of increase in flow in the South McQuesten River was also measured on August $21-22$, ranging from $3.4 \mathrm{~m}^{3} / \mathrm{sec}$ at Station 4 to $6.7 \mathrm{~m}^{3} / \mathrm{sec}$ at Station 11 although total flow was much lower in August.

The flow at Station 7, Flat Creek, ranged from $0.07 \mathrm{~m}^{3} / \mathrm{sec}$ in July to $0.06 \mathrm{~m}^{3} / \mathrm{sec}$ in August.

Flows measured at Station 2, Christal Creek, showed no change between the sample periods.
4.1.3 Dissolved Oxygen. Percent dissolved oxygen saturation (\%D0) was slightly higher on August 21-22 than on July 9-10. The South McQuesten River ranged from $82 \%$ to $97 \%$ during the July sample period. In August, \%DO ranged from $95 \%$ to $104 \%$.
\%D0 at Stations 1 and 2, Christal Creek and at Stations 6 and 7, Flat Creek, ranged from $80 \%$ to $99 \%$ over the two sample periods.

The high \%DO at Station 5 on August $21-22$ (119\%) is a result of aeration occurring where the decant water discharges from the decant culvert into a small pool before flowing into Flat Creek.
4.1.4 pH. The slightly alkaline pH of waters in the study area are characteristic of this area (Environmental Protection Service, 1978). The South McQuesten River had a mean pH value of 8.15 ( $+/-0.16$ ) in July and 7.89 ( $+/-0.18$ ) in August. No significant change could be detected at Stations 4 and 9 immediately downstream of Christal Creek Flat Creek respectively. The lowest field pH recorded (7.65) on the South McQuesten River was at Station 9 on August $21-22$ but this is not considered to be representative, as shown by the upstream pH of 7.90 at Station 8 and at Station 7, Flat Creek.
4.1.5 Conductivity. Conductivity measurements varied considerably throughout the study area. As Figure 3 demonstrates, conductivity in the receiving waters was elevated by decant from the tailing pond (Station 5). In situ values at Station 5 were 1190 umhos/cm on July 9-10 and 920 umhos/cm


FIGURE 4 MEAN ALKALINITY AND HARDNESS
on August 21-22. This high input can be traced downstream at Station 7 and to a lesser extent at Station 9. The dilution effect of the South McQuesten River is more prevalent in July, when flows were higher, than in August as shown by the difference in conductivity at Station 9 from the two sampling periods. Further downstream at Station 10 and 11 conductivity is similar to background levels of 240 umhos/cm found at Station 3.

High conductivity was also detected in Christal Creek during both sample periods as shown by Figure 3, Stations land 2. These elevated levels are suspected to reflect the influence of drainage from the Galkeno 900 adit which enters upstream of Station 1 at Christal Lake.

Data from each triplicate sample collected in August at Station 4 and 9 show the presence of a plume immediately downstream of where Christal Creek and Flat Creek join the South McQuesten River. At Station 4, lab conductivity across the river decreased from 310 umhos/cm along the left bank to 255 umhos/cm along the right bank. Lab conductivity at Station 3 was 250 umhos/cm across the full width of the river. A similar plume was detected at Station 9, where lab conductivity decreased from 680 umhos/cm along the left bank to 310 umhos/cm along the right bank. Lab conductivity at Stations 8,10 and 11 show no indications of lateral gradient in conductivity and are similar to background levels.

Historically the Environmental Protection Service (1978) reported lower conductivity in receiving waters. In July, 1974 and June and July of 1975 conductivity of decant from the tailings pond was 380, 360 and 950 umhos/cm, respectively. Conductivity of the South McQuesten River upstream of Christal Creek was very similar to that found by the present study. Bethel and Soroka (1981) reported results similar to the present report except at Station 8 where lab conductivity averaged 530 umhos/cm over a three day period in August, 1980.
4.1.6 Colour. During the July sample period, colour measured 20 Relative Units (RU) at most stations except Stations 1 and 2 where it was 10 RU. Values during the August sample period were slightly lower with most stations averaging 10 RU except Stations 1 and 2 which were 5 RU. The higher values in July at most stations no doubt reflects increased organic loading commonly associated with peak flow periods, as observed.
4.1.7 Turbidity. Turbidity readings at Station 1, 5, and 7 during the July sample period and at Station 7 during the August sample period were slightly elevated relative to background turbidity at Station 3 . The highest turbidity reading was 2.50 FTU at Station 7 July 9-10. Turbidity at Station 7 during both sample periods exceeded that found at Station 5 suggesting the source is other than the tailings pond decant. Background turbidity at Station 3 during July and August was 0.45 and 0.16 FTU, respectively.
4.1.8 Non-filterable Residue (NFR). NFR values were low or below the 5mg/L detection limit in most of the samples collected on both visits to the study area. The highest NFR level detected was $14 \mathrm{mg} / \mathrm{L}$ at Station 1 during the August sample period. NFR values in the 5 to $14 \mathrm{mg} / \mathrm{L}$ range were also detected at Station 1 on July $9-10$ and at Stations 2, 4, 6, 7, 10 and 11 on August 21-22.
4.1.9 Hardness and Alkalinity. Water hardness changed little between the two sample periods but exceeded the $100 \mathrm{mg} / \mathrm{L}$ level recommended for drinking water at all stations. Elevated levels, ranging from 238 to 671 $\mathrm{mg} / \mathrm{L}$, were detected at Stations $1,2,5,6$ and 7 . Background levels at Station 3 during July and August, 132 and $136 \mathrm{mg} / \mathrm{L}$ respectively, reflect the geology of the drainage area.

The elevated hardness found in Flat Creek, Station 7, can be detected at Station 9 on the South McQuesten River but, as Figure 4 shows, it returns to near background levels at Station 10 and 11 . Water samples collected at Station 9 on August 21-22 show elevated hardness along the left bank indicating the presence of the Flat Creek plume.

Changes in water hardness also show a direct correlation with the changes in conductivity recorded during both sample periods.

Figure 4 also shows high alkalinity in Flat Creek at Station 6 during both sample periods but it is not directly related to the increase in hardness. During July and August alkalinity was 85.7 and $86.0 \mathrm{mg} / \mathrm{L}$, respectively, at Station 3 and, as Figure 4 shows, it returns to near background levels at Station 10 and 11 after only a slight increase at Station 9. Samples collected August 21-22 at Station 9 confirm the presence of the Flat Creek plume as shown by the higher alkalinity along the left bank of the South McQuesten River.

Results from Environmental Protection Service (1978) and Bethel and Soroka (1981) compared with results from the present study show hardness was similar in the South McQuesten River but generally lower in Flat Creek and Christal Creek in past years at corresponding locations, as shown by Table 2.

TABLE 2 HISTORICAL COMPARISON OF ALKALINTTY AND HARDNESS LEVELS(mg/L)

| STATION | EPS |  | EPS |  | EPS |  | EPS |  | EPS |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | JULY | 1974 | JULY | 1975 | *AUG. | 1980 | JULY | 1985 | AUG. | 1985 |
|  | ALK. | HARD. | ALK. | HARD. | ALK. | HARD. | ALK. | HARD. | ALK. | HARD |
| 1 | N/A | 160 | 79.2 | 260 | N/A | N/A | 82.9 | 432 | 103.3 | 468 |
| 2 | N/A | 190 | 10.4 | 210 | N/A | N/A | 110.7 | 313 | 120.3 | 326 |
| 3 | N/A | 100 | 73.3 | 100 | N/A | N/A | 85.7 | 132 | 86.0 | 136 |
| 4 | N/A | N/A | N/A | N/A | N/A | N/A | 86.2 | 141 | 89.6 | 154 |
| 5 | N/A | 440 | 80.2 | 440 | 52.1 | 739 | 93.0 | 591 | 92.6 | 671 |
| 6 | N/A | 820 | N/A | N/A | N/A | N/A | 168.3 | 238 | 187.3 | 254 |
| 7 | N/A | 240 | 117.8 | 220 | 130.0 | 342 | 143.3 | 494 | 163.0 | 509 |
| 8 | N/A | 110 | 78.2 | 110 | 118.0 | 255 | 89.7 | 140 | 100.6 | 156 |
| 9 | N/A | 290 | 86.1 | 120 | 85.0 | 126 | 92.1 | 158 | 119.7 | 260 |
| 10 | N/A | N/A | N/A | N/A | N/A | N/A | 93.0 | 148 | 106.0 | 170 |
| 11 | N/A | N/A | N/A | N/A | N/A | N/A | 96.6 | 155 | 108.7 | 168 |

* Three day mean. N/A - Not Available

In July, 1974 and July, 1975 hardness ranged from 160 to $260 \mathrm{mg} / \mathrm{L}$ at Stations 1 and 2 while in the present study hardness was consistently higher, averaging 313 and $468 \mathrm{mg} / \mathrm{L}$ at the same locations. Water hardness similarly has increased at Stations 5 and 7 over the past 11 years. Since 1974 levels have increased from 440 to $671 \mathrm{mg} / \mathrm{L}$ at Station 5 and from 240 to $509 \mathrm{mg} / \mathrm{L}$ at Station 7. Hardness at Station 6 was found by the present study to be lower than Stations 5 and 7 but in July, 1974 it reached $820 \mathrm{mg} / \mathrm{L}$ (Environmental Protection Service, 1978).

Historical comparison of water hardness at Station 3 shows little change, suggesting that sustained mining activity in the area may be associated with the increases observed in receiving waters.

Alkalinity has increased slightly at all comparable stations since 1975. In July, 1975 it ranged from 73.3 to $117.8 \mathrm{mg} / \mathrm{L}$ while in August, 1985 it ranged from 86 to $187 \mathrm{mg} / \mathrm{L}$.
4.1.10 Sulfates. Sulfates remained well below the $500 \mathrm{mg} / \mathrm{L}$ level recommended for drinking water at all stations except Station 5. In the present study sulfates at Station 5 averaged $497 \mathrm{mg} / \mathrm{L}$ in July and $540 \mathrm{mg} / \mathrm{L}$ in August. Average background levels found at Station 3 ranged from $30 \mathrm{mg} / \mathrm{L}$ in July to $42 \mathrm{mg} / \mathrm{L}$ in August. Bethel and Soroka (1981) reported sulfates at each station sampled but exceeded $500 \mathrm{mg} / \mathrm{L}$ only at Station 5 ( $663 \mathrm{mg} / \mathrm{L}$ ).
4.1.11 Chlorides. Chlorides were elevated at Station 5 as compared with levels found at other stations in the study area. They averaged 18.9 and $14.3 \mathrm{mg} / \mathrm{L}$, respectively, during the July and August sample periods but were well below the $250 \mathrm{mg} / \mathrm{L}$ level recommended for drinking water. It is believed the elevated chlorides at Station 5 are residuals of the calcium hypochlorite reagent used in a cyanide flotation circuit previously used at the mill. Bethel and Soroka (1981) reported similar levels at Stations 5, 7 , 8 and 9. At Station 3 chloride levels averaged 0.7 and $0.2 \mathrm{mg} / \mathrm{L}$ during the sample periods.
4.1.12 Phosphates. Total phosphate, which was detected at several locations, generally is not considered to be toxic to aquatic organisms but levels as low as $0.002 \mathrm{mg} / \mathrm{L}$ have promoted algae growth under controlled conditions (Bothewll, 1985). Although mean concentrations found during both sample periods ranged from $<0.002$ to $0.016 \mathrm{mg} / \mathrm{L}$, unusually high values of 0.049 and $0.059 \mathrm{mg} / \mathrm{L}$ were found in individual samples at Station 8 and Station 9, respectively, August 21/22. Since these values are not comparable to other samples collected at the same time and locations, it is believed they are a result of sample contamination or analytical error.

### 4.1.13 Nitrite and Nitrate. Nitrite was below detection limit (0.005

 $\mathrm{mg} / \mathrm{L}$ ) in the July and August sample periods of the present study. In comparison, levels ranging from 0.014 to $0.075 \mathrm{mg} / \mathrm{L}$ exceeded the $0.001 \mathrm{mg} / \mathrm{L}$ criteria recommended for drinking water in August, 1980 at Stations 5, 7 and 8 (Bethel and Soroka, 1981).Nitrate ranged from 0.003 to $0.16 \mathrm{mg} / \mathrm{L}$ in the present study. These levels are well below the $10 \mathrm{mg} / \mathrm{L}$ limit recommended for drinking water and do not pose a threat to aquatic life from the perspective of stimulating algal growth.

Bethel and Soroka (1981) reported nitrate at Stations 5, 7 and 9 within the range found in the present study.
4.1.14 Total Ammonia. Toxicity of ammonia has been attributed primarily to the unionized portion of total ammonia present (Thurston et al, 1974). Unionized ammonia concentrations increase with increasing pH , temperature and total ammonia.

Levels of total ammonia in the present study ranged from (0.005 to $0.13 \mathrm{mg} / \mathrm{L}$ in July and ( 0.005 to $0.16 \mathrm{mg} / \mathrm{L}$ in August. The unionized portion calculated from the highest total ammonia value found ( $0.16 \mathrm{mg} / \mathrm{L}$ ) was 0.0009 $\mathrm{mg} / \mathrm{L}$, well below the criteria recommended for drinking water ( $0.5 \mathrm{mg} / \mathrm{L}$ ) and the protection of aquatic life ( $0.02 \mathrm{mg} / \mathrm{L})$. Bethel and Soroka (1981) reported three day averages of $1.45,0.450,0.040$ and $0.007 \mathrm{mg} / \mathrm{L}$ total
dissolved ammonia at Stations 5, 7, 8 and 9, respectively. The unionized portion of that found by Bethel and Soroka (1981) at Station 5 exceeded the $0.02 \mathrm{mg} / \mathrm{L}$ limit recommended for aquatic life.

### 4.2 Water Quality - Extractable Metals

Results of the extractable metals analysis for each sample is presented in Appendix II, Tables 1 and 2. Appendix I, Table 1 gives the detection limits for each parameter.

The following metals were below detection limits at all stations sampled July 9-10 and August 21-22:

July 9-10
arsenic (As)
beryllium (Be)
chromium (Cr)
antimony (Sb)
selenium (Se)
tin ( Sn )
titanium ( Ti )
vanadium (V)

August 21-22
arsenic (As)
beryllium (Be)
cobalt (Co)
chromium (Cr)
antimony ( Sb )
selenium (Se)
titanium (Ti)
vanadium (V)

Boron (B), barium (Ba), calcium (Ca), molybdenum (Mo), nickel (Ni), phosphorous ( P ), silica ( Si ) and strontium ( Sr ) were detected in samples collected in July and August but were below the recommended levels for drinking water and aquatic life.

Magnesium (Mg), although not considered an environmental concern, was detected at all stations during both of the sample periods and graphically traces the tailings pond decant downstream of the point of discharge. As Figure 5 shows, the elevated levels discharged at Station 5 remain elevated at Station 7 . At Station 9 the levels were higher during the August sample period because the South McQuesten River had less of a dilution affect than in July when flows were higher.


FIGURE 5 MEAN EXTRACTABLE Mg IN WATER

Plumes from Christal Creek and Flat Creek in the South McQuesten River were detected at Stations 4 and 9 respectively during the August sample period. This is shown by the higher levels of magnesium found along the left bank than what was found at midstream and near the right bank (see Appendix II, Table 2). Magnesium returned to near background levels downstream of Station 9.
4.2.1 Silver (Ag). Silver averaged $0.0024 \mathrm{mg} / \mathrm{L}$ at Station 5, on July 9-10. The analysis for this period also shows levels decreasing from 0.0024 $\mathrm{mg} / \mathrm{L}$ at Station 5 , to $0.0007 \mathrm{mg} / \mathrm{L}$ at Station 7 , to $0.0005 \mathrm{mg} / \mathrm{L}$ along the left bank at Station 9. In August, silver was $0.0010 \mathrm{mg} / \mathrm{L}$ at Station 5 and below detection limit ( $0.0005 \mathrm{mg} / \mathrm{L}$ ) at stations further downstream.

In all cases, silver was below the criteria of $0.05 \mathrm{mg} / \mathrm{L}$ recommended for drinking water. However, all levels detected by this survey did exceed the criteria of $0.0001 \mathrm{mg} / \mathrm{L}$ recommended for the protection of aquatic life.

Environmental Protection Service (1978) reported silver to be less than the detection levels of $0.01 \mathrm{mg} / \mathrm{L}$ in June, 1975 and $0.03 \mathrm{mg} / \mathrm{L}$ in July, 1975 at all stations sampled.
4.2.2 Cadmium (Cd). Table 3 compares cadmium levels from previous surveys with those found in the present study.

Cadmium has been detected at the present Station 5 on each visit shown except in August, 1980 where it was below the detection limit of $0.01 \mathrm{mg} / \mathrm{L}$.

The results of the present study show cadmium higher in Christal and Flat Creek than in the South McQuesten River.

The elevated levels found in Christal Creek during the July sample period can be detected at Station 4 on the South McQuesten River. DIAND reported (pers. comm., 1985) high concentrations of heavy metals in drainage from an inactive adit, Galkeno 900, which flows into Christal lake. On July

17, 1985 cadmium was $0.012 \mathrm{mg} / \mathrm{L}$ at the adit and $0.003 \mathrm{mg} / \mathrm{L}$ in Christal Creek near Station l. This clearly identifies one source of cadmium found at Station 1 and 2 of the present survey.

TABLE 3 HISTORICAL COMPARISON OF CADMIUM LEVELS IN WATER (mg/L)

|  | EPS | EPS | EPS | *EPS | EPS | EPS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | JULY | JUNE | JULY | AUGUST | JULY 9-10 | AUG. 21-22 |
|  | 1974 | 1975 | 1975 | 1980 | 1985 | 1985 |
|  | STATION |  |  |  |  |  |
| 1 | $<0.01$ | <0.01 | $<0.01$ | N/A | 0.0032 | 0.0016 |
| 2 | $<0.01$ | N/A | $<0.01$ | N/A | 0.0030 | 0.0020 |
| 3 | <0.01 | N/A | <0.01 | N/A | 0.0014 | $<0.0005$ |
| 4 | N/A | N/A | N/A | N/A | 0.0018 | <0.0005 |
| 5 | 0.06 | 0.03 | 0.05 | $<0.01$ | 0.0024 | 0.0023 |
| 6 | $<0.01$ | N/A | N/A | N/A | 0.0017 | $<0.0005$ |
| 7 | $<0.01$ | 0.01 | $<0.01$ | $<0.01$ | 0.0043 | 0.0018 |
| 8 | $<0.01$ | N/A | $<0.01$ | <0.01 | 0.0019 | <0.0005 |
| 9 | <0.01 | <0.01 | <0.01 | <0.01 | 0.0018 | *0.0003 |
| 10 | N/A | N/A | N/A | N/A | 0.0019 | <0.0005 |
| 11 | N/A | N/A | N/A | N/A | 0.0015 | <0.0005 |

* Three day mean. N/A - Not available

Cadmium was elevated at Station 5 during both sample periods but Station 7 on July 9-10 reflects an unusually high concentration which does not appear to originate from the tailings pond decant. The level found at Station 6 on the same day is lower than what was found in the decant, therefore this rules out the headwaters of Flat Creek as a potential source. There is insufficient information to determine the source but several small intermittent drainages entering Flat Creek between Station 5 and 7 are suspected.

Cadmium levels on August 21-22 show a well defined trend originating at the decant. Station 5 was $0.0023 \mathrm{mg} / \mathrm{L}$, Station 7, Flat Creek, was $0.0018 \mathrm{mg} / \mathrm{L}$ and station 9, South McQuesten River, was $0.0008 \mathrm{mg} / \mathrm{L}$ along the left bank. Samples collected at midstream and along the right bank at Station 9 and at Station 10 and 11 were below the detection limit ( $0.0005 \mathrm{mg} / \mathrm{L}$ ) .

All detectable levels of cadmium were below the $0.005 \mathrm{mg} / \mathrm{L}$ criteria recommended for drinking water but exceeded the $0.0002 \mathrm{mg} / \mathrm{L}$ criteria recommended for aquatic life.
4.2.3 Copper (Cu). Table 4 compares copper levels reported by previous studies with that found by the present study.

TABLE 4 HISTORICAL COMPARISON OF COPPER LEVELS IN VATER (mg/L)

|  | EPS <br> JULY <br> 1974 | EPS <br> JUNE <br> 1975 | EPS <br> JULY <br> 1975 | *EPS <br> AUGUST <br> 1980 | EPS <br> JULY 9-10 <br> 1985 | EPS <br> AUG. 21-22. <br> 1985 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| STATION |  |  |  |  |  |  |
| 1 | $<0.01$ | $<0.01$ | $<0.01$ | N/A | 0.003 | $<0.001$ |
| 2 | $<0.01$ | N/A | $<0.01$ | N/A | 0.002 | 0.003 |
| 3 | $<0.01$ | N/A | $<0.01$ | N/A | 0.003 | $<0.001$ |
| 4 | N/A | N/A | N/A | N/A | 0.003 | 0.002 |
| 5 | 0.60 | 0.19 | 0.22 | 0.047 | 0.012 | 0.012 |
| 6 | $<0.01$ | N/A | N/A | N/A | 0.004 | $<0.001$ |
| 7 | $<0.01$ | 0.20 | $<0.01$ | $<0.010$ | 0.006 | 0.005 |
| 8 | $<0.01$ | N/A | $<0.01$ | $<0.010$ | 0.003 | 0.002 |
| 9 | $<0.01$ | $<0.01$ | $<0.01$ | $<0.010$ | 0.002 | 0.001 |
| 10 | N/A | N/A | N/A | N/A | 0.002 | $<0.001$ |
| 11 | N/A | N/A | N/A | N/A | 0.003 | $<0.001$ |

* Three day mean.

N/A - Not available

Environmental Protection Service (1978) and Bethel and Soroka (1981) both reported copper as less than the detection limit ( $0.01 \mathrm{mg} / \mathrm{L}$ ) at most stations sampled except at Station 5 , the tailings pond decant.

The present study detected copper at most stations during each of the sample periods due to improved detection limits. The mean levels ranged from ( 0.001 to $0.012 \mathrm{mg} / \mathrm{L}$. The maximum mean ( $0.012 \mathrm{mg} / \mathrm{L}$ ) was detected at Station 5 but levels decrease downstream at Station 7 and return to near background levels in the South McQuesten River at Stations 9, 10 and 11.

All detected levels were below the criteria recommended for drinking water ( $1.0 \mathrm{mg} / \mathrm{L}$ ). Station 7 on July $9-10(0.006 \mathrm{mg} / \mathrm{L})$ was the only station other than Station 5 where copper exceeded the criteria recommended for aquatic life ( $0.005 \mathrm{mg} / \mathrm{L}$ ).
4.2.4 Iron (Fe). The levels of iron detected by the present survey were similar to those reported by Environmental Protection Service (1978) and Bethel and Soroka (1981). They exceeded the criteria of $0.3 \mathrm{mg} / \mathrm{L}$ recommended for drinking water in the current study at Station 1 in July ( $0.39 \mathrm{mg} / \mathrm{L}$ ) and at Stations 5 and 7 during both sample periods (ranging from 0.30 to $0.52 \mathrm{mg} / \mathrm{L})$.

Iron levels found in samples collected at Stations 4 and 9 on August $21-22$ were higher along the left bank than at midstream and the right bank indicating the presence of plumes from Christal and Flat Creek in the South McQuesten River.
4.2.5. Manganese (Mn). Manganese detected by the present study ranged between 0.001 and $2.10 \mathrm{mg} / \mathrm{L}$ during the two sample periods. In July levels exceeded the $0.05 \mathrm{mg} / \mathrm{L}$ criteria recommended for aquatic life at most stations except Stations 3, 4 and 6. In August, levels decreased slightly and only exceed the $0.05 \mathrm{mg} / \mathrm{L}$ criteria at Stations $1,2,5$ and 7.

Although elevated levels of manganese can be tolerated by some forms of freshwater aquatic life, available information suggests levels exceeding $0.1 \mathrm{mg} / \mathrm{L}$ may constitute an environmental hazard (Thurston et al, 1979). At Station 1, manganese ranged from $1.89 \mathrm{mg} / \mathrm{L}$ in July to $1.81 \mathrm{mg} / \mathrm{L}$ in August. DIAND conducted an adit survey July 16,1985 (personal comm.), which detected $13.2 \mathrm{mg} / \mathrm{l}$ manganese in water draining from the Galkeno 900 adit upstream of Station l. As well, elevated levels were detected at Station 7 in July ( $2.10 \mathrm{mg} / \mathrm{L}$ ) and at Station 5 in August ( $1.04 \mathrm{mg} / \mathrm{L}$ ). Since Station 5 was not sampled properly in July, it is unclear whether or not the high levels found at Station 7 originate from the tailings pond decant. Manganese was slightly elevated on both sample dates at Station 9 but returned to near background levels at Station 10 and 11.

Historically, manganese has been elevated in discharge from the mine tailings. In July of 1974 it was $16.0 \mathrm{mg} / \mathrm{L}$ (Environmental Protection Service, 1978) while in August, 1980 it averaged $1.95 \mathrm{mg} / \mathrm{L}$ over a three day period (Bethel and Soroka , 1981). The 1974 survey also reported elevated levels near Station 1 ( $6.1 \mathrm{mg} / \mathrm{L}$ ).
4.2.6 Lead. During the July 9-10 sample period, lead was detected at all stations with means ranging from $0.018 \mathrm{mg} / \mathrm{L}$ to $0.026 \mathrm{mg} / \mathrm{L}$. Nevertheless, one sample collected at Station 10 contained only $0.009 \mathrm{mg} / \mathrm{L}$. These results were much higher than those found on August 21-22 and because there is no known reasons for the elevated levels, it is suspected that sample contamination or analytical error occurred. There is insufficient information available to suggest the levels found were characteristic of the study area during the month of July. Bethel and Soroka (1981), reported Pb levels as less than $0.08 \mathrm{mg} / \mathrm{L}$ in July 1978. However, Pb levels of up to $0.84 \mathrm{mg} / \mathrm{L}$ were observed at the decant in June 1975 (EPS, 1978). All mean levels were below the $0.05 \mathrm{mg} / \mathrm{L}$ criteria recommended for drinking but above the $0.01 \mathrm{mg} / \mathrm{L}$ criteria recommended for aquatic life.

On August 21-22, lead levels were at or below the detection limit of $0.001 \mathrm{mg} / \mathrm{L}$ at most stations except Stations 5 and 7 where they averaged $0.003 \mathrm{mg} / \mathrm{L}$. In all cases in August, the levels present were well below the criteria recommended for drinking water and aquatic life.
4.2.7 Zinc. Table 5 compares results reported by previous surveys with that found in the present study.

Zinc at Station 5 has decreased over the period shown in Table 5. In July 1974, zinc was reported at $2.00 \mathrm{mg} / \mathrm{L}$, whereas in the present survey, mean zinc levels were $0.22 \mathrm{mg} / \mathrm{L}$ in July and $0.09 \mathrm{mg} / \mathrm{L}$ in August.

Mean zinc exceeded the $0.03 \mathrm{mg} / \mathrm{L}$ criteria recommended for aquatic life at Stations $1,2,5,7$ and 9 during the July sample period and at Stations l, 2, 4, 5, 7 and 9 during the August sample period. The highest
mean levels detected, $0.928 \mathrm{mg} / \mathrm{L}$ in July and $0.825 \mathrm{mg} / \mathrm{L}$ in August at Station l, exceeded the mine's water licence (Yukon Territory Water Board, 1985) requirement of $0.5 \mathrm{mg} / \mathrm{L}$. The high zinc levels found at Station loriginated at the Galkeno 900 adit where, on July 17,1985 it was $25.8 \mathrm{mg} / \mathrm{L}$ (personal communication, 1985). Background mean levels at Station 3 were $0.008 \mathrm{mg} / \mathrm{L}$ in July and $<0.002 \mathrm{mg} / \mathrm{L}$ in August.

## TABLE 5 HISTORICAL COMPARISON OF ZINC LEVELS IN VATER (mg/L)

|  | EPS | EPS | EPS | *EPS | EPS | EPS |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | JULY | JUNE | JULY | AUGUST | JULY 9-10 | AUG. 21-22. |
|  | 1974 | 1975 | 1975 | 1980 | 1985 | 1985 |
| STATI |  |  |  |  |  |  |
| 1 | $<0.17$ | 0.28 | 0.22 | N/A | 0.928 | 0.825 |
| 2 | N/A | N/A | 0.16 | N/A | 0.327 | 0.433 |
| 3 | 0.01 |  | 0.01 | N/A | 0.008 | $<0.002$ |
| 4 | N/A | N/A | N/A | N/A | 0.023 | 0.047 |
| 5 | 2.00 | 1.90 | 1.60 | 0.349 | 0.219 | 0.090 |
| 6 | 1.50 | N/A | N/A | N/A | 0.004 | **0.003 |
| 7 | 0.73 | 0.77 | 0.60 | 0.349 | 0.231 | 0.215 |
| 8 | 0.02 |  | 0.07 | 0.244 | 0.019 | 0.028 |
| 9 | 0.59 | 0.74 | 0.09 | 0.052 | 0.029 | 0.065 |
| 10 | N/A | N/A | N/A | N/A | 0.022 | 0.027 |
| 11 | N/A | N/A | N/A | N/A | 0.016 | 0.017 |

* Three day mean.
** One of three samples contained $0.003 \mathrm{mg} / \mathrm{L}$.
N/A - Not available

Although zinc was elevated at Station 5 during both sample periods, levels detected at Station 7 were higher, indicating, as have other parameters, that a source other than the tailings pond decant was influencing Flat Creek at the time of sampling. Levels at Station 6, upstream of Station 7, were well below that found at Stations 5 or 7.

During the August sample period, zinc levels were higher along the left bank of the South McQuesten River than at mid stream or along the right bank at Stations 4 and 9. This clearly identifies the plumes from Christal Creek and Flat Creek. The range at Station 4 was 0.084 to $0.003 \mathrm{mg} / \mathrm{L}$, from left to right bank, while at Station 9, the range was 0.106 to $0.038 \mathrm{mg} / \mathrm{L}$.

### 4.3 Stream Sediments

Particle size distribution (\%) and leachable metals results are presented in Appendix III, Tables l through 4. Size distribution data is available only for Stations 7 through ll. Particle size analysis was not carried out by the laboratory on samples from Station 1 through 4 due to misinterpreted instructions. Leachable metals analysis results are available for all stations sampled.
4.3.1 Particle Size Distribution. Overall the South McQuesten River sediments were observed to be comprised mainly of coarse material underlain with small amounts of sand and silt material. The particle size class most abundant in samples collected July 9-10 and August $21-22$ was in the gravel size and larger ( $>2.0 \mathrm{~mm}$ ) range. It was observed during sample collection that this range included material up to 40 mm in size, greatly influencing the overall weight distribution in each of the samples. Larger cobble material ( $>40 \mathrm{~mm}$ ) was present but was removed during sample collection. Sand material ( $<2.0 \mathrm{~mm}$ to $>0.063 \mathrm{~mm}$ ) represented from less than $1 \%$ to $16 \%$ of the composition of samples collected. The percentage of silt and clays ( $<0.063 \mathrm{~mm}$ ) ranged from less than $1 \%$ to $6 \%$.

Similar to South McQuesten River sediment samples, Flat Creek sediments at Station 7 were also comprised mainly of material in the $>2.0 \mathrm{~mm}$ size range with small percentages of sand between 0.25 mm and 1.0 mm in size.
4.3.2 Sediment Metal Analysis. Significant changes in certain sediment metals were detected between the two tributaries sampled, Christal Creek and Flat Creek, and the South McQuesten River. Figure 6 show mean Cd, $\mathrm{Cu}, \mathrm{Pb}, \mathrm{Zn}$ and As concentrations to be much higher at Stations 2 and 7 as compared with Station 3. Changes in sediment metals downstream of the tributaries was also detected as shown by the increase of the above metals at Station 4, downstream of Christal Creek, and more so at Station 9, downstream of Flat Creek. This can be explained, in part, by a tailings dam failure in August of 1978 which resulted in the deposition of significant




FIGURE 6 MEANS OF SELECTED METALS IN SEDIMENTS
amounts of tailings material into Flat Creek and the South McQuesten River downstream of Flat Creek. (Environmental Protection Service - Environmental Emergencies Significant Events Reports, 1978). However, the reasons for high metal concentrations in stream sediments at Station 2, Christal Creek, are unknown. It is speculated that mining activity in the Keno City area prior to the development of United Keno Hill Mines may have contributed but this cannot be substantiated.

### 4.4 Stream Benthic Pauna

Appendix IV, Table 1 shows the taxonomic classifications and distribution of invertebrates identified from samples collected in the present study.

Most invertebrates were keyed to genus and species where possible, or to genus or family level only, if full identification was not possible. Some individuals of the Orders Acari, Coleoptera, Tricoptera, Arachnida, Suborder Cyclopoida and Phylum Nematoda were not possible to identify beyond the taxonomic level given. Genera or species shown in brackets indicate that the identification was tentative.

Resh and Rosenberg (1984) review the ecology of aquatic insects and identify the difficulty of making generalizations about the relationship between aquatic invertebrates and substrate composition and the difficulty of generalizing about the effects of heavy metal pollution on aquatic insects.
4.4.1 Taxonomic Features. A total of 4,098 individuals, comprised of 67 different taxa, were collected from nine stations. The majority of the invertebrates collected were of the Class Insecta although several Nematodes, Molluscs and Copepods were also found. The artificial substrate samplers placed at Station 10 were vandalized, so the invertebrate information presented was collected August $21-22$ using a Surber sampler ( 0.93 m 2 ). No samples were collected at Station 5 , the tailings pond decant, or at Station 6 on Flat Creek.

The greatest number of individuals were found at Station $9(1,492)$ on the South McQuesten River. A total of 25 different taxa were identified of which 21 were keyed to the genus or species level. The most abundant organism, Simulium sp. represented $91 \%$ of the total number of individuals collected. Genera and species of Orders Plecoptera, Tricoptera and Ephemeroptera were also found but in low numbers.

Invertebrate abundance at other stations on the South McQuesten River varied considerably, ranging from 140 to 683 in number, but were much lower than Station 9.

At Station 3 where 683 individuals were collected, 21 of the 26 different taxa identified were keyed to the genus or species level. The dominant invertebrate was a Simulium sp. larvae (51\%) and pupae (25\%), representing $76 \%$ of the sample collected. The remaining sample included small percentages of organisms from Orders Ephemeroptera, Plecoptera, Tricoptera and Gastropoda and Phylum Nematoda.

At Station 4, immediately downstream of Christal Creek, 206 individuals were collected. A total of 27 different taxa were identified, 22 of them to the genus or species level. The dominant invertebrates were several genera of the 0rder Plecoptera representing $65 \%$ of the sample collected. They included Malenka sp. (18\%), Acroneuria sp. (16\%), Zapada sp. ( $11 \%$ ), Alloperla sp. ( $8 \%$ ), Utaperla sp. ( $6 \%$ ) , and Arcynopteryx sp . ( $2 \%$ ). The remaining sample was comprised of 2 genera of the Order Tricoptera (14\%) and 5 genera of the Order Ephemeroptera ( $10 \%$ ). Simulium sp. larvae and pupae, which were predominant at other South McQuesten River stations (except station 10 ), represented only $5 \%$ of the invertebrates collected at Station 4.

The lowest abundance was found at Station 8 where 140 individuals comprised of 20 different taxa were collected. Eighteen of the taxa were identified to the genus or species level. Simulium sp, of the Order Diptera was the most abundant, representing $31 \%$ of the sample. Members of
the Orders Ephemeroptera, Plecoptera and Tricoptera were also significant in numbers representing $16 \%, 26 \%$ and $23 \%$, respectively, of the sample.

Since a different sampling method was required at Station 10 the results will not be compared with those found at other South McQuesten River stations. At Station 10, a total of 237 individuals were collected of which 16 of the 24 different taxa identified were identified to the genus or species level. Density was calculated at 850 individuals per square metre. The dominant invertebrate was of the Class 0ligochaeta which represented $70 \%$ of the sample collected. The composition of the remaining sample was represented by invertebrates from the Orders Ephemeroptera, Plecoptera, Tricoptera, Diptera and the Phylum Nematoda.

At Station ll, Simulium sp. was the most abundant invertebrate found, making up $65 \%$ of the 500 individuals collected. The composition of the remaining sample was comprised of invertebrates from the orders Plecoptera (23\%), Ephemeroptera (5\%) and Tricoptera (3\%).

Invertebrate abundance varied considerably between the two stations on Christal Creek. At Station 1,655 individuals were collected whereas at Station 2, only 90 were collected. The reason for this difference in abundance is suspected to be partly because of the finer substate found at Station 2 which provided a less suitable habitat for invertebrates than the larger material found at Station 1 . This is based on observations of the sediment characteristics made during the first visit to the study area as there is no sediment particle size data for these two stations. At Station 1,31 different taxa were identified, 23 to the species or genus level. The sample was dominated by several genera of the Order Diptera. Those found included Simulium sp. (36\%), Cricotopus sp. (14\%), Eukiefferiella sp. (6\%) and Diplocladius sp. (9\%). Chironomidae pupae represented $5 \%$ of the sample. Individuals from the Class Oligochaeta represented $12 \%$ of the sample collected. Similar composition was found at Station 2 but, as previously stated, in much lower numbers. Only 18 different taxa were identified, 11 to the species or genus level. The sample collected was dominated by
several genera of the Order Diptera including Cricotopus sp. (21\%), Heterotrissocladius sp. (11\%), Cardiocladius sp. (2\%), Brillia sp. (17\%) and Chironomidae pupae (10\%).

At Station 7, 95 individuals were collected which were classified into 14 different taxa. Only 9 species or genera were identified. The dominant group was of the Order Diptera, representing $42 \%$ of the sample. The genera found included Simulium sp. ( $14 \%$ ), Cricotopus sp. (14\%), Heterotrissocladius sp. (5\%), Procladius sp. $1 \%$ ) and Tipula sp. (2\%). Chironomidae pupae ( $6 \%$ ) were also found. The most abundant genus, Podmosta sp., of the Order Plecoptera represented $34 \%$ of the sample. The genus Arctopsyche sp . of the order Tricoptera represented $11 \%$ of the sample.

The above results describe the aquatic invertebrate populations at the respective stations, and although there is a degree of similarity among the stations, it is not possible to isolate individual parameters as causes for population variation. Simulium sp. is noted as being present and very abundant at most stations. This genus is characteristic of clear, fast flowing water in riffle areas where it is successful as a filterer. The abundance of Simulium sp. at the various stations is also correlated with higher zinc ( Zn ) and cadmium ( Cd ) levels so it may be showing a tolerance to Zn and Cd which other organisms tolerate less readily. Similar tolerance to heavy metals is referred to by Wiederholm (1984).

### 4.4.2 Percent Similarity Index. The benthic invertebrate communities

 found at all South McQuesten River stations, except Station 10 , were compared using a Percent Similarity Index (PSC) formula described by Brock (1977):k

$$
P s c=100-0.5 \quad|a-b|
$$

where $a$ and $b$ are, for a given genus, percentages of the total samples $A$ and $B$ which that genus represents. The absolute value of their difference is summed over all genera, $k$. The PSC Index compares the percentage of genera present at two different locations but is not a comparison of total invertebrate abundance.

The greatest similarity was found between stations 9 and 11 where the index was $71 \%$. Similarity between Station 3 and Stations 4, 8, 9 and 11 ranged from $28 \%$ to $44 \%$. Similarity comparisons between all other stations never exceeded $48 \%$.

TABLE 6 INVERTEBRATE POPULATION PERCENT SIMILARITY FOR SOUTH McQUESTEN RIVER STATIONS
(EXCEPT STATION 10)

| STATION | 11 | 9 | 8 | 4 |
| :---: | :---: | :---: | :---: | :---: |
| 3 | 42 | 32 | 44 | 28 |
| 4 | 35 | 15 | 48 |  |
| 8 | 48 | 39 |  |  |

## SUMMARY

The present survey shows an improvement in mine effluent water quality at Station 5 when compared with historical data previously reported by the Environmental Protection Service. Several events, such as changes in ore type, mill processes and improved mill tailings treatment have taken place in the past 12 years and undoubtedly have influenced effluent quality but the degree of influence cannot be determined by this report.

Results of the present survey have shown that mine drainage from the Galkeno 900 adit is the primary source of high metals concentrations found in Christal Creek. The highest extractable Zn concentration found in the study area on the dates sampled was at Station 1 . This combined with other elevated metals concentrations exceed the standards recommended for drinking water and, in some cases, that recommended for the protection of aquatic life.

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## APPENDICES

## APPENDIX I

COLLECTION, PRESERVATION, ANALYSIS OR IDENTIFICATION METHODS AND WATER QUALITY CRITERIA
appendix I table 1

| PARAMETER | DETECTION <br> LIMIT | COLLECTION AND PRESERVATION <br> PROCEDURE |  | ANALYTICAL PROCEDURE |
| :--- | :--- | :--- | :--- | :--- |

APPENDIX I TABLE 1

| APPENDIX I | table 1 | WATER SAMPLE COLLECTION, pres | TION AND ANALYSIS METHODS (conti | ued) |
| :---: | :---: | :---: | :---: | :---: |
| PARAMETER | DETECTION LIMIT | COLLECTION AND PRESERVATION procedure | ANALYTICAL PROCEDURE | $\begin{aligned} & \text { METHOD } \\ & \text { SECTION } 2 \end{aligned}$ |
| Ammonia $\mathrm{NH}_{3}-\mathrm{N}$ | $0.005 \mathrm{mg} / 1$ | Single samples collected in 2 litre linear polyethylene containers. Each container was rinsed 3 times with sample before it was filled. No preservatives. stored at $4^{\circ} \mathrm{C}$. | Phenol hypochlorite-colori-metric-automated | 058 |
| Colour | 5 (colour units) | Same sample as $\mathrm{NH}_{3}$. | $\frac{\text { Platinum-cobalt visual compar- }}{\text { ison }}$ | 040 |
| Turbibity | 1.0 (FTU) | Same sample as $\mathrm{NH}_{3}$. | Nephelometric turbidity | 130 |
| Non-Filterable Residue (NFR) | $5.0 \mathrm{mg} / 1$ | Same sample as $\mathrm{NH}_{3}$. | Filtration, drying and weighing of residue on filter | 104 |
| Filterable Residue (FR) | $10.0 \mathrm{mg} / 1$ | Same sample as $\mathrm{NH}_{3}$. | Filtration, drying and weighting of filtrate | 100 |
| Total Alkalinity | $1.0 \mathrm{mg} / \mathrm{l}$ as $\mathrm{CaCO}_{3}$ | Same sample as $\mathrm{NH}_{3}$. | Potentiometric titration | 006 |
| Total Phosphate $\mathrm{TPO} 4^{-\mathrm{P}}$ | $0.005 \mathrm{mg} / 1$ | Same sample as $\mathrm{NH}_{3}$. | Ascorbic acid-persulphate, automated autoclave digestion | 086 |
| $\begin{aligned} & \text { Nitrate } \\ & \mathrm{NO}_{2}-\mathrm{N} \end{aligned}$ | $0.005 \mathrm{mg} / 1$ | Same sample as $\mathrm{NH}_{3}$. | $\xrightarrow{\text { Diazotization-colorimetric- }}$ | 070 |

WATER SAMPLE COLLECTION, PRESERVATION AND ANALYSIS METHODS (Continued)


APPENDIX

| PARAMETER | DETECTION LIMIT | COLLECTION AND PRESERVATION PROCEDURE ${ }^{1}$ | ANALYTICAL PROCEDURE $\quad \begin{gathered}\text { METHOD } \\ \text { SECTION }\end{gathered}$ |
| :---: | :---: | :---: | :---: |
| Cd | 0.0005 | Same as sample metals. | Graphite furnace-atomic absorption 330 |
| Cu | 0.001 | Same sample as metals. | spectrometry |
| Pb | 0.001 | Same sample as metals. |  |
| Ag | 0.0005 | Same sample as metals. | $\begin{aligned} & \text { Graphite furnace-atomic absorption } 330 \\ & \underline{\text { Spectrometry }} \end{aligned}$ |
| Total Hardness | $\begin{aligned} & 0.030 \mathrm{mg} / 1 \quad \text { Same sample as metals } \\ & 4.116 \mathrm{Mg}+2.497 \mathrm{Ca}+1.142 \mathrm{Sr}+1.792 \mathrm{Fe}+5.564 \mathrm{Al}+1.531 \mathrm{zn}+1.822 \mathrm{Mn} \end{aligned}$ |  |  |
| Actual Hardness $=$ |  |  |  |
| $\mathrm{Ca} / \mathrm{Mg}$ Hardness $=$ | $4.116 \mathrm{Mg}+2.497 \mathrm{Ca}$ |  |  |
| As described in Environment Canada (1976). <br> As described in Department of Environment (1979). |  |  |  |


| APPENDIX I | TABLE 3 |  |
| :--- | :--- | :--- |
| FIELD COLLECTION, SAMPLING |  |  |
| PROCEDURES AND PRESERVATION |  |  |

APPENDIX I

APPENDIX I TABLE 4 WATER QUALITY CRITERIA FOR DRINKING WATER AND AQUATIC LIFE (continued)

APPENDIX I TABLE 4 WATER QUALITY CRITERIA FOR DRINKING WATER AND AQUATIC LIFE (continued)

appendix I table 4 water quality criteria for drinking water and aquatic life

| SUBStance | RECOMMENDED LEVEL(S) FOR DRINKING WATER | Reference (S) | recommended cevel(s) FOR AQUATIC LIFE | REFERENCE (S) |
| :---: | :---: | :---: | :---: | :---: |
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## APPENDIX II

WATER QUALITY
APPENDIX II TABLE 1 Water quality results for july 9-10, 1985.

| STATION | SAMPLE NUMBER | AVG. DEPTH (m) | WIDTH <br> ( m ) | $\begin{gathered} \text { FLOW } \\ (\mathrm{m} 3 / \mathrm{s}) \end{gathered}$ | TEMP. <br> (C) | INSITU | pH <br> LAB |  | COND. os/cm) LAB | DISOLVED OXYGEN (mg/L) | $\begin{aligned} & \text { \%DO } \\ & \text { SATURA- } \\ & \text { TION } \end{aligned}$ | $\begin{aligned} & \text { COLOR } \\ & \text { (FTU) } \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 0.2 | 1.8 | 0.065 | 10.1 | 7.91 | 7.8 | 457 | 750 | 9.7 | 99 | 10 |
|  | 2 |  |  |  |  |  | 7.9 |  | $\begin{aligned} & 750 \\ & 750 \end{aligned}$ |  |  | 10 |
|  | 3 |  |  |  |  |  | 7.9 |  |  |  |  | 10 |
|  | $\bar{x}$ |  |  |  |  |  | 7.9 |  | 750 |  |  | 10 |
|  | S.D. |  |  |  |  |  | 0.1 |  | 0 |  |  | 0 |
| 2 | 4 | 0.4 | 4.5 | 0.12 | 9.5 | 8.23 | 8.2 | 348 | 550 | 10.2 | 94 | 10 |
|  | 5 |  |  |  |  |  | 8.2 |  | 550 |  |  | 10 |
|  | 6 |  |  |  |  |  | 8.1 |  | 550 |  |  | 10 |
|  | $\overline{\mathrm{x}}$ |  |  |  |  |  | 8.2 |  | 550 |  |  | 10 |
|  | S.D. |  |  |  |  |  | 0.1 |  | 0 |  |  | 0 |
| 3 | 7 | 0.4 | 12.8 | 5.9 | 16.0 | 8.40 | 8.0 | 170 | 240 | 8.3 | 86 | 20 |
|  | 8 |  |  |  |  |  | 8.0 |  | $240$ |  |  | 20 |
|  | 9 |  |  |  |  |  | 8.0 |  |  |  |  | 20 |
|  | $\overline{\mathrm{x}}$ |  |  |  |  |  | 8.0 |  | 240 |  |  | 10 |
|  | S.D. |  |  |  |  |  | 0 |  | 0 |  |  | 0 |
| 4 | 10 | N/A | N/A | 6.0 | 16.0 | 8.30 | 8.0 | 180 | 250 | 8.6 | 96 | 20 |
|  | 11 |  |  |  |  |  | 8.1 |  | 255 |  |  | 20 |
|  | 12 |  |  |  |  |  | 8.1 |  | 250 |  |  | 20 |
|  | $\overline{\mathrm{x}}$ |  |  |  |  |  | 8.1 |  | 252 |  |  | 20 |
|  | S.D. |  |  |  |  |  | 0.1 |  | 3 |  |  | 0 |

* Determined from the sum of Station 3 and Station 2.
appendix il table 1 Water quality results for july 9-10, 1985 (continued)

APPENDIX II TABLE 1 WATER QUALITY RESULTS FOR JULY 9-10, 1985 (continued)

| STATION | SAMPLE NUMBER | AVG. DEPTH <br> (m) | WIDIH (m) | $\begin{aligned} & \text { FLOW } \\ & (\mathrm{m} 3 / \mathrm{s}) \end{aligned}$ | TEMP. <br> (C) | pH |  | $\begin{gathered} \text { COND. } \\ \text { (umhos/cm) } \end{gathered}$ |  | $\begin{aligned} & \text { DISOLVED } \\ & \text { OXYGEN } \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\begin{aligned} & \text { \$DO } \\ & \text { SATURA- } \\ & \text { TION } \end{aligned}$ | COLOR <br> (FTU) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  |  | 15.0 | 8.3 | 15.0 | 8.06 | 8.0 | 250 | 335 | 8.0 | 84 | 20 |
| 9 | 25 | 0.5 | 15.0 | 8.3 | 15.0 | 8.06 | 8.1 |  | 255 |  |  | 20 |
|  | 26 27 |  |  |  |  |  | 8.2 |  | 255 |  |  | 20 |
| 10 | $\bar{z}$ |  |  |  |  |  | 8.1 |  | 255 |  |  | 20 |
|  | S |  |  |  |  |  | 0.1 |  | 46 |  |  | 0 |
|  | 28 | 0.4 | 19.9 | 8.1 | 15.0 | 7.95 | 8.1 | 190 | 270 | 9.3 | 82 | 20 |
|  | 29 |  |  |  |  |  | 8.1 |  | 270 |  |  | 20 |
|  | 30 |  |  |  |  |  | 8.1 |  | 270 |  |  | 2 |
| 11 |  |  |  |  |  |  | 8.1 |  | 270 |  |  | 20 |
|  | S |  |  |  |  |  | 0.0 |  | 0 |  |  | 0 |
|  | 31 | 0.5 | 30.0 | 9.7 | 13.5 | 8.10 | 8.2 | 185 | 275 | 9.3 | 92 | 20 |
|  | 32 |  |  |  |  |  | 8.1 |  | 275 |  |  | 20 |
|  | 33 |  |  |  |  |  | 8.1 |  | 280 |  |  |  |
|  |  |  |  |  |  |  | 8.1 |  | 270 |  |  | 20 |
|  | x |  |  |  |  |  | 0.0 |  | 0 |  |  | 0 |

appendix II table 1 Water quality results for July 9-10, 1985

| Station | SAMPLE NUMBER | TURB. (FTU) | $\begin{gathered} \text { T.ALK. } \\ \text { (as CACO3) } \\ \text { (mg/L) } \end{gathered}$ | $\begin{aligned} & \text { T. HARD } \\ & \text { (as CaCO3) } \\ & \text { (mg/ } \mathrm{L} \text { ) } \end{aligned}$ | SULFATE (mg/L) | $\begin{aligned} & \text { CHLORIDE } \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\begin{gathered} \text { PHOSPHATE } \\ \text { ( } \mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{aligned} & \text { NITRITE } \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | NITRATE (mg/L) | AMMONIA (mg/L) | $\begin{aligned} & \mathrm{NFR} \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 1.30 | 80.3 | 419 | 280 | 0.9 | 0.012 | $<0.005$ | 0.037 | 0.011 | 9 |
|  | 2 | 1.30 | 80.7 | 425 | 290 | 0.9 | 0.022 | <0.005 | 0.037 | 0.008 | 9 |
|  | 3 | 1.30 | 87.8 | 451 | 240 | 0.8 | 0.014 | <0.005 | 0.035 | 0.009 | 7 |
|  | $\bar{x}$ | 1.30 | 82.9 | 432 | 270 | 0.9 | 0.016 | 0.005 | 0.036 | 0.009 | 8 |
|  | s.D. | 0.00 | 4.2 | 17 | 26 | 0.1 | 0.005 | 0.0 | 0.001 | 0.002 | 1 |
| 2 | 4 | 0.63 | 110.0 | 309 | 160 | 0.7 | 0.002 | $<0.005$ | 0.018 | 0.005 | <5 |
|  | 5 | 0.53 | 111.0 | 314 | 140 | 0.7 | 0.006 | $<0.005$ | 0.017 | 0.005 | <5 |
|  | 6 | 0.53 | 111.0 | 317 | 160 | 0.7 | $<0.002$ | <0.005 | 0.017 | 0.005 | <5 |
|  | $\overline{\mathrm{x}}$ | 0.56 | 110.7 | 313 | 153 | 0.7 | <0.003 | $<0.005$ | 0.017 | 0.000 | <5 |
|  | s.d. | 0.06 | 0.6 | 4 | 12 | 0.0 | 0.000 | 0.0 | 0.001 | 0.000 | 0 |
| 3 | 7 | 0.45 | 85.7 | 132 | 30 | 0.7 | 0.007 | $<0.005$ | 0.011 | 0.009 | <5 |
|  | 8 | 0.43 | 85.7 | 132 | 30 | 0.8 | 0.007 | $<0.005$ | 0.010 | 0.009 | <5 |
|  | 9 | 0.48 | 85.7 | 133 | 30 | 0.7 | 0.007 | <0.005 | 0.010 | 0.010 | <5 |
|  | $\overline{\mathrm{x}}$ | 0.45 | 85.7 | 132 | 30 | 0.7 | 0.007 | $<0.005$ | 0.010 | 0.009 | <5 |
|  | s.D. | 0.03 | 0.0 |  | 0 | 0.1 | 0.000 | 0.0 | 0.001 | 0.001 | 0 |
| 4 | 10 | 0.48 | 85.7 | 141 | 40 | 0.8 | 0.010 | <0.005 | 0.010 | 0.007 | <5 |
|  | 11 | 0.45 | 86.4 | 143 | 30 | 0.8 | 0.010 | $<0.005$ | 0.011 | 0.008 | <5 |
|  | 12 | 0.45 | 86.4 | 139 | 30 | 0.9 | 0.008 | $<0.005$ | 0.010 | 0.008 | <5 |
|  | $\overline{\mathbf{x}}$ | 0.46 | 86.2 | 141 | 33 | 0.8 | 0.009 | $<0.005$ | 0.010 | 0.008 | < 5 |
|  | s.D. | 0.02 | 0.4 | 2 | 6 | 0.1 | 0.001 | 0.0 | 0.001 | 0.001 | 0 |

appendix if table 1 Water quality results for july 9-10, 1985 (continued)

| Station | SAMPLE NUMBER | $\begin{aligned} & \text { TURB. } \\ & \text { (FTU) } \end{aligned}$ | $\begin{aligned} & \text { T.ALK. } \\ & \text { (as CAC03) } \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\begin{aligned} & \text { T.HARD } \\ & \text { (as CaCO3) } \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\begin{aligned} & \text { SULIFATE } \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | CHLDRIDE (mg/L) | $\underset{(\mathrm{mg} / \mathrm{L})}{\text { PHOSPATE }}$ | $\begin{aligned} & \text { NITRITE } \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | nITRATE ( $\mathrm{mg} / \mathrm{L}$ ) | $\begin{aligned} & \text { AMMONLA } \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\begin{gathered} \mathrm{NFR} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 13 | 2.30 | 93.5 | 592 | 480 | 18.7 | 0.007 | $<0.005$ | 0.060 | 0.126 | <5 |
|  | 14 | 2.30 | 92.8 | 583 | 510 | 19.0 | 0.006 | $<0.005$ | 0.064 | 0.130 | <5 |
|  | 15 | 2.30 | 92.8 | 599 | 500 | 18.9 | 0.007 | <0.005 | 0.065 | 0.136 | <5 |
|  | $\overline{\mathrm{x}}$ | 2.30 | 93.0 | 591 | 497 | 18.9 | 0.007 | <0.005 | 0.063 | 0.131 | < 5 |
|  | s.D. | 0.00 | 0.4 | 8 | 15 | 0.2 | 0.001 | 0.0 | 0.003 | 0.005 | 0 |
| 6 | 16 | 0.15 | 168.0 | 242 | 40 | 0.5 | 0.005 | <0.005 | 0.070 | 0.005 | <5 |
|  | 17 | 0.15 | 168.0 | 237 | 50 | 0.7 | 0.005 | $<0.005$ | 0.073 | 0.005 | $<5$ |
|  | 18 | 0.15 | 169.0 | 236 | 50 | 0.7 | 0.005 | <0.005 | 0.071 | 0.005 | <5 |
|  | $\overline{\mathrm{x}}$ | 0.15 | 168.3 | 238 | 47 | 0.7 | 0.005 | <0.005 | 0.071 | 0.000 | < 5 |
|  | s.D. | 0.00 | 0.6 | 3 | 6 | 0.1 | 0.000 | 0.0 | 0.002 | 0.000 | 0 |
| 7 | 19 |  |  |  | 300 | 5.1 | 0.004 | $<0.005$ | 0.087 | 0.013 | <5 |
|  | 20 | 2.50 | 142.0 | 490 | 310 | 5.1 | 0.007 | <0.005 | 0.080 | 0.015 | <5 |
|  | 21 | 2.80 | 142.0 | 499 | 310 | 4.7 | 0.004 | <0.005 | 0.088 | 0.010 | <5 |
|  | $\stackrel{\rightharpoonup}{\mathbf{x}}$ | 2.60 | 143.3 | 494 | 307 | 5.0 | 0.005 | <0.005 | 0.088 | 0.013 | < 5 |
|  | s.D. | 0.17 | 2.3 | 5 | 6 | 0.2 | 0.002 | 0.0 | 0.001 | 0.003 | 0 |
| 8 | 22 | 1.30 | 90.0 | 140 | 40 | 0.7 | 0.007 | <0.005 | 0.016 | 0.011 | <5 |
|  | 23 | 1.00 | 89.0 | 140 | 40 | 0.7 | 0.008 | <0.005 | 0.013 | 0.011 | <5 |
|  | 24 | 0.55 | 90.0 | 140 | 30 | 0.7 | 0.008 | <0.005 | 0.013 | 0.012 | <5 |
|  | $\overline{\mathrm{x}}$ | 0.95 | 89.7 | 140 | 37 | 0.7 | 0.008 | <0.005 | 0.014 | 0.011 | <5 |
|  | S.D. | 0.38 | 0.6 | 0 | 6 | 0.0 | 0.001 | 0.0 | 0.002 | 0.001 | 0 |

APPENDIX II TABLE 1 WATER QUALITY RESULTS FOR JULY 9-10, 1985 (continued)

| station | SAMPLE NUMBER | $\begin{aligned} & \text { TURB. } \\ & \text { (FTU). } \end{aligned}$ | $\begin{aligned} & \text { T.ALK. } \\ & \text { (as CACO3) } \\ & \text { (mg/L) } \end{aligned}$ | $\begin{gathered} \text { T.HARD } \\ (\mathrm{as} \mathrm{caco3}) \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | sulfate <br> ( $\mathrm{mg} / \mathrm{L}$ ) | $\begin{aligned} & \text { CHLORIDE } \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | phospate ( $\mathrm{mg} / \mathrm{L}$ ) | $\begin{aligned} & \text { NITRITE } \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\begin{gathered} \text { NITRATE } \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \text { AMMONLA } \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{NFR} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 13 | 2.30 | 93.5 | 592 | 480 | 18.7 | 0.007 | $<0.005$ | 0.060 | 0.126 | <5 |
|  | 14 | 2.30 | 92.8 | 583 | 510 | 19.0 | 0.006 | <0.005 | 0.064 | 0.130 | <5 |
|  | 15 | 2.30 | 92.8 | 599 | 500 | 18.9 | 0.007 | <0.005 | 0.065 | 0.136 | < 5 |
|  | $\overline{\mathrm{x}}$ | 2.30 | 93.0 | 591 | 497 | 18.9 | 0.007 | <0.005 | 0.063 | 0.131 | < 5 |
|  | s.d. | 0.00 | 0.4 |  | 15 | 0.2 | 0.001 | 0.0 | 0.003 | 0.005 | 0 |
| 6 | 16 | 0.15 | 168.0 | 242 | 40 | 0.5 | 0.005 | <0.005 | 0.070 | 0.005 | <5 |
|  | 17 | 0.15 | 168.0 | 237 | 50 | 0.7 | 0.005 | <0.005 | 0.073 | 0.005 | <5 |
|  | 18 | 0.15 | 169.0 | 236 | 50 | 0.7 | 0.005 | <0.005 | 0.071 | 0.005 | <5 |
|  | $\overline{\mathrm{x}}$ | 0.15 | 168.3 | 238 | 47 | 0.7 | 0.005 | <0.005 | 0.071 | 0.000 | < |
|  | S.D. | 0.00 | 0.6 | , | 6 | 0.1 | 0.000 | 0.0 | 0.002 | 0.000 |  |
| 7 | 19 | 2.50 | 146.0 | 492 | 300 | 5.1 | 0.004 | <0.005 | 0.087 | 0.013 | <5 |
|  | 20 | 2.50 | 142.0 | 490 | 310 | 5.1 | 0.007 | <0.005 | 0.080 | 0.015 | <5 |
|  | 21 | 2.80 | 142.0 | 499 | 310 | 4.7 | 0.004 | <0.005 | 0.088 | 0.010 | < |
|  | $\overline{\mathrm{x}}$ | 2.60 | 143.3 | 494 | 307 | 5.0 | 0.005 | <0.005 | 0.088 | 0.013 | < 5 |
|  | s.d. | 0.17 | 2.3 | 5 | 6 | 0.2 | 0.002 | 0.0 | 0.001 | 0.003 | 0 |
| 8 | 22 | 1.30 | 90.0 | 140 | 40 | 0.7 | 0.007 | <0.005 | 0.016 | 0.011 | <5 |
|  | 23 | 1.00 | 89.0 | 140 | 40 | 0.7 | 0.008 | <0.005 | 0.013 | 0.011 | <5 |
|  | 24 | 0.55 | 90.0 | 140 | 30 | 0.7 | 0.008 | <0.005 | 0.013 | 0.012 | <5 |
|  | $\overline{\mathrm{x}}$ | 0.95 | 89.7 | 140 | 37 | 0.7 | 0.008 | <0.005 | 0.014 | 0.011 | <5 |
|  | s.D. | 0.38 | 0.6 | - | 6 | 0.0 | 0.001 | 0.0 | 0.002 | 0.001 | 0 |

appendix If table 1 WATER quality results for goly 9-10, 1985 (continued)

| station | SAMPLE | $\underset{\text { (FTU) }}{\text { TURB }}$ | $\begin{gathered} \text { T.ALK. } \\ (\mathrm{as} \mathrm{CAco3}) \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | T. HARD (as Caco3) (mg/L) | $\underset{(\mathrm{mg} / \mathrm{L})}{\text { SULFATE }}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\text { CHLRID }}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\text { phospare }}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{NITRIE}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\operatorname{NITRTE}}$ | $\begin{aligned} & \text { AMMONLA } \\ & (\mathrm{mg} / \mathrm{L} \end{aligned}$ | $\begin{aligned} & \mathrm{NFR} \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 9 | 25 | 0.68 | 97.1 | 186 | 60 | 1.2 | 0.006 | <0.005 | 0.021 | 0.015 | <5 |
|  | 26 | 0.55 | 90.0 | 144 | 40 | 0.6 | 0.007 | <0.005 | 0.014 | 0.013 | <5 |
|  | 27 | 0.48 | 89.3 | 143 | 20 | 0.5 | 0.007 | <0.005 | 0.013 | 0.013 | <5 |
|  | $\overline{\mathrm{x}}$ | 0.57 | 92.1 | 158 | 40 | 0.8 | 0.007 | <0.005 | 0.016 | 0.014 | <5 |
|  | S.D. | 0.10 | 4.3 | 25 | 20 | 0.4 | 0.001 | 0.0 | 0.004 | 0.001 | 0.0 |
| 10 | 28 | 0.53 | 92.1 | 148 | 30 | 0.6 | 0.007 | <0.005 | 0.015 | 0.012 | <5 |
|  | 29 | 0.53 | 93.5 | 148 | 30 | 0.9 | 0.007 | <0.005 | 0.015 | 0.012 | <5 |
|  | 30 | 0.48 | 93.5 | 149 | 40 | 1.0 | 0.009 | <0.005 | 0.015 | 0.013 | < |
|  | $\overline{\mathrm{x}}$ | 0.51 | 93.0 | 148 | 33 | 0.8 | 0.008 | <0.005 | 0.015 | 0.012 | <5 |
|  | s.d. | 0.03 | 0.8 | 1 | 6 | 0.2 | 0.001 | 0.0 | . 000 | 0.001 | 0.0 |
| 11 | 31 | 0.63 | 95.7 | 157 | 37 | 1.0 | 0.007 | <0.005 | 0.017 | 0.010 | <5 |
|  | 32 | 0.68 | 95.7 | 153 | 38 | 1.0 | 0.007 | <0.005 | 0.017 | 0.008 | <5 |
|  | 33 | 0.65 | 98.5 | 156 | 36 | 1.1 | 0.009 | <0.005 | 0.016 | 0.010 | <5 |
|  | S.D. | 0.65 0.03 | 96.6 1.6 | 155 2 | 37 1 | 1.0 0.1 | 0.008 0.001 | $<0.005$ 0.0 | 0.017 0.001 | 0.009 0.001 | 0.5 |

appendix II table 1 Water quality results for July 9-10, 1985

| station | $\underset{\text { SAMPLE }}{\text { NMBER }}$ | $\begin{gathered} \mathrm{Ag} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{As} \\ (\mathrm{mq} / \mathrm{L}) \end{gathered}$ | $\begin{aligned} & \mathrm{B} \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\begin{aligned} & \begin{array}{l} \mathrm{Ba} \\ (\mathrm{mg} / \mathrm{L}) \end{array} \end{aligned}$ | $\begin{aligned} & \left.\begin{array}{l} \mathrm{Be} \\ (\mathrm{mg} / \mathrm{L}) \end{array}\right) \end{aligned}$ | $\begin{aligned} & \begin{array}{l} \mathrm{Ca} \\ (\mathrm{mg} / \mathrm{L}) \end{array} \end{aligned}$ | $\begin{aligned} & \text { cd } \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\begin{aligned} & c_{(m g / L)}^{c o s} \end{aligned}$ | $\begin{aligned} & \hline \mathrm{Cr} \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{cu}}$ | $\begin{aligned} & \mathrm{Fe} \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Mg}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\substack{\mathrm{Mn}}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | ¢0.0005 | 80.05 | 0.015 | 0.045 | ${ }^{2} 0.001$ | 134.0 | 0.0037 | $\bigcirc 0.005$ | <0.005 | 0.002 | 0.381 | 19.0 | 1.840 |
|  | 2 | ¢0.0005 | <0.05 | 0.000 | 0.045 | <0.001 | 136.0 | 0.0031 | <0.005 | <0.005 | 0.003 | 0.383 | 19.2 | 1.850 |
|  | 3 | <0.0005 | <0.05 | 0.009 | 0.049 | <0.001 | 144.0 | 0.0029 | <0.005 | <0.005 | 0.004 | 0.416 | 20.8 | 1.970 |
|  | x | <0.0005 | ${ }^{0} 0.05$ | 0.008 | 0.046 | <0.001 | 138.0 | 0.0032 | <0.005 | <0.005 | 0.003 | 0.393 | 19.7 | 1.887 |
|  | s.d. | 0.0 | 0.0 | 0.008 | 0.002 | 0.0 | 5.3 | 0.0004 | 0.0 | 0.0 | 0.001 | 0.020 | 1.0 | 0.072 |
| 2 | 4 | <0.0005 | <0.05 | 0.006 | 0.043 | <0.001 | 95.0 | 0.0029 | <0.005 | <0.005 | 0.002 | 0.089 | 17.4 | 0.087 |
|  | 5 | <0.0005 | <0.05 | 0.011 | 0.044 | <0.001 | 96.2 | 0.0031 | <0.005 | <0.005 | 0.002 | 0.091 | 17.8 | 0.086 |
|  | 6 | <0.0005 | <0.05 | 0.014 | 0.044 | <0.001 | 96.9 | 0.0031 | <0.005 | <0.005 | 0.002 | 0.086 | 18.0 | 0.086 |
|  | $\bar{x}$ | <0.0005 | $<0.05$ | 0.010 | 0.044 | <0.001 | 96.0 | 0.0030 | <0.005 | <0.005 | 0.002 | 0.089 | 17.7 | 0.086 |
|  | s.d. | 0.0 | 0.0 | 0.004 | 0.001 | 0.0 | 1.0 | 0.0001 | 0.0 | 0.0 | 0.000 | 0.000 | 0.3 | 0.001 |
| 3 | 7 | <0.0005 | <0.05 | 0.020 | 0.044 | <0.001 | 36.8 | 0.0018 | <0.005 | <0.005 | 0.003 | 0.182 | 9.6 | 0.030 |
|  | 8 | ¢0.0005 | <0.05 | 0.002 | 0.044 | <0.001 | 36.7 | 0.0016 | ¢0.005 | co.005 | 0.003 | 0.228 | 9.7 | 0.050 |
|  | 9 | <0.0005 | <0.05 | 0.011 | 0.044 | <0.001 | 36.9 | 0.0009 | <0.005 | <0.005 | 0.002 | 0.186 | 9.8 | 0.030 |
|  | $\overline{\mathrm{x}}$ | ¢0.0005 | <0.05 | 0.011 | 0.044 | <0.001 | 36.8 | 0.0014 | <0.005 | <0.005 | 0.003 | 0.199 | 9.7 | 0.07 |
|  | S.D. | 0.0 | 0.0 | 0.009 | 0.000 | 0.0 | 0.1 | 0.0005 | 0.0 | 0.0 | 0.001 | 0.025 | 0.1 | 0.012 |
| 4 | 10 | <0.0005 | <0.05 | 0.014 | 0.045 | <0.001 | 39.3 | 0.0020 | <0.005 | <0.005 | 0.003 | 0.223 | 10.1 | 0.041 |
|  | 11 | <0.0005 | <0.05 | 0.007 | 0.045 | <0.001 | 40.1 | 0.0015 | <0.005 | <0.005 | 0.003 | 0.218 | 10.2 | 0.041 |
|  | 12 | <0.0005 | <0.05 | 0.015 | 0.045 | <0.001 | 39.0 | 0.0018 | <0.005 | <0.005 | 0.003 | 0.89 | 10.0 | 0.035 |
|  | $\bar{s}$ | <0.0005 | <0.05 | 0.012 | 0.045 | <0.001 | 39.5 | 0.0018 | <0.005 | <0.005 | 0.003 | 0.10 | 10.1 | 0.039 |
|  | s.d. | 0.0 | 0.0 | 0.004 | 0.000 | 0.0 | 0.6 | 0.0003 | 0.0 | 0.0 | 0.00 | 0.018 | 0.1 | 0.003 |

appendix in table 1

| station | SAMPLE NUMBER | $\begin{gathered} \mathrm{Ag} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | ${ }_{(\mathrm{mg} / \mathrm{L})}^{\mathrm{As}}$ | $\underset{(\operatorname{mg} / \mathrm{L})}{\mathrm{B}}$ | $\begin{gathered} \mathrm{Ba} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{aligned} & \mathrm{Be} \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\begin{gathered} \mathrm{Ca}_{(\mathrm{mg} / \mathrm{L})} \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{cd}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{cos}^{2}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{cr}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{cu}}$ | $\begin{gathered} \mathrm{Fe} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Mg}}$ | $\begin{gathered} \mathrm{Mn} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 13 | 0.0022 | $<0.05$ | 0.026 | 0.026 | <0.001 | 161.0 | 0.0017 | <0.005 | <0.005 | 0.012 | 0.384 | 45.8 | 0.563 |
|  | 14 | 0.0026 | <0.05 | 0.028 | 0.028 | <0.001 | 158.0 | 0.0030 | 0.007 | <0.005 | 0.013 | 0.652 | 44.6 | 0.863 |
|  | 15 | 0.0023 | <0.05 | 0.027 | 0.027 | <0.001 | 163.0 | 0.0025 | <0.005 | <0.005 | 0.012 | 0.529 | 46.3 | 0.740 |
|  | $\times$ | 0.0024 | <0.05 | 0.027 | 0.027 | <0.001 | 160.7 | 0.0024 | <0.007 | <0.005 | 0.012 | 0.522 | 45.6 | 0.722 |
|  | s.d. | 0.0002 | 0.0 | 0.001 | 0.001 | 0.0 | 2.5 | 0.0007 | 0.0 | 0.0 | 0.001 | 0.134 | 0.9 | 0.151 |
| 6 | 16 | <0.0005 | $<0.05$ | 0.075 | 0.075 | <0.001 | 75.8 | 0.0017 | <0.005 | <0.005 | 0.003 | 0.070 | 12.8 | 0.003 |
|  | 17 | <0.0005 | <0.05 | 0.027 | 0.074 | <0.001 | 74.2 | 0.0018 | <0.005 | <0.005 | 0.003 | 0.069 | 12.5 | 0.003 |
|  | 18 | <0.0005 | <0.05 | 0.012 | 0.072 | <0.001 | 74.0 | 0.0016 | <0.005 | <0.005 | 0.005 | 0.066 | 12.2 | 0.002 |
|  | $\overline{\mathrm{x}}$ | <0.0005 | $<0.05$ | 0.038 | 0.074 | <0.001 | 74.7 | 0.0017 | <0.005 | <0.005 | 0.004 | 0.068 | 12.5 | 0.003 |
|  | s.d. | 0.00 | 0.0 | 0.033 | 0.002 | 0.0 | 1.0 | 0.0001 | 0.0 | 0.0 | 0.001 | 0.002 | 0.3 | 0.001 |
| 7 | 19 | 0.0008 | <0.05 | 0.031 | 0.032 | <0.001 | 140.0 | 0.0044 | <0.005 | <0.005 | 0.006 | 0.300 | 33.6 | 2.090 |
|  | 20 | 0.0006 | <0.05 | 0.013 | 0.032 | <0.001 | 139.0 | 0.0043 | <0.005 | <0.005 | 0.006 | 0.295 | 33.3 | 2.080 |
|  | 21 | 0.0007 | $<0.05$ | 0.000 | 0.033 | <0.001 | 142.0 | 0.0042 | <0.005 | <0.005 | 0.005 | 0.305 | 34.2 | 2.120 |
|  | $\overline{\mathrm{x}}$ | 0.0007 | $<0.05$ | 0.015 | 0.032 | 0.001 | 140.3 | 0.0043 | <0.005 | <0.005 | 0.006 | 0.300 | 33.7 | 2.097 |
|  | s.d. | 0.0001 | 0.0 | 0.016 | 0.001 | 0.0 | 1.5 | 0.0001 | 0.0 | 0.0 | 0.001 | 0.005 | 0.5 | 0.021 |
| 8 | 22 | <0.0005 | $<0.05$ | 0.006 | 0.050 | <0.001 | 38.9 | 0.0018 | <0.005 | <0.005 | 0.005 | 2.209 | 10.2 | 0.051 |
|  | 23 | <0.0005 | <0.05 | 0.012 | 0.050 | <0.001 | 39.1 | 0.0018 | <0.005 | <0.005 | 0.002 | 0.206 | 10.2 | 0.048 |
|  | 24 | <0.0005 | <0.05 | 0.015 | 0.050 | <0.001 | 38.8 | 0.0020 | <0.005 | <0.005 | 0.003 | 0.219 | 10.3 | 0.051 |
|  | , | <0.0005 | $<0.05$ | 0.011 | 0.050 | <0.001 | 38.9 | 0.0019 | <0.005 | <0.005 | 0.003 | 0.211 | 10.2 | 0.050 |
|  | s.d. | 0.0 | 0.0 | 0.005 | 0.0 | 0.0 | 0.2 | 0.0001 | 0.0 | 0.0 | 0.002 | 0.007 | 0.1 | 0.002 |

APPENDIX II


| Station | SAMPLE NUMBER | $\begin{gathered} \mathrm{Ag} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{As} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{B} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Ba} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Be} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Ca} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \hline \mathrm{Cd} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Co}_{0} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \hline \mathrm{Cr} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Cu} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Fe} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Mg} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Mn} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 9 | 25 | 0.0005 | <0.05 | 0.019 | 0.051 | <0.001 | 52.0 | 0.0019 | $<0.005$ | $<0.005$ | 0.002 | 0.207 | 13.3 | 0.227 |
|  | 26 | 0.0005 | <0.05 | 0.009 | 0.050 | <0.001 | 40.0 | 0.0016 | <0.005 | <0.005 | 0.001 | 0.202 | 10.5 | 0.057 |
|  | 27 | <0.0005 | <0.05 | 0.020 | 0.050 | <0.001 | 39.8 | 0.0019 | <0.005 | <0.005 | 0.002 | 0.217 | 10.5 | 0.057 |
|  | $\overline{\mathrm{x}}$ | <0.0005 | <0.05 | 0.016 | 0.050 | <0.001 | 43.9 | 0.0018 | <0.005 | <0.005 | 0.002 | 0.209 | 11.4 | 0.113 |
|  | s.D. | 0.0 | 0.0 | 0.006 | 0.001 | 0.0 | 7.0 | 0.0002 | 0.0 | 0.0 | 0.001 | 0.008 | 1.6 | 0.099 |
| 10 | 28 | <0.0005 | <0.05 | 0.030 | 0.052 | <0.001 | 41.3 | 0.0019 | <0.005 | <0.005 | 0.002 | 0.202 | 10.9 | 0.067 |
|  | 29 | <0.0005 | <0.05 | 0.018 | 0.052 | <0.001 | 41.2 | 0.0019 | <0.005 | <0.005 | 0.002 | 0.202 | 10.9 | 0.067 |
|  | 30 | $<0.0005$ | <0.05 | 0.000 | 0.052 | <0.001 | 41.3 | 0.0020 | <0.005 | <0.005 | 0.002 | 0.202 | 10.9 | 0.065 |
| 11 | $\overline{\mathbf{x}}$ | <0.0005 | $<0.05$ | 0.016 | 0.052 | <0.001 | 41.3 | 0.0019 | <0.005 | $<0.005$ | 0.002 | 0.202 | 10.9 | 0.066 |
|  | S.D. | 0.0 | 0.0 | 0.015 | . 000 | 0.0 | 0.1 | 0.0001 | 0.0 | 0.0 | 0.0 | 0.000 | 0.0 | 0.001 |
|  | 31 | $<0.0005$ | $<0.05$ | 0.012 | 0.056 | <0.001 | 43.2 | 0.0012 | <0.005 | <0.005 | 0.002 | 0.226 | 11.6 | 0.071 |
|  | 32 | <0.0005 | <0.05 | 0.026 | 0.056 | <0.001 | 42.3 | 0.0016 | <0.005 | <0.005 | 0.003 | 0.269 | 11.4 | 0.072 |
|  | 33 | <0.0005 | <0.05 | 0.000 | 0.058 | <0.001 | 43.2 | 0.0016 | <0.005 | <0.005 | 0.003 | 0.240 | 11.4 | 0.078 |
|  | $\overline{\mathrm{x}}$ | <0.0005 | <0.05 | 0.013 | 0.057 | <0.001 | 42.9 | 0.0015 | <0.005 | <0.005 | 0.003 | 0.245 | 11.5 | 0.074 |
|  | s.D. | 0.0 | 0.0 | 0.013 | 0.001 | 0.0 | 0.5 | 0.0002 | 0.0 | 0.0 | 0.0 | 0.022 | 0.1 | 0.004 |

appendix II table 1

| Station | SAMPLE NUMBER | $\begin{gathered} \mathrm{Mo} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Na} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Ni} \\ (\mathrm{mq} / \mathrm{L}) \end{gathered}$ | $\begin{aligned} & \hline \mathrm{P} \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\begin{aligned} & \hline \mathrm{Pb} \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\begin{gathered} \mathrm{Sb} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{aligned} & \mathrm{Se} \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\begin{aligned} & \hline \mathrm{Si} \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | $\begin{gathered} \mathrm{Sn} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \hline \mathrm{Sr} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Ti}}$ | $\begin{gathered} \hline \mathrm{V} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\frac{\mathrm{zn}}{(\mathrm{mg} / \mathrm{L})}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | $<0.005$ | 1.4 | $<0.02$ | 0.09 | 0.028 | $<0.05$ | <0.05 | 2.1 | $\bigcirc 0.01$ | 0.188 | $<0.002$ | $<0.005$ | 0.901 |
|  | 2 | $<0.005$ | 1.4 | <0.02 | 0.08 | 0.024 | $<0.05$ | <0.05 | 2.1 | <0.01 | 0.190 | $<0.002$ | <0.005 | 0.907 |
|  | 3 | <0.005 | 1.5 | <0.02 | 0.08 | 0.020 | <0.05 | <0.05 | 2.1 | <0.01 | 0.205 | <0.002 | <0.005 | 0.975 |
|  | $\overline{\mathrm{x}}$ | <0.005 | 1.4 | <0.02 | 0.08 | 0.024 | <0.05 | <0.05 | 2.1 | <0.01 | 0.194 | <0.002 | <0.005 | 0.928 |
|  | s.d. | 0.0 | 0.0 | 0. | 0.01 | 0.004 | 0.0 | 0.0 | 0.0 | 0.0 | 0.009 | 0.0 | 0.0 | 0.041 |
| 2 | 4 | <0.005 | 1.3 | $<0.02$ | $<0.05$ | 0.019 | $<0.05$ | <0.05 | 2.1 | <0.01 | 0.176 | <0.002 | $<0.005$ | 0.323 |
|  | 5 | $<0.005$ | 1.3 | <0.02 | <0.05 | 0.018 | <0.05 | <0.05 | 2.1 | <0.01 | 0.178 | <0.002 | <0.005 | 0.327 |
|  | 6 | <0.005 | 1.4 | <0.02 | <0.05 | 0.019 | <0.05 | <0.05 | 2.1 | <0.01 | 0.181 | <0.002 | <0.005 | 0.330 |
|  | $\overline{\mathrm{x}}$ | <0.005 | 1.3 | <0.02 | <0.05 | 0.019 | <0.05 | <0.05 | 2.1 | <0.01 | 0.178 | <0.002 | <0.005 | 0.327 |
|  | s.d. | 0.0 | 0.1 | 0. | 0.0 | 0.001 | 0.0 | 0.0 | 0.0 | 0.0 | 0.003 | 0.0 | 0.0 | 0.004 |
| 3 | 7 | $<0.005$ | 1.0 | $<0.02$ | $<0.05$ | 0.021 | $<0.05$ | <0.05 | 1.3 | <0.01 | 0.143 | $<0.002$ | <0.005 | 0.006 |
|  | 8 | $<0.005$ | 1.0 | $<0.02$ | <0.05 | 0.015 | $<0.05$ | <0.05 | 1.3 | <0.01 | 0.146 | $<0.002$ | <0.005 | 0.014 |
|  | 9 | <0.005 | 1.0 | <0.02 | <0.05 | 0.017 | <0.05 | <0.05 | 1.3 | <0.01 | 0.145 | <0.002 | <0.005 | 0.004 |
|  | $\overline{\mathrm{x}}$ | <0.005 | 1.0 | <0.02 | <0.05 | 0:018 | <0.05 | <0.05 | 1.3 | <0.01 | 0.145 | <0.002 | <0.005 | 0.008 |
|  | s.d. | 0.0 | 0.0 | 0.0 | 0.0 | 0.003 | 0.0 | 0.0 | 0.0 | 0.0 | 0.002 | 0.0 | 0.0 | 0.005 |
| 4 | 10 | <0.005 | 1.0 | $<0.02$ | <0.05 | 0.023 | $<0.05$ | <0.05 | 1.3 | <0.01 | 0.148 | <0.002 | <0.005 | 0.023 |
|  | 11 | <0.005 | 1.0 | $<0.02$ | <0.05 | 0.017 | <0.05 | <0.05 | 1.3 | <0.01 | 0.150 | <0.002 | <0.005 | 0.024 |
|  | 12 | <0.005 | 1.0 | <0.02 | <0.05 | 0.021 | <0.05 | <0.05 | 1.3 | <0.01 | 0.145 | <0.002 | <0.005 | 0.021 |
|  | $\bar{x}$ | <0.005 | 1.0 | <0.02 | <0.05 | 0.020 | $<0.05$ | <0.05 | 1.3 | <0.01 | 0.148 | <0.002 | <0.005 | 0.023 |
|  | s.D. | 0.0 | 0.0 | 0.0 | 0.0 | 0.003 | 0.0 | 0.0 | 0.0 | 0.0 | 0.003 | 0.0 | 0.0 | 0.002 |



| station | SAMPLE NUMBER | $\begin{gathered} \mathrm{Mo} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Na} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Ni}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{p}}$ | $\underset{\substack{\mathrm{pb} \\(\mathrm{mg} / \mathrm{L})}}{ }$ | $\begin{gathered} \mathrm{sb} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Se} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\underset{\substack{\mathrm{sig} \\(\mathrm{mg} / \mathrm{L})}}{\mathrm{s}}$ | $\begin{gathered} \mathrm{Sn} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\underset{\substack{\mathrm{srg} / \mathrm{L})}}{ }$ | $\begin{gathered} \mathrm{Ti} \\ (\mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{aligned} & \mathrm{V} / \mathrm{m} / \mathrm{L}) \end{aligned}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{zn}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 13 | ＜0．005 | 57.8 | ＜0．02 | ＜0．05 | 0.020 | ＜0．05 | $<0.05$ | 0.4 | ＜0．01 | 0.340 | $<0.002$ | ＜0．005 | 0.187 |
|  | 14 | ＜0．005 | 56.0 | く0．02 | ＜0．05 | 0.027 | ＜0．05 | ＜0．05 | 0.4 | ＜0．01 | 0.334 | ＜0．002 | ＜0．005 | 0.247 |
|  | 15 | ＜0．005 | 58.7 | ＜0．02 | ＜0．05 | 0.024 | ＜0．05 | ＜0．05 | 0.4 | ＜0．01 | 0.345 | ＜0．002 | ＜0．005 | 0.222 |
|  | $\overline{\mathrm{x}}$ | 0.005 | 57.5 | ＜0．02 | ¢0．05 | 0.024 | ＜0．05 | ＜0．05 | 0.4 | ＜0．01 | 0.340 | ＜0．002 | ＜0．005 | 0.219 |
|  | s．D． | 0.0 | 1.4 | 0.0 | 0.0 | 0.004 | 0.0 | 0.0 | ． 0 | 0.0 | 0.006 | 0.0 | 0.0 | 0.030 |
| 6 | 16 | ＜0．005 | 1.0 | ＜0．02 | ＜0．05 | 0.026 | ＜0．05 | ＜0．05 | 2.1 | ＜0．01 | 0.311 | ＜0．002 | ＜0．005 | 0.004 |
|  | 17 | ＜0．005 | 1.0 | ＜0．02 | く0．05 | 0.022 | ＜0．05 | ＜0．05 | 2.1 | ＜0．01 | 0.306 | ＜0．002 | ＜0．005 | 0.004 |
|  | 18 | ＜0．005 | 1.0 | ＜0．02 | ${ }^{0} 0.05$ | 0.015 | ＜0．05 | ＜0．05 | 2.1 | ＜0．01 | 0.297 | ＜0．002 | ＜0．005 | 0.003 |
|  | $\overline{\mathrm{x}}$ | ＜0．005 | 1.0 | ＜0．02 | ＜0．05 | 0.021 | ＜0．05 | ＜0．05 | 2.1 | ＜0．01 | 0.305 | ＜0．002 | ＜0．005 | 0.004 |
|  | s．d． | 0.0 | 0.0 | 0.0 | 0.0 | 0.006 | 0.0 | 0.0 | 0.0 | 0.0 | 0.007 | 0.0 | 0.0 | 0.001 |
| 7 | 19 | 0.006 | 12.0 | ＜0．02 | ＜0．05 | 0.023 | ＜0．05 | ＜0．05 | 1.7 | ＜0．01 | 0.257 | ＜0．002 | ＜0．005 | 0.330 |
|  | 20 | 0.006 | 11.9 | ＜0．02 | ＜0．05 | 0.022 | ＜0．05 | ＜0．05 | 1.7 | ＜0．01 | 0.256 | ＜0．002 | ＜0．005 | 0.327 |
|  | 21 | ＜0．005 | 12.3 | ＜0．02 | ＜0．05 | 0.019 | ＜0．05 | ＜0．05 | 1.7 | ＜0．01 | 0.263 | ＜0．002 | ¢0．005 | 0.335 |
|  | $x$ | ＜0．006 | 12.1 | く0．02 | ＜0．05 | 0.021 | ＜0．05 | ＜0．05 | 1.7 | ＜0．01 | 0.259 | ＜0．002 | ＜0．005 | 0.331 |
|  | s．d． | 0.0 | 0.2 | 0.0 | 0.0 | 0.002 | 0.0 | 0.0 | 0.0 | 0.0 | 0.004 | 0.0 | 0.0 | 0.004 |
| 8 | 22 | ＜0．005 | 1.1 | ＜0．02 | ＜0．05 | 0.025 | ＜0．05 | ＜0．05 | 1.2 | ＜0．01 | 0.146 | ＜0．002 | ＜0．005 | 0.019 |
|  | 23 | ¢0．005 | 1.1 | ＜0．02 | ＜0．05 | 0.027 | ＜0．05 | ＜0．05 | 1.2 | ＜0．01 | 0.146 | ＜0．002 | ＜0．005 | 0.019 |
|  | 24 | 0.007 | 1.1 | ＜0．02 | ＜0．05 | 0.027 | ＜0．05 | ＜0．05 | 1.2 | ＜0．01 | 0.147 | ＜0．002 | ＜0．005 | 0.020 |
|  | $\overline{\mathrm{x}}$ | ¢0．007 | 1.1 | く0．02 | ＜0．05 | 0.026 | ＜0．05 | ＜0．05 | 1.2 | ＜0．01 | 0.146 | ＜0．002 | $\langle 0.005$ | 0.019 |
|  | s．D． | 0.0 | 0.0 | 0.0 | 0.0 | 0.001 | 0.0 | 0.0 | ． 0 | 0.0 | 0.001 | 0.0 | 0.0 | 0.001 |


| STATION | SAMPLE NUMBER | $\begin{gathered} \mathrm{Mo} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Na} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Ni} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} P \\ (m g / L) \end{gathered}$ | $\begin{gathered} \mathrm{Pb} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Sb} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Se} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Si} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Sn} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\frac{\mathrm{Sr}}{(\mathrm{mg} / L)}$ | $\frac{\mathrm{Ti}}{(\mathrm{mg} / \mathrm{L})}$ | $\begin{gathered} \mathrm{V} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{array}{r} 2 \mathrm{n} \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 9 | 25 | 0.005 | 2.4 | $\langle 0.02$ | $<0.05$ | 0.026 | $<0.05$ | $<0.05$ | 1.2 | $<0.01$ | 0.160 | $<0.002$ | <0.005 | 0.049 |
|  | 26 | 0.006 | 1.2 | <0.02 | <0.05 | 0.019 | <0.05 | <0.05 | 1.2 | <0.01 | 0.148 | <0.002 | <0.005 | 0.020 |
|  | 27 | 0.005 | 1.2 | <0.02 | <0.05 | 0.026 | <0.05 | <0.05 | 1.2 | <0.01 | 0.148 | <0.002 | <0.005 | 0.019 |
|  | $\overline{\mathrm{x}}$ | 0.005 | 1.6 | $\langle 0.02$ | <0.05 | 0.024 | <0.05 | <0.05 | 1.2 | <0.01 | 0.152 | <0.002 | <0.005 | 0.029 |
|  | s.d. | 0.001 | 0.7 | 0.0 | 0.0 | 0.004 | 0.0 | 0.0 | . 0 | 0.0 | 0.007 | 0.0 | 0.0 | 0.017 |
| 10 | 28 | <0.005 | 1.3 | <0.02 | $<0.05$ | 0.027 | <0.05 | $<0.05$ | 1.4 | $<0.01$ | 0.152 | $<0.002$ | <0.005 | 0.024 |
|  | 29 | 0.009 | 1.4 | <0.02 | <0.05 | 0.026 | <0.05 | <0.05 | 1.4 | <0.01 | 0.152 | $<0.002$ | <0.005 | 0.022 |
|  | 30 | $<0.005$ | 1.3 | <0.02 | <0.05 | 0.009 | <0.05 | <0.05 | 1.4 | 0.01 | 0.152 | <0.002 | <0.005 | 0.021 |
|  | $\overline{\mathrm{x}}$ | <0.009 | 1.3 | <0.02 | <0.05 | 0.021 | <0.05 | <0.05 | 1.4 | <0.01 | 0.152 | <0.002 | <0.005 | 0.022 |
|  | s.D. | 0.0 | 0.1 | 0.0 | 0.0 | 0.010 | 0.0 | 0.0 | 0.0 | 0.0 | 0.000 | 0.0 | 0.0 | 0.002 |
| 11 | 31 | 0.006 | 1.6 | <0.02 | <0.05 | 0.017 | <0.05 | <0.05 | 2.0 | <0.01 | 0.164 | <0.002 | <0.005 | 0.016 |
|  | 32 | <0.005 | 1.5 | <0.02 | <0.05 | 0.022 | <0.05 | <0.05 | 1.7 | $<0.01$ | 0.161 | $<0.002$ | <0.005 | 0.015 |
|  | 33 | 0.009 | 1.5 | <0.02 | <0.05 | 0.022 | <0.05 | <0.05 | 1.6 | $<0.01$ | 0.160 | <0.002 | <0.005 | 0.018 |
|  | $\overline{\mathrm{x}}$ | <0.008 | 1.5 | $<0.02$ | <0.05 | 0.020 | <0.05 | <0.05 | 1.8 | <0.01 | 0.162 | <0.002 | <0.005 | 0.016 |
|  | s.D. | 0.0 | 0.1 | 0.0 | 0.0 | 0.003 | 0.0 | 0.0 | 0.2 | 0.0 | 0.002 | 0.0 | 0.0 | 0.002 |

APPENDIX II TABLE 2 WATER QUALITY RESULTS FOR AUGUST 21-22, 1985 Note: rriplicate values for each station are shown in order of

| station | SAMPLE NUMBER | $\begin{aligned} & \text { AVG. DEPTH } \\ & (\mathbf{m}) \end{aligned}$ | WIDTH | $\underset{(\mathrm{m} 3 / \mathrm{s})}{\mathrm{FLWW}}$ | $\begin{gathered} \text { TEMP. } \\ \text { (C). } \end{gathered}$ | insitu | ${ }_{\text {pH }}^{\text {LAB }}$ |  | ND. <br> /cm) <br> LAB | DISOLVED oxygen (mg/L) | $\begin{gathered} \text { \&DD } \\ \text { SATVRA- } \\ \text { TION } \end{gathered}$ | $\begin{gathered} \text { COLOR } \\ \text { (FTU) } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | $\begin{aligned} & 2 \\ & 3 \end{aligned}$ | N/A | N/A | N/A | 6.5 | 7.60 | 7.4 7.5 6.9 | 500 | $\begin{aligned} & 800 \\ & 800 \\ & 800 \end{aligned}$ | 10.4 | 89 | $\begin{aligned} & 5 \\ & 5 \\ & 5 \end{aligned}$ |
|  | $\begin{gathered} \overline{\bar{x}} \\ \text { s.D. } \end{gathered}$ |  |  |  |  |  | 7.3 0.3 |  | 800 0 |  |  | $\begin{aligned} & 5 \\ & 0 \end{aligned}$ |
| 2 | $\begin{aligned} & 4 \\ & 5 \\ & 6 \end{aligned}$ | 0.3 | 3.0 | 0.12 | 5.0 | 7.90 | 7.5 6.9 7.3 | 312 | $\begin{aligned} & 580 \\ & 580 \\ & 580 \end{aligned}$ | 11.3 | 92 | $\begin{aligned} & 5 \\ & 5 \\ & 5 \end{aligned}$ |
|  | $\begin{gathered} \overline{\mathrm{x}} \\ \text { s.D. } \end{gathered}$ |  |  |  |  |  | 7.2 0.3 |  | 580 0 |  |  | 5 0 |
| 3 | $\begin{aligned} & 7 \\ & 8 \\ & 9 \end{aligned}$ | N/A | n/A | 3.3 | 11.0 | 8.10 | 7.6 7.5 7.5 | 165 | $\begin{aligned} & 250 \\ & 250 \\ & 250 \end{aligned}$ | 10.4 | 97 | $\begin{aligned} & 10 \\ & 10 \\ & 10 \end{aligned}$ |
|  | $\begin{gathered} \overline{\mathrm{x}} \\ \text { s.d. } \end{gathered}$ |  |  |  |  |  | 7.5 0.1 |  | 250 0 |  |  | $\begin{array}{r} 10 \\ 0 \end{array}$ |
| 4 | $\begin{aligned} & 10 \\ & 11 \\ & 12 \end{aligned}$ | 0.6 | 9.6 | 3.451 | 10.0 | 8.00 | 6.9 5. 9.0 | 180 | $\begin{aligned} & 310 \\ & 290 \\ & 255 \end{aligned}$ | 10.3 | 95 | $\begin{aligned} & 10 \\ & 10 \\ & 10 \end{aligned}$ |
|  | $\begin{gathered} \bar{x} \\ \text { s.D. } \end{gathered}$ |  |  |  |  |  | 7.1 0. |  | 285 28 |  |  | ${ }^{10} 0$ |

APPENDIX II TABLE 2 WATER QUALITY RESULTS FOR AUGUST 21-22, 1985 left bank, midstream and right bank

appendix II table 2 Water quality results for august 21-22, 1985 left bank, midstream and right bank

TABLE 2 WATER QUALITY RESULTS FOR AUGUST 21-22, 1985
NOTE: Triplicate values for each station are shown in order of left bank, midstream and right bank

| STATION | SAMPLE NUMBER | turb. (FTU) | $\begin{gathered} \text { T.ALK. } \\ \text { (as CACO3) } \\ \text { (mg/L) } \end{gathered}$ | $\begin{aligned} & \text { T. HARD } \\ & \text { (as CaCO3) } \\ & \text { (mg/L) } \end{aligned}$ | $\begin{gathered} \text { SUFFATE } \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{array}{r} \text { CHLORIDE } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{gathered} \text { PHOSPATE } \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\underset{\substack{\text { NITRITE } \\(\mathrm{mg} / \mathrm{L})}}{ }$ | NITRATE (mg/L) | ammonia (mg/L) | $\begin{aligned} & \mathrm{NFR} \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 0.30 | 105.0 | 463 | 270 | 0.7 | 0.005 | <0.005 | 0.075 | 0.020 | 14 |
|  | 2 | 0.30 | 102.0 | 471 | 310 | 0.8 | 0.005 | <0.005 | 0.077 | 0.012 | 7 |
|  | 3 | 0.30 | 103.0 | 469 | 300 | 0.8 | 0.003 | <0.005 | 0.077 | 0.011 | < 5 |
|  | $\overline{\mathrm{x}}$ | 0.30 | 103.3 | 468 | 293 | 0.8 | 0.004 | <0.005 | 0.076 | 0.014 | <11 |
|  | s.D. | 0.00 | 1.5 | 4 | 21 | 0.1 | 0.001 | 0.0 | 0.001 | 0.005 | 0 |
| 2 | 4 | 0.40 | 119.0 | 325 | 170 | 0.6 | <0.002 | <0.005 | 0.055 | 0.005 | 10 |
|  | 5 | 0.40 | 121.0 | 327 | 160 | 0.5 | $<0.002$ | <0.005 | 0.054 | 0.005 | 5 |
|  | 6 | 0.40 | 121.0 | 326 | 160.0 | 0.5 | <0.002 | <0.005 | 0.055 | <0.005 | 6 |
|  | $x$ | 0.40 | 120.3 | 326 | 163 | <0.6 | $<0.002$ | <0.005 | 0.055 | <0.005 | 7 |
|  | S.D. | . 00 | 1.2 | 1 | 6 | 0.0 | 0.000 | 0.0 | 0.001 | 0.0 | 3 |
| 3 | 7 | 0.15 | 86.2 | 135 | 53 | <0.5 | <0.002 | $<0.005$ | 0.008 | 0.006 | <5 |
|  | 8 | 0.15 | 85.7 | 136 | 37 | <0.5 | <0.002 | <0.005 | $<0.005$ | 0.006 | < 5 |
|  | 9 | 0.18 | 86.2 | 136 | 36 | 0.5 | <0.002 | <0.005 | <0.005 | 0.006 | < 5 |
|  | x | 0.16 | 86.0 | 136 | 42 | 0.5 | $<0.000$ | <0.002 | 0.008 | 0.006 | $<5$ |
|  | s.D. | 0.02 | 0.3 | 1 | 10 | 0.0 | 0.000 | 0.0 | 0.0 | 0.0 | 0 |
| 4 | 10 | 0.23 | 91.3 | 169 | 56 | <0.5 | $<0.002$ | $<0.005$ | 0.012 | 0.007 | <5 |
|  | 11 | 0.23 | 91.3 | 157 | 52 | <0.5 | 0.005 | <0.005 | 0.008 | 0.008 | 6 |
|  | 12 | 0.15 | 86.2 | 136 | 37 | 0.6 | <0.002 | <0.005 | $<0.005$ | <0.005 | $<5$ |
|  | $\overline{\mathrm{x}}$ | 0.20 | 89.6 | 154 | 48 | $<0.6$ | <0.002 | $<0.005$ | $<0.010$ | <0.008 | <6 |
|  | s.D. | 0.05 | 2.9 | 17 | 10 | 0.0 | 0.000 | 0.0 | 0.0 | 0.0 | 0 |

appendix II table 2 water quality results for august 21-22, 1985 NOTE: Triplicate Values for each station are shown in order of
left bank, midstream and right bank left bank, widstream and right bank

| STATION | SAMPLE NUMBER | TURB. <br> (FTU) | $\begin{gathered} \text { T. ALK. } \\ \text { (as CACO3) } \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | T. HARD (as CaCO3) ( $\mathrm{mg} / \mathrm{L}$ ) | $\begin{aligned} & \text { SULFATE } \\ & (\mathrm{mg} / L) \end{aligned}$ | CHLORIDE $(\mathrm{mg} / \mathrm{L})$ | $\begin{gathered} \text { PHOSPATE } \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{aligned} & \text { NITRITE } \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ | NITRATE $(\mathrm{mg} / \mathrm{L})$ | AMMONIA (mg/L) | NFR (mg/L) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 13 | 0.63 | 92.8 | 663 | 530 | 14.3 | 0.003 | $<0.005$ | 0.074 | 0.074 | < 5 |
|  | 14 | 0.58 | 93.1 | 680 | 550 | 14.2 | 0.002 | <0.005 | 0.078 | 0.078 | <5 |
|  | 15 | 0.68 | 91.9 | 671 | 540 | 14.4 | 0.004 | <0.005 | 0.079 | 0.079 | - 5 |
|  | $\bar{x}$ | 0.63 | 92.6 | 671 | 540 | 14.3 | 0.003 | <0.005 | 0.077 | 0.077 | < 5 |
|  | S.D. | 0.05 | 0.6 | 9 | 10 | 0.1 | 0.001 | 0.0 | 0.003 | 0.003 | 0 |
| 6 | 16 | <0.01 | 188.0 | 255 | 55 | $<0.5$ | <0.002 | $<0.005$ | 0.078 | 0.078 | < 5 |
|  | 17 | $<0.01$ | 187.0 | 254 | 48 | <0.5 | <0.002 | $<0.005$ | 0.078 | 0.078 | < 5 |
|  | 18 | <0.01 | 187.0 | 254 | 48 | <0.5 | <0.002 | <0.005 | 0.079 | 0.079 | <5 |
|  | $\bar{x}$ | <0.01 | 187.3 | 254 | 50 | $<0.5$ | $<0.002$ | <0.005 | 0.078 | 0.078 | <5 |
|  | S.D. | 0.00 | 0.6 | 1 | 4 | 0.0 | 0.000 | 0.0 | 0.001 | 0.001 | 0 |
| 7 | 19 | 1.80 | 163.0 | 507 | 300 | 6.6 | 0.006 | $<0.005$ | 0.160 | 0.160 | 6 |
|  | 20 | 2.00 | 163.0 | 508 | 290 | 6.9 | 0.004 | $<0.005$ | 0.160 | 0.160 | 13 |
|  | 21 | 2.00 | 163.0 | 511 | 300 | 6.9 | <0.002 | <0.005 | 0.160 | 0.160 | < 5 |
|  | $\bar{x}$ | 1.93 | 163.0 | 509 | 297 | 6.8 | $<0.005$ | <0.005 | 0.160 | 0.160 | <10 |
|  | S.D. | 0.12 | 0.0 | 2 | 6 | 0.2 | 0.000 | 0.0 | 0.000 | 0.000 | 0 |
| 8 | 22 | 0.20 | 98.8 | 157 | 45 | 0.7 | 0.059 | <0.005 | 0.016 | 0.016 | <5 |
|  | 23 | 0.18 | 102.0 | 156 | 44 | 0.7 | 0.003 | <0.005 | 0.008 | 0.008 | $<5$ |
|  | 24 | 0.18 | 101.0 | 155 | 41 | 0.7 | 0.002 | <0.005 | 0.053 | 0.053 | < 5 |
|  | $\overline{\mathrm{x}}$ | 0.19 | 100.6 | 156 | 43 | 0.7 | 0.021 | $<0.005$ | 0.026 | 0.026 | <5 |
|  | S.D. | 0.01 | 1.6 | 1 | 2 | 0.0 | 0.033 | 0.0 | 0.024 | 0.024 | 0 |

appendix II TABLE 2 WATER QUALITY RESULTS FOR AUGUST 21-22, 1985 left bank, midstream and right bank

| STATION | SAMPLE NUMBER | TURB. (FTU) | $\begin{gathered} \text { T.ALK. } \\ (\mathrm{as} \text { CACO }) \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{aligned} & \text { T. HARD } \\ & \text { (as CaCO3) } \\ & \text { (mg/L) } \end{aligned}$ | SULFATE (mg/L) | CHLORIDE (mg/L) | PHOSPATE $(\mathrm{mg} / \mathrm{L})$ | $\begin{aligned} & \text { NITRATE } \\ & \text { (mg/L) } \end{aligned}$ | $\underset{\substack{\text { NITRATE } \\(\mathrm{mg} / \mathrm{L})}}{ }$ | AMMONIA ( $\mathrm{mg} / \mathrm{L}$ ) | $\begin{aligned} & \mathrm{NFR} \\ & (\mathrm{mg} / \mathrm{L}) \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 9 | 25 | 0.53 | 145.0 | 380 | 200 | 4.0 | <0.002 | <0.005 | 0.090 | 0.090 | < |
|  | 26 | 0.28 | 113.0 | 227 | 80 | 1.5 | <0.002 | <0.005 | 0.036 | 0.036 | < |
|  | 27 | 0.33 | 101.0 | 173 | 125 | 0.8 | 0.049 | <0.005 | 0.014 | 0.014 | <5 |
|  | $\overline{\mathrm{x}}$ | 0.38 | 119.7 | 260 | 135 | 2.1 | <0.049 | <0.005 | 0.047 | 0.047 | <5 |
|  | S.D. | 0.13 | 22.7 | 107 | 61 | 1.7 | 0.000 | 0.0 | 0.039 | 0.039 | 0 |
| 10 | 28 | 0.28 | 106.0 | 170 | 48 | 0.7 | 0.006 | <0.005 | 0.009 | 0.009 | 7 |
|  | 29 | 0.25 | 106.0 | 170 | 47 | 0.7 | 0.003 | <0.005 | 0.009 | 0.009 | <5 |
|  | 30 | 0.15 | 106.0 | 169 | 46 | 0.7 | 0.003 | <0.005 | 0.008 | 0.008 | <5 |
|  | $\overline{\mathrm{x}}$ | 0.23 | 106.0 | 170 | 47 | 0.7 | 0.004 | <0.005 | 0.009 | 0.009 | <7 |
|  | S.D. | 0.07 | 0.0 | 1 | 1 | 0.0 | 0.002 | 0.0 | 0.001 | 0.001 |  |
| 11 | 31 | 0.38 | 113.0 | 173 | 43 | 0.8 | <0.002 | <0.005 | 0.006 | 0.010 | 7 |
|  | 32 | 0.35 | 108.0 | 167 | 44 | 0.6 | <0.002 | <0.005 | 0.016 | 0.009 | <5 |
|  | 33 | 0.48 | 105.0 | 165 | 46 | 0.6 | <0.002 | <0.005 | 0.021 | 0.009 | <5 |
|  | $\overline{\mathrm{x}}$ | 0.40 | 108.7 | 168 | 44 | 0.7 | <0.002 | <0.005 | 0.014 | 0.009 | <7 |
|  | s.D. | 0.07 | 4.0 | 4 | 2 | 0.1 | 0.000 | 0.0 | 0.008 | 0.001 |  |

appendix II table 2 hater quality results for august 21-22, 1985 NOTE: Triplicate Values for each station are shown in order of left bank, midstream and right bank

| Station | SAMPLE NUMBER | $\begin{gathered} \mathrm{Ag} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{As} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{B} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{Ba} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{Be} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{Ca} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{cd} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{Co} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\underset{\mathrm{mg} / \mathrm{L}}{\mathrm{Cr}}$ | $\begin{gathered} \mathrm{Cu} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{Fe} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\underset{\mathrm{mg} / \mathrm{L}}{\mathrm{Mg}}$ | $\underset{\mathrm{mg} / \mathrm{L}}{\mathrm{Mn}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | <0.0005 | <0.05 | 0.022 | 0.056 | <0.001 | 150.0 | 0.0017 | <0.005 | <0.005 | <0.001 | 0.188 | 20.3 | 1.790 |
|  | 2 | <0.0005 | <0.05 | 0.000 | 0.058 | <0.001 | 152.0 | 0.0015 | <0.005 | <0.005 | <0.001 | 0.204 | 20.7 | 1.830 |
|  | 3 | <0.0005 | <0.05 | 0.010 | 0.057 | <0.001 | 152.0 | 0.0015 | <0.005 | <0.005 | <0.001 | 0.199 | 20.6 | 1.800 |
| 2 | $\overline{\mathrm{x}}$ | <0.0005 | <0.05 | 0.011 | 0.057 | <0.001 | 151.3 | 0.0016 | <0.005 | <0.005 | $<0.001$ | 0.197 | 20.5 | 1.807 |
|  | s.D. | 0.0 | 0.0 | 0.011 | 0.001 | 0.0 | 1.2 | 0.0001 | 0.0 | 0.0 | 0.0 | 0.008 | 0.2 | 0.021 |
|  | 4 | <0.0005 | <0.05 | 0.028 | 0.052 | <0.001 | 99.7 | 0.0019 | <0.005 | <0.005 | <0.001 | 0.145 | 18.0 | 0.278 |
|  | 5 | $<0.0005$ | <0.05 | 0.022 | 0.053 | <0.001 | 101.0 | 0.0021 | <0.005 | <0.005 | <0.001 | 0.147 | 18. | 0.280 |
|  | 6 | <0.0005 | <0.05 | 0.028 | 0.052 | <0.001 | 100.0 | 0.0020 | <0.005 | <0.005 | 0.008 | 0.154 | 18.1 | 0.281 |
| 3 | $\overline{\mathrm{x}}$ | <0.0005 | <0.05 | 0.026 | 0.052 | <0.001 | 100.2 | 0.0020 | $<0.005$ | <0.005 | $<0.008$ | 0.149 | 18.1 | 0.280 |
|  | s.D. | 0.0 | 0.0 | 0.003 | 0.001 | 0.0 | 0.7 | 0.0001 | 0.0 | 0.0 | 0.0 | 0.005 | 0.1 | 0.002 |
|  | 7 | <0.0005 | <0.05 | 0.019 | 0.043 | <0.001 | 36.6 | <0.0005 | <0.005 | <0.005 | <0.001 | 0.101 | 10.3 | 0.018 |
|  | 8 | <0.0005 | <0.05 | 0.022 | 0.044 | <0.001 | 37.1 | <0.0005 | <0.005 | <0.005 | $<0.001$ | 0.109 | 10.5 | 0.018 |
|  | 9 | <0.0005 | <0.05 | 0.015 | 0.044 | <0.001 | 36.9 | N/A | <0.005 | <0.005 | $<0.001$ | 0.107 | 10.4 | 0.018 |
| 4 | $\overline{\mathrm{x}}$ | $<0.0005$ | $<0.05$ | 0.019 | 0.044 | <0.001 | 36.9 | <0.0005 | <0.005 | $<0.005$ | <0.001 | 0.103 | 10.4 | 0.018 |
|  | s.d. | 0.0 | 0.0 | 0.004 | 0.001 | 0.0 | 0.3 | 0.0 | 0.0 | 0.0 | 0.0 | 0.002 | 0.1 | 0.000 |
|  | 10 | <0.0005 | $<0.05$ | 0.006 | 0.046 | <0.001 | 47.7 | 0.0005 | <0.005 | <0.005 | 0.001 | 0.149 | 11.8 | 0.073 |
|  | 11 | <0.0005 | $<0.05$ | 0.006 | 0.045 | <0.001 | 44.0 | 0.0005 | <0.005 | <0.005 | <0.001 | 0.118 | 11.3 | 0.048 |
|  | 12 | <0.0005 | <0.05 | 0.025 | 0.044 | <0.001 | 37.1 | 0.0005 | <0.005 | ¢0.005 | 0.004 | 0.107 | 10.5 | 0.019 |
| $\begin{aligned} & \bar{x} \\ & \text { s.D. } \end{aligned}$ |  | <0.0005 | $<0.05$ | 0.012 | 0.045 | <0.001 | 42.9 | 0.0005 | $<0.005$ | <0.005 | $<0.002$ | 0.125 | 11.2 | 0.047 |
|  |  | 0.0 | 0.0 | 0.011 | 0.001 | 0.0 | 5.4 | 0.0000 | 0.0 | 0.0 | 0.0 | 0.022 | 0.7 | 0.027 |

APPENDIX II TABLE 2 WATER QUALITY RESULTS FOR AUGUST 21-22, 1985 NOTE: Triplicate Values for each station are shown in order of
left bank, midstream and right bank

| STATION | SAMPLE NUMBER | $\begin{gathered} \mathrm{Ag} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{As} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{B} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{Ba} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{Be} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{Ca} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{Cd} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \mathrm{Co} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\underset{\operatorname{mg} / \mathrm{L}^{\mathrm{Cr}}}{ }$ | $\begin{gathered} \mathrm{Cu} \\ \mathrm{mg} / \mathrm{L} \end{gathered}$ | $\underset{m g / L}{\mathrm{Fe}}$ | $\stackrel{M g}{\mathrm{mg} / \mathrm{L}}$ | $\underset{\mathrm{mg} / \mathrm{L}}{\mathrm{Mn}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 13 | $<0.0010$ | $<0.05$ | 0.019 | 0.017 | $<0.001$ | 193.0 | 0.0024 | $<0.005$ | $<0.005$ | 0.013 | 0.466 | 43.0 | 1.050 |
|  | 14 | $<0.0010$ | $<0.05$ | 0.032 | 0.017 | $<0.001$ | 197.0 | 0.0023 | <0.005 | $<0.005$ | 0.012 | 0.441 | 44.6 | 1.050 |
|  | 15 | <0.0010 | $<0.05$ | 0.013 | 0.017 | <0.001 | 195.0 | 0.0023 | <0.005 | $<0.005$ | 0.012 | 0.442 | 43.8 | 1.030 |
|  | $\bar{x}$ | $<0.0010$ | $<0.05$ | 0.021 | 0.017 | $<0.001$ | 195.0 | 0.0023 | <0.005 | $<0.005$ | 0.012 | 0.443 | 43.8 | 1.043 |
|  | S.D. | 0.0 | 0.0 | 0.010 | 0.000 | 0.0 | 2.0 | 0.0001 | 0.0 | 0.0 | 0.001 | 0.022 | 0.8 | 0.012 |
| 6 | 16 | <0.0005 | <0.05 | 0.025 | 0.079 | <0.001 | 79.8 | <0.0005 | <0.005 | $<0.005$ | $<0.001$ | 0.030 | 13.3 | 0.002 |
|  | 17 | <0.0005 | $<0.05$ | 0.014 | 0.080 | $<0.001$ | 79.4 | <0.0005 | <0.005 | $<0.005$ | $<0.001$ | 0.028 | 13.4 | 0.001 |
|  | 18 | <0.0005 | $<0.05$ | 0.000 | 0.081 | <0.001 | 79.5 | <0.0005 | <0.005 | $<0.005$ | $<0.001$ | 0.032 | 13.4 | 0.001 |
|  | $\overline{\mathrm{x}}$ | <0.0005 | $<0.05$ | 0.013 | 0.080 | <0.001 | 79.6 | <0.0005 | <0.005 | $<0.005$ | $<0.001$ | 0.030 | 13.4 | 0.001 |
|  | S.D. | 0.0 | 0.0 | 0.013 | 0.001 | 0.0 | 0.2 | 0.00 | 0.0 | 0.0 | 0.0 | 0.002 | 0.1 | 0.001 |
| 7 | 19 | <0.0005 | $<0.05$ | 0.014 | 0.027 | $<0.001$ | 145.0 | 0.0018 | <0.005 | $<0.005$ | 0.005 | 0.549 | 34.8 | 0.598 |
|  | 20 | $<0.0005$ | $<0.05$ | 0.001 | 0.027 | $<0.001$ | 145.0 | 0.0018 | $<0.005$ | $<0.005$ | 0.005 | 0.512 | 35.0 | 0.596 |
|  | 21 | <0.0005 | $<0.05$ | 0.001 | 0.026 | $<0.001$ | 146.0 | 0.0017 | <0.005 | <0.005 | 0.004 | 0.437 | 35.0 | 0.524 |
|  | $\overline{\mathrm{x}}$ | <0.0005 | $<0.05$ | 0.005 | 0.027 | <0.001 | 145.0 | 0.0018 | <0.005 | $<0.005$ | 0.005 | 0.499 | 34.9 | 0.573 |
|  | S.D. | 0.00 | 0.0 | 0.008 | 0.001 | 0.0 | 0.6 | 0.0001 | 0.0 | 0.0 | 0.001 | 0.057 | 0.1 | 0.042 |
| 8 | 22 | $<0.0005$ | $<0.05$ | 0.017 | 0.054 | $<0.001$ | 43.1 | <0.0005 | <0.005 | $<0.005$ | 0.006 | 0.129 | 11.7 | 0.042 |
|  | 23 | <0.0005 | $<0.05$ | 0.017 | 0.054 | $<0.001$ | 43.0 | <0.0005 | <0.005 | $<0.005$ | $<0.001$ | 0.125 | 11.6 | 0.041 |
|  | 24 | <0.0005 | $<0.05$ | 0.005 | 0.054 | $<0.001$ | 42.8 | <0.0005 | <0.005 | $<0.005$ | $<0.001$ | 0.124 | 11.6 | 0.041 |
|  | $\overline{\mathrm{x}}$ | <0.0005 | <0.05 | 0.013 | 0.054 | $<0.001$ | 43.0 | <0.0005 | <0.005 | $<0.005$ | $<0.006$ | 0.126 | 11.6 | 0.041 |
|  | S.D. | 0.00 | 0.0 | 007 | 0.000 | 0.0 | 0.2 | 0.00 | 0.0 | 0.0 | 0.0 | 0.003 | 0.1 | 0.001 |

APPENDIX II TABLE 2 WATER QUALITY RESULTS FOR AUGUST 21-22, 1985 left bank, midstream and right bank

| Station | SAMPLE NUMBER | $\begin{aligned} & \mathrm{Ag} \\ & \mathrm{mg} / \mathrm{L} \end{aligned}$ | $\begin{aligned} & \mathrm{As} \\ & \mathrm{mg} / \mathrm{L} \end{aligned}$ | $\begin{aligned} & \mathrm{B} \\ & \mathrm{mg} / \mathrm{L} \end{aligned}$ | $\begin{aligned} & \mathrm{Ba} \\ & \mathrm{mg} / \mathrm{L} \end{aligned}$ | $\begin{aligned} & \mathrm{Be} \\ & \mathrm{mg} / \mathrm{L} \end{aligned}$ | $\begin{aligned} & \mathrm{Ca} \\ & \mathrm{mg} / \mathrm{L} \end{aligned}$ | $\begin{aligned} & \mathrm{Cd} \\ & \mathrm{mg} / \mathrm{L} \end{aligned}$ | $\begin{aligned} & \mathrm{Co} \\ & \mathrm{mg} / \mathrm{L} \end{aligned}$ | $\begin{aligned} & \mathrm{Cr} \\ & \mathrm{mg} / \mathrm{L} \end{aligned}$ | $\begin{aligned} & \mathrm{Cu} \\ & \mathrm{mg} / \mathrm{L} \end{aligned}$ | $\begin{aligned} & \mathrm{Fe} \\ & \mathrm{mg} / \mathrm{L} \end{aligned}$ | $\begin{aligned} & \mathrm{Mg} \\ & \mathrm{mg} / \mathrm{L} \end{aligned}$ | $\begin{aligned} & \mathrm{Mn} \\ & \mathrm{mg} / \mathrm{L} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 9 | 25 | <0.0005 | <0.05 | 0.008 | 0.060 | $<0.001$ | 108.0 | <0.0005 | <0.005 | <0.005 | 0.002 | 0.248 | 26.3 | 0.234 |
|  | 26 | <0.0005 | <0.05 | 0.027 | 0.056 | $<0.001$ | 63.7 | <0.0005 | $<0.005$ | <0.005 | 0.001 | 0.166 | 16.3 | 0.104 |
|  | 27 | <0.0005 | <0.05 | <0.001 | 0.055 | <0.001 | 47.8 | <0.0005 | <0.005 | <0.005 | <0.001 | 0.134 | 12.8 | 0.055 |
|  | $\overline{\mathrm{x}}$ | $<0.0005$ | $<0.05$ | 0.017 | 0.057 | $<0.001$ | 73.2 | <0.0005 | <0.005 | <0.005 | <0.002 | 0.183 | 18.5 | 0.055 |
|  | s.D. | 0.00 | 0.0 | 0.0 | 0.003 | 0.0 | 31.2 | 0.00 | 0.0 | 0.0 | 0.0 | 0.059 | 7.0 | 0.055 |
| 10 | 28 | <0.0005 | <0.05 | 0.008 | 0.057 | <0.001 | 47.0 | <0.0005 | <0.005 | <0.005 | 0.001 | 0.163 | 12.5 | 0.048 |
|  | 29 | <0.0005 | <0.05 | 0.027 | 0.058 | <0.001 | 47.0 | <0.0005 | <0.005 | <0.005 | 0.001 | 0.167 | 12.6 | 0.049 |
|  | 30 | <0.0005 | <0.05 | 0.001 | 0.058 | <0.001 | 46.9 | <0.0005 | <0.005 | <0.005 | 0.001 | 0.156 | 12.6 | 0.048 |
|  | $\overline{\mathrm{x}}$ | <0.0005 | <0.05 | 0.012 | 0.058 | $<0.001$ | 47.0 | <0.0005 | <0.005 | <0.005 | 0.001 | 0.162 | 12.6 | 0.048 |
|  | s.D. | 0.00 | 0.0 | 0.013 | 0.001 | 0.0 | 0.1 | 0.00 | 0.0 | 0.0 | 0.0 | 0.006 | 0.1 | 0.001 |
| 11 | 31 | <0.0005 | <0.05 | 0.014 | 0.068 | <0.001 | 47.8 | <0.0005 | <0.005 | <0.005 | $<0.001$ | 0.269 | 12.7 | 0.055 |
|  | 32 | $<0.0005$ | <0.05 | 0.023 | 0.062 | <0.001 | 45.7 | <0.0005 | <0.005 | <0.005 | $<0.001$ | 0.225 | 12.7 | 0.042 |
|  | 33 | <0.0005 | <0.05 | 0.017 | 0.059 | <0.001 | 45.1 | <0.0005 | <0.005 | <0.005 | <0.001 | 0.214 | 12.6 | 0.038 |
|  | $\bar{x}$ | <0.0005 | <0.05 | 0.018 | 0.063 | $<0.001$ | 46.2 | <0.0005 | <0.005 | <0.005 | $<0.001$ | 0.236 | 12.7 | 0.045 |
|  | s.D. | 0.00 | 0.0 | 0.005 | 0.005 | 0.0 | 1.4 | 0.00 | 0.0 | 0.0 | 0.0 | 0.029 | 0.1 | 0.009 |

appendix II TABLE 2 water quality results for august 21-22, 1985 left bank, midstream and right bank

| Statton | SAMPLE NUMBER | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Mo}}$ | $\underset{(\mathrm{ma} / \mathrm{L})}{\substack{\mathrm{Ng}}}$ | $\underset{\substack{\mathrm{Ni} \\(\mathrm{mg} / \mathrm{L}}}{ }$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{P}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{pb}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{sb}}$ | $\underset{\substack{\text { Se } \\(\mathrm{mg} / \mathrm{L})}}{\text { (2)}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{si}_{2}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{sn}}$ | $\begin{gathered} \mathrm{Sr} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Ti}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{v}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{zn}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | <0.005 | 1.5 | 0.02 | $<0.05$ | <0.001 | <0.05 | $<0.05$ | 3.2 | <0.01 | 0.223 | <0.002 | <0.005 | 0.816 |
|  | 2 | <0.005 | 1.5 | 0.02 | <0.05 | <0.001 | <0.05 | <0.05 | 3.4 | <0.01 | 0.225 | <0.002 | <0.005 | 0.836 |
|  | 3 | <0.005 | 1.4 | <0.02 | <0.05 | 0.001 | <0.05 | <0.05 | 3.2 | <0.01 | 0.225 | <0.002 | <0.005 | 0.824 |
|  | $\overline{\mathrm{x}}$ | <0.005 | 1.5 | <0.02 | <0.05 | <0.001 | $<0.05$ | <0.05 | 3.3 | <0.01 | 0.224 | <0.002 | <0.005 | 0.825 |
|  | s.d. | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.001 | 0.0 | 0.0 | 0.010 |
| 2 | 4 | <0.005 | 1.3 | <0.02 | <0.05 | 0.002 | <0.05 | <0.05 | 3.0 | <0.01 | 0.192 | <0.002 | <0.005 | 0.429 |
|  | 5 | <0.005 | 1.3 | <0.02 | <0.05 | 0.001 | <0.05 | <0.05 | 3.1 | <0.01 | 0.194 | <0.002 | <0.005 | 0.435 |
|  | 6 | <0.005 | 1.3 | <0.02 | 0.06 | 0.001 | <0.05 | <0.05 | 2.9 | <0.01 | 0.192 | <0.002 | <0.005 | 0.436 |
|  | $\overline{\mathrm{x}}$ | <0.005 | 1.3 | <0.02 | $<0.06$ | 0.001 | 0.05 | 0.05 | 3.0 | <0.01 | 0.193 | <0.002 | <0.005 | 0.433 |
|  | s.d. | 0.0 | 0.0 | 0.0 | 0.0 | 0.001 | 0.0 | 0.0 | 0.1 | 0.0 | 0.001 | 0.0 | 0.0 | 0.004 |
| 3 | 7 | <0.005 | 1.2 | <0.02 | <0.05 | 0.002 | <0.05 | <0.05 | 1.8 | 0.03 | 0.151 | <0.002 | <0.005 | <0.002 |
|  | 8 | ¢0.005 | 1.2 | <0.02 | <0.05 | 0.001 | <0.05 | <0.05 | 1.9 | <0.01 | 0.153 | <0.002 | <0.005 | <0.002 |
|  | 9 | <0.005 | 1.2 | <0.02 | <0.05 | 0.001 | <0.05 | <0.0 | 1.6 | <0.01 | 0.153 | <0.002 | <0.005 | <0.002 |
|  | $\overline{\mathrm{x}}$ | <0.005 | 1.2 | <0.02 | <0.05 | 0.001 | <0.05 | <0.05 | 1.8 | <0.03 | 0.152 | <0.002 | <0.005 | <0.002 |
|  | s.d. | 0.0 | 0.0 | 0.0 | 0.0 | 0.001 | 0.0 | 0.0 | 0.2 | 0.0 | 0.001 | 0.0 | 0.0 | 0.0 |
| 4 | 10 | <0.005 | 1.3 | <0.02 | 0.05 | 0.001 | <0.05 | <0.05 | 1.8 | <0.01 | 0.162 | ¢0.002 | <0.005 | 0.084 |
|  | 11 | <0.005 | 1.3 | <0.02 | <0.05 | 0.001 | <0.05 | <0.05 | 1.7 | <0.01 | 0.158 | ¢0.002 | <0.005 | 0.053 |
|  | 12 | <0.005 | 1.2 | <0.02 | <0.05 | <0.001 | $<0.05$ | <0.05 | 1.6 | <0.01 | 0.153 | <0.002 | ¢0.005 | 0.003 |
|  | $\overline{\mathbf{x}}$ | <0.005 | 1.3 | <0.02 | <0.05 | ${ }^{0.001}$ | $<0.05$ | <0.05 | 1.7 | <0.01 | 0.158 | <0.002 | <0.005 | 0.047 |
|  | s.d. | 0.0 | 0.1 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.1 | 0.0 | 0.005 | 0.0 | 0.0 | 0.041 |

APPENDIX II TABLE 2 WATER QUALITY RESULTS FOR AUGUST 21-22, 1985 NOTE: Triplicate Values for each station are shown in order of
left bank, midstream and right bank

| STATION | SAMPLE NUMBER | $\begin{gathered} \text { Mo } \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Na} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Ni}}$ | $\underset{(m g / L)}{P}$ | $\begin{gathered} \mathrm{Pb} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Sb}}$ | $\begin{gathered} \mathrm{Se} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Si}}$ | $\begin{gathered} \mathrm{Sn} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Sr}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Ti}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{v}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{zn}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 5 | 13 | $<0.005$ | 44.3 | <0.02 | 0.10 | 0.003 | <0.05 | <0.05 | 0.9 | <0.01 | 0.370 | <0.002 | <0.005 | 0.092 |
|  | 14 | <0.005 | 46.5 | <0.02 | 0.07 | 0.003 | <0.05 | <0.05 | 0.9 | $<0.01$ | 0.387 | <0.002 | <0.005 | 0.090 |
|  | 15 | <0.005 | 45.5 | <0.02 | 0.06 | 0.003 | $\bigcirc 0.05$ | <0.05 | 0.9 | <0.01 | 0.380 | <0.002 | ¢0.005 | 0.089 |
|  | $\bar{x}$ | <0.005 | 45.4 | <0.02 | 0.08 | 0.003 | <0.05 | <0.05 | 0.9 | <0.01 | 0.379 | <0.02 | <0.005 | 0.090 |
|  | s.D. | 0.0 | 1.1 | 0.0 | 0.02 | 0.000 | 0.0 | 0.0 | 0.0 | 0.0 | 0.009 | 0.0 | 0.0 | 0.002 |
| 6 | 16 | <0.005 | 1.1 | $<0.02$ | <0.05 | 0.002 | $<0.05$ | <0.05 | 2.5 | 0.03 | 0.333 | <0.002 | <0.005 | <0.002 |
|  | 17 | <0.005 | 1.1 | <0.02 | <0.05 | 0.001 | $<0.05$ | <0.05 | 2.4 | 0.03 | 0.336 | <0.002 | <0.005 | <0.002 |
|  | 18 | <0.005 | 1.1 | <0.02 | <0.05 | 0.001 | <0.05 | <0.05 | 2.4 | 0.02 | 0.337 | <0.002 | <0.005 | 0.003 |
|  | $\overline{\mathrm{x}}$ | $<0.005$ | 1.1 | $<0.02$ | 0.05 | 0.001 | <0.05 | <0.05 | 2.4 | 0.03 | 0.335 | <0.002 | <0.005 | $<0.003$ |
|  | s.D. | 0.0 | 0.0 | 0.0 | 0.0 | 0.001 | 0.0 | 0.0 | 0.1 | 0.01 | 0.002 | 0.0 | 0.0 | 0.0 |
| 7 | 19 | $<0.005$ | 16.8 | <0.02 | 0.09 | 0.003 | <0.05 | $<0.05$ | 2.4 | <0.01 | 0.290 | <0.002 | <0.005 | 0.218 |
|  | 20 | <0.005 | 17.1 | $<0.02$ | 0.06 | 0.003 | $<0.05$ | <0.05 | 2.3 | $<0.01$ | 0.292 | <0.002 | <0.005 | 0.215 |
|  | 21 | <0.005 | 17.0 | <0.02 | 0.10 | 0.002 | <0.05 | <0.05 | 2.3 | <0.01 | 0.292 | <0.002 | <0.005 | 0.212 |
|  | $\bar{x}$ | $<0.005$ | 17.0 | $<0.02$ | 0.08 | 0.003 | <0.05 | <0.05 | 2.3 | <0.01 | 0.291 | <0.002 | <0.005 | 0.215 |
|  | s.D. | 0.0 | 0.2 | 0.0 | 0.02 | 0.001 | 0.0 | 0.0 | 0.1 | 0.0 | 0.001 | 0.0 | 0.0 | 0.003 |
| 8 | 22 | <0.005 | 1.3 | $<0.02$ | 0.06 | 0.001 | <0.05 | <0.05 | 1.6 | <0.01 | 0.158 | <0.002 | <0.005 | 0.031 |
|  | 23 | <0.005 | 1.3 | <0.02 | 0.07 | <0.001 | $<0.05$ | <0.05 | 1.6 | $<0.01$ | 0.157 | <0.002 | <0.005 | 0.028 |
|  | 24 | <0.005 | 1.3 | $<0.02$ | 0.05 | <0.001 | <0.05 | <0.05 | 1.6 | <0.01 | 0.156 | <0.002 | <0.005 | 0.026 |
|  | $\overline{\mathrm{x}}$ | <0.005 | 1.3 | <0.02 | 0.06 | < . 001 | <0.05 | <0.05 | 1.6 | <0.01 | 0.157 | <0.002 | <0.005 | 0.028 |
|  | S.D. | 0.0 | 0.0 | 0.0 | 0.01 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.001 | 0.0 | 0.0 | 0.003 |

appendix II table 2 water quality results for august 21-22, 1985 left bank, midstream and right bank

| Station | SAMPLE NUMBER | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Mo}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Na}}$ | $\underset{(\mathrm{mi} / \mathrm{L})}{\mathrm{Ni}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{P}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Pb}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{sb}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Se}}$ | $\begin{gathered} \mathrm{si} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \mathrm{Sn} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Sr}}$ | $\begin{gathered} \mathrm{Ti} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{v}}$ | $\underset{(\mathrm{mg} / \mathrm{L})}{\mathrm{Zn}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 9 | 25 | <0.005 | 9.8 | <0.02 | 0.06 | 0.001 | <0.05 | $<0.05$ | 2.4 | <0.01 | 0.236 | <0.002 | <0.005 | 0.052 |
|  | 26 | <0.005 | 4.0 | <0.02 | 0.06 | 0.001 | <0.05 | <0.05 | 1.7 | <0.01 | 0.182 | <0.002 | <0.005 | 0.052 |
|  | 27 | <0.005 | 1.9 | <0.02 | 0.06 | <0.001 | <0.05 | <0.05 | 1.7 | 0.02 | 0.165 | <0.002 | <0.005 | 0.038 |
|  | $\overline{\mathrm{x}}$ | <0.005 | 5.2 | <0.02 | 0.06 | <0.001 | <0.05 | <0.05 | 1.9 | <0.02 | 0.194 | <0.002 | <0.005 | 0.065 |
|  | s.d. | 0.0 | 4.1 | 0.0 | 0.00 | 0.0 | 0.0 | 0.0 | 0.4 | 0.0 | 0.037 | 0.0 | 0.0 | 0.036 |
| 10 | 28 | <0.005 | 1.7 | $<0.02$ | <0.05 | <0.001 | <0.05 | <0.05 | 1.7 | 0.02 | 0.167 | <0.002 | <0.005 | 0.024 |
|  | 29 | <0.005 | 1.7 | <0.02 | <0.05 | 0.001 | <0.05 | <0.05 | 1.7 | 0.02 | 0.169 | <0.002 | <0.005 | 0.027 |
|  | 30 | <0.005 | 1.7 | <0.02 | <0.05 | <0.001 | <0.05 | <0.05 | 1.7 | <0.01 | 0.169 | <0.002 | <0.005 | 0.031 |
|  | $\overline{\mathrm{x}}$ | <0.005 | 1.7 | $<0.02$ | <0.05 | <0.001 | <0.05 | <0.05 | 1.7 | <0.02 | 0.168 | <0.002 | 0.005 | 0.027 |
|  | s.d. | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.0 | 0.001 | 0.0 | 0.0 | 0.004 |
| 11 | 31 | <0.005 | 1.9 | <0.02 | <0.05 | <0.001 | <0.05 | <0.05 | 1.9 | <0.01 | 0.176 | <0.002 | <0.005 | 0.020 |
|  | 32 | $<0.005$ | 1.8 | <0.02 | <0.05 | <0.001 | <0.05 | <0.05 | 2.5 | 0.01 | 0.183 | <0.002 | <0.005 | 0.016 |
|  | 33 | <0.005 | 1.8 | <0.02 | <0.05 | <0.001 | <0.05 | <0.05 | 2.5 | <0.01 | 0.184 | <0.002 | <0.005 | 0.014 |
|  | ¢ ${ }_{\text {x }}^{\text {s.d. }}$ | $\xrightarrow{<0.005}$ | ${ }_{0.1}^{1.8}$ | ${ }_{0.0}^{20.02}$ | $\stackrel{0}{0.05}$ | $<0.001$ 0.0 | <0.05 0.0 | ${ }_{0}^{0.05} 0$ | 2.3 0.3 | $\stackrel{0.01}{0.0}$ | 0.181 0.004 | ${ }_{\substack{\text { co. } \\ 0.0 \\ 0.0}}$ | $\bigcirc \substack{<0.005 \\ 0.0}$ | 0.017 0.003 |

## APPENDIX III



APPENDIX III TABLE 3 SEDIMENT METALS ANALYSIS FOR JULY 9-10, 1985

| STATION NUMBER | SAMPLE NUMBER | $\begin{gathered} \mathrm{Al} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \text { As } \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{Kg})}{\mathrm{Ba}}$ | $\begin{gathered} \mathrm{Be} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Ca} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Cd} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} C o \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Cr} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Cu} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Fe} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Mg} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{aligned} & \mathrm{Mn} \\ & (\mathrm{mg} / \mathrm{kg}) \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 10100 | 72 | 467 | <2 | 19700 | 14.1 | 16.9 | 60.1 | 42.2 | 30400 | 8520 | 2760 |
|  | 2 | 8720 | 70 | 393 | <2 | 18800 | 10.5 | 12.8 | 27.9 | 37.2 | 28400 | 8350 | 2990 |
|  | 3 | 9160 | 82 | 511 | <2 | 16800 | 20.0 | 19.5 | 33.0 | 41.0 | 31400 | 8180 | 1990 |
|  | $\overline{\mathrm{x}}$ | 9330 | 75 | 457 | $<2$ | 18430 | 14.9 | 16.4 | 40.3 | 40.1 | 30100 | 8350 | 2580 |
|  | S.D. | 705 | 6 | 60 | 0 | 1480 | 4.8 | 3.4 | 17.3 | 2.6 | 1530 | 170 | 524 |
| 2 | 4 | 5500 | 1730 | 140 | <2 | 7260 | 130 | 3.0 | 16.0 | 121 | 116000 | 5720 | <0.2 |
|  | 5 | 6480 | 1520 | 178 | <2 | 7440 | 123 | 7.2 | 17.7 | 112 | 102000 | 5660 | <0.2 |
|  | 6 | 5170 | 1900 | 126 | <2 | 7300 | 124 | 7.8 | 16.3 | 126 | 125000 | 5880 | <0.2 |
|  | $\bar{x}$ | 5720 | 1720 | 148 | <2 | 7330 | 126 | 6.0 | 16.7 | 120 | 114000 | 5750 | <0.2 |
|  | s.d. | 681 | 190 | 27 | 0 | 95 | 3.8 | 2.6 | 0.9 | 7.1 | 11600 | 114 | 0 |
| 3 | 7 | 10900 | <8 | 199 | <2 | 12200 | 0.8 | 9.0 | 23.2 | 24.4 | 26300 | 8420 | 604 |
|  | 8 | 10800 | $<8$ | 202 | <2 | 13800 | 0.6 | 8.4 | 23.7 | 23.2 | 25400 | 9070 | 556 |
|  | 9 | 11500 | <8 | 215 | <2 | 13600 | 0.4 | 7.8 | 25.4 | 26.6 | 26600 | 8880 | 594 |
|  | $\overline{\mathrm{x}}$ | 11100 | <8 | 205 | <2 | 13200 | 0.6 | 8.4 | 24.1 | 24.7 | 26100 | 8790 | 585 |
|  | s.D. | 379 | 0 | 9 | 0 | 872 | 0.2 | 0.6 | 1.2 | 1.7 | 624 | 334 | 25 |

appendix III


| STATION NUMBER | SAMPLE NUMBER | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Al}}$ | $\begin{gathered} \mathrm{As} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Ba} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Be} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} c a \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} C d \\ (m g / k g) \end{gathered}$ | $\begin{gathered} \mathrm{Co} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Cr} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Cu}}$ | $\begin{gathered} \mathrm{Fe} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Mg} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Mn} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | 10 | 10200 | 174 | 216 | <2 | 15400 | 24.8 | 8.2 | 23.4 | 42.6 | 34900 | 9800 | 3130 |
|  | 11 | 10800 | 188 | 228 | <2 | 10500 | 28.3 | 9.3 | 24.6 | 44.5 | 36000 | 6770 | 3120 |
|  | 12 | 9930 | 163 | 231 | <2 | 8670 | 27.4 | 14.4 | 25.0 | 64.4 | 31600 | 5640 | 3170 |
|  | $\overline{\mathrm{x}}$ | 10310 | 175 | 225 | <2 | 11500 | 26.8 | 10.6 | 24.3 | 50.5 | 34200 | 7400 | 3140 |
|  | S.D. | 445 | 13 | 8 | 0 | 3480 | 1.8 | 3.3 | 0.8 | 12.1 | 2290 | 2150 | 26 |
| 7 | 19 | 6510 | 242 | 331 | <. 2 | 7010 | 84.6 | 14.7 | 22.8 | 136 | 45000 | 3990 | <. 2 |
|  | 20 | 6940 | 267 | 309 | <. 2 | 8020 | 82.7 | 11.3 | 21.8 | 134 | 48800 | 4340 | <. 2 |
|  | 21 | 4670 | 197 | 275 | <. 2 | 5260 | 72.3 | 14.5 | 18.7 | 112 | 38300 | 2830 | $<.2$ |
|  | $\overline{\mathrm{x}}$ | 6040 | 235 | 305 | <. 2 | 6760 | 79.9 | 13.5 | 21.1 | 127 | 44000 | 3720 | <. 2 |
|  | S.D. | 1210 | 35 | 28 | 0 | 1400 | 6.6 | 1.9 | 2.1 | 13.3 | 5320 | 790 | 0 |
| 8 | 22 | 9460 | 9 | 194 | <. 2 | 10700 | 5.2 | 8.4 | 26.3 | 29.6 | 22900 | 6110 | 1670 |
|  | 23 | 8720 | 51 | 284 | <. 2 | 10800 | 14.7 | 12.3 | 24.6 | 33.1 | 24600 | 6910 | 2630 |
|  | 24 | 6100 | 22 | 186 | <. 2 | 6640 | 7.4 | 9.0 | 24.0 | 22.6 | 17300 | 4710 | 2340 |
|  | $\overline{\mathbf{x}}$ | 8100 | 27 | 221 | <. 2 | 9380 | 9.1 | 9.9 | 25.0 | 28.4 | 21600 | 5910 | 2210 |
|  | s.D. | 1760 | 22 | 54 | 0 | 2370 | 5.0 | 2.1 | 1.2 | 5.3 | 3820 | 1110 | 492 |


APPENDIX III

| STATTON NUMBER | $\begin{aligned} & \text { SAMPLE } \\ & \text { NUMBER } \end{aligned}$ | $\frac{\mathrm{Al}}{(\mathrm{mg} / \mathrm{kg})}$ | ${ }_{(\mathrm{mg} / \mathrm{kg})}^{\mathrm{As}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Ba}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Be}}$ | $\begin{gathered} \mathrm{ca} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |  | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Co}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{cr}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{cu}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Fe}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Mg}}$ | $\sum_{(\mathrm{mg} / \mathrm{kg})}^{\mathrm{Mn}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 9 | 25 | 6280 | 230 | 1560 | ¢. 2 | 8430 | 49.7 | 15.1 | 21.8 | 145.0 | 39800 | 4310 | <. 2 |
|  | 26 | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A |
|  | 27 | 4870 | 275 | 346 | <. 2 | 5720 | 63.6 | 7.4 | 19.7 | 156.0 | 45800 | 3620 | 488 |
|  | $x$ | 5575 | 253 | 953 | <0.2 | 7080 | 56.7 | 11.3 | 20.8 | 151 | 42800 | 3965 | $<488$ |
|  | s.D. | - |  | - | - | - | - | - | - | - |  |  |  |
| 10 | 28 | 9500 | 77 | 365 | <. 2 | 10700 | 9.5 | 13.5 | 26.7 | 47.3 | 28200 | 7480 | 2670 |
|  | 29 | 8210 | 53 | 245 | <. 2 | 5770 | 6.6 | 9.9 | 27.1 | 36.5 | 24200 | 4780 | 2260 |
|  | 30 | 8540 | 79 | 226 | <. 2 | 12500 | 13.2 | 9.5 | 35.6 | 32.6 | 24700 | 8340 | 2680 |
|  | $x$ | 8750 | 70 | 279 | $<0.2$ | 9660 | 9.8 | 11.0 | 29.8 | 38.8 | 25700 | 6870 | 2540 |
|  | s.D. | 670 | 14 | 75 | 0 | 3484 | 3.3 | 2.2 | 5.0 | 7.6 | 2180 | 1860 | 240 |
| 11 | 31 | 7330 | 89 | 105 | <. 2 | 3070 | 1.3 | 15.0 | 26.6 | 23.9 | 26700 | 3100 | 583 |
|  | 32 | 7310 | 86 | 95 | <. 2 | 3150 | 0.3 | 16.0 | 29.4 | 22.3 | 30100 | 3150 | 509 |
|  | 33 | 7320 | 82 | 74 | <. 2 | 2650 | 0.5 | 16.4 | 23.7 | 22.8 | 27100 | 3030 | 465 |
|  | $\overline{\mathrm{x}}$ | 7320 | 86 | 91 | $<.2$ | 2960 | 0.7 | 15.5 | 26.7 | 23 | 28000 | 3090 | 519 |
|  | s.D. | 10 | 4 | 16 | 0 | 269 | 0.5 | 0.9 | 2.9 | 0.8 | 1860 | 60 | 60 |

APPENDIX III

| $\begin{aligned} & \text { STATION } \\ & \text { NUMBER } \end{aligned}$ | SAMPLE NUMBER | $\begin{gathered} \mathrm{Mo} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\frac{\mathrm{Na}}{(\mathrm{mg} / \mathrm{kg})}$ | $\begin{gathered} \mathrm{Ni} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{P} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Pb} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Si}}$ | $\begin{gathered} \mathrm{Sn} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Sr} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Ti}}$ | $\frac{V}{(\mathrm{mg} / \mathrm{kg})}$ | $\frac{\mathrm{zn}}{(\mathrm{mg} / \mathrm{kg})}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 13.2 | 285 | 83 | 1520 | 204 | 60 | <2 | 56.2 | 780 | 44.4 | 2190 |
|  | 2 | 6.0 | 180 | 55 | 1220 | 178 | 60 | <2 | 47.4 | 710 | 39.4 | 1730 |
|  | 3 | 9.8 | 180 | 90 | 1120 | 260 | 60 | <2 | 46.9 | 582 | 40.6 | 3580 |
|  | $\overline{\mathrm{x}}$ | 9.7 | 215 | 76 | 1290 | 214 | 60 | $<2$ | 50.2 | 691 | 41 | 2500 |
|  | s.D. | 3.6 | 61 | 19 | 208 | 42 | 0 |  | 5.2 | 100 | 3 | 963 |
| 2 | 4 | 25.2 | 125 | 11 | 702 | 7160 | 70 | $<2$ | 19.0 | 270 | 33.1 | 7860 |
|  | 5 | 21.4 | 130 | 16 | 768 | 6300 | 80 | $<2$ | 21.6 | 287 | 33.8 | 7640 |
|  | 6 | 29.2 | 115 | 11 | 684 | 8220 | 60 | <2 | 18.9 | 248 | 33.4 | 7600 |
|  | $\overline{\mathrm{x}}$ | 25.3 | 123 | 13 | 718 | 7230 | 70 | ¢2 | 19.8 | 268 | 33 | 7700 |
|  | s.c | 3.9 | 8 | 3 | 44 | 962 | 10 | 0 | 1.5 | 20 | 0 | 140 |
| 3 | 7 | 5.3 | 285 | 28 | 1050 | 37 | 60 | 2 | 40.2 | 692 | 46.0 | 169 |
|  | 8 | 4.6 | 255 | 27 | 1060 | 19 | 70 | $\stackrel{1}{ }$ | 42.1 | 740 | 45.4 | 144 |
|  | 9 | 4.1 | 265 | 30 | 1120 | 19 | 60 | 2 | 44.0 | 770 | 48.2 | 155 |
|  | $\bar{x}$ | 4.7 | 268 | 28 | 1080 | 25 | 63 | <2 | 42.1 | 734 | 47 | 156 |
|  | s.D. | 0.6 | 15 | 2 | 38 | 10 | 6 | 0 | 1.9 | 39 | 1 | 13 |

appendix int table 3 sediment metals antalysis for july 9-10, 1985 (continued)

| STATION NUMBER | SAMPLE NUMBER | $\begin{gathered} \mathrm{Mo} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Na} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Ni} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\frac{p}{(\mathrm{mg} / \mathrm{kg})}$ | $\begin{gathered} \mathrm{Pb} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Si} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Sn} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Sr} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Ti}}$ | $\frac{V}{(\mathrm{mg} / \mathrm{kg})}$ | $\begin{gathered} \mathrm{Zn} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | 10 | 7.2 | 220 | 30 | 972 | 929 | 60 | 2 | 36.0 | 520 | 39.3 | 1600 |
|  | 11 | 6.6 | 215 | 32 | 995 | 1090 | 60 | 6 | 38.9 | 574 | 42.7 | 1730 |
|  | 12 | 7.9 | 190 | 35 | 1010 | 1010 | 70 | 9 | 33.9 | 523 | 38.1 | 1670 |
|  | $\overline{\mathrm{x}}$ | 7.2 | 208 | 32 | 992 | 1010 | 63 | 6 | 36.3 | 539 | 40 | 1670 |
|  | S.D. | 0.7 | 16 | 3 | 19 | 81 | 6 | 4 | 2.5 | 30 | 2 | 65 |
| 7 | 19 | $<.8$ | 180 | 49 | 505 | 4420 | 800 | 27 | 29.1 | 346 | 28.2 | 5080 |
|  | 20 | <. 8 | 170 | 51 | 549 | 4350 | 810 | 11 | 31.0 | 385 | 31.4 | 5070 |
|  | 21 | <. 8 | 120 | 38 | 388 | 3480 | 480 | 3 | 22.5 | 306 | 22.7 | 4610 |
|  | x | <0.8 | 157 | 46 | 481 | 4080 | 697 | 14 | 27.5 | 345 | 27 | 4920 |
|  | s.d. | 0 | 32 | 7 | 83 | 524 | 188 | 12 | 4.5 | 39 | 4 | 269 |
| 8 | 22 | <. 8 | 340 | 25 | 672 | 258 | 700 | 8 | 28.3 | 1200 | 49.6 | 390 |
|  | 23 | <. 8 | 230 | 24 | 776 | 424 | 740 | <2 | 28.6 | 813 | 39.1 | 798 |
|  | 24 | <. 8 | 170 | 20 | 554 | 228 | 740 | 8 | 20.1 | 568 | 26.9 | 419 |
|  | $\overline{\mathrm{x}}$ | <0.8 | 247 | 23 | 667 | 303 | 727 | $<8$ | 25.7 | 860 | 39 | 536 |
|  | S.D. | 0 | 86 | 3 | 111 | 106 | 23 | 0 | 4.8 | 319 | 11 | 228 |

appendix III table 3 sediment metals analysis for July 9-10, 1985

| $\begin{aligned} & \text { STATION } \\ & \text { NUMBER } \end{aligned}$ | SAMPLE number | $\begin{gathered} \left.\mathrm{mo}_{0}\right) \\ \mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Na}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Ni}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{P}}$ | $\frac{\mathrm{pb}}{(\mathrm{mg} / \mathrm{kg})}$ | $\frac{\mathrm{si}}{(\mathrm{mg} / \mathrm{kg})}$ |  | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Sr}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Ti}}$ | $\frac{\mathrm{V}}{(\mathrm{mg} / \mathrm{kg})}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{zn}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 9 | 25 | <.8 | 190 | 50 | 610 | 4730 | 560 | <2 | 34.1 | 478 | 30.3 | 3300 |
|  | 26 | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A |
|  | 27 | <. 8 | 120 | 22 | 501 | 9120 | 770 | 13 | 19.1 | 284 | 23.0 | 3390 |
|  | x | <0.8 | 155 | 36 | 556 | 6920 | 665 | <13 | 26.6 | 381 | 27 | 3340 |
|  | s.d. | - | - | - | - |  | - | - | - |  |  |  |
| 10 | 28 | <. 8 | 360 | 40 | 959 | 393 | 650 | 25 | 33.5 | 623 | 38.3 | 695 |
|  | 29 | <. 8 | 270 | 32 | 761 | 349 | 800 | 29 | 26.0 | 589 | 33.6 | 498 |
|  | 30 | <. 8 | 190 | 37 | 695 | 716 | 810 | 24 | 30.9 | 574 | 34.8 | 852 |
|  | $\overline{\mathrm{x}}$ | <0.8 | 273 | 36 | 805 | 486 | 753 | 26 | 30.1 | 595 | 36.0 | 682 |
|  | s.d. | 0 | 85 | 4 | 137 | 200 | 90 | 3 | 3.8 | 25 | 2.0 | 177 |
| 11 | 31 | <. 8 | 220 | 25 | 516 | 99 | 740 | 35 | 25.0 | 483 | 30.5 | 122 |
|  | 32 | <. 8 | 160 | 27 | 569 | 34 | 830 | $\stackrel{2}{ }$ | 26.7 | 518 | 35.6 | 85.4 |
|  | 33 | <. 8 | 180 | 25 | 476 | 47 | 810 | 15 | 23.5 | 448 | 28.0 | 88.8 |
|  | $\overline{\mathrm{x}}$ | $<0.8$ | 187 | 26 | 520 | 60 | 793 | <20 | 25.1 | 483 | 31.0 | 99.0 |
|  | s.d. | 0 | 31 | 1 | 47 | 34 | 47 | 18 | 1.6 | 35 | 4.0 | 20.0 |

appendix iti table 4 sediment metais andalysis for august 21-22

| STATION | SAMPLE <br> \# | $\frac{\mathrm{Al}}{(\mathrm{mg} / \mathrm{kg})}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{As}}$ | $\begin{gathered} \mathrm{Ba} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Be}}$ | $\frac{\mathrm{ca}}{(\mathrm{mg} / \mathrm{kg})}$ |  | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{co}}$ |  |  | $\begin{gathered} F e \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Mg}}$ | $\mathrm{m}_{(\mathrm{mg} / \mathrm{kg})}^{\mathrm{Mn}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | 6860 | 49 | 286 | <2 | 17900 | 5.8 | 10.6 | 19.6 | 40.9 | 28700 | 8190 | 2830 |
|  | 2 | 7150 | 69 | 468 | <2 | 16300 | 15.8 | 18.7 | 25.9 | 40.4 | 28900 | 7900 | 1670 |
|  | 3 | 7510 | 79 | 395 | <2 | 17100 | 15.9 | 17.9 | 24.0 | 48.0 | 31300 | 8390 | 1760 |
|  | x | 7170 | 66 | 383 | <2 | 17100 | 12.5 | 15.7 | 23.2 | 43.1 | 29600 | 8160 | 2090 |
|  | s.d. | 326 | 15 | 92 | 0 | 800 | 5.8 | 4.5 | 3.2 | 4.3 | 1450 | 246 | 645 |
| 2 | 4 | 7340 | 1030 | 186 | $<2$ | 7440 | 93.9 | 14.9 | 16.2 | 90.8 | 73500 | 5760 | 230 |
|  | 5 | 7640 | 1040 | 183 | <2 | 7340 | 95.4 | 9.2 | 15.9 | 95.4 | 78900 | 6040 | 7 |
|  | 6 | 8440 | 963 | 201 | <2 | 8260 | 99.5 | 8.8 | 17.4 | 94.0 | 72400 | 6110 | 163 |
|  | $\overline{\mathbf{x}}$ | 7810 | 1010 | 190 | <2 | 7680 | 96.3 | 11.0 | 16.5 | 93.4 | 74900 | 5970 | 133 |
|  | s.d. | 569 | 42 | 10 | 0 | 505 | 2.9 | 3.4 | 0.8 | 2.4 | 3480 | 185 | 114 |
| 3 | 7 | 10300 | <8 | 226 | <2 | 9390 | 1.0 | 10.0 | 20.2 | 25.8 | 31500 | 7030 | 612 |
|  | 8 | 9570 | <8 | 198 | <2 | 9920 | $<.3$ | 1.6 | 18.6 | 21.4 | 23900 | 7550 | 456 |
|  | 9 | 9620 | <8 | 170 | <2 | 10100 | 0.9 | 7.1 | 19.4 | 24.6 | 23500 | 7670 | 424 |
|  | $\overline{\mathrm{x}}$ | 9830 | <8 | 198 | <2 | 9800 | 0.6 | 6.2 | 19.4 | 23.9 | 26300 | 7420 | 497 |
|  | s.d. | 408 | 0 | 28 | 0 | 369 | 0.6 | 4.3 | 0.8 | 2.3 | 4510 | 340 | 101 |



| Station | SAMPLE | $\frac{\mathrm{Al}}{(\mathrm{mg} / \mathrm{kg})}$ | $\stackrel{\mathrm{As}}{(\mathrm{mg} / \mathrm{kg})}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Ba}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Be})}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{ca}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Cd}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Co}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Cr}}$ | ${\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{cu}}}_{\text {( }}^{\text {( }}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Fe}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Mg}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Mn}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | 10 | 9790 | 126 | 177 | <2 | 9190 | 14.3 | 8.3 | 18.9 | 33.1 | 30600 | 6760 | 3020 |
|  | 11 | 10900 | 143 | 266 | <2 | 10200 | 19.6 | 8.9 | 21.6 | 46.2 | 36400 | 7040 | 2900 |
|  | 12 | 9630 | 27 | 225 | <2 | 11000 | 6.5 | 10.1 | 20.3 | 26.9 | 28300 | 7850 | 1730 |
|  | $\overline{\mathrm{x}}$ | 10100 | 99 | 223 | <2 | 10100 | 13.5 | 9.1 | 20.3 | 35.4 | 31800 | 7220 | 2550 |
|  | s.d. | 692 | 63 | 45 | 0 | 907 | 6.6 | 1.9 | 1.4 | 9.9 | 4170 | 566 | 713 |
| 7 | 19 | 6170 | 291 | 376 | <2 | 10200 | 99.9 | 16.4 | 21.8 | 156.0 | 49800 | 500 | <. 2 |
|  | 20 | 5620 | 278 | 348 | <2 | 7340 | 87.6 | 13.0 | 23.1 | 149.0 | 51900 | 3690 | <. 2 |
|  | 21 | 7410 | 387 | 521 | <2 | 10800 | 123.0 | 24.2 | 24.7 | 179.0 | 56200 | 4500 | <. 2 |
|  | $\bar{x}$ | 6400 | 319 | 415 | <2 | 9450 | 104 | 17.9 | 23.2 | 161.3 | 52600 | 2900 | <0.2 |
|  | s.d. | 917 | 60 | 93 | 0 | 1850 | 18.0 | 5.7 | 1.5 | 15.7 | 3260 | 2120 | 0 |
| 8 | 22 | 7260 | 119 | 116 | <2 | 10100 | 31.0 | 9.0 | 18.4 | 34.6 | 31600 | 6760 | 2130 |
|  | 23 | 7550 | 78 | 105 | <2 | 7100 | 29.6 | 8.1 | 17.7 | 36.5 | 22600 | 5050 | 2530 |
|  | 24 | 7840 | 110 | 125 | <2 | 8880 | 26.6 | 7.6 | 20.1 | 36.2 | 28600 | 6150 | 2400 |
|  | ¢ s.d. | 7550 290 | 102 22 | 115 10 | <2 | 8690 1510 | 29.1 2.2 | 8.2 0.7 | 18.7 1.2 | 35.8 1.0 | 27600 4580 | 5990 867 | 2350 204 |

appendix IIt

| Station | SAMPIE | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Al}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{As}}$ | $\begin{gathered} \mathrm{Ba} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Be} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Ca} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{cd}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Co}}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{cr})}$ | $\frac{\mathrm{Cu}}{(\mathrm{mg} / \mathrm{kg})}$ | $\begin{gathered} \mathrm{Fe} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Mg})}$ | $\begin{gathered} \mathrm{mn} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 9 | 25 | 7520 | 106 | 797 | $<2$ | 33900 | 54.9 | 18.8 | 23.9 | 85.8 | 24700 | 20200 | 627 |
|  | 26 | 5070 | 117 | 814 | $<2$ | 7740 | 48.4 | 13.2 | 28.6 | 132.0 | 35800 | 3980 | <. 2 |
|  | 27 | 8480 | 173 | 1080 | <2 | 10600 | 85.0 | 23.2 | 26.6 | 132.0 | 32900 | 5360 | <. 2 |
|  | $\stackrel{\rightharpoonup}{x}$ | 7020 | 132 | 897 | <2 | 17400 | 62.8 | 18.4 | 26.4 | 117 | 31100 | 9850 | <627 |
|  | s.d. | 1760 | 36 | 159 | 0 | 14300 | 19.5 | 5.0 | 2.4 | 26.7 | 5750 | 8990 | 0 |
| 10 | 28 | 11100 | 67 | 392 | <2 | 7830 | 11.2 | 11.7 | 34.7 | 47.9 | 38100 | 6690 | 2610 |
|  | 29 | 6970 | 85 | 288 | <2 | 4740 | 7.6 | 9.6 | 23.0 | 47.8 | 27600 | 3510 | 2520 |
|  | 30 | 8810 | 78 | 333 | <2 | 6620 | 8.0 | 9.0 | 34.1 | 56.1 | 30200 | 5010 | 2730 |
|  | к | 8960 | 77 | 338 | く2 | 6400 | 8.9 | 10.1 | 30.6 | 50.6 | 32000 | 5070 | 2620 |
|  | S.D. | 2070 | , | 52 | - | 1560 | 2.0 | 1.4 | 6.6 | 4.8 | 5470 | 1590 | 105 |
| 11 | 31 | 7750 | 88 | 122 | ${ }^{2}$ | 2950 | 0.7 | 9.4 | 24.8 | 27.8 | 26700 | 3120 | 505 |
|  | 32 | 8330 | 93 | 129 | ${ }^{2}$ | 3560 | 1.1 | 13.2 | 34.2 | 30.4 | 29700 | 3300 | 875 |
|  | 13 | 7500 | 81 | 93 | <2 | 3420 | 0.5 | 9.8 | 32.4 | 26.4 | 33300 | 3110 | 503 |
|  | $\begin{aligned} & \overline{\mathrm{x}} \\ & \text { S.D. } \end{aligned}$ | 7860 426 | 87 6 | 115 19 | ¢ 2 | 3310 320 | 0.8 0.3 | 10.8 2.1 | 30.5 5.0 | 28.2 2.0 | 29900 3310 | 3177 107 | 628 214 |

table 4 SEdiment metals analysis for august 21 - 22 (continued)
APPENDIX III

| STATION | SAMPLE \# | $\begin{gathered} \mathrm{Mo} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Na} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Ni}}$ | $\begin{gathered} \mathrm{P} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Pb} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Si} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Sn} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Sr} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Ti} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\frac{V}{(\mathrm{mg} / \mathrm{kg})}$ | $\frac{\mathrm{Zn}}{(\mathrm{mg} / \mathrm{kg})}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 1 | $<0.8$ | 130 | 42 | 1160 | 68 | 460 | $<2$ | 45.3 | 520 | 35.5 | 1180 |
|  | 2 | $<0.8$ | 140 | 69 | 1190 | 224 | 690 | <2 | 47.5 | 506 | 34.1 | 2990 |
|  | 3 | $<0.8$ | 120 | 67 | 1050 | 154 | 770 | <2 | 48.8 | 500 | 33.8 | 2800 |
|  | $\bar{x}$ | $<0.8$ | 130 | 59 | 1130 | 149 | 640 | <2 | 47.2 | 509 | 34.5 | 2320 |
|  | S.D. | 0.0 | 10 | 15 | 74 | 78 | 161 | 0 | 1.8 | 10 | 0.9 | 995 |
| 2 | 4 | $<0.8$ | 100 | 29 | 871 | 4090 | 800 | <2 | 24.1 | 234 | 32.9 | 5940 |
|  | 5 | $<0.8$ | 110 | 25 | 861 | 4320 | 520 | <2 | 24.0 | 247 | 34.4 | 5920 |
|  | 6 | $<0.8$ | 120 | 34 | 888 | 3710 | 660 | <2 | 27.4 | 245 | 34.4 | 6430 |
|  | $\overline{\mathbf{x}}$ | <0.8 | 110 | 29 | 873 | 4040 | 660 | <2 | 25.2 | 242 | 33.9 | 6100 |
|  | S.D. | 0.0 | 10 | 5 | 14 | 308 | 140 | 0 | 1.9 | 7 | 0.9 | 289 |
| 3 | 7 | $<0.8$ | 210 | 31 | 1080 | 37 | 500 | 3 | 35.5 | 497 | 43.4 | 201 |
|  | 8 | $<0.8$ | 180 | 27 | 1110 | 14 | 460 | <2 | 36.6 | 511 | 40.6 | 137 |
|  | 9 | $<0.8$ | 160 | 29 | 1090 | 20 | 230 | <2 | 34.6 | 471 | 39.5 | 168 |
|  | $\overline{\mathbf{x}}$ | $<0.8$ | 183 | 29 | 1090 | 24 | 397 | <3 | 35.6 | 493 | 41.2 | 169 |
|  | S.D. | 0.0 | 25 | 2 | 15 | 12 | 146 | 2 | 1.0 | 20 | 2.0 | 32 |

appendix ili table 4 sediment metais analysis for august 21 - 22

| STATION | $\begin{gathered} \text { SAMPLE } \\ \# \end{gathered}$ | $\begin{gathered} M o \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Na} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} N i \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} P \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{pb} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{si} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Sn} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Sr} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Ti} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\frac{V}{(\mathrm{mg} / \mathrm{kg})}$ | $\begin{gathered} \mathrm{zn} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 4 | 10 | $<0.8$ | 160 | 30 | 1000 | 538 | 350 | <2 | 34.0 | 396 | 38.1 | 1060 |
|  | 11 | $<0.8$ | 190 | 35 | 1060 | 753 | 530 | <2 | 40.2 | 396 | 42.7 | 1480 |
|  | 12 | $<0.8$ | 170 | 30 | 1130 | 98 | 460 | <2 | 36.8 | 533 | 43.1 | 633 |
|  | $\bar{x}$ | $<0.8$ | 173 | 32 | 1063 | 463 | 447 | <2 | 37.0 | 442 | 41.3 | 1060 |
|  | S.D. | 0.0 | 15 | 3 | 65 | 334 | 91 | 0 | 3.1 | 79 | 2.8 | 424 |
| 7 | 19 | $<0.8$ | 140 | 56 | 490 | 4810 | 890 | 11 | 38.6 | 338 | 28.9 | 6010 |
|  | 20 | $<0.8$ | 120 | 48 | 484 | 4660 | 800 | 32 | 33.4 | 278 | 27.0 | 5500 |
|  | 21 | $<0.8$ | 180 | 79 | 550 | 5720 | 1140 | 7 | 52.0 | 284 | 30.9 | 8090 |
|  | $\overline{\mathrm{x}}$ | $<0.8$ | 147 | 61 | 508 | 5060 | 943 | 17 | 41.3 | 300 | 28.9 | 6530 |
|  | S.D. | 0.0 | 31 | 16 | 36 | 574 | 176 | 13 | 9.6 | 33 | 2.0 | 1370 |
| 8 | 22 | $<0.8$ | 180 | 20 | 719 | 804 | 750 | 6 | 23.7 | 753 | 35.1 | 1940 |
|  | 23 | $<0.8$ | 180 | 18 | 704 | 1010 | 740 | <2 | 21.7 | 636 | 33.4 | 1820 |
|  | 24 | $<0.8$ | 190 | 25 | 748 | 816 | 660 | 11 | 25.9 | 660 | 36.0 | 1730 |
|  | $\overline{\mathbf{x}}$ | $<0.8$ | 183 | 21 | 724 | 877 | 717 | 9 | 23.8 | 683 | 34.8 | 1830 |
|  | S.D. | 0.0 | 6 | 4 | 22 | 116 | 49 | 6 | 2.1 | 62 | 1.3 | 105 |

appendix ili table 4 SEDIment metals analysis for august 21 - 22

| STATION | SAMPLE | $\mathrm{Mo}_{(\mathrm{mg} / \mathrm{kg})}$ | $\frac{\mathrm{Na}}{(\mathrm{mg} / \mathrm{kg})}$ | $(\mathrm{mg} / \mathrm{kg})$ | $\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{P}}$ | $\begin{gathered} \mathrm{Pb} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{si} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{Sn} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | $\begin{gathered} \mathrm{sr} \\ (\mathrm{mg} / \mathrm{kg}) \end{gathered}$ | ${\underset{(\mathrm{mg} / \mathrm{kg})}{\mathrm{Ti}}}^{\text {and }}$ | $\frac{v}{(\mathrm{mg} / \mathrm{kg})}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 9 | 25 | ${ }^{0} 0.8$ | 210 | 61 | 683 | 2290 | 900 | 8 | 37.6 | 538 | 32.6 | 3110 |
|  | 26 | <0.8 | 150 | 58 | 507 | 5270 | 600 | 34 | 28.7 | 556 | 30.4 | 2750 |
|  | 27 | <0.8 | 190 | 82 | 717 | 3610 | 750 | 16 | 37.8 | 534 | 36.2 | 4280 |
|  | $\overline{\mathrm{x}}$ | $<0.8$ | 183 | 67 | 636 | 3720 | 750 | 19 | 34.7 | 543 | 33.1 | 3380 |
|  | s.d. | 0.0 | 31 | 13 | 113 | 1490 | 150 | 13 | 5.2 | 12 | 2.9 | 800 |
| 10 | 28 | <0.8 | 230 | 53 | 1050 | 282 | 1000 | 25 | 33.2 | 552 | 38.0 | 764 |
|  | 29 | <0.8 | 210 | 41 | 971 | 198 | 810 | 27 | 25.6 | 360 | 30.7 | 522 |
|  | 30 | <0.8 | 230 | 43 | 927 | 246 | 820 | 25 | 29.9 | 621 | 44.0 | 639 |
|  | $\overline{\mathbf{x}}$ | <0.8 | 223 | 46 | 983 | 242 | 877 | 26 | 29.6 | 511 | 37.6 | 642 |
|  | s.d. | 0.0 | 12 | 6 | 62 | 42 | 107 | 1 | 3.8 | 135 | 6.7 | 121 |
| 11 | 31 | <0.8 | 220 | 30 | 494 | 41 | 890 | 9 | 24.9 | 509 | 29.2 | 105 |
|  | 32 | <0.8 | 180 | 30 | 585 | 73 | 900 | 16 | 29 | 573 | 37.1 | 151 |
|  | 33 | <0.8 | 180 | 26 | 613 | 43 | 900 | 6 | 27.4 | 721 | 44.8 | 93 |
|  | $\overline{\mathbf{x}}$ | ¢0.8 0.0 | 193 23 | 29 | 564 62 | 52 18 | 897 6 | 10 5 | 27.1 2.1 | 601 109 | 37.0 7.8 | 216 31 |

## APPENDIX IV

BOTTOM FAUNA DATA

APPENDIX IV TABLE 1 TAXANOMIC LIST OF BOTTOM FAUNA

| NUMBER |  | INVERTEBRATE |
| :---: | :---: | :---: |
| 1 | Order: | Acari |
|  | Class: | Insecta |
| 2 | Family: | Aphididae, unid, nymph |
| 3 | order: | Coleoptera, unid., larva |
| 4 | Family: | Haliplidae, unid., adult |
|  | order: | Hymenoptera |
| 5 | Family: | Formicidae |
|  | order: | Ephemeroptera |
|  | Family: | Baetidae |
| 6 |  | Baetis sp. |
|  | Family: | Heptageniidae |
| 7 |  | Rithrogena sp. |
| 8 |  | Cinygmula sp. |
|  | Family: | Leptophlebiidae |
|  | Family: | Ephemerellidae |
| 9 |  | Ephemerella sp. unid., damaged |
| 10 |  | Ephemerella (drunclla) |
| 11 |  | Ephemerella doddsi |
| 12 |  | Ephemerella infrequens |
|  | Family: | Siphlonuridae |
| 13 |  | Ameletus sp. |
|  | Order: | Plecoptera |
|  | Family: | Pteronarcidae |
| 14 |  | Pteronarcys dorsata |
| 15 |  | Pteronarcys californica |
| 16 |  | Pteronarcys regularis |
|  | Family: | Perlidae |
| 17 |  | Acroneuria sp. |
|  | Family: | Perlodidae |
| 18 |  | Cultus sp. |
| 19 |  | Arcynoptery (sompacta) |
| 20 |  | Isoperla sp. |


| Family: | Chloroperlidae |
| :---: | :---: |
|  | Alloperla sp. |
|  | Utaperla sp. |
| Family: | Capniidae, unid. |
| Family: | Nemouridae |
|  | Zapada (oregonensis) |
|  | Podmosta sp. |
|  | Malenka sp. |
| Order: | Trichoptera |
|  | Trichoptera pupa, unid. |
| Family: | Hydropsychidae |
|  | Arctopsyche sp. |
| Family: | Hydroptilidae |
|  | Oxyethira sp. |
|  | Hydroptila sp. |
| Family: | Brachycentridae |
|  | Brachycentrus sp. |
| Family: | Limnephilidae |
|  | (Clostonca) sp . |
|  | Onocosmoecus sp. |
|  | Phyaccophila (acropedes) |
|  | Phyaccophila vaccua |
| Order: | Diptera |
| Family: | Simulidae |
| Family: | Simulidae, adult, dam. |
|  | Simulium sp. pupae |
|  | Simulium sp. larvae |
| Family: | Chironomidae, adult |
|  | Chironomidae pupae |
| Subfamily: | Orthocladiinae |


| APPENDIX | IV TABLE 1 | TAXANOMIC LIST OF BOTTOM FAUNA |
| :---: | :---: | :---: |
| NUMBER |  | INVERTEBRATE |
| 41 |  | Cricotopus sp. |
| 42 |  | Heterotrissocladius sp. |
| 43 |  | Cardiocladius sp. |
| 44 |  | Eukiefferiella sp. |
| 45 |  | Diplocladius sp. |
| 46 |  | Brillia sp. |
|  | Subfamily: | Chironominae |
| 47 |  | Micropsectra sp. |
| 48 |  | Rheotanytarsus sp. |
| 49 |  | Stenochironomus sp. |
|  | Subfamily: | Diamesinae |
| 50 |  | Diamesa sp. |
| 51 |  | Procladius sp. |
|  | Family: | Tipulidae |
| 52 |  | Tipulidae pupae |
| 53 |  | Tipula sp. |
|  | Family: | Psychodidae |
| 54 |  | Psychoda sp. |
|  | Family: | Empididae |
| 55 |  | Chelifera sp. |
| 56 | Phylum: | Nematoda |
|  | Phylum: | Annelida |
|  | Class: | Oligochaeta |
| 57 | Family: | Enchytraeidae |
| 58 | Family: | Tubificidae, unid., uv. |
| 59 |  | Tubifex sp. |
|  | Order: | Lumbriculida |
|  | Family: | Lumbriculidae |
| 60 |  | Kincaidiana hexatheca |

cricotopus Heterotrissocladius sp. Cardiocladius sp. Eukiefferiella sp. Diplocladius sp. Brillia sp. Chironominae Micropsectra sp. eotanytarsus sp. stenochironous sp.

Diamesa sp. Procladius sp.

| NUMBER |  | INVERTEBRATE |
| :---: | :---: | :---: |
|  | Phylum: | Mollusca |
|  | Order: | Gastropoda |
| 61 | Family: | Lymnasidas |
|  |  | Stagnicola (kennicotti) |
|  | Family: | Valvatidae |
| 62 |  | Valvata sincera |
|  | Family: | Coelenterata |
| 63 |  | Hydra sp. |
|  | Phylum: | Copepoda |
|  | Suborder: | Calanoida |
| 64 |  | Diaptomus sp. |
| 65 | Suborder: | Cyclopoida |
|  | Order: | Amphipoda |
|  | Suborder: | Haustoridae |
| 66 |  | pontoporela sp. |
| 67 | Order: | Arachnida |

INVERTEBRATE DISTRIBUTION
TABLE 2
appendix IV

| number |  | invertebrate | STA. 1 | $\begin{gathered} \frac{\%}{\%} \text { of } \\ \text { Total } \end{gathered}$ | STA. 2 | $\begin{gathered} \text { q of } \\ \text { Total } \end{gathered}$ | STA. 3 | $\begin{gathered} \begin{array}{c} \text { \% of } \\ \text { Total } \end{array} \end{gathered}$ | STA. 4 | $\begin{gathered} \substack{q_{0} \text { of } \\ \text { Total }} \end{gathered}$ | STA. 7 | $\begin{aligned} & \hline \frac{\%}{} \text { of } \\ & \text { Total } \end{aligned}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Order: | Acari | 3 | 0.5 |  | 0.0 | 1 | 0.1 | 1 | 0.5 |  | 0.0 |  |
| 2 | Family: | Aphididae, unid, nymph |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 | 2 | 2.1 |  |
| 3 | order: | coleoptera, unid., larva |  | 0.0 |  | 0.0 |  | 0.0 | 4 | 1.9 |  | 0.0 |  |
| 4 | Family: | Haliplidae, unid., adult |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 5 | Family: | Formicidae |  | 0.0 | 1 | 1.1 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| ${ }^{6}$ |  | Baetis sp. | 2 | 0.3 |  | 0.0 | 9 | 1.3 | 7 | 3.4 |  | 0.0 |  |
| 7 |  | Rithrogena sp. |  | 0.0 |  | 0.0 | 11 | 1.6 | 1 | 0.5 |  | 0.0 |  |
| 8 |  | Cinygmula sp. |  | 0.0 |  | 0.0 | 4 | 0.6 | 1 | 0.5 |  | 0.0 |  |
| 9 |  | Ephemerella sp. unid., dam. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 10 |  | Ephemerolla (Drunella) |  | 0.0 |  | 0.0 |  | 0.0 | 1 | 0.5 | 1 | 1.1 |  |
| 11 |  | Ephemerella doddsi |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 12 |  | Ephemerella infrequens |  | 0.0 |  | 0.0 | 8 | 1.2 | 11 | 5.3 |  | 0.0 |  |
| 13 |  | Ameletus sp. | 3 | 0.5 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 14 |  | Pteronarcys dorsata |  | 0.0 |  | 0.0 |  | 0.0 | 1 | 0.5 |  | 0.0 |  |
| 15 |  | Pteronarcys californica |  | 0.0 |  | 0.0 | 5 | 0.7 | 5 | 2.4 |  | 0.0 |  |
| 16 |  | Pteronarcys |  | 0.0 |  | 0.0 |  | 0.0 | 32 | $\begin{array}{r}0.5 \\ \hline 15.5\end{array}$ |  | 0.0 |  |
| 17 |  | Acroneuria sp. |  | 0.0 |  | 0.0 | 26 | 3.8 | 32 | 15.5 |  | 0.0 |  |
| 18 |  | cultus sp. |  | 0.0 |  | 0.0 | 1 | 0.1 |  | 0.0 |  | 0.0 |  |
| 19 |  | Arcynopteryx (compacta) | 4 | 0.6 | 3 | 3.3 |  | 0.0 | 4 | 1.9 |  | 0.0 |  |
| 20 |  | Isoperla sp. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 21 | Family: | Alloperia sp. |  | 0.0 |  | 0.0 |  | 0.0 | 16 | 7.8 |  | 0.0 |  |
| 22 |  | Utaperla sp . | 1 | 0.2 |  | 0.0 | 3 | 0.4 | 13 | 6.3 |  | 0.0 |  |
| 23 | Family: | Capniidae, unid. | 5 | 0.8 | 7 | 7.8 |  | 0.0 | 1 | 0.5 | 7 | 7.4 |  |
| 24 |  | Zapada (oregonensis) | 49 | 7.5 | 3 | 3.3 | 20 | 2.9 | 23 | 11.2 |  | 0.0 |  |
| 25 |  | Podmosta sp. | 7 | 1.1 |  | 0.0 | 1 | 0.1 |  | 0.0 | 32 | 33.7 |  |
| 26 |  | Malenka sp. |  | 0.0 |  | 0.0 | 8 | 1.2 | 38 | 18.4 |  | 0.0 |  |
| 27 |  | Trichoptera pupa, unid. | ${ }^{2}$ | 0.2 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 26 |  | Arctopsyche sp. | 1 | 0.2 |  | 0.0 | 44 | 6.4 | 22 | 10.7 | 11 | 11.6 |  |
| 29 |  | Oxyethira sp. | 1 | 0.2 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 30 |  | Hydroptilia sp. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 31 |  | Brachycentrus sp. |  | 0.0 |  | 0.0 | 1 | 0.1 |  | 0.0 |  | 0.0 |  |
| 32 33 |  |  | 4 | 0.6 0.0 | 2 | 2.2 0.0 |  | 0.0 0.0 |  | 0.0 0.0 |  | 0.0 0.0 |  |
| 34 |  | Ehyaccophila (acropedes) | 1 | 0.2 |  | 0.0 | 4 | 0.6 | 6 | 2.9 |  | 0.0 |  |
| 35 |  | Rhyaccophila vaccua | 1 | 0.2 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| ${ }^{36}$ | Family: | Simulidae, adult, dam. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  |
| 37 |  | Simulium sp. pupae | 93 | 14.2 | 10 | 1.1 | ${ }^{347}$ | 50.8 | ${ }^{2}$ | 1.0 | 4 | 4.2 |  |
| 38 |  | Simulium sp. larvae | 144 | 22.0 | 10 | 11.1 | 169 | 24.7 | 9 | 4.4 | 9 | 9.5 |  |
| $\begin{aligned} & 39 \\ & 40 \end{aligned}$ | Family: | Chironomidae, adult Chironomidae pupae | 30 | 0.0 4.6 | , | 0.0 10.0 | 1 | 0.1 0.0 | 1 | 0.0 0.5 | 6 | 0.0 6.3 |  |

APPENDIX IV TABLE 2 INVERTEBRATE DISTRIBUTION

| NUMBER |  | INVERTEBRATE | STA. 8 | $\begin{gathered} \text { \% of } \\ \text { Total } \end{gathered}$ | STA. 9 | $\begin{gathered} \text { \% of } \\ \text { Total } \end{gathered}$ | STA. 10 | \% of Total | STA. 11 | $\begin{gathered} \text { gof of } \\ \text { Total } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Order: | Acari |  | 0.0 |  | 0.0 | 1 | 0.4 |  | 0.0 |
| 2 | Family: | Aphididae, unid, nymph |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 3 | Order: | Coleoptera, unid., larva |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 4 | Family: | Haliplidae, unid., adult |  | 0.0 | 1 | 0.1 |  | 0.0 |  | 0.0 |
| 5 | Family: | Formicidae |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 6 |  | Baetis sp. | 8 | 5.7 | 4 | 0.3 | 10 | 4.2 | 8 | 1.6 |
| 7 |  | Rithrogena sp. | 5 | 3.6 | 1 | 0.1 |  | 0.0 | 2 | 0.4 |
| 8 |  | Cinygmula sp. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 9 |  | Ephemerella sp. unid., dam. |  | 0.0 |  | 0.0 |  | 0.0 | 2 | 0.4 |
| 10 |  | Ephemerella (Drunella) | 4 | 2.9 | 3 | 0.2 | 1 | 0.4 | 10 | 2.0 |
| 11 |  | Ephemerella doddsi |  | 0.0 |  | 0.0 |  | 0.0 | 3 | 0.6 |
| 12 |  | Ephemerella infrequens | 5 | 3.6 | 12 | 0.8 |  | 0.0 |  | 0.0 |
| 13 |  | Ameletus sp. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 14 |  | Pteronarcys dorsata | 5 | 3.6 | 10 | 0.7 |  | 0.0 |  | 0.0 |
| 15 |  | Pteronarcys californica |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 16 |  | Pteronarcys regularis | 4 | 2.9 | 12 | 0.8 |  | 0.0 | 9 | 1.8 |
| 17 |  | Acronouria sp. | 21 | 15.0 | 12 | 0.8 | 1 | 0.4 | 18 | 3.6 |
| 18 |  | cultus sp. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 19 |  | Arcynopteryx (compacta) | 2 | 1.4 | 4 | 0.3 | 2 | 0.8 | 2 | 0.4 |
| 20 |  | Isoperla sp. |  | 0.0 |  | 0.0 |  | 0.0 | 6 | 1.2 |
| 21 |  | Alloperla sp. |  | 0.0 | 1 | 0.1 | 10 | 4.2 | 34 | 0.0 |
| 22 |  | Utaperla sp. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 23 |  | Capnisdae, unid. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 24 |  | Zapada (oregonensis) |  | 0.0 | 1 | 0.1 |  | 0.0 | 34 | 6.8 |
| 25 |  | Podmosta sp. |  | 0.0 | 2 | 0.1 |  | 0.0 |  | 0.0 |
| 26 |  | Malenka sp. | 4 | 2.9 | 4 | 0.3 | 1 | 0.4 | 45 | 9.0 |
| 27 |  | Trichoptera pupa, unid. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 28 |  | Arctopsyche sp. | 30 | 21.4 | 36 | 2.4 | 7 | 3.0 | 13 | 2.6 |
| 29 |  | Oxyethira sp. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 30 |  | Hydroptila sp. |  | 0.0 |  | 0.0 | 2 | 0.8 | 1 | 0.2 |
| 31 |  | Brachycentrus sp. |  | 0.0 |  | 0.0 | 2 | 0.8 | 1 | 0.2 |
| 32 |  | (Clostoeca) sp. |  | 0.0 | 1 | 0.1 |  | 0.0 |  | 0.0 |
| 33 |  | Onocosmoecus sp. | 1 | 0.7 |  | 0.0 |  | 0.0 |  | 0.0 |
| 34 |  | Rhyaccophila (acropedes) | 1 | 0.7 | 1 | 0.1 |  | 0.0 |  | 0.0 |
| 35 |  | Rhyaccophila vaccua |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 36 | Family: | Simulidae, adult, dam. |  | 0.0 |  | 0.0 | 1 | 0.4 |  | 0.0 |
| 37 |  | Simuliun sp. pupae | 1 | 0.7 | 15 | 1.0 |  | 0.0 | 13 | 2.6 |
| 38 |  | Simulium sp. larvae | 42 | 30.0 | 1347 | 90.3 | 2 | 0.8 | 312 | 62.4 |
| 39 | Family: | Chironomidae, adult |  | 0.0 | 1 | 0.1 |  | 0.0 |  | 0.0 |
| 40 |  | Chironomidae pupae | 1 | 0.7 | 3 | 0.2 | 2 | 0.8 | 2 | 0.4 |

APPENDIX IV TABLE 2 INVERTEBRATE DISTRIBUTION

| NUMBER |  | INVERTEBRATE | STA. 1 | of Total | STA. 2 | $\begin{aligned} & \text { ? of } \\ & \text { Total } \end{aligned}$ | STA. 3 | $\begin{aligned} & \text { \% of } \\ & \text { Total } \end{aligned}$ | STA. 4 | $\begin{gathered} \text { of of } \\ \text { Total } \end{gathered}$ | STA. 7 | $\begin{gathered} \text { \% of } \\ \text { Total } \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 41 |  | Cricotopus sp. | 94 | 14.4 | 19 | 21.1 | 1 | 0.1 | 1 | 0.5 | 13 | 13.7 |
| 42 |  | Heterotrissocladius sp. |  | 0.0 | 10 | 11.1 |  | 0.0 | 1 | 0.5 | 5 | 5.3 |
| 43 |  | Cardiocladius sp. | 2 | 0.3 | 2 | 2.2 | 1 | 0.1 | 2 | 1.0 |  | 0.0 |
| 44 |  | Eukiefferiella sp. | 39 | 6.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 45 |  | Diplocladius sp. | 61 | 9.3 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 46 |  | Brillia sp. | 3 | 0.5 | 15 | 16.7 |  | 0.0 |  | 0.0 |  | 0.0 |
| 47 |  | Micropsectra sp. |  | 0.0 |  | 0.0 | 1 | 0.1 | 1 | 0.5 |  | 0.0 |
| 48 |  | Rheotanytarsus sp. |  | 0.0 |  | 0.0 | 1 | 0.1 |  | 0.0 |  | 0.0 |
| 49 |  | Stenochironomus sp. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 50 |  | Diamesa sp. | 12 | 1.8 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 51 |  | Procladius sp. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 | 1 | 0.0 |
| 52 |  | Tipulidae pupae | 1 | 0.2 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 53 |  | Tipula sp. | 7 | 1.1 | 1 | 1.1 |  | 0.0 |  | 0.0 | 2 | 2.1 |
| 54 |  | Psychoda sp. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 55 |  | Chelifera sp. |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 56 | Phylum: | Nematoda | 4 | 0.6 | 1 | 1.1 | 3 | 0.4 | 1 | 0.5 |  | 0.0 |
| 57 | Family: | Enchytraeidae | 75 | 11.5 | 1 | 1.1 | 2 | 0.3 |  | 0.0 |  | 0.0 |
| 58 | Family: | Tubificidae, unid., uv. |  | 0.0 | 1 | 1.1 | 5 | 0.7 |  | 0.0 |  | 0.0 |
| 59 |  | Tubifex sp. |  | 0.0 | 3 | 3.3 |  | 0.0 |  | 0.0 |  | 0.0 |
| 60 |  | Kincaidiana hexatheca | 2 | 0.3 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 61 |  | Stagnicola (kennicotti) |  | 0.0 |  | 0.0 | 6 | 0.9 |  | 0.0 |  | 0.0 |
| 62 |  | Valvata sincera |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 63 |  | Hydra sp |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 64 |  | Diaptomus sp. | 3 | 0.5 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 65 | Suborder: | Cyclopoida | 1 | 0.2 |  | 0.0 |  | 0.0 |  | 0.0 | 1 | 1.1 |
| 66 |  | Pontoporeia sp. | 1 | 0.2 |  | 0.0 |  | 0.0 |  | 0.0 |  | 0.0 |
| 67 | Order: | Arachnida |  | 0.0 | 1 | 1.1 |  | 0.0 |  | 0.0 | 1 | 1.11 |
| TOTAL NUMBER PER STATION |  |  | 655 |  | 90 | 683 |  | 206 |  | 95 |  |  |


*Station 10 samples collected with Surber Sampler.

