MOUNT HUNDERE BNSELINE BTUDY
June 1988 and June 1990

DATA REPORT NO, 94-02
by
Environmental Protection Branch Yukon Division

November, 1994


#### Abstract

Baseline investigations of environmental quality in the Mt. Hundere study area were carried out in June, 1988, and again in June 1990. These studies were in response to mineral exploration and eventual development of the Mt. Hundere (Sa Dena Hes) Lead/Zinc mine. The surveys investigated water quality, sediment chemistry and percent particle size distribution, and benthic invertebrates in the drainages surrounding the mineral development area.

Extractable lead and zinc concentrations found in water samples were at maximum levels recommended for aquatic life. Stream sediment chemistry was comparable to sediment compositions found in other mineralized areas in the Yukon. Benthic invertebrate populations appeared significant in numbers and diversity compared with other recent surveys.


## RÉsunci

Une étude de base de la qualité de l'environnement dans la région du Mont Hundere a été conduite en juin 1988 et de nouveau en juin 1990. Ces études répondent à l'exploration minière et l'éventuel dévelopment d'une mine de plomb/zinc (Sa Dena Hes) au Mont Hundere. Les investigations portaient sur la qualité de l'eau, la composition chimique des sédiments, la distribution du pourcentage des dimensions des particules, et des invertébrés benthiques des drainages adjacents au développement minéral de la région.

Les concentrations de plomb et zinc extractable dans les echantillons d'eau étaient au taux recomendé maximum pour la protection de la vie aquatique. La composition chimique des sédiments était comparable a celle des sédiments échantillonnés dans d'autre région minéralisés du Yukon. Les populations benthiques semblent étre plus abondantes et diversifiées que d'autre études récentes.

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#### Abstract

1.0 INTRODUCTION

The Mt. Hundere lead and zinc deposits were discovered in 1962. Significant exploratory work was done by CIMA Resources between 1979 and 1982, delineating 250,000 tonnes of ore reserves.


Canamax Resources acquired the property in 1984 and completed over 20,000 metres of diamond drilling by 1988. As a result, estimated ore reserves were increased dramatically in a number of zones surrounding Mt. Hundere. A baseline water quality survey was carried out by the Department of Indian and Northern Affairs in August, 1985 (INAC unpublished data).

Frame Mining Corporation and Hillsborough Resources Limited purchased the property in 1989 and conducted definition drilling and an environmental baseline study (Mt. Hundere Joint Venture Initial Environmental Evaluation, May 1990). Frame Mining transferred their interest to Curragh Resources Inc., which became the manager of the joint venture.

Environmental protection conducted the first of two baseline investigations in June, 1988 which included sampling of water, sediments and benthic organisms. The Department of Indian and Northern Affairs carried out a second water quality survey in September, 1988 (INAC unpublished data).

Environmental protection carried out a second baseline survey in June 1990 covering the False Canyon Creek and Tom Creek drainages. Meanwhile, the Mt. Hundere Joint Venture passed through an environmental assessment and review, and was granted a water licence in 1991. The Mt. Hundere (Sa Dena Hes) lead/zinc mine and mill began producing concentrate in August 1991.

Mt. Hundere is approximately 50 km north of Watson Lake, Yukon Territory and has an elevation of 1574 m above sea level (Eigure 1). Mineral exploration has occurred above 1219 m . The property is situated on the drainage divide between the Tom Creek and Ealse Canyon Creek catchments. Ealse Canyon Creek drains into the Erances River 55 km upstream of its confluence with the Liard River. Tom Creek Elows directly into the Liard River. The average annual precipitation for the region is between 400-600 m. The average annual daily temperature lies in the range of -9 to $-11^{\circ} \mathrm{C}$ (Wahl, 1987).

Eighteen sample stations were set up within the study area (see Table 1). Stations 1, 2 and 3 were located on an upper circumference of Mt. Hundere, in unnamed creeks. The remaining stations are located in False Creek, Tom Creek, and their tributaries, as well as in the Erances and Liard Rivers (Figure 2).



| TABLE 1 STATION | DESCRIPTIONS OF SAMPLE STATIONS IN LOCATION | MT. EUNDERE SHUDY AREA. <br> REMARKS |
| :---: | :---: | :---: |
| 1 | $60^{\circ} 31^{\prime} \mathrm{N}$ by $128^{\circ} 56^{\prime} \mathrm{W}$, headwaters of Tom Creek, Tributary 5, 0.5 km from exploration camp, at water supply pump station. Elevation 1190 m . | Samples were taken downstream of camp and upstream of pump station. Banks 10\% overhung with Willow. Substrate of mixed cobble. |
| 2 | $60^{\circ} 32^{\prime} \mathrm{N}$ by $128^{\circ} 57^{\prime} \mathrm{W}$, headwaters of False Canyon Creek 1.0 km north of camp, on mine road 30 m downstream of fill from road construction. Elevation 1220 m . | Substrate of large boulders to 60 cm diameter and smaller cobble. Steep gradient (3.5: 1) has stream cascading in series of pools and falls. Ice at 4 m downstream of station. |
| 3 | $60^{\circ} 33^{\prime} \mathrm{N}$ by $129^{\circ} \mathrm{O} 1^{\prime} \mathrm{W}$, headwaters of False Canyon Creek Tributary B at road crossing. | Bank heights 0.25 to 0.5 m , and are erodible and undercut; overhung by Willow. Samples taken below bridge. |
| 4 | $60^{\circ} 28^{\prime} \mathrm{N}$ by $129^{\circ} 07^{\prime} \mathrm{W}$, Frances River Tributary A 1.75 km from Campbell Hwy, upstream of bridge crossing. Elevation 740 m . | Bank heights, 0.2 to 0.4 m , erodible and undercut but with good grass cover to the bank's edge. |
| 5 | $60^{\circ} 18^{\prime} \mathrm{N}$ by $129^{\circ} 00^{\prime} \mathrm{W}$, Tom Creek, upstream of Campbell Hwy. Elevation 762 m . | Creek approximately 15m wide. and approximately 0.30 m deep. Willow along banks. |
| 6 | $60^{\circ} 20^{\prime} \mathrm{N}$ by $128^{\circ} 56^{\prime} \mathrm{W}$, Tom Creek Tributary 2, 200 m upstream of Tom Creek. Elevation 780 m . | Creek approximately 3m wide and about .4m deep. Willow along bank. |
| 7 | $60^{\circ} 22$ 'N by $128^{\circ} 51^{\prime} \mathrm{W}$, Tom Creek upstream of Tributary 2 approximately 6 km . Elevation 810m. | Creek approximately 9m wide and approximately 0.8 m deep. Willow along bank. |
| 8 | $60^{\circ} 23^{\prime} \mathrm{N}$ by $128^{\circ} 48^{\prime} \mathrm{W}$, Tom Creek Tributary 5. Elevation 830 m . | Creek approximately 5.5m wide and approximately 0.5 m deep. Willow and grasses along bank. |
| 9 | $60^{\circ} 26^{\prime} \mathrm{N}$ by $128^{\circ} 44^{\prime} \mathrm{W}$, Tom Creek approximately 2.5 km downstream of Tom Lake. Elevation 840 m . | Creek approximately 3m wide and approximately 0.3 m deep. Willow, grasses and spruce along bank. |
| 10 | $60^{\circ} 26^{\prime} \mathrm{N}$ by $128^{\circ} 45^{\prime} \mathrm{W}, \quad$ Tom Creek Tributary 4, upstream of Tom Creek $\quad$ approximately Elevation 840 m. | Creek width and depth not measured. Willow and grasses along bank. |

$60^{\circ} 31^{\prime} \mathrm{N}$ by $128^{\circ} 46^{\prime} \mathrm{W}$, Upper False
Canyon Creek approximately 9 km
downstream of Station
2. Elevation 880 m .
$60^{\circ} 36^{\prime} \mathrm{N}$ by $128^{\circ} 46^{\prime} \mathrm{W}$, False Canyon Creek. Elevation 790m.
$60^{\circ} 38^{\prime} \mathrm{N}$ by $128^{\circ} 49^{\prime} \mathrm{W}$, Oscar Creek. Elevation. 730m.
$60^{\circ} 39^{\prime} \mathrm{N}$ by $128^{\circ} 53^{\prime} \mathrm{W}$, False Canyon Creek. Elevation 720 m .
$60^{\circ} 39^{\prime} \mathrm{N}$ by $128^{\circ} 53^{\prime} \mathrm{W}$, False Canyon Creek Tributary B. Elevation 720 m .
$60^{\circ} 41^{\prime} \mathrm{N}$ by $129^{\circ} 03^{\prime} \mathrm{W}$, False Canyon Creek approximately 60m upstream of Frances River. Elevation 680 m .
$60^{\circ} 42^{\prime} \mathrm{N}$ by $129^{\circ} 03^{\prime} \mathrm{W}$, Frances River upstream of False Canyon Creek approximately 1.5 km .
$60^{\circ} 28^{\prime} \mathrm{N}$ by $129^{\circ} 07^{\prime} \mathrm{W}$, Frances River upstream of the Campbell Hwy. approximately 200m.

Creek approximately 3.5 m wide and approximately 0.4m deep. Bank height 1 m with grass and willow cover.

Creek approximately 5 m wide and approximately 0.7 m deep. Bank height 0.5 m with grass and willow cover.

Site not sampled. No suitable landing locations nearby for helicopter.

Creek width and depth not measured. No invertebrates were collected because of depth. Bank height 0.5 m with grasses and willow cover.

Creek approximately 6.5 m wide and approximately 0.8 m deep. Bank height 1.0 m with grasses and willow cover.

Creek width and depth not measured. Creek water levels greatly influenced by Frances River levels. No benthic invertebrates or sediments collected because of water depth. bank height 0.2 m with spruce, willow and grass cover.

Stream width and depth not measured. Benthic invertebrates and sediments were not sampled.

Stream width and depth not measured. Benthic invertebrates and sediments were not sampled.

$$
2.0 \text { METHOD }
$$


#### Abstract

Samples and field measurements at Mt. Hundere were taken on June 21-22, 1988, and June 19-21, 1990. At each sampling site, three replicates samples of water and sediments were taken and a composite of 3 samples for benthic invertebrates. Methods of sample collection, preservation, and analysis or identification are listed in Appendix $I$.


### 2.1 Water Chemistey

Water samples were collected at 15 sample sites. A description of water sample collection, preparation, analytical methods, and detection limits are given in Appendix $I$, Table 1. Measurements of water temperature, conductivity, pH , dissolved oxygen and flow were made at each site. Samples sent to the laboratory were analysed for alkalinity, chloride, fluoride, true colour, conductivity, pH , total hardness and hardness as $\mathrm{Ca+Mg}$, ammonia-N, nitrite+nitrate, total phosphorous, filterable (FR) and non-filterable residues (NFR), sulphate and turbidity. For the 1988 survey, extractable metals analyses were performed (ICP scan), with a request for the higher sensitivity Graphite Furnace procedure on samples where lead and zinc were below ICP detection. The 1990 survey included dissolved, extractable, and total metals analyses, with Graphite Furnace procedure requested on samples where lead, copper, cadmium, or silver were below ICF detection. Both surveys included the following metals:

| Aluminum (Al) | Copper (Cu) | Silver (A) |
| :--- | :--- | :--- |
| Antimony (Sb) | Iron (Fe) | Sodium (Na) |
| Arsenic (As) | Lead (Pb) | Strontium (Sr) |
| Barium (Ba) | Magnesium (Mg) | Tin (Sn) |
| Beryllium (Be) | Manganese (Mn) | Titanium (Ti) |
| Boron (B) | Molybdenum (Mo) | Vanadium (V) |
| Cadmium (Cd) | Nickel (Ni) | Zinc (Zn) |
| Calcium (Ca) | Phosphorous |  |
| Chromium (Cr) | (P)Selenium (Se) |  |
| Cobalt (Ca) | Silicon (Si) |  |

The analyses were completed at the Environmental Protection Service Laboratory, 4195 Marine Drive, West Vancouver, B.C.

The percent dissolved oxygen saturation point was determined by calculating the dissolved oxygen saturation point (S') from the formula:

$$
\begin{aligned}
& S^{\prime}=S \frac{P}{760} \quad(A L P H A \text { et al 1980) } \\
& \text { where, } \text { S' }^{\prime}=\text { dissolved oxygen (DO) saturation } \\
& \text { concentration at the in situ } \\
& \text { temperature and atmospheric pressure } \\
& \mathrm{S} \quad=\text { dissolved oxygen (DO) saturation } \\
& \text { concentration at sea level for the in } \\
& \text { situ temperature. } \\
& \text { P = atmospheric pressure (mm Hg) at } \\
& \text { site elevation. }
\end{aligned}
$$

The percent dissolved oxygen saturation is the ratio of field DO to the in situ saturation concentration (S'):

```
field DO * 
```

where, field $D O=$ dissolved oxygen measured in the field and adjusted for field conditions.

### 2.2 Sediments

One set of triplicate sediment samples were taken at each station. The samples were shipped to the Environmental Protection Laboratory, 4195 Marine Drive, West Vancouver, B.C. and analysed for leachable metals and percent particle size distribution according to the Wentworth Classification System. A description of sediment collection, preparation and analysis methods are given in Appendix I, Table 2.

### 2.3 Bottom Eauna

Three benthic invertebrate samples were taken with a Surber sampler (500um mesh size). The invertebrates collected from three replicate Surber samples, taken on a short reach of stream at each station, were combined in a 1 litre bottle and considered as the sample for that station. Each sample, therefore consisted of 3 grabs combined. Samples were sorted, identified and enumerated by Dr. C. Low, consulting invertebrate biologist from Nanaimo, British Columbia. The methods of sample collection and preservation are described in Appendix I, Table 3.

Indices of benthic community diversity and evenness were calculated using the following formulae (Pielou 1975):

$$
\text { Species Diversity }\left(H^{\prime}\right)=-\sum_{i=1}^{n}\left(P_{i} \log _{10} P_{i}\right)
$$

where, $\mathrm{Pi}=\mathrm{ni} / \mathrm{N}$
$n i=n u m b e r$ of individuals in the ith most specific taxonomic group (ie. genus) at one sample location.
$N=$ total number of individuals identified to specific taxonomic group (ie. genus) at one sample location.
$\mathrm{n}=$ total number of taxonomic groups (ie. genus) identified at one sample location.

Evenness $\left(J^{\prime}\right)=H^{\prime} / \log _{10} n$

Percent Similarity Index: The benthic invertebrate communities collected during the two surveys were compared using a percent similarity index (Psc) formula described by Brock (1977):

$$
\operatorname{PsC}=100-0.5 \sum_{i=1}^{k}|a-b|
$$

where $a$ and $b$ are, for a given genus, percentage of the total samples $A$ and $B$ which that genus represents. The absolute value of their difference is summed over all genera, $k$. The Psc compares the percentage of genera present at two different locations but is not a comparison of total invertebrate abundance. The information produced by the percent similarity index was plotted into a cluster using the nearest neighbour clustering method.

Table 2 shows the different field activities performed during the surveys of June 1988 and June 1990.

TABLE 2 SAMPLING PROGRAM SUMMARY

3.0 RESULTS AND DISCUSSION

### 3.1 Hater ouality

The results of water quality analyses are listed in Appendix II, Tables 1 (June, 1988) and 2 (June, 1990). Note that seasonal variability is not accounted for in the data generated by the two surveys.

Water temperatures ranged from 0.5 in the small alpine tributary to $12.5{ }^{\circ} \mathrm{C}$ in Tom Creek. Alkalinity ranged from $32 \mathrm{mg} / 1$ as $\mathrm{CaCO}_{3}$, to $192 \mathrm{mg} / 1$ as $\mathrm{CaCO}_{3}$ at station 4 , and dissolved calcium was present in concentrations over 20 $\mathrm{mg} / \mathrm{l}$ in False Canyon and Tom Creeks; indicating good buffering capacity. pH was slightly alkaline throughout the study area, ranging from 7.5 at station 3 , to 8.5 at station 4.

Tom Creek and False Canyon Creek contained moderately to very hard water. The Tom Creek catchment ranged in hardness from 81 to $147 \mathrm{mg} / 1$ as $\mathrm{CaCO}_{3}$.

The False Canyon Creek catchment ranged from 107 to $169 \mathrm{mg} / 1$ as CaCO . Conductivity readings were all less than $250 \mu \mathrm{mhos} / \mathrm{cm}$. Suspended solids and turbidity were generally very low in both catchments during the June sampling period.

Nutrient levels were at low concentrations throughout False Canyon and Tom Creek catchments. Nitrite was undetectable, and nitrite + nitrate was below $0.050 \mathrm{mg} / \mathrm{l}$ throughout the survey area. Two exceptions should however be noted. Station 2 on June 1988 had nitrite + nitrate mean levels of $0.288 \mathrm{mg} / 1$. Also one replicate at Station 5 which was $3.06 \mathrm{mg} / \mathrm{l}$, and appears to be erroneous, as the other two replicates were both below the detection limit of $0.005 \mathrm{mg} / 1$. Ammonia was generally below or near the detection limit, except for a mean value of $0.010 \mathrm{mg} / 1$ at the mouth of Tom Creek in the 1988 survey.

Metals concentrations were generally low in both catchments. The metals of interest in the combined data set are primarily copper, lead, and zinc.

Total copper ranged from $<0.0009 \mathrm{mg} / \mathrm{l}$ in the Frances River to 0.0023 mg/l in False Canyon Creek at station 12 in 1990. The recommended guideline for total copper is $0.002 \mathrm{mg} / 1$ for medium hardness, to $0.004 \mathrm{mg} / 1$ for hard water (Appendix 1, Table 4).

Total lead ranged from $0.0010 \mathrm{mg} / 1$ in Upper False Canyon Creek to $0.0096 \mathrm{mg} / \mathrm{l}$ at station 12 on False Canyon Creek in 1990. The recommended guideline for lead concentration is identical to that of copper. Sub-lethal effects of low levels of lead on various fish species have been demonstrated in lab tests, but generally in soft water for early life stages (Moore and Ramamoorthy, 1984).

Total zinc measurements were all below detection limit of $0.002 \mathrm{mg} / 1$, except for Station 2 in 1998 where the mean zinc concentration was $0.032 \mathrm{mg} / \mathrm{I}$ (S.D. 0.007). However, extractable levels were found at station 3, 11, 15 in 1990 with the values of $0.016,0.004$ and $0.003 \mathrm{mg} / 1$ respectively. These could be related to contamination since dissolved metals are also below detection limit. The recomended guideline for zinc concentration is $0.03 \mathrm{mg} / 1$ total zinc.

The acute toxicity of the above metals is modified by hardness, but
chronic toxicity is not (CCREM, 1987). Given the low volumes available for dilution in False Canyon Creek combined with background concentrations of metals, particularly lead, contributions from mine effluent would have to be low if the recommended guidelines are to be maintained instream.

## $\because \quad 3.2$ Sediments

Results of the sediment survey are presented in Appendix III, with percent particle size in Table 1 and sediment chemistry in Table 2 . Metals were listed as dry weights and recorded in $\mu \mathrm{g} / \mathrm{g}$.

Metals appear to exist in high concentrations only in sediments from station 2 collected in 1988. This site was not samples in 1990. All other sediment metals data are comparable to other mineralized areas in the Yukon (see Table 3). Concentrations of lead and zinc in sediments were one to two orders of magnitude lower than those observed in the Ketza River study area (unpublished report, Environmental Protection, 1994).

TABLE 3 COMPARATIVE SEDIMENT VALUES OF OTEER YUKON MINERALIZED AREAS ( $\mu \mathrm{g} / \mathrm{g}$ )
Grew Creek
Mount Nansen
Retza River

| Matal | Avarage | standard Dariation | Matal | Avarage | standard Deviation | Motal | Averaqe | standard Deviation |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 18 | 28.64 | 11.0 | 38 | 384 | 729 | As | 868 | 1129 |
| Cd | 40.8 | $\cdots$ | Cd | 4.2 | 7.8 | cal | 2.6 | 2.4 |
| cos | 34.5 | 12.3 | Cris | 21 | 15 | cos | 73 | 43 |
| \% | 140 | 182 | Pb | 115 | 224 | Fb | 820 | 1500 |
| 8n | 184 | 85 | 20 | 548 | 1081 | \%n | 1156 | 1012 |

### 3.3 Benthic Fauna

A taxonomic list of the benthic organisms found in the study area is presented in Appendix IV, Table 1. Sample site enumeration is in Appendix IV, Table 2. Taxonomic orders have been grouped and presented as total organisms counted and the percentage of the total at each sample station and these results are listed below in Table 4.

A total of fifty taxa were identified in the 1988 survey, which covered 5 sample stations. The 1990 survey identified 89 taxa and covered 10 stations. Both in 1988 and 1990, the dominant orders of the overall sample population were Ephemeroptera, Plecoptera and Diptera (see Table 5).
TABLE 4 SUMMARY TABLF OF INVERYGBRATE RBUNDANCF AND TAXONOMIC DISTRIBUTION FOR 1988 AND 1990

| TAXONOMIC GROUP | $\begin{array}{r} 1988 \\ \text { STA. } 1 \end{array}$ | \% | $\begin{array}{r} 1990 \\ \text { STA. } 3 \end{array}$ | \% | $\begin{array}{r} 1988 \\ \text { STA. } 4 \end{array}$ | \% | $\begin{array}{r} 1990 \\ \text { STA. } 4 \end{array}$ | \% | $\begin{array}{r} 1988 \\ \text { STA. } 5 \end{array}$ | \% | $\begin{array}{r} 1990 \\ \text { STA. } 5 \end{array}$ | \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Plecoptera | 21 | 18 | 142 | 47 | 93 | 22 | 92 | 20 | 20 | 7 | 19 | 5 |
| Ephemeroptera | 33 | 28 | 105 | 35 | 223 | 53 | 267 | 59 | 215 | 72 | 19 136 | 37 |
| Trichoptera | 10 | 8 | 0 | 0 | 6 | 1 | 26 | 6 | 9 | 3 | 136 48 | 13 |
| Diptera | 45 | 38 | 55 | 18 | 69 | 16 | 64 | 14 | 52 | 17 | 150 | 41 |
| Collembola | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 10 | 150 | 41 0 |
| Hymenoptera | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Homoptera | 0 | 0 | 1 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| Hemiptera | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| Lepidoptera | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydracrina | 5 | 4 | 0 | 0 | 14 | 3 | 0 | 0 | 1 | 0 | 0 | 0 |
| Oribatei | 0 | 0 | 0 | 0 | - 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 |
| Copepoda | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cyclopoida | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Calanoida | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Cladocera | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Nematoda | 0 | 0 | 0 | 0 | 0. | 0 | 1 | 0 | 0 | 0 | 2 | 1 |
| Turbellaria | 0 | 0 | 0 | 0 | 3. | 1 | 2 | 0 | 0 | 0 | 0 | 0 |
| Oligochaeta | 5 | 4 | 0 | 0 | 10 | 2 | 1 | 0 | 1 | 0 | 6 | 2 |
| TOTAL NUMBERS | 119 |  | 304 |  | 422 |  | 454 |  | 298 |  | 365 |  |

TABLE 4 (Cont'd)

| TAXONOMIC GROUP | $\begin{array}{r} 1990 \\ \text { STA. } 6 \end{array}$ | 8 | $\begin{array}{r} 1990 \\ \text { STA. } 7 \end{array}$ | \% | $\begin{array}{r} 1990 \\ \text { STA. } 8 \end{array}$ | \% | $\begin{array}{r} 1990 \\ \text { STA. } 9 \end{array}$ | \% | $\begin{array}{r} 1990 \\ \text { STA. } 11 \end{array}$ | \% | $\begin{array}{r} 1990 \\ \text { STA. } 12 \end{array}$ | $\%$ | $\begin{array}{r} 1990 \\ \text { STA. } 15 \end{array}$ | \% |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Plecoptera | 1 | 1 | 52 | 31 | 91 | 20 | 45 | 13 |  |  |  |  |  |  |
| Ephemeroptera | 7 | 4 | 45 | 27 | 262 | 58 | 78 | 13 | 27 239 | 9 79 | 8 | 3 | 3 | 8 |
| Trichoptera | 2 | 1 | 0 | 0 | 22 | 5 | 0 | - | 239 1 | 7 | 109 | 45 | 3 | 8 |
| Diptera | 124 | 76 | 64 | 38 | 56 | 12 | 209 | 60 | 24 | 0 | 5 | 2 | 1 | 3 |
| Collembola | 0 | 0 | 0 | 0 | 0 | 120 | 20 | 0 | 24 | 8 | 70 | 29 | 27 | 75 |
| Hymenoptera | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Homoptera | 1 | 1 | 1 | 1 | 3 | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| Hemiptera | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 3 | 1 | 0 | 0 |
| Lepidoptera | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Hydracrina | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Oribatei | 3 | 2 | 1 | 1 | 4 | 1 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Copepoda | 0 | 0 | 0 | 0 | . 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| Cyclopoida | 3 | 2 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| Calanoida | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 0 | 1 | 1 | 0 | 1 | 3 |
| Cladocera | 0 | 0 | 1 | 1 | 0 | 0 | 3 | 0 | 0 | 0 | 2 | 1 | 0 | 0 |
| Nematoda | 15 | 9 | 0 | 0 | 3 | 1 | 3 | 1 | 0 | 0 | 9 | 4 | 0 | 0 |
| Turbellaria | 0 | 0 | 1 | 1 | 11 |  | 0 | 2 | 2 | 1 | 25 | 10 | 0 | 0 |
| Oligochaeta | 7 | 4 | 1 | 1 | 11 1 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 |
| TOTAL NUMBERS | 164 |  | 167 |  | 453 |  |  |  |  |  |  |  |  |  |

TABLE 5 SUMMARY OF BENTEIC INVERTEBRATES SAMPLING EOR 1988 AND 1990


Diptera were clearly the dominant order at station 6 (76\%) and 15 (75\%) on June 1990 survey while Ephemeroptera were dominant at station 5, 1988 (72\%); station 4, 1990 (59\%); station 8, 1990 (58\%); and station 11, 1990 (79\%). Co-dominance between the above groups was found at the other stations. Nematoda were a contributing order of the community at station 6 and 12 in June 1990 with 9 and $10 \%$ respectively of the sampled population.

Benthic invertebrate analysis summary can be found in Table 5. It is composed of the total taxa, density of invertebrates, diversity and evenness indicates for each stations. Stations 1 and 15 had the highest evenness level but station 15 had the lowest density of invertebrates with only $129 / \mathrm{m}^{2}$. The stations 4 and 8 had the highest density with 1629 and 1626 respectively during the June 1990 survey. Stations 5 and 9 showed also high density with both greater than 1200 individuals $/ \mathrm{m}^{2}$. Station 5 had the greatest number of taxa (43) followed by station 12 (37).

In comparison with other baseline studies carried out by the department in 1988, which employed the same sampling method, the Mt. Hundere benthic samples are above average in total number of individuals per station, diversity and evenness (see Table 6).

TABLE 6 DIVERSITY AND EVENAESS AT OTEHER SITUDY AREAS
Grew Creek

| STATION | DENSITY <br> (ind. $\mathrm{m}^{2}$ ) | H | $\mathrm{H}^{\prime}$ |
| :---: | :---: | :---: | :---: |
| 1 | 506 | 0.66 | 0.61 |
| 2 | 226 | 0.72 | 0.72 |
| 3 | 186 | 0.55 | 0.65 |

Quill Creek

| STATION | DENSITY <br> (ind. $\mathrm{m}^{2}$ ) | $\mathrm{H}^{\prime}$ | $\mathrm{J}^{\prime}$ |
| :---: | :---: | :---: | :---: |
| 3 | 331 | 0.576 | 0.74 |
| 4 | 1168 | 0.744 | 0.618 |
| 5 | 530 | 0.180 | 0.257 |
| 6 | 530 | 0.808 | 0.808 |
| 7 | 584 | 0.594 | 0.763 |
| 8 | 693 | 0.722 | 0.669 |
| 11 | 2781 | 0.842 | 0.734 |


#### Abstract

The nearest neighbour clustering method was used to show the percent similarity between stations during the two surveys (Table 7, Figure 3). The cluster showed a grouping of stations with similarities greater than $60 \%$. These are composed of stations 8, 4,5 and 11. A second grouping consists of stations 6, 9, and 15 with similarities between 42 and $56 \%$.


$-18$

4.0 CONCLUSIONS

1. Extractable levels of lead and zinc are near guideline recommendations. Given the low volumes available for dilution in False Canyon Creek, combined with background concentrations of metals (particularly Pb), contributions from mine effiuent would have to be low if the recommended guidelines are to be maintained.
2. Stream sediment chemistry is comparable to other mineralized sites in the Yukon.
3. Benthic invertebrate populations appear significant in numbers and the diversity compared with other recent surveys.

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## ACKNOWLEDGEMENTE

The Environmental Protection Branch - Yukon Division would like to acknowledge the following people for their contributions towards the completion of this report: Debbie Rene, Doug Davidge and Steve Arrell for data collection and field work, Ken Nordin and Bonnie Burns and Doug Davidge for original draft preparations, Benoit Godin, Peri Mehling and George Derksen for final editing and technical review and Linda Profeit for administrative assistance.

## APPENDIX 1

COLLECTION, PRESERVATION, ANALYSIS OR INDENTIFICATION METEODS AND HATER QUALITY CRITERIA
WATER SAMPLE COLLECTION, PRESEKVATION AND ANALYSIS METHODS

| farameter | detection LIMIT | COLLECTION AND PRESERVATION procedure | ANALYTICAL PROCEDURE | METHOD SECTION |
| :---: | :---: | :---: | :---: | :---: |
| Temperature | $0.1{ }^{\circ} \mathrm{C}$ | In situ temperature reading. | YSI Conductivity and Temperature Meter. Model 33. |  |
| Elow |  | In situ flow measurements using a prbeetype cuisent meter. | Cross-section of the stream was measured and the velocity of flow was calculated using the standard Price-type current meter method. |  |
| Dissolved Oxygen | $1.00 \mathrm{mg} / \mathrm{l}$ | In situ measurement. The instrument was calibrated in the feld mbder wath ur-saturated air combition. | YSI Dissolved Oxygen meter (in sit') Orion model 701 pH meter \& orion o, electinde (laboratory) |  |
| pH | 0.1 pH units | Small aliquots of sample were taken ard fead soon after collection. No preservative. Instrument was calirated usinis 7.0 buffering solution. | Potentiometric | 080 |
| Conductivity |  | It situ measurement. Laboratory measurement, specific conductivity at $25+C$. No preservative. The measurement was taken from the same sample as $\mathrm{NH}_{3}$ telow. | YSI Conductivity meter model 33 <br> (in situ). Radiometer conductivity meter ICDM2D(laboratory). | 044 |
| ( 14.4 | $\begin{aligned} & \therefore \quad(6,10414 \\ & \text { (14itio) } \end{aligned}$ | (\%aus :апи, le as VH,. | Platimm-cobalt visual amfarisom | 040 |
| Tusidily | (1.) (r゙al | Bame samult as: 'IH, | Heprichemetai: latidity | 130 |
| $\begin{aligned} & \text { Wh-Fjiterathe } \\ & \text { Ke.i dir: (Ht K) } \end{aligned}$ | S.0 my/ 1 | Same samyte as \%h, | filtration, difitug atw weiging <br> (f filtrate | 104 |
| $\begin{aligned} & \text { Filterable } \\ & \text { Kesidue (Fk) } \end{aligned}$ | 10.0 mg/l | Same sample as NH , | Eiltration, diyifig and weighity of filtrate. | 100 |
| Totad Alkalinity | $\begin{aligned} & 1.0 \mathrm{mg} / 1 \\ & \text { as } \mathrm{caC} \text {, } \end{aligned}$ | Same sample as $\mathrm{NH}_{3}$ | Potentiometriotitration | 006 |


| APPENDIX 1 TABLE 1 |  | WATER SAMPLE COLLECTION, PRESERVATION AND ANALYSIS METHODS (continued) |  |  |
| :---: | :---: | :---: | :---: | :---: |
| PAKAME'TER | DETECTION <br> I.IMIT | COLLECTIGN AND presekvaition lRROCLDURE | ANALYTfCAL Procedure | METHOD SEC'TION |
| Almmolifa <br> $\mathrm{NH}_{3}-\mathrm{N}$ | u.vub $1 \mathrm{~mm} / 1$ | Bilugle samples cullected in 2 Jitre linear polyethylene contalners. Each container was rinsed 3 times with sample before it was filled. No preservatives. Stored at $4^{\circ} \mathrm{C}$. | Ehenod theocichorimetriv: automated | ט¢\% |
| Nitrate $\mathrm{NO}_{2}-\mathrm{N}$ | $0.005 \mathrm{mg} / 1$ | Same sample as $\mathrm{NH}_{3}$. | $\frac{\text { Diazotization-colorimetric- }}{\text { alltomated }}$ | 070 |
| $\begin{aligned} & \text { Nitrate } \\ & \mathrm{NO}_{3}-\mathrm{N} \end{aligned}$ | $0.005 \mathrm{mg} / 1$ | Same sample as $\mathrm{NH}_{3}$. | Cadmium-copper reduction-colorimetric-automated | 072 |
| Total Phosphate <br> T $\mathrm{PO}_{4}-\mathrm{P}$ | $0.002 \mathrm{mg} / 1$ | Same sample as $\mathrm{NH}_{3}$. | Ascorbic acid-persulphater automated autoclave digestion | 086 |
| Sulphate $\mathrm{SO}_{4}$ | $1 \mathrm{mg} / 1$ | Same sample as $\mathrm{NH}_{3}$. | Automated methylthymol-blue colorimetric | 122 |
| $\begin{aligned} & \text { Chloride } \\ & \text { Cl } \end{aligned}$ | $0.5 \mathrm{mg} / \mathrm{l}$ | Same sample as $\mathrm{NH}_{3}$. | Thiocyanate-combined reagentcolorimetric | 024 |

APPENDIK 1 TABLE 1

\begin{tabular}{|c|c|c|c|c|}
\hline farameter \& DETECEIION
LIMIT \& COLLECTION AND PRESERVATIUN PKocedure \& ANALYTICAL HROCEDURE \& METHOD SECTION <br>
\hline Extractable/Total Metals \& mg/1 \& Single or triplicate samples collected in 125 ml linear \& Inductively Coupled Agron Plasma \& 300 <br>
\hline Ag \& 0.01 \& polyethylene bottles. Each bottle \& \& <br>
\hline ${ }_{\text {Al }}^{\text {As }}$ \& 0.05 \& was rinsed 3 times with sample \& \& <br>
\hline As \& 0.05 \& before filling. Preserved to a ph \& \& <br>
\hline B
Ba

a \& 0.01 \& <1.5 using 1.0 ml concentrated HNO . \& \& <br>
\hline Be \& 0.001 \& \& \& <br>
\hline Ca \& 0.1 \& \& \& <br>
\hline Cd \& 0.005 \& \& \& <br>
\hline $\bigcirc$ \& 0.006 \& \& \& <br>
\hline $\mathrm{Cl}_{1}$ \& 0.005 \& \& \& <br>
\hline Cu \& 0.005 \& \& \& <br>
\hline Fe \& 0.005 \& \& \& <br>
\hline Ma \& 0.10 \& \& \& <br>
\hline Mn \& 0.001 \& \& \& <br>
\hline Mo \& 0.01 \& \& \& <br>
\hline Na \& 0.1 \& \& \& <br>
\hline Hi \& 0.02 \& \& \& <br>
\hline Pb \& 0.05 \& \& \& <br>
\hline Sb \& 0.05 \& \& . \& <br>
\hline Se \& 0.05 \& \& \& <br>
\hline Si \& 0.05 \& \& \& <br>
\hline Sr \& 0.05 \& \& \& <br>
\hline 31 \& 0.10101 \& \& \& <br>
\hline Ti \& a. 10 \& \& \& <br>
\hline $\checkmark$ \& (1.11) \& \& \& <br>
\hline 2.1 \& 6. 110 \& \& \& <br>
\hline
\end{tabular}

WATER SAMPLE COLLECTION, PRESERVATION AND ANALYSIS METHODS (continued)

| APPENDIX 1 TABLE 1 WATER SAMPLE COLLECTION, PR | WATER SAmple collection, preservation and analysts methods (continued) |  |
| :---: | :---: | :---: |
| PARAMETER DETECTION LIMIT $\quad$ COLLECTION AND PRESERVATION | ANALXTICAL PRUCEDURE | METHOD <br> SECTION |
| Pb (0.0005 Same sample as metals. | Graphite Eurnace Atomic Absorption Spectrometry | 330 |
| $\begin{aligned} & \text { Total Hardness } \quad 0.030 \mathrm{mg} / \mathrm{l} \text { Same sample as metals. } \\ & \mathrm{Ca} / \mathrm{Mg} \text { Hardness }=4.116 \mathrm{Mg}+2.497 \mathrm{Ca} \end{aligned}$ |  |  |
| ${ }^{1}$ As described in Environment Canada (1976). <br> ${ }^{2}$ As descriled in Department of Envjroniment (1979). |  |  |

APPENDIX 1

| FARAMETER | EKEPAKATION | ANALYSIS | METHODS CODE $_{1}$ |
| :---: | :---: | :---: | :---: |
| All Parameters | Creek and River Stations: Sediment samples were collected using a Teflon scoop to scoop stream sediments into sample bag. |  |  |
|  | Three samples were collected at each station and placed in geochemical paper sample bags. Each sample is then sealed in plastic bags and frozen or keep cool within 24 hours. |  |  |
| Metals <br> (Leachable) | Sample was freeze-dried for 48 hours to remove water. Sample was sieved through a size 100 |  |  |
| Al, B, | mesh (.l5mil stajnless steel sleve. The |  |  |
| Ba, Be, Ca, | portion passing through was analyzed for |  |  |
| $\begin{array}{ll}\mathrm{Cd}, \mathrm{Co}, \mathrm{Cr}, & \text { leachable metals. } \\ \mathrm{Cu}, \mathrm{Fe}, \mathrm{Mg}, & \end{array}$ |  |  |  |
|  |  |  |
| $\mathrm{Mn}, \mathrm{Mo}, \mathrm{Na}$, | Sample was leached with HCl and $\mathrm{HNO}^{3}$. The |  | Inductively Coupled Argon | 320 |
| Ni, P, Pb, | sample was heated for 3 hours. | Plasma (ICAP) |  |
| Si, Sn, Sr, |  |  |  |
| Ti, V, 2 n |  |  |  |
| As | Same as other metals. | Hydride Gerieration ICAP | 350 |
| A. 1 | Same as uthel metals. | Eleme Atomie Alusorftion | 330 |
| raitiole size | Sany, was fletate-diciod. | Standard Sieving ofetation | 078 |
| 1 |  |  |  |
| letarartment. of Env <br>  | nt, Department of Fisheries ard Ocearis, Labisito rvior: (1179). | Marual, Environmental Prote | Service |

APPENDIX 1 TABLE 3 BOTTOM EAUNA COLLECTION, PRESERVATION AND IDENTIEICATION METHODS

| FIELD COLLECTION, SAMFLING PROCEDURES AND PRESERVATION | LABOKATUKY PROCEDUKES | IDENTIEICATION AND ENUMERATION |
| :---: | :---: | :---: |
| Surber Sampler: Creek and river samples were taken using a Suber sampler with a net, 60 cm long (mesh size $500 \mu \mathrm{~m})$. Area sampled was 929 $\mathrm{cm}^{2}$ (1 $\mathrm{ft}^{2}$ ). The sampler was deployed three times, over a short reach of river (approx. 10 m ), at each site. Replicates were all washed from the net into a 11 bottle. A mixture of $10 \%$ formalin was added to preserve the sample. | Bottom fauna was sorted from other material and placed in a vial contalnliag $70 \%$ methanol. | Bottom fauna samples were sent to Or. C. Low, a Consulting Invertebrate Biologist, Nanaimo, B.C. for identification to genus and speries if possible and enumeration of sample. |

WATER QUALITY CRITERIA FOR DRINKING WATER AND AQUATIC LIFE (continued)

| substance | KECOMMENDED LEVEL(S) FOR DRINKING WATER | References (S) | RECOMMENDED LEVEL(S) for rquatic life | REfERENCE |
| :---: | :---: | :---: | :---: | :---: |
| *Copper ( Cu ) total mq/1 | < 1.0 aesthetic objective | 1 | 0.002 at hardiess $0-120 \mathrm{mg} / 1 \mathrm{CaCO}_{3}$ <br> 0.004 at hardness $120-180 \mathrm{mg} / 1 \mathrm{CaCO}_{3}$ <br> 0.006 at hardness $>180 \mathrm{mg} / 1 \mathrm{CaCO}_{3}$ |  |
| Dissolved oxygen (\% saturation) | Near 100\% | 4 | >5.0mg/l | 10 |
| Eluoride ( 5 ) mg/l | 1.5 | 1 | 1.5 | 7 |
| Hardness (Total) as mg/l CaCO3 | 80-100 | 1 |  |  |
| Iron (Fe) total mg/l | <0.3 aesthetic objective | 1 | 0.3 | 10 |
| Lead (Pb) total mg/l | 0.05 | 1 | 0.001 at hardness $0-60 \mathrm{mg} / 1 \mathrm{CaCO}_{3}$ <br> 0.002 at harduess $60-120 \mathrm{mg} / 1 \mathrm{CaCO}_{3}$ <br> 0.004 at hardness $120-180 \mathrm{mg} / \mathrm{l} \mathrm{CaCO}_{3}$ <br> 0.007 at hardness $>180 \mathrm{mg} / 1 \mathrm{CaCO}_{3}$ | 10 |
| Magnesium ( Mg ) mg/l | 50 | 4 |  |  |
| Manganese ( Mn ) mg/l | <0.05 aesthetic objective | 1 | 1.0 | 7 |
| Mol ybdenum (Molmg/l |  |  |  |  |
| Nickel ( Ni ) total $\mathrm{mg} / 1$ | 0.25 | 2 | 0.025 at hardness $0-60 \mathrm{mg} / 1 \mathrm{CaCO}_{3}$ 0.065 at hardness $60-120 \mathrm{mg} / 1 \mathrm{CaCO}_{3}$ 0.11 at hardness $120-180 \mathrm{mg} / 1 \mathrm{CaCO}_{3}$ 0.15 at hardness $>180$ | 10 |
| Nitrate ( $\mathrm{NO} 3-\mathrm{N} / \mathrm{mg} / 1$ | 10 | 1 |  |  |
| Nitrite ( $\mathrm{NO} 2-\mathrm{N}$ ) mg/l | 1.0 | 1 | 0.06 | 10 |
| pH units | 6.5-8.5 | 1 | 6.5-9.0 |  |
| Phosphate (PO4)mg/l | 0.2 | 8 |  |  |
| *Phosphorus ( P ) mg/l (Total) |  |  | 0.020 to prevent algae | 5 |
| Residue: Filterable mg/l (Total dissolved solids) | <500 aesthetic objective | 4 | 70-400 with a maximum of 2000 | 6 |
| Residue: Non-Filterable |  |  | Increase of $10 \mathrm{mg} / 1$ with bkgd<100mg/l | 8 |
| (mg/l) (TSS) |  |  | increase of $10 \%$ above bkgd with bkgd $>100.0 \mathrm{mg} / \mathrm{l}$ | 10 |

WATER QUALITY CRITERIA EOR DRINKING WATER AND AQUATIC LIfE

| SUBSTANCE | KECOMMENDED Level (S) EOR DRINKING WATER | REEERENCE (S) | KECOMMENDED LEVEL(S) FOR AQUATIC LIFE | REFERENCE(S) |
| :---: | :---: | :---: | :---: | :---: |
| Physical |  |  |  |  |
| Colour (TCU) | $<15$ | 1 |  |  |
| Temperature ( ${ }^{\circ} \mathrm{C}$ ) | 15 |  |  |  |
| Odour and taste | If offensive | 1 |  |  |
| Turbidity NTU | <5 | 1 |  |  |
| Collform-Total (count $/ 100 \mathrm{ml}$ ) | 10 | 1 | recreational water Total $500-1000 / 100 \mathrm{ml}$ | 9 |
| fecal coliform | 0 | 1 | 200 /100ml | 9 |
| Chemical |  |  |  |  |
| *Alkalinity mg/l (Total) | Not considered a putlic health proklem | 4 | 220 | 3 |
| *Aluminum (Al) mg/l | Not considered a putilic. health problem | 7 | 6.1 at $\mathrm{pH}>6.5$ | 5 |
| Ammonia total ( $\mathrm{NH} 3-\mathrm{N}$ ) $\mathrm{mg} / \mathrm{l}$ | 0.5 | 4 | 2.2 at pH 6.5 temp $10^{\circ} \mathrm{C}$ 1.37 at ph 8.0 temp $10^{\circ} \mathrm{C}$ | 10 |
| Antimony (SL) mg/l |  |  |  |  |
| Arsenic (As) total $\mathrm{mg} / 1$ | 0.05 | 1 | $\begin{aligned} & <0.05 \\ & 5.0 \end{aligned}$ | $10$ |
| Barium ( Ba ) mg/l | 1.0 | 1 |  |  |
| Boron (B) ing/l | 5.0 | 1 | 0.0002 for hardness $0-60 \mathrm{mg} / 1 \mathrm{CaCO}_{3}$ |  |
| *Cadmium (Cd) total mg/l | 0.005 | 1 | 0.0008 for hardness $60-120 \mathrm{mg} / \mathrm{l}$ <br> $\mathrm{CaCO}_{3}$ <br> 0.0013 for hardness $120-180 \mathrm{CaCO}_{3}$ | 10 |
| Calcium ( Ca ) mg/1 | 75-200 | 7 | 0.0018 for hardress $>180 \mathrm{CaCO}$, |  |
| Chacride (Cl)mg/ | $\bigcirc 50$ aesthetje olijectives | 1 |  |  |
|  | 0.05 | 1 | 11.02 to protect tish <br> 11.002 tis frotest ruatic life | 10 |
|  | Depernds on dissoived salts | 7 | $100-500$ | 6 |

APPENDIX 1
TABLE 4
WATER QUALITY CRITERIA FOR DRINKING WATER AND AQUATIC LIFE (continued)

| SUBStance | RECOMMENDED LEVEL(S) EOR DRINKING WATER | REFERENCES(S) | RECOMMENDED LEVEL(S) for Aguatic life | Reference |
| :---: | :---: | :---: | :---: | :---: |
| **Selenium (Se) |  |  |  |  |
| Silica (Si)mg/l |  |  |  |  |
| *Silver (Ag) total mg/l |  | 1 | 0.0001 | 10 |
| Sodium ( Na ) $\mathrm{mg} / \mathrm{l}$ |  | 1 |  |  |
| Strontium ( Sr ) mg/l | 10 | 1 |  |  |
| Sulphate ( $\mathrm{SO}^{4}$ ) $\mathrm{mg} / \mathrm{l}$ | 500 | 1 |  |  |
| Tin (Sn)mg/l | Not present in natural waters | 7 |  |  |
| Titanium (Ti)mg/l <br> Vanadium (V) |  |  |  |  |
|  |  |  |  |  |
| Zinc ( Zn ) mg/l | <5.0 aesthetic objective | 1 | 0.030 | 10 |
| * Use graphite furnace for the lab detection limit to be less than the recommended levels. <br> ** Lab detection limit > recommended levels. |  |  |  |  |
| REFERENCES: |  |  |  |  |

1. Health Welfare Canada. 1987. Guidelines for Canadian Drinkinq Water Ouality 1987. Supply and Services,
Canada.

[^0]APPENDIX I TABLE 4 WATER QUALITY CRITERIA EOR DRINKING WATER AND AQUATIC LIFE (continued)

| SUBSTANCE | KECOMMENDED LEVEL(S) FOR DRINKING WATER | ketekences $(\mathrm{S})$ | RECOMMENDED LEVEL(S) for aquatic life | keference |
| :---: | :---: | :---: | :---: | :---: |

[^1]APPENDIX II

WATER QUALITY
APPENDIX II TABLE 1 mT. hundere WATER quality DATA for June, 1988

| STATION | DESCRIPTION | $\begin{array}{r} F R \\ S T D \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} \text { NFR } \\ \text { Mean } \\ \langle m g / L\rangle \end{array}$ | $\begin{array}{r} \text { NFR } \\ \text { STD } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} \mathrm{Ag} \\ \text { Mean } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} \mathrm{Ag} \\ \mathrm{STD} \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} \text { Al } \\ \text { Mean } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} A L \\ S T D \\ (m g / L) \end{array}$ | $\begin{array}{r} \text { As } \\ \text { Mean } \\ (m g / L) \end{array}$ | $\begin{array}{r} \text { As } \\ \text { STD } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Headwaters of Tom Creek Tributary 5 | 10 | 42 | 29 | $<0.01$ | n/a | $<0.05$ | n/a | $<0.05$ | n/a |
| 2 | Headwaters of False Canyon Creek | 7 | $<5$ | n/a | $<0.01$ | $n / a$ | $<0.05$ | $n / a$ | $<0.05$ | n/a |
| 3 | Headwaters of False Canyon Creek Tributary B | 17 | 72 | 73 | $<0.01$ | $n / a$ | $<0.05$ | $\mathrm{n} / \mathrm{a}$ | $<0.05$ | n/a |
| 4 | Erances River Tributary A © bridge crossing | 43 | 34 | 3 | $<0.01$ | $\mathrm{n} / \mathrm{a}$ | $<0.05$ | $\mathrm{n} / \mathrm{a}$ | $<0.05$ | n/a |
| 5 | Tom Creek u/s Campbell Hwy bridge | 10 | 36 | 10 | $<0.01$ | n/a | $<0.05$ | n/a | $<0.05$ | n/a |

[^2]aphendix II table 1

| STATION | DESCRIPTION |  |  | $\begin{array}{r} \text { Co } \\ \text { Mean } \\ (m g / L) \end{array}$ | $\begin{array}{r} \mathrm{Co} \\ \mathrm{STD} \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ |  | $\begin{array}{r} \text { Mean } \\ (m g / L)\left(\begin{array}{rl} \mathrm{Cr} \end{array}\right. \\ \hline \end{array}$ | $\begin{array}{r} \mathrm{Cr} \\ \mathrm{STD} \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ |  | $\begin{array}{r} \mathrm{Cu} \\ \mathrm{STD} \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} \text { Fe } \\ \text { Mean } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} \mathrm{Fe} \\ \mathrm{STD} \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Headwaters of Tom Creek Tributary 5 | n/a |  | 0.005 | n/a | $<$ | 0.005 | n/a | 0.005 | 0.000 | < 0.012 | n/a |
| 2 | Headwaters of False Canyon Creek | $n / \mathrm{a}$ | $<$ | 0.005 | n/a | $<$ | 0.005 | n/a | 0.005 | 0.000 | < 0.005 | n/a |
| 3 | Headwaters of False Canyon Creek Tributary B | n/a | < | 0.005 | n/a | $<$ | 0.005 | n/a | 0.005 | 0.000 | 0.010 | 0.005 |
| 4 | Frances River Tributary A e bridge crossing | n/a | $<$ | 0.005 | $n / \mathrm{a}$ | $<$ | 0.005 | n/a | 0.005 | 0.000 | < 0.011 | n/a |
| 5 | Tom Creek u/s Campbell Hwy bridge | n/a | $<$ | 0.005 | $n / \mathrm{a}$ | $<$ | 0.005 | n/a | 0.005 | 0.000 | 0.069 | 0.010 |


| STATION | DESCRIPTION | $\begin{array}{r} M g \\ M e a n \\ \{\mathrm{mg} / \mathrm{L}\} \end{array}$ | $\begin{array}{r} \mathrm{Mg} \\ \text { STD } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ |  | $\begin{array}{r} \text { Mn } \\ \text { Mean } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} M n \\ S T D \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ |  |  | $\begin{array}{r} \text { Mo } \\ \text { ST'D } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} \mathrm{Na} \\ \mathrm{Mean} \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} \mathrm{Na} \\ \mathrm{STD} \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} \mathrm{NI} \\ \mathrm{Mean} \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Headwaters of Tom Creek Tributary 5 | 1.5 | 0.2 | $<$ | 0.001 | n/a |  | 0.01 | n/a | 0.6 | 0.1 | $<0.02$ |
| 2 | Headwaters of False Canyon Creek | 0.4 | 0.2 | $<$ | 0.001 | n/a |  | 0.01 | n/a | < 0.1 | n/a | $<0.02$ |
| 3 | Headwaters of False Canyon Creek Tributary B | 3.5 | 1.2 | < | 0.001 | n/a |  | 0.01 | $n / \mathrm{a}$ | 0.5 | 0.2 | $<0.02$ |
| 4 | Erances River Tributary A © bridge crossing | 11.3 | 3.5 | $<$ | 0.001 | n/a |  | 0.01 | n/a | 0.9 | 0.3 | $<0.02$ |
| 5 | Tom Creek u/s Campbell Hwy bridge | 6.3 | 1.2 |  | 0.004 | 0.001 |  | 0.01 | n/a | 0.9 | 0.2 | $<0.02$ |

APPENDIX II TABLE 1 MT. HUNDERE WATER QUALITY DATA FOR JUNE, 1988


Metals (ICP SCan); low detection analysis (GF) for Ag, Cd, Cu and Pbif below ICP scan. n/a $=$ analysis not done
afpendik il table 2 mt. hundere water quality data for june, 1990


APPENDIX II TABLE 2 MT. HUNDERE WATER QUALITY DATA FOR JUNE, 1990

| STATION | DESCRIPTION | $\begin{array}{r} F R \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} \text { NER } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{gathered} \text { Diss. GE } \\ \text { Ag } \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | Diss. (mg/L Al | Diss. <br> As (mg/L) | Diss. (mg/L) | Diss. Ba (mg/L) | Diss. Be (mg/L) | $\begin{array}{r} \text { Diss. } \\ \text { Ca } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | Headwaters of Ealse Canyon Cr. Trib. B | 230 | 10 | 0.0004 | $<0.05$ | <0.05 | <0.01 | 0.027 | <0.001 | 44.2 |
| 4 | Frances River Trib. A \& Bridge | 320 | $<10$ | 0.0007 | $<0.05$ | <0.05 | $<0.01$ | 0.069 | $<0.001$ | 54.2 |
| 5 | Tom Creek u/s Campbell Hwy bridge | 283 | $<10$ | 0.0003 | $<0.05$ | <0.05 | $<0.01$ | 0.088 | <0.001 | 34.1 |
| 6 | Tom Cr. Trib. $2(200 \mathrm{~m} / \mathrm{s}$ ) Tom Cr. | 180 | <10 | <0.0001 | $<0.05$ | <0.05 | $<0.01$ | 0.071 | <0.001 | 22.1 |
| 7 | Tom Cr . | 260 | $<10$ | 0.0004 | <0.05 | <0.05 | <0.01 | 0.090 | <0.001 | 38.1 |
| 8 | Tom Cr. Trib. 5 | 240 | <10 | 0.0004 | $<0.05$ | $<0.05$ | $<0.01$ | 0.101 | $<0.001$ | 38.0 |
| 9 | Tom Cr. Outflow from Tom Lake $\mathrm{u} / \mathrm{s}$ of Site 10 Trib . (150M) | 290 | <10 | 0.0005 | $<0.05$ | $<0.05$ | $<0.01$ | 0.083 | <0.001 | 47.6 |
| 10 | Tom Cr. Trib. u/s of outflow from Tom Lake | 220 | <10 | 0.0004 | <0.05 | <0.05 | <0.01 | 0.100 | <0.001 | 36.7 |
| 11 | Upper False Canyon Cr. | 210 | $<10$ | 0.0005 | <0.05 | <0.05 | <0.01 | 0.099 | <0.001 | 45.7 |
| 12 | False Canyon Cr. Trib. Air Photo Stn. | 250 | $<10$ | 0.0005 | $<0.05$ | <0.05 | <0.01 | 0.118 | <0.001 | 47.4 |
| 14 | False Canyon $\mathrm{Cr} . \mathrm{u} / \mathrm{s}$ of Trib. B. 30 m | 130 | 13 | 0.0003 | $<0.05$ | $<0.05$ | $<0.01$ | 0.072 | <0.001 | 29.6 |
| 15 | Ealse Canyon Cr. Trib. B u/s of Ealse Canyon Cr. 300 m | 193 | 25 | 0.0005 | $<0.05$ | $<0.05$ | $<0.01$ | 0.087 | <0.001 | 45.5 |
| 16 | False Canyon Cr. $60 \mathrm{~m} \mathrm{u/s} \mathrm{of} \mathrm{Erances} \mathrm{R}$. | 207 | 10 | 0.0003 | <0.05 | <0.05 | <0.01 | 0.086 | <0.001 | 35.8 |
| 17 | Frances R. u/s of False Canyon Cr. | 117 | $<10$ | <0.0001 | $<0.05$ | <0.05 | <0.01 | 0.025 | <0.001 | 18.7 |
| 18 | Frances R. A Hwy. Bridge | 157 | <10 | <0.0001 | <0.05 | <0.05 | <0.01 | 0.027 | <0.001 | 19.4 |



APPENDIX II TABLE 2 MT. HUNDERE WATER QUALITY DATA FOR JUNE, 1990

APPENDIX II TABLE 2 MT. hUNDERE WATER QUALITY DATA FOR JUNE, 1990

| STATION | DESCRIPTION | Extr. <br> Ca <br> (mg/L) | $\begin{gathered} \text { Extr. } \mathrm{Cd} \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | Extr. Co (mg/L) | Extr. (mg/L) | $\begin{gathered} \text { Extr. GF } \\ \text { Cu } \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \text { Extr. } \\ \text { Ee } \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | $\begin{gathered} \text { Extre } \\ \text { K } \\ (\mathrm{mg} / \mathrm{L}) \end{gathered}$ | Extr. <br> Mg <br> (mg/L) | Extr. Mn (mg/L) |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | Headwaters of Ealse Canyon Cr. Trib. B | 44.5 | <0.0001 | <0.005 | 0.005 | 0.0017 | 0.099 | <2 | 4.1 | 0.003 |
| 4 | Frances River Trib. A 0 Bridge | 55.6 | <0.0001 | <0.005 | 0.009 | $<0.0005$ | 0.059 | <2 | 16.0 | 0.003 |
| 5 | Tom Creek u/s Campbell Hwy bridge | 35.5 | <0.0001 | <0.005 | $<0.005$ | 0.0006 | 0.0504 | <2 | 16.0 9.9 | 0.008 |
| 6 | Tom Cr. Trib. $2(200 \mathrm{~m} \mathrm{u} / \mathrm{s}$ ) Tom Cr . | 22.4 | <0.0001 | 0.010 | 0.010 | 0.0013 | 0.240 | <2 | 9.9 | 0.014 |
| 7 | Tom Cr . | 38.7 | <0.0001 | 0.009 | 0.008 | 0.0005 | 0.078 | <2 | 10.6 | 0.029 0.006 |
| 8 | Tom Cr. Trib. 5 | 38.1 | $<0.0001$ | $<0.005$ | 0.007 | 0.0009 | 0.230 | <2 | 9.6 | 0.011 |
| 9 | Tom Cr. Outflow from Tom Lake $\mathrm{u} / \mathrm{s}$ of Site 10 Trib . 1150 M ) | 46.9 | <0.0001 | <0.005 | 0.008 | <0.0005 | 0.074 | <2 | 13.0 | 0.017 |
| 10 | Tom Cr. Trib. u/s of outflow from Tom Lake | 37.6 | $<0.0001$ | <0.005 | <0.005 | <0.0005 | 0.133 | <2 | 12.9 | 0.021 |
| 11 | Upper False Canyon Cr. | 46.0 | <0. 0001 | <0.005 | 0.005 | <0.0005 | 0.099 | <2 | 9.2 | 0.006 |
| 12 | False Canyon Cr. Trib. A Air Photo Stn. | 49.5 | <0.0001 | <0.005 | $<0.005$ | <0.0005 | 0.104 | <2 | 10.6 | 0.008 |
| 14 | False Canyon Cr. u/s of Trib. B. 30 m | 30. | <0.0001 | <0.005 | <0.005 | <0.0005 | 0.401 | <2 | 6.8 | 0.022 |
| 15 | False Canyon Cr. Trib. B u/s of False Canyon Cr. 300 m | 48.3 | <0.0001 | $<0.005$ | 0.007 | 0.0007 | 0.328 | <2 | 11.4 | 0.015 |
| 16 | False Canyon $\mathrm{Cr}, ~ 60 \mathrm{~m} \mathrm{u/s} \mathrm{of} \mathrm{Frances} \mathrm{R}$. | 36.9 | <0.0001 | <0.005 | 0.005 | 0.0005 | 0.370 | <2 | 8.6 | 0.027 |
| 17 | Frances R. u/s of False Canyon Cr. | 19.4 | <0.0001 | $<0.005$ | $<0.005$ | <0.0005 | 0.088 | <2 | 4.1 | 0.006 |
| 18 | Frances R. 0 Hwy. Bridge | 20.1 | <0.0001 | <0.005 | <0.005 | <0.0007 | 0.122 | <2 | 4.3 | 0.008 |



APPENDIX II TABLE 2 MT. HUNDERE WATER QIIALITY DATA FOR JUNE, 1990

| STATION | DESCRIPTION | $\begin{array}{r} \text { Extr } \\ \text { Sr } \\ (\mathrm{mg} / \mathrm{I}) \end{array}$ | $\begin{array}{r} \text { Extr. } \\ \mathrm{Ti} \\ (\mathrm{mg} / \mathrm{I}) \end{array}$ | $\begin{array}{r} \text { Extr. } \\ V \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} \text { Extr. } \\ 2 n \\ (\mathrm{mog} / \mathrm{L}) \end{array}$ | $\begin{gathered} \text { Total GF } \\ \text { AG } \\ (\mathrm{mg} / \mathrm{l}, \mathrm{~s} \end{gathered}$ | Total <br> Al <br> (mg/L) | $\begin{array}{r} \text { Total } \\ \text { As } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} \text { Total } \\ B \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} \text { Total } \\ \text { Ba } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | Headwaters of Ealse Canyon Cr. Trib. B | 0.120 | $<0.002$ | $<0.01$ | 0.016 | 0.0004 | 0.27 | $<0.05$ | $<0.01$ | 0.033 |
| 4 | Frances River Trib. A Bridge | 0.193 | $<0.002$ | <0.01 | $<0.002$ | 0.0005 | 0.07 | $<0.05$ | 0.02 | 0.079 |
| 5 | Tom Creek u/s Campbell Hwy bridge | 0.126 | $<0.002$ | <0. 01 | $<0.002$ | 0.0004 | 0.11 | $<0.05$ | $<0.01$ | 0.102 |
| 6 | Tom Cr. Trib. 2 (200m u/s) Tom Cr. | 0.094 | <0.002 | $<0.01$ | $<0.002$ | 0.0003 | 0.17 | $<0.05$ | 0.02 | 0.088 |
| 7 | Tom Cr. | 0.132 | $<0.002$ | <0.01 | $<0.002$ | 0.0004 | 0.15 | $<0.05$ | $<0.01$ | 0.104 |
| 8 | Tom Cr. Trib. 5 | 0.120 | <0.002 | $<0.01$ | 0.007 | 0.0003 | 0.35 | $<0.05$ | $<0.01$ | 0.131 |
| 9 | Tom Cr. Outflow from Tom Lake u/s of Site 10 Trib . (150M) | 0.178 | $<0.002$ | $<0.01$ | $<0.002$ | 0.0005 | 0.10 | $<0.05$ | $<0.01$ | 0.094 |
| 10 | Tom Cr. Trib. u/s of outflow from Tom Lake | 0.132 | $<0.002$ | $<0.01$ | $<0.002$ | 0.0005 | 0.14 | $<0.05$ | <0.01 | 0.119 |
| 11 | Upper False Canyon Cr. | 0.161 | <0.002 | $<0.01$ | 0.004 | 0.0005 | 0.09 | $<0.05$ | $<0.01$ | 0.111 |
| 12 | Ealse Canyon Cr. Trib. Air Photo Stn. | 0.184 | <0.002 | <0.01 | $<0.002$ | 0.0006 | 0.08 | $<0.05$ | $<0.01$ | 0.133 |
| 14 | Ealse Canyon Cr . $\mathrm{u} / \mathrm{s}$ of Trib. B. 30 m | 0.114 | $<0.002$ | $<0.01$ | $<0.002$ | 0.0002 | 0.60 | $<0.05$ | $<0.01$ | 0.092 |
| 15 | False Canyon Cr. Trib. Bu/s of False Canyon Cr. 300 m | 0.132 | $<0.002$ | <0.01 | 0.003 | 0.0005 | 0.58 | $<0.05$ | 0.01 | 0.117 |
| 16 | Ealse Canyon Cr . $60 \mathrm{~m} \mathrm{u/s}$ of Erances R. | 0.126 | $<0.002$ | <0.01 | $<0.002$ | 0.0004 | 0.45 | $<0.05$ | <0.01 | 0.106 |
| 17 | Erances R. u/s of False Canyon Cr. | 0.083 | <0.002 | <0.01 | <0.002 | $<0.0002$ | 0.21 | <0.05 | $<0.01$ | 0.031 |
| 18 | Erances R. Hwy. Bridge | 0.085 | <0.002 | $<0.01$ | <0.002 | 0.0001 | 0.22 | $<0.05$ | <0.02 | 0.035 |


| IZ ${ }^{*}$ ¢ | S0＊0＞ | S0＊0＞ | －G200＊0＞ | I＇0＞ | $20^{\circ} 0>$ |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $60^{\circ} \mathrm{\varepsilon}$ | So $0>$ | So＇0＞ | －100．0 | $1 \cdot 0>$ | $20^{\circ} 0>$ | $0 \cdot 1$ | $10^{\circ} 0>$ | $\begin{aligned} & 600^{\circ} 0 \\ & 900^{\circ} 0 \end{aligned}$ |  | 8 T |
| てI＊ | S0＊0＞ | So＇0＞ | $6100^{\circ} 0$ | ［－0＞ | $20^{\circ} 0>$ | $6 \cdot 0$ | 10．0＞ | $0 \varepsilon O^{\circ} 0$ | Kuep astes | $\angle I$ |
| t 1 ＊${ }^{\text {c }}$ | S0＊0＞ | S0＊0＞ | $2100^{\circ}$ | ［ ${ }^{\circ} 0>$ | 20＊0＞ | $8 \cdot 0$ | 10．0＞ |  | woor－ $\mathrm{J}_{\text {c uokuej esteg jo }}$ |  |
| I6 $\mathcal{E}$ | S0＊0＞ | So＇0＞ | －EOO 0 | $1 \cdot 0>$ | 20：0＞ | 6.0 | t0．0＞ | $\angle 100^{\circ}$ 92000 |  | GI |
| bs＊ 2 | S0．0＞ | 90．0＞ | $9600^{\circ} 0$ | $1 \cdot 0>$ | $20 \cdot 0>$ | $0 \cdot 1$ | to 0＞ | $800 \%$ |  | ET |
| T0＊E | S0＊＊ | SO＊${ }^{\circ}$ | $0100^{\circ}$ | I－0＞ | $20^{\circ} 0>$ | L． 0 | to $0>$ | $900^{\circ} 0$ | $\cdot$ uf | 2I |
| 8L＊ | S0＊ 0 | G0＊0＞ | $\varepsilon 100^{\circ} 0$ | ［＊0＞ | $20^{\circ} 0>$ | $6^{\circ} 0$ | 10＊0＞ | $\varepsilon 20 * 0$ |  | 0 T |
| ZL．E | S0＊0＞ | S0．0＞ | $6100^{\circ} 0$ | 「•0＞ | $20^{\circ} 0>$ | I＇$冖$ | $10^{\circ} 0>$ | $810^{\circ} 0$ | （WOGt）qqud ot afts jo s／n |  |
| SでE | S0\％${ }^{\circ}$ | $90^{\circ} 0>$ | $8100{ }^{\circ}$ | $1 \cdot 0>$ | 20＊0＞ | 6.0 | $10 \cdot 0>$ | $210{ }^{\circ}$ | ayet mol moxi motjino ed mod | 6 |
| して・ | S0．0＞ | S0＊0＞ | －100 0 | I－0＞ | $20 \cdot 0>$ | $z \cdot 1$ | $10 \cdot 0>$ | $1.00 \cdot 0$ |  | 8 |
| でも | $50^{\circ} 0>$ | $50 \cdot 0>$ | $9200^{\circ}$ | ［－0＞ | $20 \cdot 0>$ | 9＊ 1 | 10．0＞ | عEO＊ |  | $L$ |
| ZでE | S0＊0＞ | So＊0＞ | St00＊0 | I－0＞ | $20^{\circ} 0>$ | z＇I | ［0．0＞ | ¢ $10{ }^{\circ} 0$ |  | 9 |
| $8 \square^{\circ} \mathrm{E}$ | S0＊0＞ | S0＊0＞ | $9100{ }^{\circ}$ | I－0＞ | $20^{\circ} 0>$ | $\varepsilon \cdot \tau$ | 10．0＞ | $600^{\circ} 0$ | obpraq KMh tteqduej $5 / \mathrm{n}$ yeax uod | 5 |
| Lて＇＊ | S0＊0＞ | SO＊${ }^{\circ}$ | $9200^{\circ}$ | ［ $0>$ | 20＇0＞ | 9.0 | 10＊0＞ | $900 \cdot 0$ |  | b |
| （IT／5u） | （7／6w） | （＇T／bur） | （7／6ur） | （7／6w） | （＇T／5us） | （＇7／6w） |  |  |  | $\varepsilon$ |
| ！ | es | qS | qd | d | IN | en | CW |  | NOILdİDS3a | NOILYLS |
| ［P70山 | Te704 | Te70L | 35 Te7od | tejol | Te70l | 1e704 | Ie7ol | ［e7ol |  |  |


| STATION | DESCRIPTION | $\begin{array}{r} \text { Total } \\ \text { Sn } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | Total Sr $(\mathrm{mg} / \mathrm{L})$ | $\begin{array}{r} \text { Total } \\ \text { Ti } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ | $\begin{array}{r} \text { Total } \\ V \\ (m g / L) \end{array}$ | $\begin{array}{r} \text { Total } \\ \text { Zn } \\ (\mathrm{mg} / \mathrm{L}) \end{array}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 3 | Headwaters of Ealse Canyon Cr．Trib．B |  |  |  |  |  |
| 4 | Erances River Trib．A A Bridge | $<0.05$ $<0.05$ | 0.131 | $<0.002$ | $<0.01$ | ＜0．002 |
| 5 | Tom Creek u／s Campbell Hwy bridge | ＜0．05 | 0.209 0.135 | $<0.002$ $<0.002$ | $<0.01$ | ＜0．002 |
| 6 | Tom Cr．Trib． 2 （200m u／s）Tom Cr． | $<0.05$ | 0.103 | $<0.002$ | $<0.01$ | $<0.002$ |
| 7 | Tom Cr． | $<0.05$ | 0.146 | $<0.002$ | ＜0．01 | $<0.002$ |
| 8 | Tom Cr．Trib． 5 | $<0.05$ | 0.146 | $<0.002$ | ＜0．01 | $<0.002$ |
| 9 | Tom Cr．Out flow from Tom Lake | ＜0．05 | 0.191 | ＜0．002 | ＜0．01 | $<0.002$ |
|  | u／s of Site 10 Trib ．（150M） | ＜0．05 | 0.191 | ＜0．002 | ＜0．01 | $<0.002$ |
| 10 | Tom Cr．Trib．u／s of outflow from Tom Lake | ＜0．05 | 0.142 | ＜0．002 | $<0.01$ | $<0.002$ |
| 11 | Upper False Canyon Cr． |  |  |  |  |  |
| 12 | Ealse Canyon Cr．Trib．Air Photo Stn． | ＜0．05 | 0.169 0.177 | ＜0．002 | $<0.01$ | $<0.002$ |
| 14 | False Canyon Cr．u／s of Trib．B． 30 m ． | ＜0．05 | 0.120 | ＜0．002 | ＜0．01 | $<0.002$ |
| 15 | Ealse Canyon Cr．Trib．B u／s | ＜0．05 | 0.120 0.138 | ＜0．002 | ＜0．01 | $<0.002$ |
|  | of False Canyon Cr．300m | ＜0．05 | 0.138 | ＜0．002 | ＜0．01 | ＜0．002 |
| 16 | False Canyon Cr． $60 \mathrm{~m} \mathrm{u} / \mathrm{s}$ of Frances R． | $<0.05$ | 0.136 | $<0.002$ | $<0.01$ | $<0.002$ |
| 17 | Erances R．u／s of False Canyon Cr． | $<0.05$ | 0.090 | $<0.002$ | $<0.01$ | $<0.002$ |
| 18 | Frances R．\＆Hwy．Bridge | $<0.05$ | 0.093 | $<0.002$ | $<0.01$ | $<0.002$ |

APPENDIX III

SITREAM SEDTMENTS
appendix ili table 1 mT hundere percent sediment particie size distribution, june 1988

TABLE 2 MT. HUNDERE SEDIMENT CHEMISTRY FOR JUNE, 1988
APPENDIX III

| STATION NUMBER | $\begin{array}{r} A g \\ (\mathrm{ug} / \mathrm{g}) \end{array}$ | $\begin{array}{r} \text { Al } \\ (\mathrm{ug} / \mathrm{g}) \end{array}$ | $\begin{array}{r} \text { As } \\ (\mu \mathrm{g} / \mathrm{g}) \end{array}$ | $\begin{array}{r} \mathrm{Ba} \\ (\mu \mathrm{~g} / \mathrm{g}) \end{array}$ | $\begin{array}{r} \mathrm{Be} \\ (\mu \mathrm{~g} / \mathrm{g}) \end{array}$ | $\begin{gathered} \mathrm{Ca} \\ (\mu \mathrm{~g} / \mathrm{g}) \end{gathered}$ | $\begin{array}{r} \text { Cd } \\ (\mu \mathrm{g} / \mathrm{g}) \end{array}$ | $\begin{array}{r} C o \\ (\mu \mathrm{~g} / \mathrm{q}) \end{array}$ | $\begin{array}{r} \mathrm{Cr} \\ (\mu \mathrm{~g} / \mathrm{g}) \end{array}$ | $\underset{(\mu \mathrm{g} / \mathrm{g})}{\mathrm{Cu}}$ | $\begin{array}{r} \mathrm{Fe} \\ (\mu \mathrm{~g} / \mathrm{g}) \end{array}$ | $\underset{(\mu \mathrm{g} / \mathrm{g})}{\mathrm{Mg}}$ | ${ }_{(\mu \mathrm{g} / \mathrm{g})}^{\mathrm{Mn}}$ | $\begin{gathered} \text { Mo } \\ (\mu \mathrm{g} / \mathrm{g})^{\prime} \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | $<2$ | 21400 | 20 | 152 | 0.8 | 9880 | 3 | $<20$ | 41.6 | 28.5 | 29800 | 10900 | 1160 | 3 |
|  | $<2$ | 20300 | 23 | 138 | 0.7 | 9100 | 2.7 | $<20$ | 3.9 .5 | 28.6 | 29700 | 10900 | 1040 | 3 |
|  | $<2$ | 19400 | 20 | 182 | 0.9 | 11200 | 2.6 | < 20 | 36.4 | 37.9 | 26600 | 9340 | 658 | 4 |
| 2 | 4 | 25500 | 75 | 177 | 3.2 | 12800 | 32.7 | 20 | 53.4 | 73.1 | 45000 | 9670 | 2470 | 5 |
|  | 4 | 25600 | 80 | 176 | 3.3 | 13700 | 36.4 | < 20 | 58.7 | 87.1 | 48100 | 10000 | 2320 | 8 |
|  | 4 | 21400 | 89 | 169 | 2.9 | 12300 | 43.5 | 20 | 47.7 | 74.4 | 48400 | 7870 | 2580 | 6 |
| 3 | $<2$ | 12400 | 10 | 253 | 0.6 | 12300 | 3.1 | $<20$ | 42.8 | 26.8 | 26000 | 7410 | 565 | 5 |
|  | $<2$ | 15000 | 20 | 245 | 0.7 | 11000 | 2.2 | $<20$ | 32.1 | 35.9 | 23700 | 7470 | 593 | 3 |
|  | $<2$ | 13700 | 10 | 419 | 0.6 | 11700 | 2 | $<20$ | 48.5 | 44.1 | 23700 | 7800 | 541 | 6 |
| 4 | $<2$ | 10300 | 10 | 279 | 0.5 | 10800 | 1 | $<20$ | 35.6 | 44.9 | 22200 | 4720 | 997 | 4 |
|  | $<2$ | 10700 | 10 | 437 | 0.5 | 9560 | 0.9 | $<20$ | 43.1 | 30.7 | 23500 | 4790 | 1360 | 4 |
|  | $<2$ | 10800 | 10 | 367 | 0.5 | 10300 | 1 | $<20$ | 40.7 | 25.6 | 24000 | 4710 | 792 | 2 |
| 5 | $<2$ | 15600 | 10 | 241 | 0.4 | 5300 | $<0.8$ | $<20$ | 36.6 | 61.3 | 34600 | 6630 | 942 | 6 |
|  | $<2$ | 11800 | $<8$ | 307 | 0.4 | 5960 | 0.9 | $<20$ | 55.1 | 21.3 | 20600 | 5500 | 712 | 2 |
|  | $<2$ | 13300 | 10 | 267 | 0.5 | 5390 | 1 | $<20$ | 42.5 | 25.1 | 21300 | 5600 | 507 | 3 |


APPENDIX III

APPENDIX III TABLE 3 MT. HUNDERE SEDIMENT CHEMISTRY EOR JUNE, 1990

APPENDIX III

APPENDIX III TABLE 4 MT. hundere sediment data summary for june, 1988 and june, 1990

| STATION NUMBER | SAMPLE DATE | $\begin{aligned} & \quad \mathrm{Cd} \\ & \operatorname{STD} \\ & (\mu \mathrm{~g} / \mathrm{g}) \end{aligned}$ | $\begin{gathered} \text { Co } \\ \text { Mean } \\ \|\mu g / g\| \end{gathered}$ | $\operatorname{STD}_{(\mu \mathrm{g} / \mathrm{g})}^{\mathrm{Co}}$ | $\begin{gathered} C r \\ \text { Mean } \\ (\mu \mathrm{g} / \mathrm{g}) \end{gathered}$ | $\begin{aligned} & \operatorname{STD} \\ & (\mu g / g) \end{aligned}$ | Cu <br> Mean ( $\mu \mathrm{g} / \mathrm{g}$ ) | $\begin{gathered} \mathrm{STD}_{(\mu \mathrm{g} / \mathrm{g})}^{\mathrm{Cu}} \end{gathered}$ | $\begin{array}{r} \text { Fe } \\ \text { Mean } \\ (\mu(g / g) \end{array}$ | $\operatorname{STD}_{(\mu \mathrm{g} / \mathrm{g})}^{\mathrm{Fe}}$ | $\begin{gathered} \text { Mean }^{K} \\ \langle\mu g / g\rangle \end{gathered}$ | $\underset{(\mu \mathrm{STD} / \mathrm{g})}{\mathrm{ST}}$ | $\begin{gathered} \text { Mg } \\ \text { Mean } \\ (\mu g / g) \end{gathered}$ | $\begin{gathered} \mathrm{STD} \\ (\mu \mathrm{~g} / \mathrm{g}) \end{gathered}$ | $\begin{gathered} \text { Mn } \\ \text { Mean } \\ (\mu g / g) \end{gathered}$ | $\begin{gathered} \mathrm{Mn} \\ (\mu \mathrm{STD} / \mathrm{g}) \end{gathered}$ | $\begin{gathered} \text { Mo } \\ \text { Mean } \\ (\mu \mathrm{g} / \mathrm{g}) \end{gathered}$ |  | $\begin{aligned} & \text { Mo } \\ & \mathrm{g} \end{aligned}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 21-Jun-88 | $n / a$ | $<20$ | $\mathrm{n} / \mathrm{a}$ | 39.2 | 2.6 | 31.7 | 5.4 | 28700 | 1819 | $n / \mathrm{a}$ | $n / a$ | 10380 | 901 | 953 | 262 | 3 | 3 | 1 |
| 2 | 21-Jun-88 | $n / a$ | $<20$ | n/a | 53.3 | 5.5 | 78.2 | 7.7 | 47167 | 1882 | n/a | n/a | 9180 | 1146 | 2457 | 131 | 6 | 6 | 2 |
| 3 | 22-Jun-88 | $\mathrm{n} / \mathrm{a}$ | $<20$ | $n / a$ | 41.1 | 8.3 | 35.6 | 8.7 | 24467 | 1328 | $n / a$ | n/a | 7560 | 210 | 566 | 26 | 5 | 5 | 2 |
| 3 | 20-Jun-90 | 0.09 | $<20$ | n/a | 35.3 | 2.4 | 21.0 | 1.9 | 25167 | 1097 | 1600 | 529 | 6480 | 420 | 529 | 34 | 3 | 3 | 1. |
| 4 | 22-Jun-88 | n/a | $<20$ | $\mathrm{n} / \mathrm{a}$ | 39.8 | 3.8 | 33.7 | 10.0 | 23233 | 929 | n/a | n/a | 4740 | 44 | 1050 | 288 | 3 | 3 | 1 |
| 4 | 20-Jun-90 | 0.05 | $<20$ | n/a | 29.3 | 3.0 | 21.4 | 3.4 | 23867 | 757 | 1733 | 58 | 4287 | 90 | 941 | 20 | $<2$ |  | n/a |
| 5 | 22-Jun-88 | n/a | $<20$ | $n / a$ | 44.7 | 9.5 | 35.9 | 22.1 | 25500 | 7889 | n/a | n/a | 5910 | 626 | 720 | 218 | 4 | 4 | 2 |
| 5 | 21-Jun-90 | 0.06 | $<20$ | $n / \mathrm{a}$ | 68.3 | 17.6 | 13.7 | 2.1 | 22067 | 2312 | 1300 | 520 | 4783 | 232 | 775 | 117 | $<2$ |  | n/a |
| 6 | 19-Jun-90 | 0.11 | $<20$ | $n / \mathbf{a}$ | 57.0 | 10.2 | 11.6 | 3.0 | 16967 | 1358 | 1333 | 577 | 5160 | 303 | 349 | 136 | $<2$ | 2 | n/a |
| 7 | 19-Jun-90 | 0.03 | $<20$ | n/a | 65.0 | 3.4 | 16.5 | 0.9 | 27033 | 1041 | 1567 | 513 | 4267 | 202 | 292 | 20 | $<2$ | 2 | n/a |
| 8 | 19-Jun-90 | 0.20 | $<20$ | n/a | 37.0 | 5.3 | 20.1 | 0.9 | 22833 | 839 | 1667 | 577 | 4517 | 105 | 518 | 54 | $<3$ | 3 | n/a |
| 9 | 19-Jun-90 | 0.11 | $<20$ | n/a | 75.8 | 29.0 | 14.7 | 2.3 | 19367 | 1620 | 1967 | 289 | 4857 | 411 | 462 | 263 | $<2$ | 2 | n/a |
| 11 | 19-Jun-90 | 0.37 | $<20$ | $\mathrm{n} / \mathrm{a}$ | 34.0 | 7.3 | 21.8 | 2.7 | 26900 | 964 | 2367 | 306 | 5610 | 355 | 718 | 508 | $<2$ | 2 | n/a |
| 12 | 19-Jun-90 | 0.83 | $<20$ | n/a | 34.2 | 3.3 | 21.3 | 8.4 | 17900 | 520 | 2100 | 458 | 5393 | 93 | 141 | 30 | $<2$ | 2 | n/a |
| 14 | 19-Jun-90 | 0.01 | $<20$ | n/a | 44.0 | 2.3 | 20.9 | 1.8 | 25000 | 608 | 1267 | 462 | 7313 | 330 | 372 | 24 | $<2$ | 2 | $n / a$ |
| 15 | 19-3un-90 | 0.02 | < 20 | n/a | 25.2 | 2.7 | 13.7 | 7 1.2 | 18000 | 1114 | 967 | 58 | 6360 | 481 | 248 | 18 | 3 | 3 | 1 |

STD $=$ standard deviation $n / a=$ analysis not done
APPENDIX III TABLE 4 MT. HUNDERE SEDIMENT DATA SUMMARY FOR JUNE, 1988 AND JUNE, 1990

| STATION NUMBER | SAMPLE DATE | $\begin{gathered} \operatorname{STD}_{(\mu \mathrm{g} / \mathrm{g})}^{\mathrm{Sr}} \end{gathered}$ | $\underset{(\mu \mathrm{g} / \mathrm{g})}{\mathrm{Ti}}$ | $\operatorname{STD}_{(\mu \mathrm{g} / \mathrm{g})}^{\mathrm{TI}}$ |  | $\underset{(\mu \mathrm{grD} / \mathrm{g})}{\mathrm{V}}$ |  | ${\underset{(\mu \mathrm{g} / \mathrm{g})}{\mathrm{Zn}}}^{\mathrm{STD}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 21-Jun-88 | 2.3 | 554 | 50 | 48 | . 4 | 220.7 | 16.0 |
| 2 | 21-Jun-88 | 5.7 | 203 | 55 | 196 | 41 | 14366.7 | 1001.7 |
| 3 | 22-Jun-88 | 2.5 | 459 | 48 | 62 | 4 | 254.3 | 137.4 |
| 3 | 20-Jun-90 | 4.6 | 554 | 60 | 58 | 6 | 160.7 | 17.9 |
| 4 | 22-Jun-88 | 1.6 | 162 | 26 | 41 | 6 | 107.3 | 1.5 |
| 4 | 20-Jun-90 | 2.4 | 212 | 10 | 37 | 4 | 97.7 | 3.0 |
| 5 | 22-Jun-88 | 2.7 | 286 | 31 | 35 | 4 | 77.6 | 7.2 |
| 5 | 21-Jun-90 | 1.9 | 364 | 23 | 37 | 5 | 80.5 | 7.0 |
| 6 | 19-Jun-90 | 1.9 | 622 | 89 | 29 | 2 | 58.2 | 4.6 |
| 7 | 19-Jun-90 | 2.4 | 184 | 29 | 43 | 4 | 113.0 | 7.5 |
| 8 | 19-Jun-90 | 1.7 | 141 | 25 | 36 | 3 | 140.3 | 5.5 |
| 9 | 19-Jun-90 | 4.0 | 291 | 35 | 39 | 5 | 89.0 | 9.4 |
| 11 | 19-Jun-90 | 5.7 | 254 | 26 | 54 | 5 | 196.3 | 3.5 |
| 12 | 19-Jun-90 | 16.1 | 163 | 38 | 39 | 10 | 147.0 | 36.7 |
| 14 | 19-Jun-90 | 3.7 | 132 | 10 | 33 | 3 | 80.2 | 7.0 |
| 15 | 19-Jun-90 | 3.7 | 126 | 8 | 40 | 4 | 132.7 | 20.0 |

APPENDIX III TABLE 4 MT. HUNDERE SEDIMENT DATA SUMMARY FOR JUNE, 1988 AND JUNE, 1990

| STATION NUMBER | SAMPLE DATE | $\begin{gathered} \mathrm{Na} \\ \text { Mean } \\ (\mu \mathrm{g} / \mathrm{g}) \end{gathered}$ | $\begin{gathered} \mathrm{STD}^{\mathrm{Na}} \\ (\mu \mathrm{~g} / \mathrm{g}) \end{gathered}$ | $\begin{gathered} \mathrm{Ni} \\ \text { Mean } \\ (\mu g / g) \end{gathered}$ | $\begin{gathered} \operatorname{STD}^{N 1} \\ (\mu \mathrm{~g} / \mathrm{g}) \end{gathered}$ | $\begin{gathered} \text { Mean } \\ (\mu \mathrm{g} / \mathrm{g}) \end{gathered}$ | $\underset{(\mu \mathrm{g} / \mathrm{g})}{\mathrm{STD}}$ | $\begin{gathered} \mathrm{Pb} \\ \text { Mean } \\ (\mu \mathrm{g} / \mathrm{g}) \end{gathered}$ | $\begin{gathered} \mathrm{STD} \\ (\mu \mathrm{~g} / \mathrm{g}) \end{gathered}$ | $\begin{gathered} \mathrm{Pb} \\ \text { Mean } \\ (\mu \mathrm{g} / \mathrm{g}) \end{gathered}$ | $\operatorname{STD}_{(\mu \mathrm{g} / \mathrm{g})}^{\mathrm{Pb}}$ | $\begin{gathered} \text { Sb } \\ \text { Mean } \\ (\mu g / g) \end{gathered}$ | ${\underset{(\mu \mathrm{g} / \mathrm{g})}{\mathrm{Sb}}}^{\mathrm{STD}}$ | $\begin{array}{r} \text { Si } \\ (\mu \mathrm{g} / \mathrm{g}) \end{array}$ | $\operatorname{STD}_{(\mu \mathrm{g} / \mathrm{g})}^{\mathrm{Si}}$ |  |  | $\begin{aligned} & \text { TD } \\ & / g)^{\text {Sn }} \end{aligned}$ | $\begin{gathered} \mathrm{Sr} \\ \text { Mean } \\ (\mu \mathrm{g} / \mathrm{g}) \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 21-Jun-88 | 93 | 12 | 31 | 2 | 887 | 64 | 27 | 7 | $n / \mathrm{a}$ | $n / \mathrm{a}$ | $\mathrm{n} / \mathrm{a}$ | $n / a$ | 487 | 11 | $<$ | 8 | n/a | 47.7 |
| 2 | 21-Jun-88 | 40 | 10 | 63 | 6 | 2463 | 395 | 11367 | 1234 | n/a | n/a | n/a | n/a | 791 | 36 | $<$ | 8 | n/a | 38.7 |
| 3 | 22-Jun-88 | 100 | 0 | 35 | 9 | 987 | . 23 | 139 | 130 | $n / a$ | $n / a$ | n/a | n/a | 400 | 31 | $<$ | 9 | $\mathrm{n} / \mathrm{a}$ | 45.4 |
| 3 | 20-Jun-90 | 93 | 12 | 32 | 2 | 910 | 17 | 101 | 26 | n/a | $n / \mathrm{a}$ | $<8$ | n/a | 666 | 12 | $<$ | 8 | n/a | 37.3 |
| 4 | 22-Jun-88 | 83 | 6 | 30 | 3 | 1100 | 0 | 23 | 2 | n/a | n/a | n/a | n/a | 349 | 9 | $<$ | 8 | n/a | 39.3 |
| 4 | 20-Jun-90 | 83 | 6 | 27 | 3 | 1033 | 58 | $<10$ | n/a | 11.20 | 1.22 | $<8$ | n/a | 630 | 19 | $<$ | 8 | n/a | 35.7 |
| 5 | 22-Jun-88 | 90 | 10 | 33 | 2 | 753 | 40 | 15 | 9 | n/a | n/a | n/a | n/a | 401 | 97 | $<$ | 8 | $\mathrm{n} / \mathrm{a}$ | 31.0 |
| 5 | 21-Jun-90 | 93 | 12 | 35 | 2 | 727 | 67 | $<8$ | $n / \mathrm{a}$ | 7.99 | 1.18 | $<8$ | $n / a$ | 594 | 43 | $<$ | 8 | n/a | - 28.5 |
| 6 | 19-Jun-90 | 80 | 10 | 32 | 3 | 700 | 62 | $<8$ | n/a | 7.46 | 1.39 | $<8$ | n/a | 577 | 16 | $<$ | 8 | n/a | 28.9 |
| 7 | 19-Jun-90 | 90 | 10 | 40 | 4 | 763 | 45 | 13 | 6 | 12.13 | 2.06 | $<8$ | n/a | 609 | 53 | $<$ | 8 | n/a | 24.7 |
| $b$ | 13-5แぃ-90 | 90 | 10 | 39 | 2 | 857 | 31 | 13 | 6 | 13.90 | $2 .: 1$ | $<8$ | n/a. | 535 | 20 | $<$ | 8 | n/a | 25.8 |
| 9 | 19-J14-90 | 87 | $1:$ | 35 | 5 | 16.3 | 51 | $<8$ | n/a | 1.01 | 1.01 | $<8$ | n/a | 515 | 43 | $<$ | 8 | n/a | 33.5 |
| 11 | 19-3un-90 | 133 | 58 | 34 | 2 | 1033 | 58 | 21 | 2 | 24.07 | 3.86 | $<8$ | n/a | 558 | 60 | $<$ | 8 | n/a | 42.4 |
| 12 | 19-Jı $\mathrm{r}_{1}$-90 | 100 | 0 | 26 | 2 | 883 | 25 | 13 | 5 | 15.37 | 4.20 | $<8$ | $n / a$ | 589 | 111 | $<$ | 8 | n/a | 61.1 |
| 14 | 19-Jun-90 | 210 | 36 | 44 | 2 | 667 | 21 | $<10$ | n/a | 10.36 | 0.63 | $<8$ | n/a | 609 | 49 | $<$ | 8 | n/a | 48.8 |
| 15 | 19-Jun-90 | 70 | 26 | 29 | 7 | 963 | 25 | $<9$ | n/a | 10.60 | 0.44 | $<8$ | n/a | 527 | 25 | $<$ | 9 | n/a | 40.6 |

STD $=$ standard deviation $\quad$ a/a $=$ analysis not done


| NUMBER | INVERTEBRATE | $\begin{array}{r} 1988 \\ \text { STA. } \end{array}$ | - Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 3 \end{array}$ | 8 Tot. | $\begin{array}{r} 1988 \\ \text { STA. } 4 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 4 \end{array}$ | \% Tot. | $\begin{array}{r} 1988 . \\ \text { STA. } 5 \end{array}$ |  | Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 5 \end{array}$ | 8 Tot. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Capnia sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 1 | 0 |
| 2 | Isogenoides sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 1 | 0 |
| 3 | Isoperla sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |  | 0 | 0 | 0 |
| 4 | Kogotus sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 4 |  | 1 | 1 | 0 |
| 5 | Megarcys sp | 0 | 0 | 6 | 2 | 1 | 0 | 3 | 1 | 0 |  | 0 | 0 | 0 |
| 6 | Podmosta sp | 0 | 0 | 60 | 20 | 0 | 0 | 8 | 2 | 3 |  | 1 | 3 | 1 |
| 7 | Sweltsa sp group | 1 | 1 | 50 | 16 | 41 | 10 | 45 | 10 | 6 |  | 2 | 12 | 3 |
| 8 | Skwala (paralella) | 2 | 2 | 0 | 0 | 2 | 0 | 0 | 0 | 1 | - | 0 | 0 | 0 |
| 9 | Utaperla sp | 15 | 13 | 2 | 1 | 39 | 9 | 9 | 2 | 5 |  | 2 | 0 | 0 |
| 10 | Zapada sp | 3 | 3 | 24 | 8 | 10 | 2 | 27 | 6 | 0 |  | 0 | 1 | 0 |
| 11 | Ameletus sp | 16 | 13 | 7 | 2 | 0 | 0 | 3 | 1 | 1 |  | 0 | 7 | 2 |
| 12 | Baetis sp | 12 | 10 | 75 | 25 | 103 | 24 | 123 | 27 | 123 |  | 41 | 17 | 5 |
| 13 | Cinygmula sp | 4 | 3 | 8 | 3 | 81 | 19 | 67 | 15 | 58 |  | 19 | 51 | 14 |
| 14 | Epeorus deceptivus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 1 | 0 |
| 15 | Epeorus (albertae) | 0 | 0 | 2 | 1 | 0 | 0 | 20 | 4 | 0 |  | 0 | 3 | 1 |
| 16 | Epeorus albertae | 0 | 0 | 0 | 0 | 29 | 7 | 0 | 0 | 4 |  | 1 | 0 | 0 |
| 17 | Epeorus longimanus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 10 |  | 3 | 0 | 0 |
| 18 | Ephemerella coloradensis | 0 | 0 | 0 | 0 | 0 | 0 | 10 | 2 | 0 |  | 0 | 38 | 10 |
| 19 | Ephemerella dodds 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 20 | Ephemerella grandis | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 1 | 0 |
| 21 | Ephemerella inermis | 0 | 0 | 0 . | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 4 | 1 |
| 22 | Ephemerella mollitia | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 12 |  | 4 | 0 | 0 |
| 23 | Ephemerella spinifera | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 |  | 1 | 0 | 0 |
| 24 | Ephemerella (grandis?) | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 2 |  | 1 | 0 | 0 |
| 25 | Paraleptophlelia sp | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 26 | Rithrogena sp | 0 | 0 | 13 | 4 | 9 | 2 | 44 | 10 | 3 |  | 1 | 14 | 4 |
| 27 | Unid J/D | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 28 | Unid pupa | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |  | 0 | 0 | 0 |
| 29 | Brachycentrus sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 8 |  | 3 | 35 | 10 |
| 30 | Glossosoma sp | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 1 | 0 |  | 0 | 0 | 0 |
| 31 | Grensia sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 32 | Hydroptila sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 11 | 3 |
| 33 | Parapsyche sp | 0 | 0 | 0 | 0 | 0 | 0 | 7 | 2 | 1 |  | 0 | 0 | 0 |
| 34 | Pseudostenophylax sp | 10 | 8 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 35 | Rhyacophila angelita | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 0 |  | 0 | 0 | 0 |
| 36 | Rhyacophila sp | 0 | 0 | 0 | 0 | 0 | 0 | 4 | 1 | 0 |  | 0 | 0 | 0 |
| 37 | Rhyacophila vagrita | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 0 |  | 0 | 2 | 1 |
| 38 | Rhyacophila (vaolacropedes) | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |  | 0 | 0 | 0 |
| 39 | Ryacophila vaccua | 0 | 0 | 0 | 0 | 4 | 1 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 40 | Ryacophila (acropedes) | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 41 | Culicidae adult | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 42 | (Corynoptera sp?) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 43 | (Hydrelia sp?) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 44 | Palpomyia sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 2 | 1 |

APPENDIX IV TABLE 2 MT. HUNDERE BENTHIC INVERTEBRATE DISTRIBUTION FOR 1988 AND 1990

| NUMBER INVERTEBRATE | $\begin{array}{r} 1988 \\ \text { STA. } 1 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 3 \end{array}$ | \% Tot. | $\begin{array}{r} 1988 \\ \text { STA. } 4 \end{array}$ | 8 Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 4 \end{array}$ | $\%$ Tot. | $\begin{array}{r} 1988 \\ \text { STA. } 5 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 5 \end{array}$ | \% Tot. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 45 Antocha sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 46 Dicranota sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 1 |
| 47 Erioptera sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 48 Hesperoconopa sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 |
| 49 Hexatoma sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 50 Ormosia sp | 2 | 2 | 0 | 0 | 4 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 51 Rhatdomastix sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 52 Tipula sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 53 Pupae | 0 | 0 | $0^{\circ}$ | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 54 Chelifera sp | 0 | 0 | 1 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 3 | 1 |
| 55 Weidemannia sp | 0 | 0 | 3 | 1 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 56 Prosimulium sp | 0 | 0 | 39 | 13 | 0 | 0 | 2 | 0 | 1 | 0 | 1 | 0 |
| 57 Prosimulium sp pupa | 1 | 1 | 0 | 0 | 3 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 58 Simulium sp | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 9 | 3 | 4 | 1 |
| 59 Simulium sp pupae | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| 60 Adult | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 61 Adult | 1 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 62 Pupae | 0 | 0 | 3 | 1 | 3 | 1 | 1 | 0 | 3 | 1 | 15 | 4 |
| 63 Unid J/D | 2 | 2 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 64 Brillia sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 65 Cardiocladius sp | 13 | 11 | 3 | 1 | 13 | 3 | 19 | 4 | 10 | 3 | 11 | 3 |
| 66 Coristemfellina sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| 67 Corynoheura sp | 0 | 0 | 0 | 0 | 7 | 2 | 0 | 0 | 0 | 0 | 3 | 1 |
| -8 Cricotorus sp | 10 | $\varepsilon$ | 1. | 0 | 7 | 2 | 11 | 2 | 13 | 4 | 32 | 9 |
| 69 Dlamesa sp | 15 | 13 | 0 | 0 | 3 | 1 | 0 | 0 | 2 | 1 | 0 | 0 |
| 70 Eukiefferiella sp | 0 | 0 | 2 | 1 | 2 | 0 | 14 | 3 | 0 | 0 | 9 | 2 |
| 71 Euryhapsis sp | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 72 Gymnometriocnemus sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| 73 Heterotrissocladius sp | 1 | 1 | 0 | 0 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 74 Micropsectra sp | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 1 | 5 | 2 | 30 | 8 |
| 75 Monopelopia sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 76 Orthocladius sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| 77 Paratendipes sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 |

APPENDIX IV TABLE 2 MT. HUNDERE BENTHIC INVERTEBRATE DISTRIBUTION FOR 1988 AND 1990

| NUMBER | INVERTEBRATE | $\begin{aligned} & 1988 \\ & \text { STA. } 1 \end{aligned}$ | \% Tot. | $\begin{aligned} & 1990 \\ & \text { STA. } 3 \end{aligned}$ | \% Tot. | $\begin{aligned} & 1988 \\ & \text { STA. } 4 \end{aligned}$ | Tot. | $1990$ $\text { STA. } 4$ | \% Tot. | $\begin{aligned} & 1988 \\ & \text { STA. } 5 \end{aligned}$ | 8 | Tot. | $\begin{aligned} & 1990 \\ & \text { STA. } 5 \end{aligned}$ | \% Tot. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Polypedilum sp | 0 | 0 | 0 | 0 | 12 | 3 | 0 | 0 | 2 |  | 1 | 0 | 0 |
| 79 | Polypedilum (pentapedilum) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
|  | Potthastia sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | - | 0 | 0 | 0 |
|  | Psectrocladius sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | - | 0 | 3 | 1 |
|  | Psectrocladius sp B | 0 | 0 | 0 | 0 | 3 | 1 | 0 | 0 | 0 |  | 0 | 0 | 1 |
| 83 | Rheocricotopus sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
|  | Rheotanytarsus sp | 0 | 0 | 1 | 0 | 0 | 0 | 4 | 1 | 0 |  | 0 | 19 | 5 |
| 85 | Tanytarsus sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 7 |  | 2 | 19 | 0 |
|  | Thienemanniella sp | 0 | 0 | 0 | 0 | 0 | 0 | - 1 | 0 | 0 |  | 0 | 4 | 1 |
| 87 | Thienemannimyla sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 88 | Trichocladius sp | 0 | 0 | 0 | 0 | 7 | 2 | 0 | 0 | 0 |  | 0 | 0 | 0 |
|  | Unid. orthocladilnae | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |  | 0 | 2 | 1 |
|  | Isotomurus sp | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |  | 0 | 0 | 0 |
| 91 | Formicidae | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 92 | Cicadellidae | 0 | 0 | 1. | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 93 | Psyllidae | 0 | 0 | 0. | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 1 | 0 |
| 94 | Aphididea | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 95 | Soldidae | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 96 | Lepidoptera larvae | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | ' | 0 | 0 | 0 |
| 97 | Unid J/D | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 98 | Hydracarina sp | 5 | 4 | 0 | 0 | 14 | 3 | 0 | 0 | 1 |  | 0 | 0 | 0 |
| 99 | Lebertia sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 100 | Newmannia sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 101 | Sperchon sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 3 | 1 |
| 102 | Wandesia sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 1 |
| 103 | Diaptomus sp | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 104 | Cyclopoida | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 105 | Calanolda | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 106 | Daphnia rosea | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 107 | Eurycercus lamellatus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 108 | Polyphemus pediculus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | ' | 0 | 0 | 0 |
| 109 | Nematoda | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |  | 0 | 2 | 1 |
| 110 | Polycelis coronata | 0 | 0 | 0 | 0 | 3 | 1 | 2 | 0 | 0 |  | 0 | 0 | 0 |

APPENDIX IV table 2 MT. hUNDERE benthic invertebrate distribution for 1988 AND 1990

| NUMBER INVERTEBRATE | $\begin{array}{r} 1988 \\ \text { STA. } 1 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 3 \end{array}$ | 8 Tot. | $\begin{array}{r} 1988 \\ \text { STA. } 4 \end{array}$ | 8 Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 4 \end{array}$ | 8 Tot. | $\begin{array}{r} 1988 \\ \text { STA. } 5 \end{array}$ | 8 Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 5 \end{array}$ | \% Tot. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 111 Lumbriculidae, unid J/D | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 6 | 2 |
| 112 Kincaldiana hexatheca | 0 | 0 | 0 | 0 | 3 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| 113 Enchytraeidae | 5 | 4 | 0 | 0 | 7 | 2 | 0 | 0 | 1 | 0 | 0 | 0 |
| 114 Tubificidae | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
|  |  | 100 |  | 100 |  | 100 |  | 100 |  | 100 |  | 100 |
| totals | 119 |  | 304 |  | 422 |  | 454 |  | 298 |  | 365 |  |


| NUMBER | INVERTEBRATE | $\begin{array}{r} 1988 \\ \text { STA. } 6 \end{array}$ | 8 Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 6 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 7 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 8 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STR. } 9 \end{array}$ | 8 | Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 11 \end{array}$ | \% Tot. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Capnia sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 1 | 0 |
|  | Isogenoides sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |  | 0 | 0 | 0 |
| 3 | Isoperla sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | - | 0 | 0 | 0 |
| 4 | Kogotus sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 5 | Megarcys sp | 5 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 6 | Podmosta sp | 0 | 0 | 0 | 0 | 11 | 7 | 11 | 2 | 0 |  | 0 | 7 | 2 |
| 7 | Sweltsa sp group | 14 | 4 | 1 | 1 | 35 | 21 | 32 | 7 | 40 |  | 11 | 8 | 3 |
| 8 | Skwala (paralella) | 25 | 6 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 9 | Utaperla sp | 73 | 18 | 0 | 0 | 3 | 2 | 18 | 4 | 0 |  | 0 | 5 | 2 |
| 10 | Zapada sp | 14 | 4 | 0 | 0 | 3 | 2 | 30 | 7 | 4 |  | 1 | 6 | 2 |
| 11 | Ameletus sp | 5 | 1 | 0 | 0 | 4 | 2 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 12 | Baetis sp | 137 | 35 | 2 | 1 | 2 | 1 | 79 | 17 | 47 |  | 14 | 102 | 34 |
| 13 | Cinygmula sp | 3 | 1 | 3 | 2 | 14 | 8 | 93 | 21 | 19 |  | 5 | 106 | 35 |
| 14 | Epeorus deceptivus | 0 | 0 | 0. | 0 | 1 | 1 | 0 | 0 | 0 |  | 0 | 1 | 0 |
| 15 | Epeorus (albertae) | 0 | 0 | 1. | 1 | 8 | 5 | 55 | 12 | 1 |  | 0 | 23 | 8 |
| 16 | Epeorus albertae | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 17 | Epeorus longimanus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 18 | Ephemerella coloradensis | 0 | 0 | 1 | 1 | 11 | 7 | 5 | 1 | 9 |  | 3 | 2 | 1 |
| 19 | Ephemerella doddsi | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |  | 0 | 0 | 0 |
| 20 | Ephemerella grandis | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 1 |  | 0 | 0 | 0 |
| 21 | Ephemerella inermis | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 22 | Ephemerella mollitia | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 23 | Ephemerella spinifera | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 24 | Ephemerella (grandis?) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 25 | Paraleptophlelia sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 26 | Rithrogena sp | 0 | 0 | 0 | 0 | 3 | 2 | 28 | 6 | 1 |  | 0 | 5 | 2 |
| 27 | Unid J/D | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 28 | Unid pupa | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 29 | Brachycentrus sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 30 | Glossosoma sp | 0 | 0 | 0 | 0 | 0 | 0 | 21 | 5 | 0 |  | 0 | 1 | 0 |
| 31 | Grensia sp | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | - | 0 | 0 | 0 |
| 32 | Hydroptila sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 33 | Parapsyche sp | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 |  | 0 | 0 | 0 |

APPENDIX IV TABLE 2 MT. HUNDERE BENTHIC INVERTEBRATE DISTRIBUTION FOR 1988 AND 1990

| NUMBER INVERTEBRATE | $\begin{array}{r} 1988 \\ \text { STA. } 6 \end{array}$ | $\%$ Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 6 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 7 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 8 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 9 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 11 \end{array}$ | \% Tot. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 34 Pseudostenophylax sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 35 Rhyacophila angelita | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 36 Rhyacophila sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 37 Rhyacophila vagrita | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 38 Rhyacophila (vao\acropedes) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 39 Ryacophila vaccua | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 40 Ryacophila (acropedes) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 41 Culicidae adult | 0 | 0 | 0. | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 42 (Corynoptera sp?) | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 43 (Hydrelia sp?) | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 44 Ealpomyia sp | 0 | 0 | 13 | 8 | 21 | 13 | 3 | 1 | 9 | 3 | 0 | 0 |
| 45 Antocha sp | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 46 Dicranota sp | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 47 Erioptera sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 48 Hesperoconopa sp | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 49 Hexatoma sp | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 1 | 0 | 0 | 0 |
| 50 Ormosia sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 51 Rhabdomastix sp | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 52 Tipula sp | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 |
| 53 Eupae | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 |
| 54 Chelifera sp | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | ${ }^{1}$ | 0 |
| 55 Weidemanria sp | 3 | 1 | 0 | 0 | 0 | 0 | 2 | 0 | 0 | 0 | 1) | 0 |
| 55 Prosimulium sp | 0 | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 1 | 0 | 1) | 0 |
| 57 Fiosimulium sp pupa | 7 | 2 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 4 | 0 |
| 58 Simuliur sp | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 |
| 59 Simulitur sp pupae | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | ü | 1 | 0 |
| 6 6) Adult | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 9 | 0 |
| 61 Adult | 0 | 0 | 0 | 0 | 2 | 1 | 1 | 0 | 0 | 0 | 0 | 0 |
| 62 Pupae | 0 | 0 | 20 | 12 | 3 | 2 | 7 | 2 | 9 | 3 | 6 | 2 |
| 63 Unid J/D | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 64 Brillia sp | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 . | 0 | 0 | 0 |
| 65 Cardiocladius sp | 50 | 13 | 4 | 2 | 0 | 0 | 2 | 0 | 0 | 0 | 4 | 1 |
| 66 Constempellina sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |

APPENDIX IV

| NuMber | invertebrate | $\begin{array}{r} 1986 \\ \text { STA. } 6 \end{array}$ | Tot. | $\begin{array}{r} 1990 \\ \text { STA. } \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } \end{array}$ | ${ }^{*}$ Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 8 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 9 \end{array}$ | 8 | Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 11 \end{array}$ | \% Tot. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 67 | Corynoneura sp | ${ }^{0}$ | 0 | 0 | 0 | 12 | 7 | 21 | 5 | 1 |  | 0 | 0 | 0 |
| 68 | Cricotopus sp | 30 | 8 | 14 | 9 | 5 | 3 | 11 | 2 | 22 |  | 6 | 10 | 3 |
| 69 | Diamesa sp | 10 | 3 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 70 | Eukiefferiella sp | 2 | 1 | 3 | 2 | 4 | 2 | 3 | 1 | 6 |  | 2 | 0 | 0 |
| 71 | Euryhapsis sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | ${ }_{0}$ | 0 |  | $\bigcirc$ | 0 | 0 |
| 72 | Gymnometriocnemus sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 73 | Heterotrissocladius sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 74 | Micropsectra sp | 10 | 3 | 38 | 23 | 4 | 2 | 0 | 0 | 107 |  | 31 | 0 | 0 |
|  | Monopelopia sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 |  |
| 76 | Orthocladius sp | 0 | 0 | 2 | 1 | 0 | 0 | 0 | 0 |  |  | 0 | 0 | ${ }_{0}$ |
| 77 | Paratendipes sp | 0 | 0 | 4 | 2 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 |  |
| 78 | Polypedilum sp | 0 |  | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 79 | Poiypedilum (pentapedilum) | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 80 | Potthastia sp | 0 | 0 | 0 | 0 | 2 | 1 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 81 | Psectrocladius sp | 0 | 0 | 5 | 3 | 0 | 0 | 0 | 0 | 2 |  | 1 |  | 0 |
| 82 | Psectrocladius sp B | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 |  |
| 83 | Rheocricotopus sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 1 |  | 0 |  | 0 |
| 84 | Rheotanytarsus sp | 0 |  | 12 | 7 | 0 | 0 | 0 | 0 | 0 |  | 0 |  | 0 |
| 85 | Tanytarsus sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 |  |
| 86 | Thienemanitilla sp | 0 |  | 0 | - | 2 | 1 | 0 | 0 | 44 |  | 13 |  | 0 |
| 87 | Thienemannimyia sp | 0 | 0 | 5 | 3 | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 |  |
|  | Trichocladius sp | 2 | 1 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  | 0 |  | 0 |
| 89 90 | Unid. Orthocladilinae Isotomurus sp | 0 | 0 | 1 | 1 | 3 | 2 | 1 | 0 | 2 |  | 1 |  | 1 |
| 91 | Formicidae | 0 | 0 | 0 | 0 | 0 | 0 | 0 | $\bigcirc$ | 1 |  | 0 | 0 | 0 |
| 92 | Cicadellidae | 0 | ${ }_{0}$ | 1 | 1 | 1 |  | 1 | ${ }_{0}^{0}$ | ${ }_{0}^{0}$ |  | 0 | 0 | ${ }_{0}^{0}$ |
| 93 | Psyllidae |  | 0 | 0 | 0 | 0 | 0 | 2 | 0 | 0 |  | 0 | 0 | ${ }_{0}^{0}$ |
| 94 | Aphididea | 0 | 0 | 0 |  | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 95 | Soldidae | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |  |  | 0 |  |
| 96 | Lepidoptera larvae | 0 | 0 | 0 | 0 | 0 |  | 0 | 0 | 0 |  | 0 | 0 | 0 |
| 97 98 |  | 0 1 | 0 | 1 | 1 | $\bigcirc$ | 0 | 0 | 0 | 0 |  | 0 |  | 0 |
| 99 | Lebertia sp | 0 | 0 | 2 | 1 | 1 | 1 | 0 | 0 | 0 |  | ${ }_{0}^{0}$ | 0 | ${ }_{0}$ |

APPENDIX IV

| NUMBER INVERTEBRATE | $\begin{array}{r} 1988 \\ \text { STA. } 6 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 6 \end{array}$ | 8 Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 7 \end{array}$ | 8 Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 8 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 9 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 11 \end{array}$ | 8 Tot. |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 100 Newmannia sp | 0 | 0 | 0 | 0 | 0 | 0 | 1 |  |  |  |  |  |
| 101 Sperchon sp | 0 | 0 | 1. | 1 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 102 Wandesia sp | 0 | 0 | 0 | 0 | 0 | 0 | 1 | 0 | 0 | 0 | 0 | 0 |
| 103 Diaptomus sp | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 104 Cyclopoida | 0 | 0 | 3 | 2 | 1 | 1 | 0 | 0 | 0 |  |  | 1 |
| 105 Calanoida | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | O | 2 | 1 |
| 106 Daphnia rosea | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 3 | 1 | 0 | 0 |
| 107 Eurycercus lamellatus | 0 | 0 | 0 | 0 | 1 | 1 | 0 | 0 | 0 | 0 | 0 | 0 |
| 108 Polyphemus pediculus | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 109 Nematoda | 0 | 0 | 15 | 9 | 0 | 0 | 3 | 1 | 6 | 2 | 2 | 1 |
| 110 Polycelis coronata | 1 | 0 | 0 | 0 | 1 | 1 | 11 | 2 | 0 | 0 | 2 | 1 |
| 111 Lumbriculidae, unid J/D | 0 | 0 | 2 | 1 | 1 | 1 | 0 | 0 | 1 | 0 | 5 | 2 |
| 112 Kincaidiana hexatheca | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| 113 Enchytraeidae | 1 | 0 | 1 | 1 | 0 | 0 | 1 | 0 | 2 | 1 | 0 | 0 |
| 114 Tubifigidae | 0 | 0 | 4 | 2 | 0 | 0 | 0 | 0 | 3 | 1 | 0 | 0 |
|  |  | 100 |  | 100 |  | 100 |  | 101 |  | 100 |  | 100 |
| TOTALS | 395 |  | 164 |  | 167 |  | 453 |  | 348 |  | 303 | 100 |


| number | INVERTEBRATE | $\begin{array}{r} 1990 \\ \text { STA. } 12 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 15 \end{array}$ | 8 Tot. |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | Capnia sp | 0 | 0 | 1 | 3 |
| 2 | Isogenoides sp | 0 | 0 | 0 | 0 |
| 3 | Isoperla sp | 0 | 0 | 0 | 0 |
| 4 | Kogotus sp | 1 |  | 0 | 0 |
| 5 | Megarcys sp | 0 | - | 0 | 0 |
| 6 | Podmosta sp | 2 | 1. | 0 | 0 |
| 7 | Sweltsa sp group | 0 | 0 | 1 | 3 |
| 8 | Skwala (paralella) | 0 | 0 | 0 | 0 |
| 9 | Utaperla sp | 1 |  | 1 | 3 |
| 10 | zapada sp | 4 | - 2 | 0 | 0 |
| 11 | Ameletus sp | 0 | 0 | 0 | 0 |
| 12 | Baetis sp | 7 | 3 | 3 | 8 |
| 13 | Cinygmula sp | 53 | 22 | 0 | 0 |
| 14 | Epeorus deceptivus | 0 | 0 | 0 | 0 |
| 15 | Epeorus (albertae) | 0 | 0 | 0 | 0 |
| 16 | Epeorus albertae | 0 | 0 | 0 | 0 |
| 17 | Epeorus longimanus | 0 | 0 | 0 | 0 |
| 18 | Ephemerella coloradensis | 48 | 20 | 0 | 0 |
| 19 | Ephemerella doddsi | 0 | 0 | 0 | 0 |
| 20 | Ephemerella grandis | 0 | 0 | 0 | 0 |
| 21 | Ephemerella inermis | 0 | 0 | 0 | 0 |
| 22 | Ephemerella mollitia | 0 | 0 | 0 | 0 |
| 23 | Ephemerella spinifera | 0 | 0 | 0 | 0 |
| 24 | Ephemerella (grandis?) | 0 | 0 | 0 | 0 |
| 25 | Paraleptophlelia sp | 0 | 0 | 0 | 0 |
| 26 | Rithrogena sp | 1 | 0 | 0 | 0 |
| 27 | Unid J/D | 0 | 0 | 0 | 0 |
| 28 | Unid pupa | 0 | 0 | 0 | 0 |
| 29 | Brachycentrus sp | 0 | 0 | 0 | 0 |
|  | Glossosoma sp | 5 | 2 | 1 | 3 |
| 31 | Grensia sp | 0 | 0 | 0 | 0 |
| 32 | Hydroptila sp | 0 | 0 | 0 | 0 |
| 33 | Parapsyche sp | 0 | 0 | 0 | 0 |


| NUMBER | INVERTEERATE | $\begin{array}{r} 1990 \\ \text { STA. } 12 \end{array}$ | 8 Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 15 \end{array}$ | 8 Tot. |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 34 | Pseudostenophylax sp | 0 | 0 | 0 | 0 |
| 35 | Rhyacophila angelita | 0 | 0 | 0 | 0 |
| 36 | Rhyacophila sp | 0 | 0 | 0 | 0 |
| 37 | Rhyacophila vagrita | 0 | 0 | 0 | 0 |
| 38 | Rhyacophila (vao\acropedes) | 0 | 0 | 0 | 0 |
| 39 | Ryacophila vaccua | 0 | 0 | 0 | 0 |
| 40 | Ryacophila (acropedes) | 0 | 0 | 0 | 0 |
| 41 | Culicidae adult | 1 | 0 | 0 | 0 |
| 42 | (Corynoptera sp?) | 0 | 0 | 0. | 0 |
| 43. | (Hydrelia sp?) | 0 | 0 | 0 | 0 |
| 44 | Palpomyia sp | 0 | 0 |  | 3 |
| 45 | Antocha sp | 0 | 0 | 0 | 0 |
| 46 | Dicranota sp | 0 | 0 | 3 | 8 |
| 47 | Erioptera sp | 2 | 1 | 0 | 0 |
| 48 | Hesperoconopa sp | 0 | 0 | 0 | 0 |
| 49 | Hexatoma sp | 0 | 0 | 0 | - 0 |
| 50 | Ormosia sp | 0 | 0 | 0 | 0 |
| 51 | Rhabdomastix sp | 0 | 0 | 0 | 0. |
| 52 | Tipula sp | 1 | 0 | 1 | 3 |
| 53 | Pupae | 6 | 2 | 0 | 0 |
| 54 | Chelifera sp | 2 | 1 | 0 | 0 |
| 55 | Weidemannia sp | 0 | 0 | 0 | 0 |
| 56 | Prosimulium sp | 0 | 0 | 0 | 0 |
| 57 | Prosimuljuri sp pupa | 0 | 0 | 0 | 0 |
| 58 | Simulium sp | 0 | ${ }^{0}$ | 0. | 0 |
| 59 | Simulium sp pupae | 0 | 0 | 0 | 0 |
| 50 | Adult | 0 | 0 | 0 | 0 |
| 51 | Adult | 1 | 0 | 0 | 0 |
| 62 | Pupae | 4 | 2 | 0 | 0 |
| 63 | Unid J/D | 0 | 0 | 0 | 0 |
| 64 | Brillia sp | 0 | 0 | 0 | 0 |
| 65 | Cardiocladius sp | 28 | 12 | 0 | - 0 |
| 66 | Constempellina sp | 0 | 0 | 0 | 0 |


| Number | INVERTEBRATE | $\begin{array}{r} 1990 \\ \text { STA. } 12 \end{array}$ | \% Tot. | $\begin{array}{r} 1990 \\ \text { STA. } 15 \end{array}$ | - Tot. |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 67 | Corynoneura sp | 1 | 0 | 0 | 0 |
| 68 | Cricotopus sp | 5 | 2 | 2 | 6 |
| 69 | Diamesa sp | 0 | 0 | 0 | 0 |
| 70 | Eukiefferiella sp | 1 | 0 | 1 | 3 |
| 71 | Euryhapsis sp | 0 | 0 | 0 | 0 |
| 72 | Gymnometriocnemus sp | 0 | 0 | 0 | 0 |
| 73 | Heterotrissocladius sp | 0 | 0 | 0 | 0 |
| 74 | Micropsectra sp | 12 | 5 | 2 | 6 |
| 75 | Monopelopia sp | 2 | 1 | 0 | 0 |
| 76 | Orthocladius sp | 0 | 0 | 0 | 0 |
| 77 | Paratendipes sp | 0 | 0 | 0 | 0 |
| 78 | Polypedilum sp | 0 | 0 | 0 | 0 |
| 79 | Polypedilum (pentapedilum) | 0 | 0 | 3 | 8 |
| 80 | Potthastia sp | 0 | 0 | 0. | 0 |
| 81 | Psectrocladius sp | 1 | 0 | 0. | 0 |
| 82 | Psectrocladius sp B | 0 | 0 | 0 | 0 |
| 83 | Rheocricotopus sp | 0 | 0 | 0 | 0 |
| 84 | Rheotanytarsus sp | 0 | 0 | 0 | 0 |
| 85 | Tanytarsus sp | 0 | 0 | 0 | 0 |
| 86 | Thienemanniella sp | 2 | 1 | 13 | 36 |
| 87 | Thienemannimyia sp | 0 | 0 | 0 | 0 |
| 88 | Trichocladius sp | 0 | 0 | 0 | 0 |
| 89 | Unid. orthocladiinae | 1 | 0 | 1 | 3 |
| 90 | Isotomurus sp | 0 | 0 | 0 | 0 |
| 91 | Eormicidae | 0 | 0 | 0 | 0 |
| 92 | Clcadellidae | 1 | 0 | 0 | 0 |
| 93 | Psyllidae | 2 | 1 | 0 | 0 |
| 94 | Aphididea | 0 | 0 | 0 | 0 |
| 95 | Soldidae | 0 | 0 | 0 | 0 |
| 96 | Lepidoptera larvae | 0 | 0 | 0 | 0 |
| 97 | Unid J/D |  | 0 | 0 | 0 |
| 98 | Hydracarina sp | 0 | 0 | - | 0 |
| 99 | Lebertia sp | 0 | 0 | 0 | - 0 |




[^0]:    
     Bethesda, MD, 313 BP .

[^1]:    4. Anonymous 1977. Guldellnes for Establishing Water Quality Objectives for the Territorial Waters of the
    Yukon and Northwest Territories. Report of the Working Group on Water Quality Objectives to the
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