# Sediment Quality in Lake Superior Tributaries: A Screening-Level Survey 

Ecosystem Health Division, Ontario Region<br>Environmental Conservation Branch

## Environment Canada

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## Executive Summary

A survey of sediment quality was undertaken in the summer and fall of 2005 in the mouths of tributaries draining to Lake Superior. A total of 117 samples were obtained, representing 108 tributaries. Six (6) samples were field duplicates and given fictitious names.

The sampling program was based on the Guidelines for Collecting and Processing Samples of Stream Bed Sediment for Analysis of Trace Elements and Organic Contaminants, developed by the United States Geological Survey (USGS) for the U.S. National Water-Quality Assessment Program (NAWQA; Shelton and Capel, 1994). A number of sub-samples are combined at each site so that one sample is obtained that is representative of the overall conditions in that tributary.

The samples were analyzed for 26 organochlorine compounds plus seven (7) PCB Aroclors and Total PCBs. Sixteen (16) PAH compounds and 27 metals were analyzed, and the inorganic and organic carbon content as well as loss on ignition was determined. For many of the tributaries, this study represents the first information related to organic compounds in sediments.

Twenty three (22) organochlorine compounds were not detected in any sample. DDT and metabolites DDE and DDD were the only detected organochlorines, with only 2 occurrences, both above the PEL. One or more PCB Aroclor was detected at nine sites. Total PCB concentrations were detected at only two (2) sites both below any guidelines. Polycyclic aromatic hydrocarbons (PAHs) were found more often, with one or more of the 16 PAH compounds detected at 34 sites (i.e., detection frequency of $31 \%$ ). Exceedences of one or more federal TEL guideline for PAHs occurred at $15 \%$ of the sites and PEL exceedences occurred at a further two (2) sites. In general, PAH concentrations were lower than found on the lower great lakes. This is probably do to the lower urbanization found along Lake Huron.

At most sites, the detections of metals are likely related to the natural occurrence of trace elements in stream sediments. For some metals, however, concentrations appear to be elevated to a degree that is considered to be toxic to aquatic biota. These metals include: chromium, zinc. Copper has a natural occurrence in Michigan which reflects levels of greater than ten times the provincial serious effect level. Other metals, including manganese and iron, appeared to be elevated at certain sites but these higher levels might be related to natural sources.

### 1.0 Introduction and Purpose

The Ecosystem Health Division (EHD) of Environment Canada (EC), Ontario Region, conducted a screening-level survey of sediment quality in both Canadian and American tributaries to Lake Superior in the fall of 2005. The sampling represents the final screening stage of a track-down program to identify potential sources of contamination to the upper Great Lakes that are not being addressed by other Great Lakes programs. The program constitutes a portion of Environment Canada's commitment towards the Great Lakes Water Quality Agreement (GLWQA).

The purpose of the sampling was to assess sediment quality in deposition zones in each tributary prior to discharge to Lake Superior. One sediment sample, consisting of many sub samples, was taken from each tributary in a manner that is representative of the overall sediment quality in that tributary.

The study was designed to maximize the probability of detecting persistent toxic substances entering the lake, if they exist. The intent of the program is to identify remaining sources of contamination for subsequent follow-up work. It is not the intent at this stage to quantify the loadings of contaminants entering Lake Superior. Instead, the results from this program will be combined with existing water quality, fish, benthic and sediment contaminant information, using a weight-of-evidence approach, to prioritize subsequent track-down efforts.

As with the Lake Erie, Lake Ontario and Lake Huron screening studies, this study also targeted parameters for the sediment screening identified in their respective Lakewide Management Plan (LaMP) as impairing lake-wide beneficial uses. In addition, a suite of contaminants targeted for virtual elimination in the Canada-U.S. Binational Toxics Strategy (BTS) were considered in order to assess Canada's commitments towards that Strategy. Additional parameters were included for contextual information (such as particle size and total organic carbon) and to improve our understanding of the contaminant status of Lake Superior tributaries (e.g., metals, pesticides, contaminants of emerging concern).

### 2.0 Methodology

To achieve the study objectives, the sampling program consisted of a survey-level, screening assessment of recently deposited sediment quality near the mouths of tributaries entering Lake Superior. The targeted substances are relatively insoluble in water (i.e., hydrophobic) and are therefore typically found at higher concentrations in sediments than in water. In addition, bed sediments in depositional environments provide a time-integrated sample of particulate matter transported by a stream. Analysis of bed sediments alleviated problems associated with detecting trace levels of substances in water samples. Bed sediment sampling can overcome problems detecting periodic or intermittent sources of contaminants in water from non-point pollution sources.

### 2.1 Field Program

## Tributary Selection

A reconnaissance survey was conducted in May and June 2005 to identify tributaries and select sampling sites. Sediment deposition zones were sought near the mouths of the tributaries such that they were likely downstream from potential contaminant sources yet sufficiently far upstream not to be influenced by the water body into which it drains. In other words, sites were selected to be outside of the zone of lake influence.

During the reconnaissance survey, the method of access was also identified. Most sites were accessed by wading or were sampled from a bridge crossing. In certain, larger tributaries, sampling sites were accessed by boat. In the majority of cases, the sample site coincided with the most downstream road crossing of the tributary.

## Number of Sites

Significantly less tributaries, containing sediment, draining the Lake Superior watersheds was sampled in this program as compared to previous tributary screening studies. Lake Superior represents the least urban developed Great Lake and therefore presented many challenges in accessing tributaries. Many sites were not sampled as access was prohibitable, 29 sites were attempted without success to sample and many other sites that were accessed ( 27 sites) did not have sediment. As well a further 4 sites were on First Nation Land and permission to sample was not given. As a result a total of 57 Canadian and 51 American sites were sampled. For many sites, this program has provided its first information about organic contaminants in sediments. The geographic extent of the program was restricted to Lake Superior from just east of the St. Marys River on the northern tip of the North Channel at Echo Bay. The tributaries sampled during the project are shown in Figure 1.

A total of 117 samples were obtained, representing 108 tributaries. As mentioned above, a single site sediment sample was generally taken from depositional zones upstream of the tributary mouth.

Of the 117 samples, 9 were blind duplicate samples; that is, they were split samples that were assigned a fictitious name in the field (usually a name of a common bird, unless a more appropriate name was conceived). The blind duplicates were obtained to assess variability due to sample handling and laboratory precision. A list of blind duplicates and the corresponding tributary is provided in Table 1, below.

Figure 1. Lake Superior tributaries sampled


Table 1. Blind Duplicate Sample Listing

| Tributary | Blind Duplicate Sample |
| :---: | :---: |
| Big Trout Creek | Goshawk Creek |
| Goulais River | Vulture Creek |
| Little Pic River | Kestrel Creek |
| Neebing River | Heron Creek |
| Prairie River | Redtail Creek |
| Sawmill Creek | Harrier Creek |
| Temperance River | Boa Creek |
| Grand Marais Creek | Rattler Creek |
| Sturgeon River | Copperhead Creek |

## Sampling Methodology

The sampling program was based on the Guidelines for Collecting and Processing Samples of Stream Bed Sediment for Analysis of Trace Elements and Organic Contaminants, developed by the United States Geological Survey (USGS) for the U.S. National Water-Quality Assessment Program (NAWQA; Shelton and Capel, 1994). In the NAWQA program, downstream locations in watersheds are selected to provide a coarse-scale network of sites. At these "integrator" sites, large-scale problems that may not be detected in smaller basins have a reasonable chance of being detected. A number of sub-samples are combined so that one sample is obtained that is representative of the overall conditions in that tributary.

## Field Campaign

Sampling was conducted between August 15 - Sept 16 and October 3, 2005 for Canadian tributaries and from Sept 17 - Oct 26 for the American tributaries. One or more depositional reach was sampled upstream of the mouth in each tributary. Only the very fine-grained surface deposits, to a maximum depth of approximately 1 or 2 cm , depending on the site, were collected. These surface sediments better represent relatively recent rather than historic deposition. Sites were selected to be representative of the variety of locations (i.e., mid-channel, left bank, right bank) and habitat types (pools, different depths of water, and depositional zones behind obstacles such as boulders or sand bars) present at each site. Only wetted depositional zones were sampled.

Where water depths permitted wading and water velocities were slow enough to permit sample retrieval, samples were obtained using a stainless steel spoon and collected in a glass bowl. At sites where the water depth was too great for wading or water velocities were swift enough to wash the fine particles from the spoons during sample retrieval, an all-stainless steel Wildco Petite Ponar sampler was used.

Upon arrival at each site, the sampling equipment was thoroughly rinsed in the ambient river water. The surface sediments were collected (either by spoon or Ponar, as described above) and combined in a glass bowl. The sediments were sieved through a $2-\mathrm{mm}$ stainless steel sieve to
remove the larger size fractions and to assist with homogenization of the sample. The sample was further homogenized by mixing with a spoon for approximately two minutes.

Several sample jars were filled at each site. In general, four jars were used:

- one $125-\mathrm{mL}$ polyethylene container filled with approximately 2 cm of sediment for metals analysis;
- one $125-\mathrm{mL}$ polyethylene container filled approximately $1 / 2$ full for total organic carbon and grain size analysis;
- one $250-\mathrm{mL}$ glass container with Teflon-lined screw cap filled approximately $3 / 4$ full for organochlorine (OC) and polyaromatic hydrocarbon (PAH) analysis, and;
- one $250-\mathrm{mL}$ Teflon or glass container filled approximately $3 / 4$ full for archiving purposes.

Blind duplicate samples were obtained at nine sites, including blind duplicate archive samples at each of these.

Sample jars were labeled with permanent marker on both the lids and on laboratory tape affixed to the side of the jars. The recorded information included the site name, date, organization (EHD/OR), and parameters for analysis (e.g. OCs and PAHs, metals, TOC and grain size, Archive). After the appropriate sample jars were filled, the sampling equipment was thoroughly rinsed in the ambient river water.

A field drawing was made and digital photos were taken at each site. A sketch of each tributary reach was made to include its major features, habitat types, approximate dimensions, surrounding land uses, major road crossings, etc. The locations and number of sampling sites were identified on each sketch, and the method of sediment retrieval was noted. A Lowrance Global Map 100 geographic positioning system (GPS) device was used to obtain each location using the position averaging function. The GPS location within the site was included on the sketch.

Samples were kept on ice in portable coolers while in the field. Upon return to the Canada Centre for Inland Waters in Burlington, the samples were decanted then frozen at $-10^{\circ} \mathrm{C}$. Samples in glass bottles were frozen on their sides to prevent bottle breakage.

### 2.2 Laboratory Methods

The samples in polyethylene containers (i.e., those for metals, TOC and grain size analysis) were freeze-dried prior to analysis. Samples were sent to Natural Resources Canada in Ottawa, Ontario. All metals samples, TOC and grain size samples were freeze-dried at these facilities. Once freeze-dried, TOC was analyzed by Leco Cr-412 and grain size fractions were determined using a Lecotrac Particle Size Analyzer LT100.

Metal analysis was performed in their entirety by Caduceon Enterprises Inc. The laboratory performed the metals analysis (including mercury) on freeze-dried sediment samples using aqua regia digestion methods.

Analysis of organochlorines (OCs), including PCBs and polycyclic aromatic hydrocarbons (PAHs) was awarded to Maxxam Analytics Inc. in Mississauga as the result of a competitive bidding process. Frozen, wet sediment samples were sent to Maxxam in the winter of 2006.

Samples were thawed and OCs were analyzed by gas chromatography/dual column electron capture detector (GC/ECD) after accelerated solvent extraction following the EPA protocol SW846 EPA 3545. Samples for PAH analysis were extracted using a sonication method. The extracts were then concentrated and analyzed by mass spectrometry (GC/MS). Sample results were reported on a dry weight basis.

The archived sediments have proven useful for a variety of purposes to date. The National Laboratory for Environmental Testing (NLET) is analyzing selected samples for selected compounds of emerging concern (e.g., polybrominated diphenyl ethers, selected musk compounds). Dioxin, furans, dioxin-like PCBs and polychlorinated napthalene analysis was conducted on 13 samples by the Ontario Ministry of the Environment.

### 2.3 Data Analysis

The laboratory results were analyzed in a spreadsheet program. Results were compared with the Federal and Provincial sediment quality objectives and with other sites in the program. The frequency of detection and frequency of exceedances of the sediment quality objectives were computed.

### 3.0 Results

Throughout this report, references and comparisons are made to the federal and provincial sediment quality guidelines.

### 3.1 Quality Assurance/Quality Control

All laboratories used for the project were CAEAL accredited for their respective analytical parameters. As mentioned above in the methodology, Maxxam Analytical Inc. performed the organochlorine and polycyclic aromatic hydrocarbon analyses. The Maxxam laboratory QA/QC program consisted of blanks, spiked blanks and duplicate samples (i.e., laboratory replicate runs).

All method blanks were within acceptable limits (below method detection limit) and spikes were within acceptable limits ( $40-130 \%$ ) for all parameters.

Paired student t -tests were also performed to assess differences between blind duplicate samples submitted to the laboratory. The majority of the inorganic parameters could be assessed this way, with the exception of parameters that were detected in fewer than three samples. Organic parameters could not be assessed do to there non-detects. There were no significant differences observed between the blind duplicate samples, for any of the parameter assessed, at the $95 \%$ confidence level.

### 3.2 Method Detection Limits

All of the analytical parameters used in the study are hydrophobic, i.e., they have a propensity for solid surfaces such as sediments as opposed to the dissolved phase. Sampling very fine, flocculent surface deposits, serves to maximize the probability of encountering these analytes, if they are present in the environment. Typical laboratory detection limits are therefore sufficient to detect these parameters at ambient concentrations. The laboratory method detection limits are
provided in Table 2, below, for both laboratories used in this study. This Table also provides a useful reference of all parameters measured in the study.

Table 2. Analytical Parameters and Laboratory Method Detection Limits a. Maxxam Analytics Inc. (Organics)

| Polychlorinated Biphenyls(PCBs) |  | Polycyclic Aromatic Hydrocarbons(PAHs) |  | $\frac{\text { Organochlorine Pesticides }}{(\mathrm{OCs})}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Parameter | MDL | Parameter | MDL | Parameter | MDL |
| Aroclor 1016 | $0.01 \mu \mathrm{~g} / \mathrm{g}$ | Naphthalene | $5 \mu \mathrm{~g} / \mathrm{kg}$ | Hexachlorobenzene | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
| Aroclor 1221 | $0.03 \mu \mathrm{~g} / \mathrm{g}$ | Acenaphthylene | $5 \mu \mathrm{~g} / \mathrm{kg}$ | o,p'-DDD | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
| Aroclor 1232 | $0.01 \mu \mathrm{~g} / \mathrm{g}$ | Acenaphthene | $10 \mu \mathrm{~g} / \mathrm{kg}$ | Endrin aldehyde | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
| Aroclor 1242 | $0.02 \mu \mathrm{~g} / \mathrm{g}$ | Fluorene | $5 \mu \mathrm{~g} / \mathrm{kg}$ | o,p'-DDT | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
| Aroclor 1248 | $0.01 \mu \mathrm{~g} / \mathrm{g}$ | Phenanthrene | $5 \mu \mathrm{~g} / \mathrm{kg}$ | Toxaphene | $0.08 \mu \mathrm{~g} / \mathrm{g}$ |
| Aroclor 1254 | $0.01 \mu \mathrm{~g} / \mathrm{g}$ | Anthracene | $5 \mu \mathrm{~g} / \mathrm{kg}$ | o,p'-DDE | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
| Aroclor 1260 | $0.01 \mu \mathrm{~g} / \mathrm{g}$ | Fluoranthene | $5 \mu \mathrm{~g} / \mathrm{kg}$ | Aldrin | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
| Total PCB | $0.01 \mu \mathrm{~g} / \mathrm{g}$ | Pyrene | $5 \mu \mathrm{~g} / \mathrm{kg}$ | $\alpha-\mathrm{HCH}$ | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  | Benz(a)anthracene | $10 \mu \mathrm{~g} / \mathrm{kg}$ | $\beta-\mathrm{HCH}$ | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  | Chrysene | $10 \mu \mathrm{~g} / \mathrm{kg}$ | $\delta-\mathrm{HCH}$ | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  | Benzo(b)fluoranthene | $10 \mu \mathrm{~g} / \mathrm{kg}$ | Lindane | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  | Benzo(k)fluoranthene | $10 \mu \mathrm{~g} / \mathrm{kg}$ | $\alpha$-Chlordane | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  | Benzo(a)pyrene | $5 \mu \mathrm{~g} / \mathrm{kg}$ | $\gamma$-Chlordane | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  | Indeno(1,2,3-cd)pyrene | $20 \mu \mathrm{~g} / \mathrm{kg}$ | p,p'-DDD | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  | Dibenzo(a,h)anthracene | $20 \mu \mathrm{~g} / \mathrm{kg}$ | $\mathrm{p}, \mathrm{p}$ '-DDE | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  | Benzo(ghi)perylene | $20 \mu \mathrm{~g} / \mathrm{kg}$ | p,p'-DDT | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  |  |  | Dieldrin | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  |  |  | $\alpha$-Endosulfan | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  |  |  | $\beta$-Endosulfan | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  |  |  | Endosulfan sulfate | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  |  |  | Endrin | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  |  |  | Heptachlor | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  |  |  | Heptachlor epoxide | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  |  |  | Methoxychlor | $0.008 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  |  |  | Mirex | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |
|  |  |  |  | Octachlorostyrene | $0.002 \mu \mathrm{~g} / \mathrm{g}$ |

All laboratory method detection limits, for organic analytes are below the federal PEL except for Lindane (PEL sediment quality guideline $0.00138 \mu \mathrm{~g} / \mathrm{g}$ ). Most of the laboratory detection limits for the organic analytes are below the TEL with the exception of total DDT, total DDE, toxaphene, heptachlor epoxide and acenaphthene.

Table 2 cont. Analytical Parameters and Laboratory Method Detection Limits b. Caduceon Enterprises (Metals)

| Parameters | Units | MDL |
| :--- | :--- | :--- |
| Aluminum | $\%$ | 0.01 |
| Antimony | $\mu \mathrm{g} / \mathrm{g}$ | 5 |
| Arsenic | $\mu \mathrm{g} / \mathrm{g}$ | 5 |
| Barium | $\mu \mathrm{g} / \mathrm{g}$ | 1 |
| Beryllium | $\mu \mathrm{g} / \mathrm{g}$ | 0.2 |
| Bismuth | $\mu \mathrm{g} / \mathrm{g}$ | 5 |
| Cadmium | $\mu \mathrm{g} / \mathrm{g}$ | 0.5 |
| Calcium | $\%$ | 0.01 |
| Chromium | $\mu \mathrm{g} / \mathrm{g}$ | 1 |
| Colbalt | $\mu \mathrm{g} / \mathrm{g}$ | 1 |
| Copper | $\mu \mathrm{g} / \mathrm{g}$ | 1 |
| Iron | $\%$ | 0.01 |
| Lead | $\mu \mathrm{g} / \mathrm{g}$ | 5 |
| Lithium | $\mu \mathrm{g} / \mathrm{g}$ | 0.5 |
| Magnesium | $\%$ | 0.01 |
| Manganese | $\mu \mathrm{g} / \mathrm{g}$ | 0.5 |
| Molybdenum | $\mu \mathrm{g} / \mathrm{g}$ | 1 |
| Nickel | $\mu \mathrm{g} / \mathrm{g}$ | 1 |
| Phosphorus | $\mu \mathrm{g} / \mathrm{g}$ | 3 |
| Potassium | $\%$ | 0.05 |
| Silver | $\mu \mathrm{g} / \mathrm{g}$ | 0.1 |
| Sodium | $\%$ | 0.01 |
| Strontium | $\mu \mathrm{g} / \mathrm{g}$ | 1 |
| Tin | $\mu \mathrm{g} / \mathrm{g}$ | 10 |
| Titanium | $\mu \mathrm{g} / \mathrm{g}$ | 0.5 |
| Vanadium | $\mu \mathrm{g} / \mathrm{g}$ | 1 |
| Yttrium | $\mu \mathrm{g} / \mathrm{g}$ | 0.5 |
| Zinc | $\mu \mathrm{g} / \mathrm{g}$ | 1 |
| Zirconium | $\mu \mathrm{g} / \mathrm{g}$ | 0.1 |
| Mercury | $\mathrm{ng} / \mathrm{g}$ | 5 |
|  |  |  |

All laboratory method detection limits for metals were below federal sediment quality guidelines.

### 3.3 Laboratory Results

A review of the detection frequency of analytical parameters and exceedences of sediment quality guidelines is provided here. A discussion of the highest observed levels is provided for selected parameters in Section 4. A full listing of the laboratory data for the 108 unique sites is
provided in Appendix A. The laboratory data for the blind duplicate samples is not provided but can be obtained from Environment Canada.

### 3.3.1 Frequency of Detection

In general, organochlorine parameters were not detected, with only a few exceptions. Of the Canadian tributaries none of the 26 organochlorine compounds were detected and of the US tributaries a total of twenty two (22) organochlorine parameters were not detected in any sample. In addition, five (5) PCB Aroclors were not detected in Canadian tributary while no PCBs were detected in any U.S. tributary. Of the sixteen (16) PAHs, all were detected in at least one sample. Four (4) of the trace metals were not detected in any sample. The parameters that were not detected are listed below in Table 3.

Table 3. Parameters Not Detected

| Organochlorines |
| :---: |
| - Hexachlorobenzene <br> - o,p'-DDE <br> - o,p'-DDD <br> - Endrin aldehyde <br> - Toxaphene <br> - Aldrin <br> - $\alpha-\mathrm{HCH}$ <br> - $\beta-\mathrm{HCH}$ <br> - $\delta-\mathrm{HCH}$ <br> - Lindane <br> - $\alpha$-Chlordane <br> - $\gamma$-Chlordane <br> - Dieldrin <br> - $\alpha$-Endosulfan <br> - $\beta$-Endosulfan <br> - Endosulfan sulfate <br> - Endrin <br> - Heptachlor <br> - Heptachlor epoxide <br> - Methoxychlor <br> - Mirex <br> - Octachlorostyrene |
| PCB Aroclors |
| - Aroclor 1016 <br> - Aroclor 1221 <br> - Aroclor 1232 <br> - Aroclor 1242 <br> - Aroclor 1248 |
| Metals |
| - Antimony <br> - Bismuth <br> - Molybdenum <br> - Tin |

### 3.3.2 Comparison of Results with Sediment Quality Guidelines

The sediment quality results were compared to the Canadian Environmental Quality Guidelines (Canadian Council of Ministers of the Environment, 2001). The CCME sediment quality guidelines provide scientific benchmarks, or reference points, for evaluating the potential for observing adverse biological effects in aquatic systems. The guidelines are derived from available toxicological information. A lower value, referred to as the threshold effect level (TEL), represents the concentration below which adverse biological effects are expected to occur rarely. The upper value, referred to as the probable effect level (PEL), represents the level above which adverse effects are expected to occur frequently. Fewer than $25 \%$ of adverse effects (in the Biological Effects Database for Sediments) occur below the TEL, and more than $50 \%$ of adverse effects occur above the PEL.

Where no federal guidelines were available, the provincial guidelines were used for comparison (Persaud et al., 1992). Provincial Severe Effect Levels for organic compounds and polycyclic aromatic hydrocarbons were calculated individually for each site using the organic carbon concentration in the sediment. However, no SEL exceedences were determined for these compounds in this study.

A special mention should be made of toxaphene. At present the only guideline that is available is an interim sediment quality guideline which the federal government has adopted from the New York State Department of Environmental Conservation (NYSDEC 1994), $0.01 \mu \mathrm{~g} / \mathrm{g}$ TOC, which has been converted to dry weight. This value is the lowest available guideline from other jurisdictions; in fact it is lower than many laboratory detection limits. This guideline is considerably lower than the $80 \mu \mathrm{~g} / \mathrm{kg}$ method detection limit reported for this program. While it is acknowledged that lower detection limits would be more beneficial for screening sediments it is pointed out that toxaphene is not a critical pollutant; and that sediment inventories of the Great Lakes support atmospheric transport as opposed to local sources (Muir et. al. 2005).

Table 5 provides a summary of the numbers of exceedences of the federal guidelines, and exceedences of the provincial guidelines for those parameters for which federal guidelines are not available. A complete list of the sediment quality guidelines relevant to this study is provided in Appendix B.

Table 5. Number of Sites Exceeding Sediment Quality Guidelines

| A. Metals | Federal Guidelines |  | Provincial Guidelines |  |
| :---: | :---: | :---: | :---: | :---: |
|  | Exceeds TEL ${ }^{1}$ <br> Below PEL | Exceeds PEL ${ }^{2}$ | Exceeds LEL ${ }^{3}$ Below SEL | Exceeds SEL ${ }^{4}$ |
| Chromium | 12 | 0 | 28 | 0 |
| Zinc | 10 | 1 | 10 | 1 |
| Lead | 2 | 1 | 3 | 0 |
| Nickel |  |  | 40 | 0 |
| Manganese |  |  | 33 | 22 |
| Iron |  |  | 30 | 5 |
| Copper | 17 | 3 | 52 | 4 |
| Cadmium | 7 | 0 | 7 | 0 |
| Arsenic | 5 | 1 | 5 | 0 |


| Mercury | 1 | 0 | 0 | 0 |
| :---: | :---: | :---: | :---: | :---: |
| B. Organochlorines |  |  |  |  |
| pp DDD |  |  | 1 |  |
| Total DDE | 0 | 2 |  |  |
| Total DDD | 0 | 1 |  |  |
| Total DDT | 0 | 2 | 2 | 0 |
| Total DDT metabolites |  |  | 1 | 0 |
| PCB Aroclor 1254 | 1 |  | 1 | 0 |
| PCB Aroclor 1260 |  |  | 1 | 0 |
| Total PCB | 1 | 0 | 1 | 0 |


| C. Polycyclic Aromatic Hydrocarbons | Exceeds TEL ${ }^{1}$ Below PEL | Exceeds PEL ${ }^{2}$ | $\begin{array}{\|c} \hline \text { Exceeds LEL } \\ \text { Below SEL } \end{array}$ | Exceeds SEL ${ }^{4}$ |
| :---: | :---: | :---: | :---: | :---: |
| Naphthalene | 1 | 0 |  |  |
| Acenaphthylene | 8 | 0 |  |  |
| Acenaphthene | 6 | 1 |  |  |
| Fluorene | 5 | 2 | 2 |  |
| Phenanthrene | 14 | 2 | 2 |  |
| Anthracene | 7 | 0 | 0 |  |
| Fluoranthene | 12 | 1 | 4 |  |
| Pyrene | 14 | 1 | 4 |  |
| Benz(a)anthracene | 11 | 2 | 3 |  |
| Chrysene | 11 | 1 |  |  |
| Benzo k fluoranthene | 1 | 0 |  |  |
| Benzo(a)pyrene | 13 | 0 | 1 |  |
| Indeno(1,2,3-cd)pyrene |  |  | 5 |  |
| Dibenzo(a,h)anthracene | 5 | 0 | 0 |  |

Notes: 1 Federal Threshold Effect Level
2 Federal Probable Effect Level
3 Provincial Lowest Effect Level
4 Provincial Severe Effect Level

### 4.0 Discussion

### 4.1 DDT and Metabolites

DDT (dichlorodiphenyltrichloroethane) is a chlorinated hydrocarbon that has broad-spectrum pesticide properties. It was used in large quantities in the 1950s and 1960s on crops. The U.S. banned the use of DDT in 1973. The use of DDT in Canada was severely restricted in the early 1970s and discontinued in 1985, with the sale and use of existing stocks permitted until the end of 1990 (CCME, 2001). DDT is still used as an insecticide in other countries.

DDT has two metabolites: DDE (dichlorodiphenyldichloroethylene) and DDD (dichlorodiphenyldichloroethane). Each DDT molecule has several isomeric forms, depending on the configurations of the chlorine atoms on the molecule. For comparison with sediment quality guidelines, the laboratory results were analyzed according to the following:

Total DDT $=$ o-p'- plus $p-$ p $^{\prime}$ DDT

$$
\begin{gathered}
\text { Total DDE }=\text { o-p'- plus p-p'DDE } \\
\text { Total DDD }=\text { o-p'- plus p-p'DDD } \\
\text { Total DDT and metabolites }=\text { Total DDT }+ \text { Total DDE }+ \text { Total DDD }
\end{gathered}
$$

DDT, including its metabolites, was the only detected organochlorine compound in the current study with only 2 of the samples reaching detectable levels of one or more isomer of DDT or its metabolites.
Total DDD exceeded sediment quality guidelines (PEL; . $00851 \mu \mathrm{~g} / \mathrm{g}$ ) in one (1) tributary while total DDE and total DDT exceeded the PEL ( $.00675 \mu \mathrm{~g} / \mathrm{g}$ and $0.00477 \mu \mathrm{~g} / \mathrm{g}$ respectively) in two (2) tributaries.

| Tributary | Total DDE <br> $(\mu \mathrm{g} / \mathbf{g})$ | Total DDD <br> $(\mu \mathrm{g} / \mathbf{g})$ | Total DDT <br> $(\mu \mathrm{g} / \mathbf{g})$ |
| :--- | ---: | ---: | ---: |
| Eagle River $(\mathrm{Am})$ | 0.012 | $<0.002$ | 0.02 |
| Silver River $(\mathrm{Am})$ | 0.028 | 0.01 | 0.017 |

### 4.2 PCBs

Polychlorinated biphenyls, or PCBs, were commonly used in electrical equipment such as transformers and capacitors due to their chemical stability. The manufacture of PCBs was halted in 1977 in the United States. PCBs were not produced in Canada but approximately 40000 t of PCBS were imported and used commercially prior to the 1980s. Like many other organochlorine compounds, PCBs are persistent, bioaccumulative and toxic. They are the cause of the majority of the fish consumption advisories in each of the Great Lakes and they are considered a priority pollutant by many authorities. The Great Lakes Water Quality Agreement calls for the virtual elimination of discharges of PCBs.

In the current study, PCBs were analyzed in the laboratory as Aroclors, the trade name that describes the complex mixture of PCB congeners under which some PCBs were manufactured. This method is much less expensive than the more elaborate congener analyses although it is also less precise and less accurate. The detection limit for Aroclor analysis was $10 \mathrm{ng} / \mathrm{g}$, which is below the sediment quality guidelines and sufficient for the purposes of detecting PCBs in sediments.

Only two of the seven PCB Aroclors were detected in this study. Aroclor 1254 and Aroclor 1260 were both detected at two sites giving rise to only two sites with total PCB values. Only one site, the Fort Creek (Can) exceeded the TEL, $34.1 \mathrm{ng} / \mathrm{g}$, with a value of $140 \mathrm{ng} / \mathrm{g}$, well below the PEL of $277 \mathrm{ng} / \mathrm{g}$.

### 4.3 PAHs

Polycyclic aromatic hydrocarbons are produced during the incomplete combustion of organic substances, most commonly the combustion of fossil fuels. As an indicator of human industrial activities, PAH contamination is relatively widespread.

PAHs were commonly detected in the current survey. One or more PAH compounds were found in $100 \%$ of the tributaries. Total PAHs exceeded the LEL of $4,000 \mu \mathrm{~g} / \mathrm{kg}$ three (3) times. A listing is provided below of the ten (10) tributaries with concentrations of total PAH (i.e., the sum of the 16 PAH compounds investigated here) greater than $1,000 \mu \mathrm{~g} / \mathrm{kg}$.

| Tributary | Total PAH concentration <br> $(\mu \mathbf{g} / \mathbf{k g})$ |
| :--- | :---: |
| Eagle River (Am) | 1,460 |
| Kaministiquia River (Can) | 1,572 |
| McIntyre River (Can) | 2,050 |
| Davignon Creek (Can) | 2,210 |
| Knowlton Creek (Am) | 2,620 |
| Fort Creek (Can) | 2,880 |
| McVicar Creek (Can) | 3,925 |
| Bluff Creek (Am) | 4,400 |
| Firesteel River (Am) | 4,864 |
| Big Trout Creek (Can) | 21,760 |

### 4.4 Metals

### 4.4.1 Arsenic

Arsenic (As) is a metalloid and a nonessential trace element. Its release from anthropogenic sources is mainly from gold and base metal production facilities, with smaller releases from the use of arsenical pesticides, wood preservatives, coal-fired power generation and disposal of domestic and industrial wastes (Environment Canada, 1993).

In the current study, arsenic was found to exceed sediment quality criteria infrequently. Of the 108 unique sites, concentrations were above the federal TEL at 5 sites, and above the PEL at one. (see Table 5). At some of these sites, the exceedences may be due to naturally elevated As levels. In the National Geochemical Reconnaissance (NGR) program of the Geological Survey of Canada, the mean concentration of As in stream sediments was determined to be $10.7 \mu \mathrm{~g} / \mathrm{g}$ (P.W.B. Friske, 1996 in CCME 2001), which is greater than the federal TEL of $5.9 \mu \mathrm{~g} / \mathrm{g}$.

### 4.4.2 Cadmium

Cadmium (Cd) is a non-essential trace element that is produced commercially from base-metal smelters and refineries especially zinc refining. It is used in batteries, coatings, pigments, stabilizers and alloys (Hoskin, 1991 in Environment Canada, 1994a). Natural, background levels of Cd may be greater than the TEL, as the NGR program determined the mean concentration of Cd in stream sediments to be $0.63 \mu \mathrm{~g} / \mathrm{g}$ (P.W.B. Friske, 1996 in CCME 2001) as compared with the federal TEL of $0.6 \mu \mathrm{~g} / \mathrm{g}$. In an assessment of the NGR data, Painter et al. (1994) found that $95 \%$ of the data were below $1.3 \mu \mathrm{~g} / \mathrm{g}$.

Cadmium concentrations in the current study were generally below the federal TEL ( $0.6 \mu \mathrm{~g} / \mathrm{g}$ ) with only 7 (seven) sites exceeding the TEL. All sites were below the federal PEL ( $3.5 \mu \mathrm{~g} / \mathrm{g}$ )

### 4.4.3 Chromium

Chromium (Cr) is an essential trace element that can be toxic to organisms at elevated levels (CCME 2001). It is not mined in Canada, but its import contributes to the production of pigments, metal finishing, leather tanning and wood preservatives (Nriagu 1988 in Environment Canada 1994b).

Chromium was found at concentrations above sediment quality guidelines at more than 25 per cent of the sites. At twelve sites chromium levels exceeded the federal TEL ( $37.3 \mu \mathrm{~g} / \mathrm{g}$ ). No sites exceeded the federal PEL ( $90 \mu \mathrm{~g} / \mathrm{g}$ ).

### 4.4.4 Copper

Copper $(\mathrm{Cu})$ is an essential trace element whose anthropogenic sources are mainly from mining and smelting operations. Naturally elevated Cu concentrations may contribute to the Cu content in streambed sediments. In an analysis of the NGR sediment database, Painter et al. (1994) found that $95 \%$ of Cu concentrations were below $76 \mu \mathrm{~g} / \mathrm{g}$. In the current study, 17 sites showed concentrations above the TEL of $35.7 \mu \mathrm{~g} / \mathrm{g}$ while an additional 3 sites exceeded the PEL of 197
$\mu \mathrm{g} / \mathrm{g}$. Each of the Rivers with PEL exceedances are located in the state of Michigan where large copper deposits are found. Mining for copper in the area began in 1845 but declined after 1870. A listing of the tributaries with PEL exceedances are listed below.

| Tributary | $\mathbf{C u}(\mathbf{\mu g} / \mathbf{g})$ |
| :--- | ---: |
| Little Cranberry River (Am) | 320 |
| Mineral River | 464 |
| Eagle River | 1200 |

### 4.4.5 Mercury

Mercury $(\mathrm{Hg})$ is a nonessential trace element that is toxic, persistent and bioaccumulative. Fish consumption advisories are in effect for mercury in much of the Great Lakes ecosystem. Current uses of mercury include some batteries, dental fillings, thermometers and switches, cathode tubes and household cleaners. Sources of mercury to the environment include mining and smelting, wastewater, fossil fuel combustion and waste incineration.

Sediment from most tributaries contained relatively low concentrations of mercury. Only one site exceeded any federal guidelines; the Current River in Thunder Bay, Ontario exceeding the LEL ( $170 \mathrm{ng} / \mathrm{g}$ ). Local, natural mercury deposits can impact environmental concentrations. The $95^{\text {th }}$ percentile for mercury in the NGR database was determined to be $190 \mathrm{ng} / \mathrm{g}$ (Painter et al., 1994). Levels above this are therefore unlikely to be of natural origin.

### 4.4.6 Nickel

Nickel $(\mathrm{Ni})$ is a trace element whose primary anthropogenic sources include primary base metal production and fossil fuel combustion (Environment Canada, 1994c). There is no federal sediment quality guideline for Ni , therefore sediment concentrations were compared with the provincial guidelines. The Lowest Effect Level (LEL) of $16 \mu \mathrm{~g} / \mathrm{g}$ was exceeded at 40 sites. However, exceedences of the LEL can occur naturally. Analysis of the NGR database of stream and lake sediment metals concentrations showed that the $95^{\text {th }}$ percentile for Ni concentration was $60 \mu \mathrm{~g} / \mathrm{g}$ (Painter et al., 1994). Levels greater than this are more likely to indicate anthropogenic impacts. In the current study, the Severe Effect Guideline (SEL) of $75 \mu \mathrm{~g} / \mathrm{g}$ was not exceeded.

### 4.4.7 Lead

Lead $(\mathrm{Pb})$ is a nonessential trace element. Its past use as an additive in gasoline has resulted in its widespread distribution in the environment above background levels. Currently, sources of lead to the environment include lead processing activities, batteries, and industrial and municipal effluents. Lead concentrations exceeded the federal TEL of $35 \mu \mathrm{~g} / \mathrm{g}$ at 2 sites and the PEL of $91.3 \mu \mathrm{~g} / \mathrm{g}$ at one site. The $95^{\text {th }}$ percentile of stream and lake sediment Pb concentration in the NGR database was $25 \mu \mathrm{~g} / \mathrm{g}$ (Painter et al., 1994), therefore even TEL exceedences are likely to be due to anthropogenic influences. The guideline exceedences are listed below.

| Tributary | Pb ( $\mu \mathrm{g} / \mathbf{g})$ |
| :--- | :---: |
| McVicar Creek (Can) | 36 |
| Fort Creek (Can) | 54 |
| Little Cranberry River (Am) | 137 |

### 4.4.8 Zinc

$\operatorname{Zinc}(\mathrm{Zn})$ is an essential trace element that is considered toxic to aquatic biota at elevated concentrations (CCME, 2001). Anthropogenic zinc sources are primarily related to metals processing, with smaller releases from fossil fuel burning and ancillary sources such as fertilizers, rubber goods and pharmaceuticals.

In the current study, the federal TEL of $124 \mu \mathrm{~g} / \mathrm{g}$ was exceeded at 10 sites. There was one PEL $(315 \mu \mathrm{~g} / \mathrm{g})$ exceedences. The $95^{\text {th }}$ percentile zinc sediment concentration in the NGR database was $191 \mu \mathrm{~g} / \mathrm{g}$; therefore TEL exceedences may not be due to anthropogenic sources. The five highest concentrations of zinc for the Lake Superior Tributaries are listed below.

| Tributary | Zn $(\mu \mathrm{g} / \mathbf{g})$ |
| :--- | :---: |
| Current River (Can) | 167 |
| Fort Creek (Can) | 193 |
| Blackbird Creek (Can) | 201 |
| Bennet Creek (Can) | 247 |
| Little Cranberry River (Am) | 1790 |

### 4.4.9 Manganese and Iron

Concentrations of the essential metals manganese and iron were compared with provincial sediment quality guidelines as no federal guidelines are available. Manganese concentrations exceeded the LEL of $460 \mu \mathrm{~g} / \mathrm{g}$ at 33 sites and the SEL of $1100 \mu \mathrm{~g} / \mathrm{g}$ at a further 22 sites. The tributaries with concentrations greater than $2000 \mu \mathrm{~g} / \mathrm{g}$ are listed below. The exceedences generally did not appear to be related to industrial impacts in the majority of cases. Mn exceedences also occurred in relatively "clean" tributaries that were expected to represent background or unimpacted conditions. The median Mn concentration in the Ontario Geological Survey stream sediment database (Fortescue, 1984) is calculated to be $850 \mu \mathrm{~g} / \mathrm{g}$, and the $95^{\text {th }}$ percentile of concentrations was $2150 \mu \mathrm{~g} / \mathrm{g}$. It might therefore be interpreted that stream sediment concentrations in this range may be attributed to natural sources. Indeed, the application of the Ontario Sediment Quality Guidelines should take the background levels of metals into account prior to any management action (Persaud et al., 1992).

| Tributary | Mn ( $\mu \mathbf{g} / \mathbf{g}$ ) |
| :--- | :---: |
| Sand River | 2070 |
| Anglers Creek (Can) | 2100 |
| Beaver River (Am) | 2310 |
| Carp River (Am) | 2900 |
| Grand Marais Creek (Am) | 3490 |
| Little Goulais River (Can) | 3580 |
| Michipicoten River (Can) | 3710 |
| Brule River (Am) | 4310 |
| Magpie River (Can) | 4450 |
| Little Carp Creek (Can) | 4960 |
| Current River (Can) | 32500 |

For iron ( Fe ), background levels may also be high due to natural sources. The Ontario Geological Survey stream sediment data (Fortescue, 1984) shows that the median Fe concentration is 3.1\% and that the $95^{\text {th }}$ percentile of Fe concentrations is $5.5 \%$. These values are comparable to the LEL of $2 \%$ and the SEL of $4 \%$. In the current study, 30 sites exceeded the LEL and 5 sites (listed below) exceeded the SEL. Several of these included sites at which contamination from anthropogenic sources would not be expected. Similar to Mn, the natural or background concentration of Fe would need to be determined in order to interpret sediment quality guideline exceedences.

| Tributary | Fe (\%) |
| :--- | :---: |
| Pigeon River (Can) | 4.24 |
| Brule River (Am) | 4.31 |
| McIntyre River (Can) | 5.09 |
| Current River (Can) | 5.5 |
| Little Goulais River (Can) | 6.39 |

### 5.0 Next Steps

This sampling represents the first stage of a track-down program to identify potential sources of contamination to the upper Great Lakes that are not being addressed by other Great Lakes programs. The program constitutes a portion of Environment Canada's commitment towards the Great Lakes Water Quality Agreement (GLWQA).

By committing to the track-down program, the federal and provincial partners have agreed to conduct follow-up work at locations where ambient data indicate potentially significant sources of persistent, bioaccumulative and toxic substances ( PBTs ) may exist. The program has, to date, focused on potential PCB sources. Three pilot projects have been initiated/conducted in Lake Ontario tributaries and another two in Lake Erie tributaries where PCB contamination is suspected based on available ambient information. Based on the experiences in these projects, the project partners have developed a decision framework guide for potential future track-down projects; in particular, it provides guidelines for the initiation and termination of such projects
and provides recommendations with respect to appropriate project design and sampling methodologies.

The parties have determined that potential projects must be prioritized based on the available information. The degree of contamination is determined for various media, and a prioritization is then made. These recommendations will require full disclosure and the sharing of ambient information between the project partners. To that end, steps have already been taken to ensure that information is freely shared in a manner that permits a broad prioritization based on the most current and reliable information.

By virtue of this document the information from the current study is being shared with other environmental authorities and partners in Ontario.

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Lake Superior Tributary Data:

| Tributary | Latitude | Longitude | Sampling Date | HCB | Endrin aldehyde | OCS | Toxaphene | Aldrin | $\alpha \mathrm{HCH}$ | $\beta \mathrm{HCH}$ | $\overline{\mathrm{HCH}}$ | Lindane |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units |  |  |  | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | ug/g | $\mu \mathrm{g} / \mathrm{g}$ |
| Agawa River | 47.3575 | -84.6318 | 31-Aug-05 | $<0.002$ | <0.002 | <0.002 | $<0.08$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Aguasabon River | 48.7728 | -87.1164 | 19-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Alona Bay Creek | 47.1637 | -84.6913 | 31-Aug-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Anglers Creek | 48.776 | -86.4073 | 15-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Barrett River | 47.4049 | -84.7024 | 30-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Batchawan River | 46.9351 | -84.5274 | 14-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Bennett Creek | 46.5234 | -84.3863 | 13-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Big Carp River | 46.516 | -84.4652 | 13-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Big Trout Creek | 48.9463 | -88.2616 | 17-Aug-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Blackbird Creek | 48.8508 | -87.02 | 15-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Black Sturgeon River | 48.9043 | -88.3784 | 17-Aug-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Carp River | 46.9517 | -84.5755 | 14-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Chippewa River | 46.9278 | -84.4265 | 31-Aug-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Coldwater River | 48.804 | -88.5401 | 17-Aug-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Cold Water River East | 47.4715 | -84.788 | 30-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Cranberry Creek | 46.69 | -84.395 | 13-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Current River | 48.4545 | -89.1876 | 15-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Cypress River | 48.9 | -87.86 | 18-Aug-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Davignon Creek | 6.5 | -84.3619 | 15-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Dead Horse Creek | 48.8173 | -86.6866 | 20-Aug-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Diversion Channe | 46.5219 | -84.4091 | 13-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Fort Creek | 46.515 | -84.3436 | 15-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Goulais River | 46.7235 | -84.3823 | 01-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Government Creek | 46.869 | -84.3539 | 01-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Gra | 48.9 | -87.768 | 18-Aug-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Harmony River | 46.8461 | -84.3706 | 01-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Havilland Creek | 46.818 | -84.4248 | 01-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Jackpine River | 48.9718 | -87.9997 | 18-Aug-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Kaministiquia River | 48.3591 | -89.2882 | 16-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Kelly Creek | 46.6698 | -84.4501 | 13-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Little Cypress River | 48.9 | -87.8 | 18-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.00 | <0.0 |


| Tributary | a Chlordane | y Chlordane | Total Chlordane | op DDD | pp DDD | op DDE | pp DDE | op DDT | pp DDT | Total DDD | Total DDE | Total DDT |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ |
| Agawa River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Aguasabon River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Alona Bay Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Anglers Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Barrett River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Batchawan Rive | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Bennett Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Big Carp River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Big Trout Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Blackbird Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Black Sturgeon River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Carp River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Chippewa River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Coldwater River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Cold Water River East | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Cranberry Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Current River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Cypress River | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Davignon Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Dead Horse Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Diversion Chan | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Fort Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Goulais River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Government Creek | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Gravel River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Harmony River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Havilland Creek | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Jackpine River | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Kaministiquia River | <0.002 | <0.002 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Kelly Creek | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Little Cypress River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |


| Tributary | Total DDT <br> Metabolites | Dieldrin | Endosulfan I | Endosulfan II | Endosulfan Sulfate | Endrin | Heptachlor | Heptachlor Epoxide | Methoxychlor | Mirex | Aroclor 1016 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ |  | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ |
| Agawa River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Aguasabon River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Alona Bay Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Anglers Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Barrett River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Batchawan River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Bennett Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Big Carp River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Big Trout Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | $<0.01$ |
| Blackbird Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Black Sturgeon River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Carp River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Chippewa River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | $<0.01$ |
| Coldwater River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Cold Water River East | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Cranberry Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Current River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Cypress River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | $<0.01$ |
| Davignon Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Dead Horse Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Diversion Channel | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Fort Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Goulais River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Government Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | $<0.01$ |
| Gravel River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Harmony River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Havilland Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Jackpine River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Kaministiquia River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Kelly Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Little Cypress River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |


|  | $\stackrel{8}{2}$ |  | ） | レ | v0 | ๑ | ¢ | ロ ¢ | $\mathfrak{V}$ ¢ | ¢ ${ }^{\text {® }}$ | จง | $\stackrel{\square}{v}$ | $\stackrel{\text { v }}{ }$ | レ | レ0 ข | ท | レ |  | レ0 | $\stackrel{\text { ® }}{ }$ | $\pm$ | V | レ0 | จ | V | จ | ง | ๑ |  | $\stackrel{10}{ }$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & \text { O } \\ & \text { 号 } \\ & 2 \end{aligned}$ | $\stackrel{\text { ® }}{ }$ | レ๐ | レ | V | ¢ | $\stackrel{\text { ® }}{ }$ | จง ง | ๑๐ท | จ ง ก | ง ¢ | ๒ | $\stackrel{\text { v }}{ }$ | レ | セ8 ロ | จู | ง | 안 | $\stackrel{1}{v}$ | V | $\stackrel{\sim}{\sim}$ | $\bigcirc$ | V | จ | レ |  | $\bigcirc$ | ท |  | $\stackrel{\sim}{0}$ |
| $\begin{aligned} & 00 \\ & 0 \\ & \stackrel{0}{0} \\ & \stackrel{\square}{0} \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \end{aligned}$ | ¢ | $\stackrel{\rightharpoonup}{\circ}$ | ¢ | ¢ | ¢ | ¢ | ¢ |  |  | $\stackrel{\rightharpoonup}{\text { V }}$ | $\stackrel{\square}{\text { ¢ }}$ | ¢ | $\begin{gathered} \overline{0} \\ \dot{v} \\ \text { in } \end{gathered}$ | $\begin{array}{ll} \bar{o} & \overline{0} \\ 0 \\ 0 & 0 \\ \hline \end{array}$ | ¢ | $\begin{gathered} \overline{0} \\ \dot{\rightharpoonup} \\ \text { v } \end{gathered}$ |  | ¢ |  | $\underset{i}{\dot{O}}$ | $\bar{\circ}$ | O－ | $\begin{gathered} \overline{0} \\ \dot{v} \end{gathered}$ | ¢ |  | ¢ | $\stackrel{\bigcirc}{\circ}$ |  |  |
| $\begin{aligned} & \stackrel{\circ}{0} \\ & \stackrel{y}{0} \\ & \stackrel{\vdots}{\circ} \\ & \stackrel{\circ}{4} \end{aligned}$ | $\begin{gathered} \text { 옹 } \\ \hline ㅁ \end{gathered}$ | ¢ | $\stackrel{\rightharpoonup}{\dot{\circ}}$ | ¢ | V＇． | $\stackrel{\rightharpoonup}{\circ}$ | $\stackrel{\bar{\circ}}{\text { V }}$ |  |  |  |  |  |  | $\begin{gathered} \overline{0} \\ \dot{v} \end{gathered}$ | $\begin{array}{ll} \bar{o} & \overline{0} \\ 0 \\ \dot{v} & 0 \\ \hline \end{array}$ |  | $\begin{gathered} \overline{0} \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ |  | ¢ | ¢ |  | － |  | $\begin{gathered} \bar{i} \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ | ¢ |  | $\stackrel{\bar{o}}{\substack{\text { b }}}$ | $\bar{\circ}$ |  | ¢ |
|  | $\begin{gathered} \frac{0}{2} \\ \frac{\mathrm{D}}{2} \end{gathered}$ | ¢ | $\stackrel{\rightharpoonup}{\circ}$ | $\begin{aligned} & \bar{o} \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & \bar{o} \\ & \dot{O} \\ & \hline \end{aligned}$ |  | $\begin{gathered} \bar{i} \\ \dot{\rightharpoonup} \end{gathered}$ | $\begin{gathered} \bar{j} \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ |  | vo |  | $\stackrel{\square}{\text { V }}$ |  |  | $\begin{array}{ll} \bar{y} & \overline{0} \\ 0 \\ 0 & 0 \\ \hline \end{array}$ | $\stackrel{\text { ¢ }}{\text { ¢ }}$ | $\stackrel{\square}{\mathrm{O}}$ |  | ¢ | ¢ |  | $\overline{5}$ | $\stackrel{\square}{\text { v }}$ | $\begin{gathered} \overline{0} \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ | $v$ |  | v | O |  |  |
|  | $\begin{aligned} & 0 \\ & \frac{0}{2} \\ & \hline 1 \end{aligned}$ | ¢ | $\begin{gathered} \bar{O}_{0} \\ \dot{v} \end{gathered}$ | $\begin{aligned} & \bar{o} \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \bar{o} \\ & \dot{0} \end{aligned}$ |  | $\begin{gathered} \bar{i} \\ \dot{v} \end{gathered}$ | $\begin{gathered} \overline{0} \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ | $\begin{aligned} & \bar{o} \\ & \stackrel{\rightharpoonup}{0} \\ & \stackrel{\rightharpoonup}{0} \\ & 0 \end{aligned}$ | $\begin{gathered} \bar{o} \\ \stackrel{\rightharpoonup}{v} \\ \mathrm{o} \\ \hline \end{gathered}$ |  |  | $\begin{aligned} & 5 . \\ & \vdots \\ & \hline \end{aligned}$ | $\begin{aligned} & \overline{0} \\ & 0 . \\ & 0 . \end{aligned}$ | $\begin{array}{l\|l} \bar{o} \\ \stackrel{\rightharpoonup}{0} \\ \stackrel{\rightharpoonup}{0} \\ \hline \end{array}$ | $\stackrel{\bar{O}}{\square}$ |  |  | ¢ |  |  | $\stackrel{\bar{c}}{\text { V }}$ | $\stackrel{\bar{\rightharpoonup}}{\stackrel{\rightharpoonup}{v}}$ | $\begin{gathered} \bar{b} \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ | ¢ |  | $\stackrel{\bar{c}}{\text { V }}$ |  |  |  |
|  | $\begin{aligned} & 0 \\ & \frac{0}{0} \end{aligned}$ | No | N | N | $\begin{gathered} \text { N } \\ \text { ó } \\ \text { in } \end{gathered}$ |  | $\begin{gathered} \text { o } \\ \text { io } \\ \text { vin } \end{gathered}$ | $\begin{gathered} \text { N. } \\ \dot{\mathrm{V}} \end{gathered}$ |  |  |  |  | $\begin{aligned} & N \\ & \substack{N \\ 0 \\ i \\ 0 \\ \hline} \end{aligned}$ | $\begin{gathered} \text { N } \\ \text { O } \\ \text { ver } \end{gathered}$ | No |  | $\begin{gathered} \text { N } \\ \text { V } \end{gathered}$ | $\begin{gathered} \text { No } \\ \text { O } \\ \dot{0} \end{gathered}$ | No． | $\underset{\text { No }}{\substack{\text { O}}}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { O } \end{aligned}$ | No | N | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { in } \end{aligned}$ | N | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | v | N |  |  |
| $\begin{aligned} & \text { N } \\ & \text { N } \\ & \text { 흘 } \\ & \text { O } \end{aligned}$ | $\begin{aligned} & 0 \\ & \frac{0}{2} \end{aligned}$ | ¢ | $\stackrel{\rightharpoonup}{\stackrel{\rightharpoonup}{*}}$ | ¢ | ¢ | ¢ | $\stackrel{\bar{c}}{\text { v }}$ | ¢ |  | vo |  | V | － | $\begin{gathered} \bar{o} \\ \dot{0} \\ \text {. } \end{gathered}$ |  | ¢ |  | ¢ | $\stackrel{\square}{\dot{\circ}}$ | $\stackrel{\bar{C}}{\text { v }}$ |  | ¢ | v | $\stackrel{\square}{\square}$ | v |  |  | V |  |  |
| $\begin{aligned} & \bar{N} \\ & \underset{y}{\prime} \\ & \dot{\bar{O}} \\ & \stackrel{i}{4} \end{aligned}$ |  | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\begin{gathered} \text { on } \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | O | $\begin{gathered} \text { O} \\ \stackrel{\rightharpoonup}{\dot{V}} \end{gathered}$ | $\begin{gathered} 0 \\ 0 \\ \dot{v} \end{gathered}$ | $\begin{aligned} & 0 \\ & 0 \\ & \dot{v} \end{aligned}$ |  | $\begin{gathered} \infty \\ 0 \\ \dot{O} \\ \dot{0} \\ 0 \end{gathered}$ | vor |  |  | $\begin{gathered} \text { M} \\ \dot{v} \\ \hline \end{gathered}$ | $\begin{array}{cc} 0 & 0 \\ 0 & 0 \\ 0 & 0 \\ 0 & 0 \end{array}$ | $\stackrel{\text { O}}{\substack{\text { ® } \\ \text { V }}}$ |  | $\begin{gathered} 0 \\ 0 \\ 0 \\ 0 \\ 0 \\ 0 \end{gathered}$ | ¢ |  |  | $\stackrel{0}{\circ}$ |  | $\stackrel{\text { O}}{\stackrel{\text { O}}{\mathrm{V}}}$ | $\stackrel{0}{\circ}$ |  | v | v |  |  |
|  | $\stackrel{0}{5}$ |  |  |  | $\begin{aligned} & \text { y } \\ & 0 \\ & 0 \\ & 0 \\ & \frac{9}{0} \\ & 0 \\ & \frac{c}{4} \end{aligned}$ |  |  |  |  |  |  |  |  | $\overline{9}$ $\frac{y}{y}$ $\bar{d}$ 0 0 0 0 |  |  | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ |  |  | Diversion Channel | $\begin{aligned} & \stackrel{\rightharpoonup}{0} \\ & \stackrel{0}{0} \\ & \stackrel{\rightharpoonup}{\circ} \\ & \stackrel{\rightharpoonup}{\circ} \\ & \hline \end{aligned}$ |  |  | $\left\{\begin{array}{l} \frac{1}{x} \\ \frac{9}{x} \\ \frac{0}{2} \\ \frac{0}{0} \end{array}\right.$ |  |  |  |  |  |  |




| Q | \％ | $\stackrel{1}{0}$ | V | $\infty$ | $\infty \infty$ | $\infty$ | ト | $\stackrel{\text { ® }}{ }$ | $\infty$ |  | $\stackrel{\circ}{0}$ |  | $\bigcirc 0$ | $\infty$ | $\stackrel{ }{\sim}$ | N | ₹ へ | F | の罗 |  | $\infty$ | $\bigcirc$ | 入 | ๑๐ | 안 | $\bar{m}$ | F |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| z | 응 | $\stackrel{\sim}{\circ}$ | $\stackrel{m}{\square}$ | 은 | $\bigcirc$ | 안 | $\stackrel{ }{\sim}$ | 入 | N | ₹ | $\stackrel{\text { N }}{ } \stackrel{1}{\square}$ | $\pm$ | N | N | $\stackrel{\square}{\square}$ | ल | $\stackrel{\varrho}{\circ}$ | N | F® | $\stackrel{\square}{\square}$ | F | 入 | 「 | の | $\stackrel{9}{ }$ | ल | N |
| z | $\bigcirc$ | $0$ | $\begin{gathered} N \\ \vdots \\ 0 \end{gathered}$ | $\begin{array}{lc} \bar{o} \\ 0 \\ 0 & 0 \end{array}$ | No | $\begin{aligned} & N \\ & \hline \end{aligned}$ | $\begin{aligned} & \hat{N} \\ & 0 \end{aligned}$ | $\begin{aligned} & \infty \\ & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\begin{array}{ll} 0 & 0 \\ 0 & 0 \end{array}$ |  | $0$ | $\begin{array}{ll} 5 & 0 \\ 0 \\ 0 & 0 \\ \hline \end{array}$ | $\begin{gathered} \pm \\ \hline \\ \hline \end{gathered}$ |  |  |  |  |  | $\begin{array}{ll} \infty \\ \vdots \\ 0 \\ 0 \\ \hline 0 \\ \hline \end{array}$ | $\begin{gathered} \infty \\ 0 \\ 0 \\ \hline \end{gathered}$ | $\begin{aligned} & \infty \\ & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $0$ | $0^{\circ}$ | $\begin{aligned} & \hat{o} \\ & \mathbf{o} \end{aligned}$ | ${ }_{0}^{0}$ |  | $\bigcirc$ |
| $\sum^{\circ}$ | \％ | v | v | $\bar{v}$ | v | v | $\bar{v}$ | v | $\overline{\mathrm{v}}$ | $\overline{\mathrm{v}}$ | $\overline{\mathrm{v}}$ | $\overline{\mathrm{v}}$ V | v | $\bar{v}$ | v | $\overline{\mathrm{v}}$ | $\overline{\mathrm{v}}$ v | $\bar{v}$ | v | $\bar{v}$ | $\overline{\mathrm{v}}$ | v | $\overline{\mathrm{v}}$ | $\overline{\mathrm{v}}$ | v | $\bar{v}$ | $\overline{\text { v }}$ |
| $\Sigma$ | 웅 | \& | bo | $\stackrel{m}{\Gamma}$ | $\stackrel{\circ}{\mathrm{N}} \stackrel{\text { n }}{\stackrel{1}{c}}$ | $\underset{\sim}{2}$ | $\stackrel{N}{f}$ | $\stackrel{\stackrel{\rightharpoonup}{0}}{\sim}$ | No | $\hat{\rho}_{\infty}^{\infty}$ | $\stackrel{\infty}{\infty}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{4}{4}$ | $\stackrel{\otimes}{\dot{q}}$ | $\stackrel{\stackrel{\sim}{N}}{ }$ | $0 \text { Bio }$ | 앙 응 |  | $\stackrel{ \pm}{\lambda} \stackrel{( }{\Gamma}$ | $\hat{m}_{n}$ | $\underset{\sim}{\underset{\sim}{2}}$ | $\bar{m}$ | $0_{i}^{\infty}$ | $\underset{N}{N}$ | $\begin{aligned} & \stackrel{\rightharpoonup}{\mathrm{y}} \end{aligned}$ | \％ | － |
| $\sum^{5}$ | ${ }^{50}$ | $0$ | $\hat{j} \hat{o}$ | $\begin{gathered} \text { N } \\ \text { Ó } \end{gathered}$ | $\underset{\sim}{\text { N }}$ | $\underset{\substack{4 \\ \hline \\ \hline}}{ }$ | $\stackrel{\stackrel{N}{0}}{\circ}$ | $\underset{0}{\underset{\sim}{N}}$ | $\underset{\sim}{\underset{\sim}{\underset{c}{c}} \underset{\sim}{c}}$ | $\stackrel{\infty}{\infty} \stackrel{\infty}{\infty}$ | － |  |  | $\overline{\mathrm{N}}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\stackrel{\substack{0 \\ 0}}{ }$ |  | $j{ }_{j}^{0}$ | $\underset{\substack{N}}{\underset{0}{N}}$ | $\bigcirc$ | \％ |  | $\stackrel{N}{\mathrm{~N}}$ | $\hat{N}$ | 단 |  | $\stackrel{8}{\square}$ |
| צ | $\bigcirc$ | $\begin{aligned} & \text { O } \\ & \mathbf{O} \end{aligned}$ | $\mathfrak{i}$ | O- | O | ${ }_{0}^{0}$ |  | $\begin{aligned} & \text { O } \\ & \text { O. } \end{aligned}$ | $\underset{0}{\pi}$ |  | $0^{\circ}$ |  | $\underset{0}{4}$ | t | $\stackrel{\circ}{0}$ | $\begin{aligned} & 3 \\ & \hline 8 \\ & \hline \end{aligned}$ | $\stackrel{\circ}{0}$ |  | $\underset{O}{\stackrel{N}{O}} \underset{\substack{0}}{ }$ | 人 | $\stackrel{Q}{0}_{0}^{0}$ | $8$ | B | $0$ | $\overline{0}$ |  | $\stackrel{\text { d }}{\text { N }}$ |
| $\stackrel{\leftrightarrow}{\llcorner }$ | $\bigcirc$ | $\stackrel{\rightharpoonup}{\mathrm{o}}$ | $\stackrel{?}{P}$ |  | $\stackrel{\sim}{\mathrm{N}}$ | $\underset{\sim}{\underset{\sim}{\sim}}$ |  | $\stackrel{8}{\mathrm{O}}$ | $\stackrel{8}{\square}$ | $\stackrel{\circ}{\circ}$ | $\stackrel{\aleph}{\Gamma} \underset{\sim}{\underset{\sim}{m}}$ | $\stackrel{m}{c} \stackrel{\underset{\sim}{r}}{\dot{r}}$ | $\stackrel{+}{\text { ¢ }}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{+}{\square}$ | ใ0 | $\stackrel{\text { f }}{\stackrel{8}{8}}$ | 「 | $\stackrel{\leftrightarrow}{\circ}$ | 「 | N |  | $\stackrel{\infty}{+}$ | $\dot{\circ}$ | $\stackrel{\text { ल }}{\text { N }}$ | $\stackrel{\text { m }}{\text { m }}$ |  |
| Ј | 잉 | $\stackrel{\infty}{\square}$ | 안 | $\bar{\sim}$ | $\stackrel{\text { ®の }}{\sim}$ | の | $\stackrel{\infty}{\sim}$ | ๑ | $\stackrel{\circ}{\sim}$ ？ | ® | N | $\stackrel{\text { m}}{ }$ | 당 | $\ulcorner$ | 아 | ¢ | ＋${ }_{\text {¢ }}$ | $\pm$ | m is | ＾ | $\stackrel{\sim}{\sim}$ | $\bigcirc$ | F | $\bigcirc$ | ¢ | \％ | $\stackrel{\sim}{\sim}$ |
| ̇ | 흥 | m | ผ | ฝ | $\stackrel{\text { ® }}{ }$ | $\stackrel{\infty}{\sim}$ | ผ๊ | $\stackrel{ }{\sim}$ | m | へ | ® | ৯ | $\stackrel{\infty}{\sim}$ | F | $\stackrel{\sim}{\sim}$ | ¢ | N へ | g | N | N | ミ | $\stackrel{6}{\square}$ | $\stackrel{\square}{-}$ | $\stackrel{\square}{\square}$ | m | \％ | \％ |
| $\bigcirc$ | 용 | の | 人 | $\infty \infty$ | $\infty$ | $\llcorner$ | $\bullet$ | $\bigcirc$ | F | ＋ | N | N | $\infty \stackrel{\text { N }}{ }$ | の | 入 | $\bar{m}$ | $\stackrel{\text { m }}{\square}$ | $\pm$ | $\bigcirc$－ | $\infty$ | 入 | m | $\wedge$ | $\llcorner$ | ¢ | $\stackrel{冂}{\square}$ | $\stackrel{\square}{-}$ |
| $\bigcirc$ | 흠 | $\hat{2}_{2}^{2} \stackrel{0}{2}$ | $\stackrel{n}{0}$ | $\stackrel{0}{v i}$ | $\begin{array}{lll} n & n \\ \stackrel{n}{0} \\ \stackrel{0}{v} \end{array}$ | $\stackrel{\sim}{\stackrel{\sim}{\circ}} \stackrel{\sim}{\bullet}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & 0 \\ & \stackrel{0}{0} \end{aligned}$ | $\stackrel{n}{\stackrel{n}{\mathrm{v}}} \stackrel{\mathrm{~m}}{\mathrm{r}}$ | $\stackrel{m}{\square} \stackrel{n}{\square}$ | $\circ \vee$ | $\stackrel{0}{v}$ | $\stackrel{\circ}{v}$ | $\stackrel{\rightharpoonup}{v}$ | $\stackrel{n}{n}_{\substack{0 \\ 0 \\ 0}}$ | $\underset{i}{i} \hat{i}$ | $\dot{c} \stackrel{n}{\circ}$ | $0$ | $\begin{array}{c\|c} n \\ \stackrel{n}{0} \\ \stackrel{\rightharpoonup}{0} \\ 0 \end{array}$ | $\dot{c}$ | $\dot{v}$ | $\stackrel{\rightharpoonup}{v}$ | io | $\begin{aligned} & \text { n } \\ & \stackrel{0}{v} \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \end{aligned}$ | v | $\stackrel{6}{\square}$ |
| ర็ | $\bigcirc$ | $\frac{9}{c}$ | $\underset{j}{2}$ |  | $\begin{array}{cc} \underset{\sim}{x} \\ \underset{\sim}{\circ} \\ \hline \end{array}$ |  | $\stackrel{\sim}{\sim}$ | $\underset{\sim}{\underset{0}{2}}$ | $\dot{c}$ | $\stackrel{\otimes}{\mathrm{q}}$ |  | － | $$ | $\stackrel{\sim}{0}$ | $\stackrel{\substack{\mathrm{o} \\ 0}}{ }$ | $\bigcirc$ |  | $\begin{array}{rl} 2 \\ j & 0 \\ j \\ \hline \end{array}$ | $\stackrel{\rightharpoonup}{\circ} \stackrel{m}{\square}$ | \％ | ¢ | － | $\bigcirc$ | $\begin{gathered} N \\ 0 \end{gathered}$ | え | \％ | $\stackrel{\square}{+}$ |
| ¢ | 흥 | $\stackrel{1}{0}$ | レ0 | V | V | レ | V | ทิ | v | V | ソ ท | ขロ | ท | V | ถ | ๒ | レセソ | ท | ท | ๒ | V | V | ท | ท | V | $\stackrel{1}{0}$ | $\stackrel{10}{ }$ |
| ® | $\frac{k}{0}$ | $\hat{y}_{2}^{2} \underset{\sim}{n}$ |  | $0$ |  | $\begin{gathered} \underset{\sim}{N} \\ \stackrel{\rightharpoonup}{v} \\ \stackrel{y}{*} \end{gathered}$ | $\begin{aligned} & \infty \\ & 0 \end{aligned}$ | $\begin{gathered} \mathrm{N} \\ \mathrm{~V} \end{gathered}$ | $\begin{array}{ll} 0 \\ 0 \\ 0 \end{array}$ | $\begin{array}{ll} \text { N } \\ \text { N } \\ \text { V } & 0 \end{array}$ | $\stackrel{N}{N}$ |  | $\underset{\sim}{c} \underset{\sim}{v}$ | No | $\stackrel{m}{0}$ | $\overbrace{0}^{\circ}$ | $$ | $\begin{aligned} & 0 \\ & 0 \end{aligned}$ | $\begin{array}{ll} \infty & 0 \\ 0 & 0 \\ 0 \end{array}$ | N | N | $0$ | $\stackrel{y}{v}$ | v | － | Ȯ | $\stackrel{\bullet}{\circ}$ |
| ๗ | 등 | $\stackrel{\sim}{\sim}$ | － | ค \＆ | 囚 フ | ¢ ハ | ¢ | กั | $\stackrel{\ominus}{\square}$ | N「 | ス | ¢ ¢ ¢ | ¢ | \％ | $\infty$ | $\stackrel{\text { 응 }}{\stackrel{0}{0}}$ | $\bigcirc \stackrel{\sim}{\square}$ | ¢ | ¢88 | 罗 | \＆ | ～ | ¢ | ₹ | $\stackrel{\text { ® }}{\sim}$ | 단 | N |
| 8 | 들 | ท | V | ค ¢ ¢ | เคู | レ゚ | ๒ | ท | ค | $\stackrel{\text { ® }}{ }$ | v | －¢ ¢ | v | V | V | v | レค ¢ | V | ทู้ | V | V | $\stackrel{\text { V }}{ }$ | ๒ | V | V | V | － |
| 『 | $\bigcirc$ | $\stackrel{n}{0}$ | $5$ |  |  | $\underset{\sim}{7}$ |  | $\dot{f}$ | ${ }_{0}^{\infty}$ | $\begin{gathered} \mathcal{F} \\ 0 \end{gathered}$ | － | $0$ | $\stackrel{8}{\circ}$ | $\begin{aligned} & \substack{\text { O} \\ \hline \\ \hline \\ \hline} \end{aligned}$ | $\bar{\square}$ |  |  |  | $\stackrel{n}{6}$ | ${ }_{0}^{\infty}$ | $\stackrel{\bigcirc}{\circ}$ |  | $0$ | $\begin{aligned} & \overline{5} \\ & 0 \end{aligned}$ | N | $\bigcirc$ | $\stackrel{\text { ¢ }}{\stackrel{-}{+}}$ |
| 8 | ¢ | $\begin{array}{\|r} \hline \dot{\bar{v}} \\ \hline \end{array}$ | $\dot{\rightharpoonup}$ | vi | $\begin{array}{l\|r} \bar{i} \\ \stackrel{\rightharpoonup}{v} \\ \hline \end{array}$ | $\dot{v}$ | $\begin{aligned} & \stackrel{r}{\dot{v}} \\ & \hline \end{aligned}$ | $\stackrel{r}{\dot{v}}$ | $\begin{array}{ll} \stackrel{\rightharpoonup}{\dot{c}} \\ \dot{v} \\ \hline \end{array}$ |  | $\begin{array}{r} 0 \\ 0.0 \\ 0 \end{array}$ | $\begin{array}{r} 0 \\ 0.0 \\ v \end{array}$ | $\underset{\sim}{v} \underset{\sim}{v}$ | $\stackrel{0}{v}$ | $0$ | \％ | $\dot{\stackrel{r}{v}} \stackrel{\rightharpoonup}{v}$ | － | $\begin{array}{r} \stackrel{\rightharpoonup}{0} \\ \stackrel{\rightharpoonup}{v} \\ \hline \end{array}$ | $\stackrel{\square}{\mathrm{v}}$ | v | v | $\stackrel{\stackrel{\rightharpoonup}{v}}{\stackrel{\rightharpoonup}{2}}$ | $\begin{array}{r} \stackrel{\rightharpoonup}{v} \\ \hline \end{array}$ | v | v | $\stackrel{\square}{\text { v }}$ |
|  | $5$ |  |  | Alona Bay Creek |  |  |  | $\frac{9}{4}$ $\frac{9}{x}$ $\frac{2}{5}$ 0 0 0 |  |  |  |  |  |  | Cranberry Creek |  |  |  |  |  |  |  |  |  |  |  |  |


| Tributary | Sb | Sn | Sr | Ti | V | Y | Zn | Mercury | Total C | Inorganic C | Organic <br> C | LOI |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units | ppm | ppm | ppm | ppm | ppm | ppm | ppm | ppm | \% | \% | \% | \% |
| Agawa River | <5 | <10 | 7 | NA | 21 | 2.6 | 32 | 0.005 | 0.7 | 0 | 0.7 | 1.7 |
| Aguasabon River | <5 | <10 | 16 | NA | 23 | 4.3 | 28 | 0.015 | 2.2 | 0.8 | 1.4 | 3.4 |
| Alona Bay Creek | <5 | <10 | 17 | NA | 35 | 10.1 | 51 | 0.089 | 9.1 | 0 | 9.1 | 19 |
| Anglers Creek | <5 | <10 | 26 | 669 | 26 | 10.5 | 84 | 0.057 | 3.9 | 1.7 | 2.2 | 5.6 |
| Barrett River | <5 | <10 | 9 | 298 | 24 | 3.2 | 27 | 0.029 | 3.5 | 0 | 3.5 | 7.7 |
| Batchawan River | <5 | <10 | 9 | 432 | 23 | 3.8 | 36 | 0.021 | 1 | 0 | 1 | 2.6 |
| Bennett Creek | <5 | <10 | 34 | 535 | 27 | 6.7 | 247 | 0.051 | 3.8 | 0.7 | 3.1 | 6.5 |
| Big Carp River | <5 | <10 | 8 | 356 | 18 | 4.2 | 38 | 0.017 | 1.6 | 0 | 1.6 | 4 |
| Big Trout Creek | <5 | <10 | 28 | 938 | 55 | 7.7 | 83 | 0.032 | 3.7 | 1.3 | 2.4 | 6 |
| Blackbird Creek | <5 | <10 | 29 | 361 | 26 | 4.1 | 201 | 0.102 | 10.8 | 1.8 | 9 | 18.6 |
| Black Sturgeon River | <5 | <10 | 14 | 986 | 95 | 5.4 | 35 | 0.014 | 1.7 | 0.3 | 1.4 | 3.2 |
| Carp River | <5 | <10 | 9 | 546 | 33 | 4.1 | 42 | 0.015 | NA | NA | NA | NA |
| Chippewa River | $<5$ | <10 | 8 | 621 | 32 | 3.5 | 37 | 0.023 | 0.9 | 0 | 0.9 | 2.3 |
| Coldwater River | <5 | <10 | 20 | 1130 | 98 | 7.1 | 44 | 0.021 | 2.2 | 0.9 | 1.3 | 3.3 |
| Cold Water River East | <5 | <10 | 14 | 372 | 23 | 5 | 67 | 0.037 | 5.8 | 0 | 5.8 | 12.5 |
| Cranberry Creek | <5 | <10 | 20 | 494 | 29 | 5.6 | 52 | 0.057 | 5.2 | 0 | 5.2 | 11.4 |
| Current River | <5 | <10 | 26 | 249 | 89 | 14 | 167 | 0.181 | 16.2 | 0.2 | 16.1 | 36.6 |
| Cypress River | <5 | $<10$ | 8 | 387 | 23 | 4.9 | 55 | 0.026 | 1.9 | 0.1 | 1.8 | 4 |
| Davignon Creek | <5 | <10 | 86 | 552 | 34 | 7 | 136 | 0.023 | 3.8 | 1.6 | 2.2 | 5.5 |
| Dead Horse Creek | <5 | $<10$ | 18 | 683 | 33 | 6.7 | 96 | 0.054 | 4 | 0 | 4 | 8.9 |
| Diversion Channel | <5 | <10 | 15 | 419 | 27 | 6.4 | 80 | 0.042 | 2.8 | 0.4 | 2.5 | 5.5 |
| Fort Creek | <5 | <10 | 31 | 576 | 37 | 8.4 | 193 | 0.115 | 3.4 | 0.7 | 2.7 | 5.7 |
| Goulais River | <5 | <10 | 14 | 776 | 34 | 5.7 | 45 | 0.015 | 1.4 | 0.1 | 1.3 | 13.5 |
| Government Creek | <5 | <10 | 15 | 681 | 35 | 4.6 | 41 | 0.022 | 2.9 | 0 | 2.9 | 6.4 |
| Gravel River | <5 | <10 | 23 | 419 | 16 | 4.3 | 19 | 0.013 | 3.2 | 2.5 | 0.7 | 2.1 |
| Harmony River | <5 | <10 | 17 | 626 | 34 | 5.1 | 46 | 0.026 | 2.6 | 0 | 2.6 | 6 |
| Havilland Creek | <5 | <10 | 7 | 346 | 19 | 4.5 | 24 | 0.019 | 1.8 | 0 | 1.8 | 4.2 |
| Jackpine River | <5 | <10 | 27 | 607 | 50 | 8.4 | 79 | 0.044 | 4.3 | 0.2 | 4.2 | 10.2 |
| Kaministiquia River | <5 | $<10$ | 18 | 859 | 70 | 7.6 | 117 | 0.103 | 2.7 | 0.2 | 2.5 | 5.9 |
| Kelly Creek | <5 | <10 | 19 | 354 | 26 | 8.8 | 41 | 0.049 | 4.6 | 0 | 4.6 | 10.3 |
| Little Cypress River | <5 | <10 | 27 | 960 | 43 | 9.1 | 78 | 0.051 | 3.4 | 0.9 | 2.5 | 6.5 |


| Tributary | Latitude | Longitude | Sampling Date | HCB | Endrin aldehyde | OCS | Toxaphene | Aldrin | $\alpha \mathrm{HCH}$ | $\beta \mathrm{HCH}$ | б HCH | Lindane |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units |  |  |  | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ |
| Little Gravel River | 48.9223 | -87.7741 | 18-Aug-15 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Little Pine River | 48.0393 | -89.534 | 16-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Little Carp Creek | 46.5097 | -84.4442 | 13-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | $<0.002$ | <0.002 | $<0.002$ | <0.002 |
| Little Goulais River | 46.8214 | -84.5006 | 01-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Little Pic River | 48.80 | -86.6 | 15-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Magpie River | 47.9393 | -84.8299 | 14-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.002 |
| McIntyre River | 48.399 | -89.2489 | 16-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| McKellar Creek | 48.8154 | -86.7107 | 19-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| McVicar Creek | 48.439 | -89.2146 | 16-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Michipicoten River | 47.9225 | -84.8055 | 14-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Neebing River | 48.3943 | -89.2466 | 16-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Nipigon River | 49. | -88.2553 | 03-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Old Woman River | 47. | -84.8941 | 30-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Ozone Creek | 49.0218 | -88.0377 | 17-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Pancake River | 46.9603 | -84.65 | 31-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Pic River | 48.607 | -86.2908 | 15-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Pigeon River | 48.0038 | -89.5927 | 16-Sep-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Pine River | . 06 | -89.5 | 16-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Prairie River | 48.803 | -86.776 | 19-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Sand River | 47.433 | -84.7304 | 30-Aug-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Sawmill Creek | 46.87 | -84.3 | 31-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Speckled Trout Creek | 47.315 | -84.5986 | 31-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Steel River | 48.7717 | -86.8905 | 19-Aug-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Stokely Creek | 46.8147 | -84.4078 | 01-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Tremblay Creek | 47.961 | -84.8903 | 30-Aug-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Wolf River | 48.8204 | -88.5316 | 17-Aug-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Amnicon River | 46.6855 | -91.8639 | 24-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Anna River | 46.4089 | -86.6402 | 18-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Au Train River | 46.4324 | -86.8269 | 05-Oct-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Beaver River | 47.2602 | -91.2958 | 18-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Big Iron River | 46.8301 | -89.5705 | 22-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.0 | <0.002 |


| $\begin{aligned} & 50 \\ & \stackrel{0}{0} \\ & \stackrel{\rightharpoonup}{0} \end{aligned}$ | $\begin{aligned} & \text { 인 } \\ & \hline 0 \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O} \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O} \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { O } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O} \\ & \text { V } \end{aligned}$ |  | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { O } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N} \\ & \text { O} \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N} \\ & \text { O} \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | O |
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|  | 기 | $0$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \hline \end{aligned}$ |  | $\begin{aligned} & \text { O. } \\ & \text { i } \end{aligned}$ | $\begin{aligned} & \text { No } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { No } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $0$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | No | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \hline \end{aligned}$ | $\begin{aligned} & 0 \\ & \text { i } \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \hline \end{aligned}$ | Ò | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | V | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | N |
| $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | \| | 2 | $\begin{aligned} & \mathrm{O} \\ & \mathrm{O} \\ & \mathrm{v} \end{aligned}$ |  | $\begin{aligned} & \text { O. } \\ & \text { v } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O} \\ & \text { V } \end{aligned}$ | $0$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \hline \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O} \\ & \text { O } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \mathrm{O} \\ & \mathrm{O} \\ & \mathrm{v} \end{aligned}$ | O. | $\begin{aligned} & 0 . \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 . \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ |  | $0$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | N O V v |
|  | 안 | $20$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & 0 \\ & 0 \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { O } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O} \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O} \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { O } \end{aligned}$ | N O v | O. | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O} \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & \text { N } \\ & 0 \\ & 0 \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { O } \end{aligned}$ | $\begin{aligned} & \text { O} \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & 0 \\ & 0 \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | N |
|  | $\frac{0}{7}$ |  | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & \text { O} \\ & \text { O } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O. } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { O } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { O} \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O} \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & 0 \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \hline \end{aligned}$ | $\begin{aligned} & 0 \\ & \text { O } \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & 0 \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { v } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | O |
| ठ | 운 | $\begin{aligned} & \text { No } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & 8 \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { No } \\ & \text { ó } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & 0 \\ & 0 \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { O } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { O } \end{aligned}$ | $\begin{aligned} & \mathrm{O} \\ & \text { o } \end{aligned}$ | O. | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { O } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & \text { O } \\ & \text { V } \end{aligned}$ | o | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { O } \\ & \text { in } \end{aligned}$ | O. | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & 0 \\ & \text { o } \\ & \text { v } \end{aligned}$ | $\begin{aligned} & \text { N } \\ & \text { O } \\ & \text { V } \end{aligned}$ | O |
| $\begin{aligned} & \text { Nan } \\ & \end{aligned}$ | $\stackrel{8}{5}$ |  |  |  |  |  |  |  | $\begin{aligned} & \frac{\mathrm{D}}{0} \\ & \frac{0}{U} \\ & \frac{1}{0} \\ & \overline{0} \\ & \frac{1}{0} \\ & \end{aligned}$ |  |  |  | $\begin{aligned} & \frac{2}{\mathrm{o}} \\ & \frac{0}{\mathrm{D}} \\ & \frac{0}{\mathrm{O}} \\ & \hline \end{aligned}$ |  | ․ <br> 0 <br> 0 <br> 0 <br>  |  |  |  |  |  |  |  |  |  |  |  | $\begin{aligned} & \frac{y}{x} \\ & \frac{1}{4} \\ & \frac{4}{3} \end{aligned}$ |  |  |  |  |  |


| Tributary | Total DDT <br> Metabolites | Dieldrin | Endosulfan <br> I | Endosulfan II | Endosulfan Sulfate | Endrin | Heptachlor | Heptachlor Epoxide | Methoxychlor | Mirex | Aroclor 1016 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ |  | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ |
| Little Gravel River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Little Pine River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Little Carp Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Little Goulais River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Little Pic River | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Magpie River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| McIntyre River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.01 |
| McKellar Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| McVicar Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Michipicoten River | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.01 |
| Neebing River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Nipigon River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.01 |
| Old Woman River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Ozone Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Pancake River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Pic River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Pigeon River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.01 |
| Pine River | <0.002 | <0.002 | <0.002 | $<0.002$ | $<0.002$ | $<0.002$ | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.01 |
| Prairie River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.01 |
| Sand River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.01 |
| Sawmill Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Speckled Trout Creek | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.01 |
| Steel River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Stokely Creek | <0.002 | <0.002 | <0.002 | $<0.002$ | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Tremblay Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Wolf River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Amnicon River | <0.002 | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Anna River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Au Train River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Beaver River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Big Iron River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |



|  | $\begin{aligned} & 0 \\ & \frac{0}{0} \\ & 0 \end{aligned}$ | $\stackrel{\sim}{N}$ | $\frac{0}{v}$ | $\stackrel{\circ}{v}$ | $\stackrel{\circ}{v}$ | $\stackrel{\circ}{\mathrm{v}}$ | 8 | $\stackrel{ }{ }$ | $\stackrel{\circ}{\mathrm{v}}$ |  | $\frac{0}{v}$ | $\bar{\sim}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{O}{\mathrm{v}}$ | $\frac{0}{\mathrm{v}}$ | $\stackrel{\text { 안 }}{ }$ | $\stackrel{\square}{v}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\circ}{\text { v }}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{O}{v}$ | $\stackrel{\circ}{v}$ | $\bigcirc$ | $\stackrel{\circ}{\mathrm{V}}$ | $\stackrel{\square}{v}$ | $\stackrel{\square}{v}$ |  | \％ |  |  | $\stackrel{\circ}{\mathrm{v}}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $5$ | $\stackrel{\sim}{\infty}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{O}{\mathrm{v}}$ | $\bigcirc$ | 8 | $\stackrel{\text { 안 }}{ }$ | $\stackrel{\square}{\mathrm{v}}$ | － | $\stackrel{\stackrel{\rightharpoonup}{v}}{ }$ | $\bigcirc$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\circ}{v}$ | 은 | $\stackrel{\circ}{v}$ | $\stackrel{\square}{v}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\circ}{v}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | $\frac{0}{v}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\stackrel{\rightharpoonup}{v}}{ }$ | $\bigcirc$ | 안 | $\stackrel{\circ}{\mathrm{v}}$ | ¢ |  |  | $\stackrel{\circ}{\mathrm{v}}$ |
|  | $\begin{aligned} & \text { 号 } \\ & \text { 오 } \end{aligned}$ | ㄱ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\circ}{v}$ | $\stackrel{\circ}{v}$ | $\frac{O}{v}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\sim}{\square}$ | $\stackrel{\square}{\mathrm{v}}$ | 유N | $\stackrel{O}{\mathrm{v}}$ | $\bigcirc$ | － | $\stackrel{\circ}{\mathrm{v}}$ | $\frac{O}{v}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\circ}{\mathrm{V}}$ | $\stackrel{\square}{v}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\text { ㄴ }}{ }$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{V}}$ | $\bigcirc$ | 앙 | $\stackrel{\rightharpoonup}{v}$ | － |  |  | $\stackrel{\text {－}}{ }$ |
| $\begin{aligned} & \mathbf{0} \\ & \stackrel{0}{\omega} \\ & \stackrel{\text { D}}{2} \end{aligned}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | ® | ค | v | ロ | $\stackrel{10}{ }$ | 8 | \％ | V | \％ | $\stackrel{0}{0}$ | ® | \％ | $\stackrel{\circ}{0}$ | $\stackrel{10}{ }$ | V | v | ® | ๒ | $\bigcirc$ | $\stackrel{10}{*}$ | V | $\stackrel{\text { ® }}{ }$ | v | レ | ® | $\stackrel{\text { V }}{ }$ |  | $\infty$ |  |  | $\stackrel{\circ}{v}$ |
|  | $\begin{aligned} & 0 \\ & \frac{0}{0} \\ & 0 \end{aligned}$ |  | $\stackrel{\sim}{\mathrm{v}}$ | $\stackrel{\text { V }}{ }$ | $\stackrel{セ}{v}$ | V | \％ | ¢ | $\stackrel{\square}{\vee}$ | $\stackrel{\circ}{\circ}$ | $\stackrel{6}{6}$ | － | \％ | $\stackrel{\text { ® }}{ }$ | v | V | ท | $\stackrel{10}{ }$ | レ | $\stackrel{1}{6}$ | $\stackrel{10}{*}$ | V | $\stackrel{\text { v }}{ }$ | $\stackrel{0}{0}$ | ๑ | $\stackrel{10}{ }$ | $\stackrel{6}{V}$ | $\stackrel{\circ}{\vee}$ | 안 |  |  | $\stackrel{\text { v }}{ }$ |
|  | $\begin{aligned} & \text { O} \\ & \text { 令 } \end{aligned}$ | $\stackrel{\bullet}{v}$ | レ | ท | ® | ロ | ） | g | レ0 | ¢ | ๑ | レ0 | ท | $\stackrel{0}{0}$ | v | $\stackrel{\text { ® }}{ }$ | V | $\stackrel{\text { ® }}{ }$ | レ0 | $\stackrel{1}{v}$ | ท | ถ | $\stackrel{\text { V }}{ }$ | ท | $\stackrel{\text { ® }}{ }$ | $\stackrel{10}{ }$ | $\stackrel{\circ}{\text { ® }}$ | $\stackrel{10}{ }$ | $\stackrel{10}{ }$ |  |  | $\stackrel{\circ}{\text { v }}$ |
|  | 을 | $\underset{\tau}{\infty}$ | $\stackrel{\sim}{\mathrm{v}}$ | $\stackrel{\text { V }}{ }$ | $\stackrel{\text { ® }}{ }$ | ロ | 8 | － | レ | 亏 | $\stackrel{\bullet}{v}$ | － | ～ | $\stackrel{\text { ® }}{ }$ | V | $\stackrel{\text { ® }}{ }$ | ๑ | $\stackrel{\ominus}{V}$ | $\stackrel{\text { V }}{ }$ | $\stackrel{\text { V }}{ }$ | v | V | $\stackrel{\bullet}{v}$ | จ | $\stackrel{\text { v }}{ }$ | ® | $\stackrel{6}{V}$ | $\stackrel{10}{ }$ | in |  |  | $\stackrel{\circ}{v}$ |
|  | $\begin{aligned} & 0 \\ & \frac{0}{0} \\ & \hline 2 \end{aligned}$ | $\stackrel{\bullet}{\vee}$ | レ | ท | ロ | V | v | － | レ0 | N | V | レ0 | V | $\stackrel{0}{0}$ | v | V | v | $\stackrel{\text { ® }}{ }$ | v | v | v | ถ | จ | ท | ง | $\stackrel{10}{ }$ | ท | ท | ๒ |  |  | レ0 |
|  | 문 | $\stackrel{\circ}{v}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\circ}{v}$ | $\stackrel{\circ}{v}$ | $\stackrel{\text { O}}{\text { v }}$ | 안 | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\circ}{\mathrm{v}}$ | － | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\circ}{v}$ | $\stackrel{\text { O}}{ }$ | $\stackrel{\circ}{v}$ | $\stackrel{\square}{v}$ | $\stackrel{\text { V }}{ }$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\circ}{\mathrm{V}}$ | $\stackrel{\text { V }}{ }$ | 안 | $\stackrel{\text { V }}{ }$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\text { O}}{ }$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\text { O}}{ }$ | $\stackrel{\circ}{\mathrm{V}}$ | 은 | $\stackrel{\text { O}}{ }$ | $\stackrel{\circ}{\mathrm{v}}$ |  |  | $\bigcirc$ |
|  | $\begin{aligned} & \frac{n}{5} \\ & 5 \end{aligned}$ |  |  |  |  |  |  |  |  | $\begin{aligned} & \stackrel{y}{0} \\ & \stackrel{0}{0} \\ & \vdots \\ & \vdots \\ & \vdots \\ & \vdots \\ & \hline \end{aligned}$ |  |  |  | $\begin{aligned} & \frac{9}{y} \\ & \substack{x \\ c \\ c \\ 0 \\ 0 \\ 0 \\ 0 \\ 0 \\ 0 \\ 0} \end{aligned}$ |  |  | $\begin{aligned} & \substack{0 \\ \\ \\ 0 \\ 0} \end{aligned}$ | － |  |  |  |  |  |  | $\stackrel{\rightharpoonup}{\omega}$ |  | $\begin{aligned} & \frac{9}{y} \\ & \frac{y}{y} \\ & \hline 0 \end{aligned}$ | － |  |  |  | － |



| Q | ह⿳亠二口欠贝） |  | $\infty$ | $\infty$ | $\stackrel{1}{\square}$ | 「 |  | $\stackrel{\text { N }}{ }$ | $\checkmark$ | $\stackrel{L}{6}$ | $\stackrel{\circ}{\mathrm{V}}$ 안 | ＋ | $\stackrel{\circ}{v}$ |  | $\infty$ | $\stackrel{セ 0}{V}$ |  | の ${ }^{\text {® }}$ | $\stackrel{\text { ® }}{\sim}$ |  | $\infty$ | 人 |  | $\stackrel{1}{6}$ | $\stackrel{\square}{\square}$ | 은 | 아 | $\stackrel{\sim}{\sim}$ |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 之 | 哀 |  | へ | $\stackrel{m}{\square}$ | N |  | $\stackrel{\square}{\square}$ | － | m | ন | 「 $\bar{\sim}$ | － | 入 | ～ | $\stackrel{\Omega}{\square}$ | $\infty$ |  | せの | N | － | $\stackrel{\square}{\square}$ | ＾ | $\stackrel{\sim}{\sim}$ | $\bigcirc \bigcirc$ | $\infty$ | － | $\sim$ | $\bigcirc$ | － |
| 20 | $\bigcirc$ |  |  |  | $\stackrel{\rightharpoonup}{\sigma}$ | $\stackrel{\rightharpoonup}{\sigma}$ | $\frac{\square}{0}$ | $\stackrel{3}{0}_{0}^{\mathbf{o}}$ | No | $\begin{array}{l\|l} 0 \\ 0 \\ 0 \\ 0 & 0 \\ 0 & 0 \\ \hline \end{array}$ | $\begin{aligned} & 88 \\ & \hline 8.8 \\ & 0.8 \\ & \hline 0 \end{aligned}$ |  | $\begin{aligned} & \circ \\ & \stackrel{1}{0} \\ & \hline \end{aligned}$ | ${ }_{3}^{\circ}$ | $\begin{aligned} & 0 \\ & \vdots \\ & 0 \\ & 0 \end{aligned}$ | $\stackrel{m}{c}_{\substack{c}}^{\substack{0}}$ |  | $\begin{array}{ll} 0 \\ 0 \\ 0.0 \\ 0 & 0 \\ 0 \end{array}$ | $\underset{\substack{\circ}}{\substack{0}}$ | N | $\underset{\substack{v} \stackrel{m}{c}}{\substack{0}}$ | $\begin{aligned} & ? \\ & \hline \end{aligned}$ |  | $$ | $\begin{aligned} & \text { f } \\ & j \\ & \vdots \end{aligned}$ | $\stackrel{6}{0}$ |  | $\stackrel{ \pm}{\tau}$ | O |
| $\sum^{\circ}$ | 䂴 | v | $\bar{v} \bar{v}$ | $\overline{\mathrm{v}}$ | $\overline{\mathrm{v}}$ | $\bar{v}$ | ¢ | v | v | v | $\overline{\mathrm{v}}$ v | v | v | v | $\overline{\mathrm{v}}$ ¢ | $\overline{\mathrm{v}}$ | $\bar{v}$ | $\overline{\mathrm{v}}$ v | V | $\overline{\mathrm{v}}$ | v | v |  | $\overline{\mathrm{v}} \overline{\mathrm{v}}$ | $\overline{\mathrm{v}}$ | $\overline{\mathrm{v}}$ | $\overline{\mathrm{v}}$ |  | v |
| $\stackrel{5}{\Sigma}$ | $\underset{i}{E}$ |  | $\stackrel{\sim}{\sim}$ | $\stackrel{\stackrel{\circ}{\circ}}{\stackrel{\rightharpoonup}{*}}$ | $\begin{aligned} & \stackrel{\circ}{\circ} \\ & \stackrel{\circ}{\mathrm{O}} \end{aligned}$ | $\hat{\tilde{\gamma}}$ | $5$ | $\stackrel{\circ}{\infty}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\stackrel{\circ}{\stackrel{\circ}{ }} \stackrel{\stackrel{1}{N}}{N}$ | $\frac{0}{\hat{m}} \stackrel{\mathrm{~N}}{\mathrm{\sigma}}$ | $\frac{\ddots}{f} \frac{n}{N}$ | $\stackrel{\text { ホ }}{\text { }}$ | $8$ | $\frac{0}{\lambda}$ | $\underset{\sim}{\underset{N}{2}}$ | ¢ | \％${ }_{\circ}^{\circ}$ N | $\underset{\sim}{\stackrel{N}{\sim}}$ | $\begin{aligned} & \text { O} \\ & \text { Nָ } \end{aligned}$ | $\underset{\sim}{\underset{\sim}{x}}$ | O | $\stackrel{\sim}{\sim}$ | $\underset{\sim}{N} \underset{\sim}{\sim}$ | パべN | ¢ |  |  | A |
| $\sum^{\infty}$ |  |  |  | $\stackrel{\star}{\circ}$ | $\stackrel{\stackrel{N}{0}}{0}$ |  | $\stackrel{\sim}{\mathrm{i}} \stackrel{\sim}{\mathrm{~N}}$ | ס | $\stackrel{\infty}{\circ}$ |  | No | $\overbrace{i}^{N} \underset{i}{i}$ | jo | $\stackrel{\infty}{\circ}$ | $\begin{aligned} & \mathrm{Z} \\ & 0 \end{aligned}$ | ণ |  |  | $\underset{\sim}{N}$ | $\stackrel{\mathbb{O}}{\circ}$ | $\stackrel{\substack{\mathrm{o} \\ \hline \\ \hline}}{\sim}$ | $\stackrel{8}{\stackrel{8}{\square}}$ |  | $\stackrel{\circ}{\circ} \stackrel{\infty}{\circ}$ | $\stackrel{\substack{\mathrm{C}}}{2}$ | Ơ |  | 8 | － |
| צ | $\bigcirc$ |  | $\stackrel{N}{\circ} \underset{\sim}{\infty}$ |  | $\mp$ | $\circ$ |  |  |  | $\begin{array}{l\|l} \infty \\ 0 \\ 0 & \text { c } \\ 0 \end{array}$ | $\begin{array}{ll} \circ & 8 \\ 0 \\ 0 & 0 \\ 0 \end{array}$ |  | $\stackrel{8}{\circ}$ | $\underset{\sim}{\infty}$ | O |  |  | $\begin{gathered} t \\ \vdots \\ 0 \end{gathered}$ | $$ | $\stackrel{n}{\square}$ |  | $8$ |  | $\bigcirc$ |  | t. |  |  | $\stackrel{\infty}{0}$ |
| $\stackrel{\sim}{\sim}$ | $\bigcirc$ |  | $\stackrel{Q}{\mathrm{~N}} \stackrel{\stackrel{L}{\mathrm{~N}}}{\stackrel{\sim}{i}}$ | $\stackrel{\square}{\square}$ | $\begin{aligned} & \underset{\sim}{0} \\ & \text { b } \end{aligned}$ | $\stackrel{8}{\stackrel{8}{+}}$ |  |  |  |  | $\stackrel{\overline{+}}{\stackrel{\infty}{\infty}}$ | $\underset{\sim}{\circ} \underset{\sim}{\underset{\sim}{n}}$ | $?$ | $\stackrel{8}{-8}$ | $\hat{\varrho}$ |  |  | $\stackrel{m}{\dot{m}}-$ | － | $\underset{\sim}{+}$ |  | \％ |  | 揭 | no |  |  |  | $\stackrel{+}{+}$ |
| 3 | ह⿳亠二口欠心 |  | N | N | の | $\infty$ | $\bigcirc$ | フ | $\stackrel{\sim}{\sim}$ | ¢ | 두 | ホ | $\bigcirc$ | ค | $\stackrel{ }{\sim}$ | へ | ¢ | O | 0 | ～ | $\stackrel{\sim}{-}$ | 「 |  | $\pm \bar{N}$ | ＇ | $\stackrel{1}{\square}$ | $\infty$ | F | $\stackrel{\circ}{\sim}$ |
| う̄ | ¢ 칭 | $\overline{\text { m }}$ | － | N | N | ํ | $\stackrel{\sim}{\circ}$ | $\stackrel{8}{8}$ | ¢ | $\stackrel{\sim}{-}$ | $\stackrel{\infty}{\sim}$ | $\stackrel{\circ}{\sim}$ | N | $\infty$ | N | $\bigcirc$ | $\overline{\text { F }}$ | ¢ ${ }_{-}^{\infty}$ | $\bigcirc$ | ल | A | － |  | へ | ₹ | $\stackrel{m}{\square}$ | $\bigcirc$ |  | $\bigcirc$ |
| $\bigcirc$ | 칭 | 안 | 앋 | の | N | ๑ | $\infty$ | $\bigcirc$ | へ | 아 | ค 안 | － | 은 | $\stackrel{m}{\square}$ | の | － | N | $\stackrel{\circ}{\sim}$ | $\bigcirc \stackrel{6}{\square}$ | $\stackrel{\square}{\square}$ | г | 人 |  | $\wedge \infty$ | － | m |  |  |  |
| J | 흥 |  | $00$ | $\mathfrak{l}$ | $\begin{aligned} & ? \\ & \vdots \\ & 0 \\ & 0 \end{aligned}$ | $0$ | $0$ | $\begin{aligned} & \text { n } \\ & \stackrel{0}{2} \end{aligned}$ | $\stackrel{0}{0} \stackrel{1}{6}$ |  |  | $\stackrel{\ominus}{\mathrm{v}} \stackrel{n}{\stackrel{n}{v}}$ | $\stackrel{n}{?} \stackrel{n}{0}$ | $\stackrel{n}{0}$ | $$ | $0$ | $\begin{aligned} & n \\ & \stackrel{n}{0} \end{aligned}$ | $\begin{array}{ll} n \\ \stackrel{n}{0} & \stackrel{n}{0} \\ 0 \end{array}$ | $\stackrel{\sim}{v} \stackrel{\wedge}{\circ}$ | $\begin{aligned} & \bullet \\ & \stackrel{0}{0} \end{aligned}$ |  | $\begin{aligned} & 2 \\ & \vdots \\ & i \\ & 0 \end{aligned}$ | V | $\begin{array}{lll} 0 & 0 \\ \stackrel{0}{0} \\ \text { vo } \\ 0 \end{array}$ | $\begin{array}{r} 0 \\ \vdots \\ i \end{array}$ | $0$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \end{aligned}$ | v | V |
| Ő | $\bigcirc$ |  | $\stackrel{\leftrightarrow}{\Gamma} \underset{0}{\underset{O}{\circ}}$ | $\stackrel{\stackrel{N}{\mathrm{O}}}{0}$ | $\hat{o}$ | $0$ | $\stackrel{P}{i} \dot{\sim}$ | $\begin{aligned} & 8 \\ & \hline 0 \\ & \hline 0 \end{aligned}$ | $\begin{aligned} & \text { to } \\ & \stackrel{y}{n} \end{aligned}$ | $\overbrace{0}^{\circ}$ | $\stackrel{\Im}{\substack{\circ}}$ | $\stackrel{\circ}{\circ} \stackrel{ \pm}{\omega}$ | $0$ | $\stackrel{\sim}{~}$ | $\ddot{O}$ | $\dot{\sigma}$ |  | $\begin{aligned} & \stackrel{\rightharpoonup}{\infty} \\ & \stackrel{\infty}{\dot{\sim}} \end{aligned}$ |  | $\stackrel{\bullet}{\circ}$ | $\stackrel{\circ}{\circ}$ | $\stackrel{\substack{\mathrm{o} \\ \hline \\ \hline}}{\substack{0}}$ |  | $\begin{aligned} & \bar{\infty} \\ & \stackrel{8}{\circ} \\ & \dot{\sim} \end{aligned}$ |  | $\stackrel{8}{\circ}$ |  | ． | $\stackrel{\circ}{\circ}$ |
| ¢ | ⿳亠二口欠2 | ${ }^{\circ}$ | $\stackrel{\text { ® }}{ }$ | ๒ | V | V | ท | $\stackrel{\bullet}{V}$ | $\stackrel{\circ}{0}$ | ๒ ¢ ¢ | ロํข | セู | V | V | v | $\stackrel{\text { v }}{ }$ | V | ¢ ¢ ¢ | V | V | ๒ | $\bigcirc$ | เ | ソูレ | $\stackrel{0}{0}$ | $\bigcirc$ | V | $\stackrel{\rightharpoonup}{*}$ | v |
| ® | 场 |  | $\underset{\circ}{\circ}$ | $\begin{gathered} m \\ 0 \end{gathered}$ | $\stackrel{\infty}{\circ}$ | N | $\underset{\sim}{N}$ | $\underset{\substack{t}}{ }$ | ${ }^{\circ}$ | Noc | $\begin{array}{c\|c} N & N \\ \underset{V}{\circ} & 0 \\ V \end{array}$ | $\begin{array}{lc} \text { N N } \\ \stackrel{y}{v} \\ \stackrel{y}{*} \end{array}$ | $\stackrel{\text { N}}{\stackrel{\rightharpoonup}{v}}$ | $\stackrel{\circ}{\circ}$ | $\stackrel{m}{0}$ | $\stackrel{N}{\mathrm{~N}}$ | $0$ |  | $\begin{aligned} & \underset{\sim}{v} \\ & \stackrel{y}{*} \\ & \hline \end{aligned}$ | $\stackrel{\circ}{\circ}$ |  | N | $v$ | $\begin{array}{c\|c} N \\ \underset{\sim}{V} & \dot{O} \\ \hline \end{array}$ | $\stackrel{\text { ¢ }}{\sim}$ | $\bigcirc$ |  | へo | $\dot{\circ}$ |
| 0 | ¢ 칭 |  | $\infty$ | $\stackrel{\sim}{\square}$ | － | ¢ | $\stackrel{\sim}{\sim}$ | $\stackrel{\square}{\circ}$ | $\stackrel{\infty}{\stackrel{\infty}{\sim}}$ | $\stackrel{\circ}{\infty}$ | $\stackrel{\sim}{\sim}$ | N | ¢ | ก | － | へ | \＆ | 8 ！¢ | 욷 | \％ | ） | $\stackrel{\infty}{+}$ | ल | 윧 | － | $\bigcirc$ | $\stackrel{\circ}{\circ}$ | － | is |
| 8 | 딩 | ${ }^{\circ}$ | ง ง | レ | V | เ | N | $\infty$ | V | เง เ ¢ | ค ¢ ¢ | ค ¢ | $\stackrel{\text { V }}{ }$ | V | V | $\stackrel{0}{ }$ | $\bigcirc$ | レ ${ }^{\text {¢ }}$ | จ ¢ | V | V | $\stackrel{\text { V }}{ }$ | ท | の し゚ | V | $๑$ | V | $\stackrel{\square}{\text { V }}$ | $\stackrel{\sim}{*}$ |
| 『 | く |  |  |  | $\hat{o}$ | $0$ | $5$ | $\underset{\sim}{8}$ |  | $\begin{array}{lll} \infty \\ \circ \\ 0 & \infty \\ 0 & 0 \end{array}$ | ${ }_{0}^{\circ} \underset{\sim}{N}$ | $\begin{array}{ll} \mathrm{V} \\ \vdots \\ 0 \\ \hline \end{array}$ |  | $\stackrel{\infty}{\stackrel{\infty}{+}}$ | $\underset{\substack{N \\ \hline}}{ }$ |  | $\stackrel{\text { N }}{\stackrel{1}{r}}$ | $\begin{array}{cc} \underset{\sim}{\sim} \\ \underset{\sim}{J} \\ \hline \end{array}$ |  |  |  |  |  |  |  |  | \％－ |  | ${ }_{0}^{\infty}$ |
| 8 | 2 |  | $\overline{\stackrel{\rightharpoonup}{v}} \stackrel{\rightharpoonup}{\dot{v}}$ | No | $\begin{gathered} c \\ \vdots \\ \hline \end{gathered}$ | $\stackrel{\rightharpoonup}{\dot{v}}$ | $\begin{gathered} m \\ 0 \end{gathered}$ | $\stackrel{\rightharpoonup}{\dot{v}}$ | $\stackrel{\square}{0}$ | $\overline{\dot{v}}$ | $\begin{array}{ll} \bar{r} \\ \dot{0} \\ \hline \end{array}$ | $\dot{\rightharpoonup}$ | $\bar{v}$ | $\stackrel{\square}{\mathrm{v}}$ |  | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | $\begin{array}{l:} \bar{\circ} \\ \dot{v} \\ \hline \end{array}$ | $\bigcirc$ | v |  | V＇ |  | $\stackrel{\rightharpoonup}{\dot{\circ}} \stackrel{\square}{\text { v }}$ | － |  | $\bigcirc$ | v | $\stackrel{\square}{\text { v }}$ |
|  | $\begin{aligned} & n \\ & \\ & \hline 1 \end{aligned}$ |  |  |  |  |  |  | $\begin{aligned} & \frac{1}{9} \\ & \frac{y}{x} \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & \hline \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  | $\stackrel{\circ}{i}$ |  |  |  |  |  | － |


| Tributary | Sb | Sn | Sr | Ti | V | Y | Zn | Mercury | Total C | Inorganic C | Organic C | LOI |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units | ppm | ppm | ppm | ppm | ppm | ppm | ppm | ppm | \% | \% | \% | \% |
| Little Gravel River | <5 | <10 | 17 | 705 | 33 | 7.4 | 60 | 0.031 | 3.7 | 0.8 | 2.8 | 7 |
| Little Pine River | <5 | <10 | 15 | 845 | 54 | 8 | 65 | 0.031 | 1.9 | 0.1 | 1.8 | 5.1 |
| Little Carp Creek | <5 | <10 | 16 | 387 | 30 | 6.7 | 79 | 0.029 | 2.1 | 0 | 2.1 | 4.9 |
| Little Goulais River | <5 | <10 | 29 | 393 | 47 | 6.7 | 71 | 0.063 | 8.4 | 0.1 | 8.3 | 19.7 |
| Little Pic River | < 5 | <10 | 42 | 536 | 19 | 6 | 31 | 0.023 | 5 | 3.8 | 1.2 | 3.5 |
| Magpie River | <5 | <10 | 17 | 387 | 23 | 5.1 | 48 | 0.034 | 4.3 | 1.1 | 3.2 | 7.5 |
| McIntyre River | <5 | <10 | 23 | 810 | 85 | 8.3 | 175 | 0.118 | 3.4 | 0.2 | 3.2 | 7.8 |
| McKellar Creek | <5 | <10 | 22 | 765 | 39 | 10 | 150 | 0.076 | 6.9 | 0.5 | 6.4 | 15.2 |
| McVicar Creek | <5 | <10 | 12 | 645 | 68 | 4.7 | 150 | 0.031 | 1.7 | 0.1 | 1.6 | 3.3 |
| Michipicoten River | <5 | <10 | 10 | 458 | 22 | 3.3 | 38 | 0.005 | 1 | 0.1 | 0.9 | 2.1 |
| Neebing River | <5 | <10 | 12 | 891 | 88 | 5.1 | 83 | 0.036 | 1.3 | 0.1 | 1.3 | 2.8 |
| Nipigon River | <5 | <10 | 19 | 730 | 45 | 5.9 | 38 | 0.016 | 2.7 | 1.5 | 1.2 | 3 |
| Old Woman River | <5 | <10 | 7 | 438 | 23 | 3.6 | 38 | 0.021 | 2.1 | 0 | 2.1 | 4.7 |
| Ozone Creek | <5 | <10 | 17 | 814 | 34 | 7 | 81 | 0.032 | 2.9 | 0.4 | 2.5 | 6.2 |
| Pancake River | <5 | <10 | 15 | 521 | 32 | 6.5 | 68 | 0.026 | 3.2 | 0 | 3.2 | 7.4 |
| Pic River | <5 | <10 | 40 | 457 | 15 | 5.4 | 37 | 0.02 | 5.3 | 4.2 | 1.1 | 2.8 |
| Pigeon River | <5 | <10 | 16 | 700 | 65 | 9 | 6.7 | 0.047 | 1.7 | 0 | 1.7 | 4 |
| Pine River | <5 | <10 | 18 | 997 | 78 | 9.4 | 89 | 0.098 | 1.9 | 0.2 | 1.7 | 4.2 |
| Prairie River | <5 | <10 | 27 | 478 | 19 | 4.5 | 30 | 0.015 | 2.2 | 1.2 | 1 | 1.9 |
| Sand River | <5 | <10 | 18 | 530 | 29 | 6.4 | 138 | 0.029 | 4.8 | 0 | 4.8 | 9.9 |
| Sawmill Creek | <5 | <10 | 27 | 725 | 48 | 10.2 | 104 | 0.07 | 9 | 0 | 9 | 19.3 |
| Speckled Trout Creek | <5 | <10 | 10 | 392 | 27 | 4.9 | 58 | 0.03 | 5.9 | 0 | 5.8 | 12.3 |
| Steel River | <5 | <10 | 17 | 590 | 22 | 6.5 | 53 | 0.029 | 1.4 | 0.1 | 1.3 | 2.2 |
| Stokely Creek | <5 | <10 | 11 | 691 | 31 | 4.9 | 41 | 0.015 | 1.8 | 0 | 1.8 | 3.9 |
| Tremblay Creek | <5 | <10 | 11 | 466 | 20 | 4.3 | 57 | 0.025 | 2.3 | 0.2 | 2.1 | 4.9 |
| Wolf River | <5 | <10 | 15 | 556 | 43 | 7.4 | 41 | 0 | 1.8 | 1.1 | 0.7 | 2.2 |
| Amnicon River | <5 | <10 | 27 | 662 | 65 | 18.6 | 79 | 0.04 | 2.6 | 0.3 | 2.3 | 7.1 |
| Anna River | <5 | <10 | 26 | 163 | 18 | 5.4 | 45 | 0.128 | 7.9 | 0.3 | 7.6 | 16.2 |
| Au Train River | <5 | <10 | 6 | 53 | 8 | 4.1 | 25 | 0.035 | 5.4 | 0.2 | 5.2 | 10.1 |
| Beaver River | <5 | <10 | 42 | 1080 | 70 | 14.3 | 85 | 0.043 | 4.8 | 0 | 4.7 | 11.8 |
| Big Iron River | <5 | <10 | 17 | 1630 | 50 | 13.8 | 33 | 0.011 | 1 | 0.1 | 0.9 | 2.6 |


| Tributary | Latitude | Longitude | Sampling Date | HCB | Endrin aldehyde | OCS | Toxaphene | Aldrin | $\alpha \mathrm{HCH}$ | $\beta \mathrm{HCH}$ | $\delta \mathrm{HCH}$ | Lindane |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units |  |  |  | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | ug/g | $\mu \mathrm{g} / \mathrm{g}$ | ug/g |
| Bluff Creek | 46.6823 | -92.0154 | 25-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.00 |
| Brule River | 47.8179 | -90.0514 | 17-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Carp River | 46.5185 | -87.384 | 18-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Chocolay River | 46.489 | -87.3285 | 18-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 |
| Cranberry Slough | 46.829 | -91.2675 | 24-Oct-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 |
| Deviltrack River | 47.7701 | -90.2616 | 17-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Eagle River | 47.4146 | -88.2986 | 20-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 |
| Falls River | 46.7 | -88.4537 | 19-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Firesteel River | 46.93 | -89.1952 | 21-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Flag River | 46.78 | -91.3841 | 24-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Flintsteel River | 46 | -89.2147 | 21-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Gooseberry River | 47.1 | -91.4685 | 18-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Gratiot River | 47.3428 | -88.4515 | 20-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Grand Marais Creek | 46.6627 | -85.9091 | 06-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Cedar Creek | 47 | -88.1497 | 20-Oct-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Huron River | 46.90 | -88.0364 | 26-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Knife River | 46.9483 | -91.7841 | 18-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Knowlton Creek | 46.7162 | -92.2002 | 25-Oct-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 |
| Laughing Whitefish River | 46.5 | -87.027 | 18-Oct-0 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Little Carp River | 46.8356 | -88.4837 | 19-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Little Cranberry River | 46.8396 | -89.4306 | 22-Oct-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 |
| Little Iron River | 46.82 | -89.5881 | 22-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Mineral River | 46.8342 | -89.5495 | 22-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Miners River | 46.4883 | -86.5407 | 18-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 |
| Misery River | 46.997 | -88.9809 | 21-Oct-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Nemadji River | 46.6857 | -92.0482 | 25-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Onion River | 47.6082 | -90.7709 | 17-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Ontonagon River | 46.8682 | -89.3186 | 21-Oct-05 | $<0.002$ | <0.002 | <0.002 | <0.08 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 |
| Pike River | 47.0193 | -88.5252 | 19-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Pokegama River | 46.666 | -92.1263 | 25-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Potatoe River | 46.851 | -89.387 | 22-Oct-05 | <0.00 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.0 | <0.0 |


| Tributary | $\alpha$ Chlordane | y Chlordane | Total Chlordane | op DDD | pp DDD | op DDE | pp DDE | op DDT | pp DDT | Total DDD | Total DDE | Total DDT |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ |
| Bluff Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Brule River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Carp River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Chocolay River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Cranberry Slough | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Deviltrack River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Eagle River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | 0.012 | <0.002 | 0.02 | <0.002 | 0.012 | 0.02 |
| Falls River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Firesteel River | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Flag River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Flintsteel River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Gooseberry River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Gratiot River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Grand Marais Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Cedar Creek | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Huron River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Knife River | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Knowlton Creek | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Laughing Whitefish River | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Little Carp River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Little Cranberry River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Little Iron River | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Mineral River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Miners River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Misery River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Nemadji River | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Onion River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Ontonagon River | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Pike River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Pokegama River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Potatoe River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |


| Tributary | Total DDT Metabolites | Dieldrin | Endosulfan | Endosulfan | Endosulfan Sulfate | Endrin | Heptachlor | Heptachlor Epoxide | Methoxychlor | Mirex | Aroclor 1016 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ |  | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ |
| Bluff Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Brule River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Carp River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Chocolay River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Cranberry Slough | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Deviltrack River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Eagle River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Falls River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Firesteel River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Flag River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Flintsteel River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Gooseberry River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Gratiot River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Grand Marais Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Cedar Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Huron River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Knife River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Knowlton Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Laughing Whitefish River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Little Carp River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Little Cranberry River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Little Iron River | <0.002 | <0.002 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Mineral River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Miners River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Misery River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Nemadji River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Onion River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Ontonagon River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Pike River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Pokegama River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Potatoe River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |



|  | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $8$ | $\frac{0}{v}$ | 9 | i | 은 | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{?}{ }$ | $\bar{\infty}$ |  | $\frac{0}{v}$ | $\frac{0}{v}$ | $\stackrel{\circ}{v}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\circ}{\mathrm{V}}$ | N | $\frac{\square}{v}$ | 은 | $\stackrel{0}{\square}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\circ}{v}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | 상 |  | $\stackrel{\square}{\mathrm{V}}$ | F |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | $\begin{aligned} & 0 \\ & 0 \\ & 0 \end{aligned}$ | pie | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | 8 | $\stackrel{\square}{v}$ | $\stackrel{\square}{\mathrm{v}}$ | 안 | is | － | $\bigcirc$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | $\frac{\square}{v}$ | $\underline{2}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | 8 | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\text { Vㅏㄴ }}{ }$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{V}}$ | $\stackrel{\square}{\text { v }}$ |  | $\stackrel{\square}{\mathrm{v}}$ | － | \％ |
|  | 皆 | on | $\stackrel{\circ}{v}$ | － | 8 | $\stackrel{O}{v}$ | $\stackrel{\square}{\mathrm{v}}$ | 은 | is | \％ | $\bigcirc$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{O}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ | 안 | $\bigcirc$ | 웃 | $\stackrel{\square}{\mathrm{v}}$ | $\bigcirc$ | 8 | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{V}}$ | 앙 | $\stackrel{\mathrm{O}}{\mathrm{V}}$ | $\stackrel{\square}{\mathrm{v}}$ | － |  | $\stackrel{\circ}{\mathrm{v}}$ | － |  |
| $\begin{aligned} & \stackrel{0}{0} \\ & \stackrel{y}{0} \\ & \hline \end{aligned}$ | $\left\lvert\, \begin{aligned} & 0 \\ & \stackrel{0}{0} \end{aligned}\right.$ | $\stackrel{\circ}{\circ}$ | ท | i | \％ | レ | ท | $\stackrel{\circ}{\sim}$ | 앋 | \％ | レ | ทจ | ® | ท๐ | V | v | ท | セ | $\stackrel{\sim}{\sim}$ | $\stackrel{\text { ® }}{ }$ | ๒ | $\stackrel{m}{\Gamma}$ | $\bigcirc$ | ๒ | セู | $\stackrel{\text { ® }}{ }$ | 앙 | q |  | $\stackrel{6}{V}$ | ก | g |
|  | 웅 | $?_{i}^{2}$ | $\stackrel{\text { ® }}{ }$ | 8 | 안 | $\stackrel{\text { V }}{ }$ | ท | － | \％ | 응 | $\stackrel{1}{6}$ | ๑๐ | $\stackrel{\text { ® }}{ }$ | ๑๐ | $\stackrel{1}{*}$ | $\stackrel{\circ}{0}$ | V | レ | $\stackrel{\infty}{0}$ | $\stackrel{1}{0}$ | $\bigcirc$ | N | $\stackrel{10}{ }$ | $\stackrel{\sim}{\sim}$ | ท | $\stackrel{\text { ® }}{ }$ | $\stackrel{\sim}{-}$ | 8 | \％ | $\stackrel{10}{ }$ | 8 |  |
|  | 문 | $\overbrace{n}^{n}$ | $\stackrel{1}{v}$ | v | ¢ | レ0 | ท | ¢ | の | 予 | レ | ท | ® | ท | レ | $\stackrel{0}{0}$ | ง | $\stackrel{\text { ® }}{ }$ | б | $\stackrel{10}{ }$ | $\stackrel{\text { V }}{ }$ | $\stackrel{\square}{\square}$ | $\stackrel{\text { ® }}{ }$ | $\stackrel{\text { ® }}{ }$ | ท | へ | $\stackrel{10}{ }$ | $\bigcirc$ |  | $\stackrel{6}{V}$ | $\pm$ |  |
|  | $\begin{aligned} & \text { 号 } \\ & \text { 오 } \end{aligned}$ | $\overbrace{0}^{0}$ | $\stackrel{\sim}{\mathrm{v}}$ | ～ | レ⁄ | $\stackrel{\sim}{\mathrm{v}}$ | ¢ | $\infty$ | 下 | \％ | レ | ๑ | ® | ๑๐ | $\stackrel{\square}{\mathrm{v}}$ | $\stackrel{\text { v }}{ }$ | $\stackrel{\sim}{V}$ | $\stackrel{\text { ® }}{ }$ | N | $\stackrel{10}{ }$ | V | g | $\stackrel{10}{ }$ | v | ท | ® | V | 2 | \＆ | $\stackrel{1}{v}$ | 5 |  |
|  | $\begin{aligned} & 0 \\ & 0 \\ & 0 \end{aligned}$ | is | レ0 | V | V | レ0 | ท | レ0 | ¢ | ¢ | － |  | งจ | ทิ | － | ท | v | セ | ¢ | ๑ | $\stackrel{\text { ® }}{ }$ | ท | $\stackrel{1}{0}$ | $\stackrel{10}{ }$ | ท | $\stackrel{\text { ® }}{ }$ | $\stackrel{10}{ }$ | V | ํㅗㄹ | V | $\wedge$ |  |
|  |  | $\stackrel{-}{v}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\text { V }}{ }$ | $\stackrel{\square}{v}$ | $\frac{0}{v}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\frac{\circ}{v}$ | $\stackrel{\text { V }}{ }$ | \％ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\text { 안 }}{ }$ | $\stackrel{\text { O}}{V}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ |  | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\sim}{\circ}$ | $\stackrel{\circ}{\mathrm{V}}$ | $\stackrel{\bigcirc}{v}$ | $\stackrel{\bigcirc}{\mathrm{v}}$ | $\stackrel{\text { V }}{ }$ | $\stackrel{\circ}{v}$ | ㅇ | $\stackrel{\circ}{v}$ | $\stackrel{\bigcirc}{\mathrm{v}}$ | － |  | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{V}}$ |  |
|  | $\stackrel{9}{5}$ |  | $\begin{aligned} & \frac{0}{y} \\ & \vdots \\ & \frac{0}{x} \\ & \frac{0}{2} \\ & \frac{0}{0} \\ & \hline \end{aligned}$ | $\begin{aligned} & \dot{y} \\ & \frac{y}{x} \\ & \frac{0}{n} \\ & 0 \\ & \hline \end{aligned}$ | $\begin{aligned} & \frac{1}{0} \\ & \frac{y}{x} \\ & \frac{1}{x} \\ & \frac{0}{0} \\ & 0 \\ & 0 \\ & 0 \\ & \hline 0 \end{aligned}$ | $\begin{aligned} & \frac{5}{0} \\ & \frac{0}{0} \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & 0 \\ & \hline \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |  | － |  |  | $\begin{aligned} & { }_{3}^{j} \end{aligned}$ | $\begin{aligned} & \frac{\omega}{4} \\ & \stackrel{y}{0} \\ & \stackrel{c}{5} \end{aligned}$ | $\underset{\Sigma}{\stackrel{e}{\sim}}$ |  |  |  |  |  |  |



|  | 칭 | 8 | $\stackrel{\sim}{1}$ | ®8 | m | ～ス | － | ल | ¢ ¢ | ¢ ¢ | へ | Z | $\stackrel{\infty}{\sim}$ |  | $\bigcirc$ in | \％ | セ⁄ セ8 | ¢ ${ }^{\text {P }}$ |  | N | \％ |  | $\pm$ |  | ¢ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| F | 흥 | ？ | N | \％ |  | ふi¢ | 令 |  | $\stackrel{0}{\text { en }}$ | ¢ | $\stackrel{8}{\square} \stackrel{0}{\square}$ | Z | $\stackrel{\infty}{\sim} \stackrel{\infty}{\square}$ | $\stackrel{\circ}{\sim} \stackrel{\circ}{\circ}$ | $\stackrel{\circ}{\circ} \stackrel{\circ}{N}$ |  | $\underset{\sim}{9}$ | $\stackrel{\text { N}}{\sim}$ | $\infty$ | 年 | 악 | N | $\stackrel{セ 8}{6}$ |  | ก |
| ら | 힝 | N | N | \％ | $\bar{\sim}$ | $\wedge \stackrel{\text { ® }}{ }$ | $\stackrel{\infty}{\sim}$ | $\stackrel{\square}{\square}$ | 웅 | $\stackrel{\sim}{\square}$ | $\stackrel{m}{\square}$ | $z$ | $\bigcirc$ | ¢ | $\bigcirc \infty$ | 은 | ® | N | $\stackrel{\sim}{\sim}$ |  | \％ |  | $\bigcirc$ | $\stackrel{\sim}{\square}$ | $\odot$ |
| ¢ | 힝 | $\frac{1}{2} \stackrel{\rightharpoonup}{2} \stackrel{\circ}{v}$ | 은 | $\stackrel{\circ}{\mathrm{v}}$ | $\bigcirc$ | $\stackrel{\circ}{\mathrm{v}} \stackrel{0}{\mathrm{v}}$ | $\stackrel{\circ}{\mathrm{v}} \stackrel{\circ}{\mathrm{v}}$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | $\bigcirc$ | Z | $\stackrel{\circ}{\mathrm{v}} \stackrel{ }{\mathrm{v}}$ | $\bigcirc$ | $\stackrel{\circ}{\mathrm{V}} \stackrel{\circ}{\mathrm{v}}$ | $\bigcirc$ | $\frac{O}{v} \frac{o}{v}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\circ}{v}$ | $\bigcirc \stackrel{\circ}{\mathrm{v}} \stackrel{\mathrm{O}}{\mathrm{v}}$ | $\stackrel{\square}{\mathrm{v}}$ |  | $\frac{\square}{v}$ |  | $\stackrel{\text { v }}{ }$ |
| ¢ | 칭 | ） | จ ¢ | $\stackrel{6}{0}$ | V | $\stackrel{\text { V }}{ }$ | $\stackrel{\text { v }}{ }$ | จ | V | セ®® | ข゚レ | Z | v | จ เ๑ | レ ¢ | v | V | $\stackrel{\sim}{v}$ | $\stackrel{\square}{\mathrm{v}}$ | v ¢ | ท | v | セ | V | v |
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| z | $\bigcirc$ | $\begin{aligned} & \overline{0} \\ & 0 \\ & 0 \end{aligned}$ | $\hat{S}_{\substack{\circ}}^{\substack{0 \\ 0}}$ | $\begin{array}{ll} 0 & n \\ \\ & 0 \\ 0 \end{array}$ | $\stackrel{N}{0}$ |  | ${ }^{0}$ | $\begin{aligned} & \dot{0} \end{aligned}$ | $\stackrel{S}{0}_{\substack{0 \\ 0 \\ 0 \\ 0}}$ | $\stackrel{\circ}{\circ} \mathrm{O}$ | $\begin{array}{ll} \mathrm{O} \\ 0 \\ \mathrm{O} \\ \hline \end{array}$ | z | $\begin{aligned} & \text { M } \\ & \hline 0.0 \\ & 0 \\ & \hline \end{aligned}$ | $0 .$ | $\begin{aligned} & 0 \\ & \hline 0 \\ & \hline 0 \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { So } \\ & \hline \end{aligned}$ | $\mathrm{O}_{\mathrm{O}}^{\circ}$ | ©. | O8 | $8$ | $\begin{aligned} & 2 \\ & \hline \end{aligned}$ |  | $\stackrel{\text { O}}{\substack{0}}$ | $0$ | $\stackrel{\square}{\square}$ |
| $\sum$ | 잉 | v | v | － $\bar{v}$ | v | v v | v v | v | ָ | $\overline{\mathrm{v}}$ v | v | K | v | v | $\bar{v} \bar{v}$ | － | v | $\overline{\mathrm{v}}$ V | v | v v | $\overline{\mathrm{v}}$ |  | $\bar{v}$ |  | v |
| $\sum_{\Sigma}^{5}$ | 힝 | $\hat{2}_{2}^{2} \stackrel{N}{f}$ | $\underset{\gamma}{\underset{\gamma}{\sim}}$ | $\stackrel{8}{子}$ | $\mathscr{\infty}$ | テ | $\stackrel{\circ}{\circ}$ | $\stackrel{\leftrightarrow}{0}$ | $\stackrel{\circ}{3} \stackrel{\circ}{\mathrm{~N}}$ | $\stackrel{\infty}{\circ} \stackrel{\circ}{\circ}$ | Fi | Z | ¢ | $\circ \div$ | $8$ | $\underset{\circ}{\mathrm{O}}$ | $\stackrel{\circ}{\sim}$ | $\stackrel{\circ}{\circ}$ | 尔 | \％ | 大 | N | ก | er | ® |
| $\Sigma$ | $\bigcirc$ | $\stackrel{\varrho}{\stackrel{\circ}{r}}$ | $\underset{\sim}{f} \underset{\sim}{\circ}$ | No | co | $$ |  |  | $0$ | $\stackrel{\substack{\mathrm{N}}}{\stackrel{\circ}{\mathrm{O}}}$ |  | Z | $\stackrel{\square}{\circ}$ |  | $\underset{0}{ \pm} \underset{\sim}{\circ}$ | $\begin{aligned} & 0 \\ & j \end{aligned}$ | $\stackrel{\infty}{\infty}$ | ${ }_{6}^{\circ} \mathrm{O}$ | mi |  | N |  |  | $\stackrel{8}{\circ}$ |  |
| צ | $\bigcirc \bigcirc$ | $\underset{\substack{\text { N } \\ \sim}}{ }$ | $\underset{\sim}{N}$ | $\underset{0}{0}$ |  | $\begin{array}{ll} \circ \\ \hline \end{array}$ | $\begin{gathered} 0 \\ 0 \\ 0 \\ \hline 0 \end{gathered}$ |  | $1$ | $\begin{aligned} & \text { O. N } \\ & 0 \\ & 0 \end{aligned}$ | $\begin{array}{ll} 0 \\ 0 \\ 0 & 0 \\ 0 \end{array}$ |  |  |  | $\stackrel{N}{\mathrm{~N}}$ |  | do | $\underset{\sim}{N}$ | St |  | $\stackrel{\circ}{\circ}$ |  |  | $\stackrel{\text { º }}{ }$ | 8 |
| $\stackrel{\circ}{\sim}$ | $\bigcirc$ | $\circ \frac{m}{c}$ | $\underset{r}{\stackrel{m}{r}} \underset{\dot{\sigma}}{\bar{m}}$ | $\underset{\sim}{\sim} \underset{\sim}{c}$ | $\stackrel{m}{\sim}$ |  |  |  | $\stackrel{\text { ¢ }}{\sim}$ | $\stackrel{N}{N}$ | $\stackrel{\text { m }}{\text { c }}$ | z | －0 |  | $\stackrel{i}{-}$ | $\xrightarrow[\substack { \infty \\ \begin{subarray}{c}{\infty \\ \\ \hline{ \infty \\ \begin{subarray} { c } { \infty \\ \\ \hline } } \\ {\hline} \\ {\hline}\end{subarray}]{ }$ | $\stackrel{\&}{\square}$ | ® | $\underbrace{\infty}_{i}$ |  | N／ |  | $\stackrel{\circ}{\circ}$ | $\stackrel{\text { m}}{\text { i }}$ | － |
| $\bigcirc$ | $\frac{\varepsilon}{a}$ | ）${ }^{\circ}$ | \％ | ホ | 암 | の® | $\stackrel{\text { O}}{\text { N }}$ | $\stackrel{\square}{\square}$ | $\bigcirc$ | ¢ 9 | 운 | z | の | － | \％セ ロ | ¢ | $\bar{m}$ | ＋ | $\stackrel{\sim}{\sim}$ | $\infty$ | m | \％ | $\pm$ | ๓ | $\stackrel{\sim}{\circ}$ |
| － | 틍 | 친 | － | $\stackrel{\infty}{\sim}$ | $\stackrel{\sim}{\sim}$ | $\stackrel{\text { ® }}{ }$ | ®の | $\stackrel{\infty}{\sim}$ | ロー | －98 | ก N | Z | N | ल | 운 | $\stackrel{ }{ }$ | $\stackrel{\square}{\square}$ | N | $\stackrel{ }{\sim}$ | N | ̇ | N | $\infty$ | ̇ | $\stackrel{\square}{\circ}$ |
| $\bigcirc$ | 층 | ${ }^{\circ}$ | － | 0 | ＋ | －${ }_{-}$ | に m | 入 | － | $\bigcirc$ | の～ | Z | $\checkmark$ | ＾ | N | N | $\bigcirc$ | の | $\bigcirc$ | $\cdots$ | $\stackrel{m}{\square}$ |  | m |  |  |
| O | 딤 |  | $\stackrel{\sim}{v} \stackrel{N}{\circ}$ | $\dot{5} \stackrel{n}{0}$ | $\stackrel{0}{v}$ |  | $\begin{array}{cc} n \\ \stackrel{n}{n} \\ \stackrel{0}{0} \\ \stackrel{\rightharpoonup}{0} \end{array}$ |  | $\mathfrak{c}$ | $$ | $\begin{array}{lll} n & n \\ \stackrel{n}{0} \\ \stackrel{\rightharpoonup}{0} \\ 0 \end{array}$ | $\mathbb{Z}$ |  | $\stackrel{\rightharpoonup}{v}$ |  | $\underset{\sim}{?} \stackrel{n}{0}$ | $\begin{array}{lll} n & 1 \\ \stackrel{L}{2} \\ \stackrel{v}{v} & \mathrm{v} \end{array}$ | $\begin{aligned} & 0 \\ & 0 \\ & 0 \end{aligned}$ | $01$ | v | v | v | $\stackrel{0}{0}$ | $\stackrel{0}{\circ}$ | $\stackrel{0}{0}$ |
| © | $\bigcirc$ | $0$ | $\underset{\substack{t \\ \hline \\ \hline \\ \hline \\ \hline \\ \hline \\ \hline}}{ }$ | $\stackrel{\text { O}}{\substack{0}}$ | $\stackrel{0}{0}$ | $\stackrel{\circ}{0} \underset{0}{\circ}$ | $\stackrel{\infty}{\stackrel{\infty}{\sim}} \stackrel{\sim}{0}$ | $\stackrel{\circ}{\circ}$ | $\begin{aligned} & \text { In } \\ & \text { did } \\ & \hline \end{aligned}$ | $\underset{\sim}{\mathrm{O}} \underset{\sim}{\underset{\sim}{N}}$ | $\stackrel{\circ}{\circ}$ | $\Sigma$ | $0^{\circ}$ |  | M. | $\stackrel{\substack{\underset{\sim}{2} \\ \hline}}{ }$ | $\stackrel{m}{\circ}$ | $\stackrel{\circ}{\circ}$ |  |  | － |  | N |  | － |
| ¢ | $\frac{\varepsilon}{2}$ | 칠 | V | v | V | $\bigcirc$ |  | V | v | ๑ู | レ®ソ | z | $\stackrel{\text { ® }}{ }$ | ง | ๑ง | V | ท | ๒ | ท ¢ | จ ¢ | V | v | ๒ | v | $\stackrel{\sim}{*}$ |
| $\stackrel{\oplus}{\infty}$ | 츨 | 칠 | \％ | ${ }^{\circ}$ | $\bigcirc$ | $\stackrel{m}{0} 0 \stackrel{0}{0}$ | $\mathfrak{j} \underset{O}{\sim}$ | ${ }_{0}^{\text {J }}$ | － | ${ }_{\circ}^{+}{ }_{0}^{\infty}$ | No Nó | Z | Nom | $\bigcirc \bigcirc$ | $\stackrel{\circ}{\circ} \stackrel{0}{\circ}-$ | － | No | ${ }_{0}^{\infty}$ | Mo co | N | ¢ | $\bigcirc$ | ${ }_{0}^{\circ}$ | $\bigcirc$ | $\stackrel{\text { o }}{ }$ |
| ¢ | $\begin{gathered} \varepsilon \\ \stackrel{\vdots}{\alpha} \\ \hline \end{gathered}$ | Ean | － | $\stackrel{\text { ¢ }}{+}$ | 꾼 | $\bigcirc$ | ¢ | $\bar{\square}$ | $\bigcirc$ |  | ¢ ${ }_{\circ}^{\infty}$ | Z | 边 | $\stackrel{ }{\sim}$ | ¢ ¢ | ल | － | \％ | ¢ ${ }^{\circ}$ | \％ | is |  | あ | $\stackrel{\infty}{\sim}$ | $\stackrel{\infty}{\sim}$ |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  | Little Cranberry River |  |  |  |  |  |  |  |  |  |


| Tributary | Y | Zn | Mercury | Total C | Inorganic C | Organic C | LOI |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units | ppm | ppm | ppm | \% | \% | \% | \% |
| Bluff Creek | 13.7 | 91 | 0.051 | 6.2 | 0.4 | 5.7 | 12.2 |
| Brule River | 20.5 | 147 | 0.091 | 10.2 | 0 | 10.1 | 23 |
| Carp River | 7.9 | 56 | 0.069 | 5.5 | 0.1 | 5.4 | 11.3 |
| Chocolay River | 6.9 | 64 | 0.055 | 9.8 | 0.1 | 8.7 | 20.6 |
| Cranberry Slough | 6 | 20 | 0.009 | 1.4 | 0 | 1.4 | 3.3 |
| Deviltrack River | 25.3 | 84 | 0.05 | 6.3 | 0.1 | 6.2 | 13.5 |
| Eagle River | 8.8 | 64 | 0.096 | 5.6 | 0.1 | 5.4 | 11.9 |
| Falls River | 3.7 | 26 | 0.024 | 1.7 | 0.1 | 1.6 | 3.4 |
| Firesteel River | 11.3 | 33 | 0.024 | 2 | 0 | 2 | 4.4 |
| Flag River | 7.4 | 26 | 0.019 | 2.6 | 0.1 | 2.6 | 5.7 |
| Flintsteel River | 9.2 | 107 | 0.027 | 2.4 | 0.1 | 2.4 | 5.2 |
| Gooseberry River | 15.7 | 91 | 0.041 | 5.6 | 0.2 | 5.5 | 13.6 |
| Gratiot River | 7.4 | 40 | 0.033 | 4.6 | 0 | 4.6 | 9.8 |
| Grand Marais Creek | 5.1 | 43 | 0.115 | 20.8 | 0.5 | 20.3 | 42.6 |
| Cedar Creek | NA | NA | 0.082 | 8.6 | 0.1 | 5.4 | 11.3 |
| Huron River | 7.6 | 36 | 0.066 | 2.5 | 0 | 2.5 | 5.3 |
| Knife River | 7.4 | 41 | 0.034 | 5.5 | 0.2 | 5.3 | 12.9 |
| Knowlton Creek | 12.9 | 68 | 0.044 | 1.5 | 0.2 | 1.3 | 3.7 |
| Laughing Whitefish River | 11.7 | 58 | 0.025 | 5.4 | 0.2 | 5.3 | 11.1 |
| Little Carp River | 5 | 26 | 0.026 | 3.4 | 0.1 | 3.4 | 8 |
| Little Cranberry River | 8.3 | 1790 | 0.017 | 0.9 | 0 | 0.9 | 2.3 |
| Little Iron River | 7.1 | 25 | 0.008 | 0.6 | 0 | 0.6 | 1.9 |
| Mineral River | 14.2 | 80 | 0.045 | 2 | 0.2 | 1.7 | 4.2 |
| Miners River | 4.7 | 70 | 0.139 | 13.9 | 0.4 | 13.6 | 28.6 |
| Misery River | 6.6 | 19 | 0.012 | 1.8 | 0.1 | 1.7 | 3.7 |
| Nemadji River | 8.5 | 37 | 0.015 | 1.1 | 0.4 | 0.7 | 1.5 |
| Onion River | 10.6 | 60 | 0.027 | 3.2 | 0 | 3.2 | 7.3 |
| Ontonagon River | 13.2 | 70 | 0.122 | 2.2 | 0.2 | 1.9 | 4.3 |
| Pike River | 5 | 29 | 0.036 | 2.2 | 0 | 2.2 | 4.6 |
| Pokegama River | 8.9 | 61 | 0.019 | 0.9 | 0.2 | 0.7 | 1.8 |
| Potatoe River | 11.3 | 54 | 0.019 | 1.3 | 0 | 1.2 | 2.8 |


| Tributary | Latitude | Longitude | Sampling Date | HCB | Endrin aldehyde | OCS | Toxaphene | Aldrin | $\alpha \mathrm{HCH}$ | $\beta \mathrm{HCH}$ | $\delta \mathrm{HCH}$ | Lindane |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units |  |  |  | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ |
| Ravine River | 46.8401 | -88.2513 | 26-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Rock River | 46.4637 | -86.9161 | 05-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Sand River | 46.4947 | -87.1078 | 05-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.002 |
| Shelldrake River | 46.6752 | -85.0335 | 06-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Silver River | 47.4629 | -88.0725 | 20-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Silver River B | 46.8162 | -88.2981 | 26-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | $<0.002$ | <0.002 |
| Sioux River | 46.7301 | -90.8992 | 23-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Splitrock Creek | 47.1831 | -91.4092 | 18-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | $<0.002$ | <0.002 | <0.002 | $<0.002$ | <0.002 |
| St Louis River | 46.6534 | -92.2263 | 25-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 |
| Sturgeon River | 47.0175 | -88.5066 | 19-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Sturgeon River Slough | 47.0126 | -88.4968 | 19-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Sucker River | 46.6617 | -85.869 | 06-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Tahquamenon River | 46.562 | -85.0311 | 06-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Temperance River | 47.5545 | -90.8746 | 17-Sep-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Waiskey River | 46.3971 | -84.5326 | 18-Oct-05 | <0.002 | <0.002 | <0.002 | <0.08 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |


| Tributary | $\alpha$ Chlordane | Y Chlordane | Total Chlordane | op DDD | pp DDD | op DDE | pp DDE | op DDT | pp DDT | Total DDD | Total DDE | Total DDT |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ |
| Ravine River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Rock River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Sand River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Shelldrake River | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Silver River | $<0.002$ | $<0.002$ | <0.002 | <0.002 | 0.01 | <0.002 | 0.028 | 0.009 | 0.008 | 0.01 | 0.028 | 0.017 |
| Silver River B | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Sioux River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Splitrock Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| St Louis River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Sturgeon River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Sturgeon River Slough | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Sucker River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Tahquamenon River | <0.002 | <0.002 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Temperance River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |
| Waiskey River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 |


| Tributary | Total DDT Metabolites | Dieldrin | Endosulfan I | Endosulfan II | Endosulfan Sulfate | Endrin | Heptachlor | Heptachlor Epoxide | Methoxychlor | Mirex | Aroclor 1016 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ |  | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ | $\mu \mathrm{g} / \mathrm{g}$ |
| Ravine River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Rock River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Sand River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Shelldrake River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Silver River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Silver River B | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Sioux River | <0.002 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Splitrock Creek | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| St Louis River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Sturgeon River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Sturgeon River Slough | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Sucker River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Tahquamenon River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Temperance River | <0.002 | $<0.002$ | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |
| Waiskey River | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.002 | <0.01 |


|  | $\begin{aligned} & 0 \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\stackrel{\circ}{V}$ | $\stackrel{10}{ }$ | $\stackrel{\square}{\text { ® }}$ | $\stackrel{10}{ }$ | レ0 | $\stackrel{10}{ }$ | ท | เจ | ๑レヤ | $\stackrel{10}{ }$ | $\stackrel{10}{ }$ | レ0 | $\stackrel{\square}{\text { ® }}$ | $\stackrel{10}{ }$ | ๑ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | $\stackrel{\bullet}{v}$ | $\stackrel{\circ}{V}$ | $\stackrel{\text { ® }}{ }$ | $\stackrel{10}{ }$ | レ0 | $\stackrel{\text { ® }}{ }$ | ท | ® | $\bigcirc$ | $\stackrel{10}{ }$ | $\stackrel{\text { ® }}{ }$ | レ | レก | レ | ท |
|  |  |  | $\stackrel{\bar{c}}{\text { V }}$ | $\begin{gathered} \bar{\circ} \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ | $\begin{aligned} & \bar{o} \\ & 0 \\ & 0 \\ & \hline \end{aligned}$ | $\begin{aligned} & \text { ס. } \\ & \text { V } \end{aligned}$ | $\begin{aligned} & \bar{o} \\ & \stackrel{\rightharpoonup}{v} \end{aligned}$ |  |  |  |  |  |  | $\begin{gathered} \bar{o} \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ | $\begin{gathered} \overleftarrow{O}_{\dot{\circ}} \end{gathered}$ |  |
|  |  | $\begin{array}{ll} \substack{0 \\ \hline \\ 2} & \overline{0} \\ \hline \end{array}$ | $\begin{gathered} \overline{0} \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ | $\begin{gathered} \overline{0} \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ | $\begin{aligned} & \overline{0} \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & \dot{C}_{\dot{0}} \\ & \mathrm{~V} \end{aligned}$ | $\begin{aligned} & \dot{o} \\ & \stackrel{\rightharpoonup}{0} \end{aligned}$ | $\begin{aligned} & \stackrel{\rightharpoonup}{\mathrm{o}} \\ & \stackrel{\rightharpoonup}{\mathrm{O}} \end{aligned}$ | $\stackrel{\rightharpoonup}{\stackrel{\rightharpoonup}{c}} \stackrel{\rightharpoonup}{\mathrm{o}}$ | $\begin{gathered} c_{i} \\ \stackrel{\rightharpoonup}{0} \\ \hline \end{gathered}$ | $\begin{aligned} & \overline{0} \\ & \stackrel{\rightharpoonup}{v} \end{aligned}$ | $\begin{gathered} \overline{0} \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ | $\begin{gathered} \bar{\circ} \\ \stackrel{\rightharpoonup}{0} \end{gathered}$ | $\begin{gathered} \overline{0} \\ \dot{0} \end{gathered}$ | $\begin{aligned} & \bar{o}_{\dot{~}} \end{aligned}$ | ¢ |
| $\begin{aligned} & \text { H} \\ & \stackrel{N}{N} \\ & \text { 흠 } \\ & \text { 우 } \end{aligned}$ |  | $\begin{array}{l\|l} \substack{0 \\ \hline \\ 2} & \overline{0} \\ \hline \end{array}$ | $\stackrel{\bar{c}}{\text { V }}$ | ס | $\begin{aligned} & \bar{\delta} \\ & 0 \\ & 0 \end{aligned}$ |  | $\begin{aligned} & \bar{\circ} \\ & \dot{\circ} \end{aligned}$ |  | $\stackrel{\rightharpoonup}{\stackrel{\rightharpoonup}{v}} \stackrel{\rightharpoonup}{\mathrm{o}}$ | $\begin{gathered} \stackrel{\rightharpoonup}{c} \\ \stackrel{\rightharpoonup}{0} \\ \hline \end{gathered}$ | $\begin{aligned} & \bar{o} \\ & \stackrel{\rightharpoonup}{0} \end{aligned}$ | $\begin{gathered} \bar{o} \\ \stackrel{\rightharpoonup}{0} \end{gathered}$ | $\begin{aligned} & \overline{0} \\ & \dot{\circ} \end{aligned}$ | $\begin{gathered} \overline{0} \\ \dot{v} \end{gathered}$ |  | $\begin{aligned} & \dot{o} \\ & \dot{O} \\ & \dot{O} \end{aligned}$ |
|  |  | $\begin{array}{ll} 0 \\ 0 & \vdots \\ 0 & 0 \\ 0 \end{array}$ | $\begin{aligned} & \bar{o} \\ & \dot{0} \end{aligned}$ | $\begin{aligned} & \bar{o} \\ & \dot{0} \end{aligned}$ | $\begin{aligned} & \bar{\delta} \\ & 0 \\ & 0 \end{aligned}$ |  | $\begin{aligned} & \dot{o} \\ & \stackrel{\rightharpoonup}{0} \end{aligned}$ |  |  |  | $\begin{aligned} & 5 \\ & \stackrel{\rightharpoonup}{x} \\ & \stackrel{\rightharpoonup}{x} \end{aligned}$ | $\begin{gathered} \bar{o} \\ \stackrel{\rightharpoonup}{0} \end{gathered}$ | $\begin{aligned} & 5 \\ & \hline \\ & \hline \\ & 0 \\ & 0 \end{aligned}$ | $\begin{aligned} & \overline{0} \\ & \dot{0} \end{aligned}$ |  | $\begin{aligned} & 5 \\ & \dot{c} \\ & \dot{0} \\ & \hline \end{aligned}$ |
| N N 흥 운 |  | $\begin{array}{l\|l} 0 \\ 0 & \text { y } \\ 0 & 0 \\ 0 \end{array}$ | $\begin{gathered} \text { No } \\ \text { V } \end{gathered}$ | $\begin{gathered} \text { O} \\ \text { O } \end{gathered}$ | $\begin{aligned} & \text { N } \\ & 0 \\ & 0 \\ & 0 \end{aligned}$ | $\begin{gathered} \text { O} \\ \text { O } \end{gathered}$ | $\begin{gathered} \text { No } \\ \text { V } \end{gathered}$ | $\begin{gathered} \stackrel{N}{O} \\ \underset{\sim}{\circ} \\ \underset{\sim}{2} \end{gathered}$ |  | $\underset{\sim}{N}$ | $\begin{gathered} \underset{\sim}{N} \\ \underset{\sim}{\circ} \\ \underset{\sim}{\circ} \end{gathered}$ | $\begin{aligned} & \text { No } \\ & \substack{\circ \\ \text { O- }} \end{aligned}$ |  | $\begin{aligned} & \text { No } \\ & \text { O. } \end{aligned}$ |  | $\begin{gathered} \stackrel{N}{\circ} \\ \substack{\circ \\ \hline \\ \hline} \end{gathered}$ |
|  |  |  |  | $\begin{aligned} & \bar{o} \\ & \dot{0} \end{aligned}$ |  |  | $\begin{gathered} \overline{0} \\ \stackrel{\rightharpoonup}{0} \end{gathered}$ |  | V |  |  |  |  | $\begin{aligned} & \overline{0} \\ & \dot{\rightharpoonup} \end{aligned}$ |  | $\begin{aligned} & 5 \\ & \dot{c} \\ & \hline \end{aligned}$ |
|  |  | $\begin{aligned} & 0 \\ & 0 \\ & 1 \\ & 0 \\ & 0 \\ & \vdots \end{aligned}$ | $\begin{gathered} \text { O} \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ | $\begin{gathered} \text { O} \\ \stackrel{\rightharpoonup}{\mathrm{V}} \end{gathered}$ | $\begin{gathered} 0 \\ 0 \\ 0 \\ 0 \end{gathered}$ | $\begin{gathered} \text { O} \\ \stackrel{\rightharpoonup}{\mathrm{V}} \end{gathered}$ | $\begin{gathered} \text { ơ } \\ \stackrel{\rightharpoonup}{\mathrm{V}} \end{gathered}$ |  |  |  |  | $\begin{aligned} & 3 \\ & \substack{0 \\ 0 \\ 0 \\ \hline} \end{aligned}$ |  |  |  |  |
|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Tributary | Acenaphthene | Fluorene | Phenanthrene | Anthracene | Fluoranthene | Pyrene | Benzoanthracene | Chrysene | Benzobfluoranthene |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units | $\mu \mathrm{g} / \mathrm{kg}$ | $\mu \mathrm{g} / \mathrm{kg}$ | $\mu \mathrm{g} / \mathrm{kg}$ | $\mu \mathrm{g} / \mathrm{kg}$ | $\mu \mathrm{g} / \mathrm{kg}$ | $\mu \mathrm{g} / \mathrm{kg}$ | $\mu \mathrm{g} / \mathrm{kg}$ | $\mu \mathrm{g} / \mathrm{kg}$ | $\mu \mathrm{g} / \mathrm{kg}$ |
| Ravine River | <10 | <5 | <5 | <5 | <5 | <5 | <10 | <10 | <10 |
| Rock River | <10 | <5 | <5 | <5 | <5 | <5 | <10 | <10 | <10 |
| Sand River | <10 | <5 | <5 | <5 | <5 | <5 | <10 | <10 | <10 |
| Shelldrake River | <10 | <5 | <5 | <5 | 40 | 30 | 20 | <10 | 20 |
| Silver River | <10 | <5 | 20 | <5 | 60 | 50 | 30 | <10 | 40 |
| Silver River B | <10 | <5 | <5 | <5 | <5 | <5 | <10 | <10 | <10 |
| Sioux River | <10 | <5 | <5 | <5 | <5 | <5 | <10 | <10 | <10 |
| Splitrock Creek | <10 | <5 | 30 | <5 | 60 | 50 | <10 | <10 | 50 |
| St Louis River | <10 | <5 | 18 | <5 | 38 | 32 | 20 | 20 | 27 |
| Sturgeon River | <10 | <5 | 32 | 6 | 65 | 52 | 30 | 20 | 29 |
| Sturgeon River Slough | <10 | <5 | <5 | <5 | <5 | <5 | <10 | <10 | <10 |
| Sucker River | <10 | <5 | <5 | <5 | <5 | <5 | <10 | <10 | <10 |
| Tahquamenon River | <10 | <5 | <5 | <5 | <5 | <5 | <10 | <10 | <10 |
| Temperance River | <10 | <5 | <5 | <5 | <5 | <5 | <10 | <10 | <10 |
| Waiskey River | <10 | <5 | <5 | <5 | <5 | <5 | <10 | <10 | <10 |


| Tributary | Benzokfluoranthene | Benzoapyrene | Indeno123cdpyrene | Dibenzoahanthracene | Benzoghiperylene | Total PAH | Ag | Al | As |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units | $\mu \mathrm{g} / \mathrm{kg}$ | $\mu \mathrm{g} / \mathrm{kg}$ | $\mu \mathrm{g} / \mathrm{kg}$ | $\mu \mathrm{g} / \mathrm{kg}$ | $\mu \mathrm{g} / \mathrm{kg}$ | $\mu \mathrm{g} / \mathrm{kg}$ | ppm | \% | ppm |
| Ravine River | <10 | <5 | <20 | <20 | <20 | <20 | <0.1 | 0.18 | <5 |
| Rock River | <10 | <5 | <20 | <20 | <20 | <20 | <0.1 | 0.34 | <5 |
| Sand River | <10 | <5 | <20 | <20 | <20 | <20 | <0.1 | 0.41 | <5 |
| Shelldrake River | <10 | <5 | <20 | <20 | <20 | 110 | <0.1 | 0.11 | <5 |
| Silver River | <10 | 30 | <20 | <20 | <20 | 230 | <0.1 | 0.92 | <5 |
| Silver River B | <10 | <5 | <20 | <20 | <20 | <20 | <0.1 | 0.39 | <5 |
| Sioux River | <10 | <5 | <20 | <20 | <20 | <20 | <0.1 | 0.59 | <5 |
| Splitrock Creek | <10 | 30 | $<20$ | $<20$ | $<20$ | 220 | <0.1 | 1.06 | <5 |
| St Louis River | <10 | 17 | <20 | <20 | <20 | 172 | <0.1 | 0.46 | <5 |
| Sturgeon River | 10 | 21 | <20 | <20 | <20 | 265 | <0.1 | 0.49 | <5 |
| Sturgeon River Slough | <10 | <5 | <20 | <20 | $<20$ | <20 | <0.1 | 0.19 | <5 |
| Sucker River | <10 | <5 | <20 | <20 | <20 | <20 | <0.1 | 0.25 | <5 |
| Tahquamenon River | <10 | <5 | $<20$ | $<20$ | <20 | <20 | <0.1 | 0.26 | <5 |
| Temperance River | <10 | <5 | $<20$ | <20 | <20 | <20 | <0.1 | 1 | <5 |
| Waiskey River | <10 | <5 | $<20$ | $<20$ | $<20$ | $<20$ | <0.1 | 0.44 | <5 |


|  | 칭 | $\stackrel{\text { F－}}{\text {－}}$ | $\stackrel{\square}{\circ}$ |  | N | F | － | $\stackrel{\rightharpoonup}{\circ}$ | $\stackrel{+}{\sim}$ | $\pm{ }^{\wedge}$ | $\stackrel{\sim}{\infty}$ | － | $\stackrel{\stackrel{\rightharpoonup}{\circ}}{\stackrel{1}{2}}$ | m | m | 2 | $\stackrel{\sim}{8}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| $>$ | 틍 | $\stackrel{\square}{\sim}$ | $\stackrel{\square}{\square}$ | N | F | O | ¢ | \％ | $\bigcirc$ | \％ | － | $\bigcirc$ | $\stackrel{\rightharpoonup}{*}$ | $\sim$ | N |  | $\stackrel{\square}{\square}$ |
| F | $\overline{ }$ | ¢ | ब̀ | ळ |  | $\stackrel{\otimes}{\underset{\sim}{\circ}}$ | － | $\stackrel{\text { O}}{\mathbf{\gamma}}$ | 추N | $\underset{\sim}{2}$ | প্ల | $\underset{\sim}{N}$ | $\stackrel{\sim}{N}$ | $\bar{f}$ | － | O | $\stackrel{\bigcirc}{\mathrm{N}}$ |
| ら | $\stackrel{\varepsilon}{0}$ | － | $\bar{N}$ | $\bigcirc$ | m | え | N | $\stackrel{\square}{+}$ | m | $\pm$ | $\bigcirc$ | $\checkmark$ | $\infty$ | 人 | － | $\bigcirc$ | の |
| ¢ | 등 | $\left.\frac{1}{2} \right\rvert\, \frac{0}{2}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\text { V }}{ }$ | $\frac{\circ}{\mathrm{V}}$ | $\frac{\circ}{\mathrm{v}}$ | $\frac{0}{v} \frac{O}{v}$ | $\stackrel{\circ}{v}$ | $\frac{\circ}{v} \stackrel{0}{v}$ | $\frac{2}{v} \frac{O}{v}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{\circ}{\mathrm{v}}$ | $\stackrel{O}{v}$ | $\frac{\circ}{\mathrm{v}}$ | 안 | $\stackrel{\text { 은 }}{ }$ |
| क | 들 | ） | $\stackrel{1}{v}$ | V | V | $\stackrel{\square}{v}$ | V゚ | ๒ | V | $\stackrel{1}{v}$ | レ | レ | $\stackrel{1}{v}$ | $\stackrel{0}{ }$ | レ |  | $\stackrel{\sim}{\circ}$ |
| 요 | $\frac{\varepsilon}{0}$ | ） | の | ㅇ | の | $\stackrel{ }{ }$ | － | － | $\bigcirc$ | N | $\bigcirc$ | レ0 | $\llcorner$ | $\stackrel{\sim}{0}$ | の |  | $\stackrel{10}{ }$ |
| z | $\overline{\hat{0}}$ | ） | $\infty$ | $\checkmark$ | － | ¢ | ○の | N | № | $\stackrel{6}{\square}$ | 안 | $\checkmark$ | $\sim$ | m | N | N | $\wedge$ |
| z | $\bigcirc$ | $\bigcirc \begin{gathered} \circ \\ \circ \\ \stackrel{\rightharpoonup}{\mathrm{O}} \end{gathered}$ | $0$ | $0$ | $\stackrel{\rightharpoonup}{v}$ | $\stackrel{N}{0}$ | No |  |  | $0$ | $0$ | $\begin{gathered} \bar{O} \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ | $\begin{aligned} & \bar{o} \\ & \dot{0} \\ & \dot{v} \end{aligned}$ | $8$ | Ơ |  | O |
| ${ }^{\circ}$ | 층 | －$\overline{\mathrm{V}}$ | v | $\bar{v}$ | v | v | v | $\bar{v}$ | v | v | v | v | v | v | v |  | $\overline{\mathrm{v}}$ |
| $\sum^{¢}$ | 팅 |  | $\infty$ | ส | $\bigcirc$ | $\stackrel{m}{\wedge}$ | ${ }^{2} \stackrel{\infty}{6}$ | $\underset{\text { Nै }}{ }$ | $\underset{\sim}{\text { ci }}$ | $\stackrel{\circ}{n} \stackrel{\circ}{9}$ | $\widehat{\infty}$ | $5 \mathbb{N}$ | $\stackrel{\circ}{\circ}$ |  | \％ |  | $\stackrel{\sim}{\sim}$ |
| $\sum^{0}$ |  | － | $\stackrel{6}{\square}$ | － | $\hat{O}_{0}$ | $\stackrel{\bullet}{\square}$ | $\mathfrak{B}$ |  | $\underset{0}{\circ}{ }_{0}^{\infty}$ | $00$ | $\stackrel{\sim}{0}$ | $\underset{0}{\dot{O}}$ | O | $\underset{i}{3} \underset{\sim}{3}$ | N |  | $\stackrel{m}{0}$ |
| $צ$ | \％ | $\underset{O}{\mathrm{O}}$ | $0_{0}^{0}$ | － | $\begin{aligned} & { }_{5}^{2} \\ & \hline \end{aligned}$ | $\overbrace{0}^{\circ}$ | $00$ |  | $\underset{\sigma}{ }$ |  | $0$ | $\stackrel{0}{0}$ | N | ${ }_{5}^{5}$ | $\stackrel{0}{0}$ |  | ${ }^{\circ}$ |
| $\stackrel{\text { ® }}{ }$ | $\bigcirc$ | $\stackrel{6}{\square}$ | $\stackrel{+}{+}$ | $\stackrel{\square}{+}$ | － | $\stackrel{\stackrel{\sim}{\mathrm{N}}}{ }$ | $\stackrel{\sim}{\sim}$ |  | $\stackrel{\infty}{\square}$ | $\stackrel{8}{\mathrm{P}} \stackrel{\infty}{\mathrm{o}}$ | 炧 | － | $\underset{\sim}{\text { o }}$ | $\stackrel{\infty}{\circ}$ | $\stackrel{\sim}{\infty}$ |  | $\stackrel{\otimes}{\circ}$ |
| 3 | 팅 | ） | ก | N | $\sim$ | $\stackrel{\square}{\square}$ | $\stackrel{\sim}{\sim}$ |  | ミ | む | $\pm$ | $\circ$ | － | ल | ल |  | $\bigcirc$ |
| $\bar{\square}$ | 팅 | $\sim$ | の | $\bullet$ | $\bigcirc$ | ¢ | さ |  | N | $\stackrel{\infty}{\sim}$ | $\stackrel{\square}{\bullet}$ | － | $\wedge$ | － | $\stackrel{\sim}{\sim}$ |  | $\stackrel{m}{\sim}$ |
| $\bigcirc$ | 등 | m | F | $\checkmark$ | － | $\pm$ | ๒ ↔ |  | $\bigcirc$ | 入 | $\bigcirc$ | N | N | m | $\bigcirc$ |  | \％ |
| J | 팅 |  | $0$ | ${ }^{\circ}$ | $\stackrel{0}{v}$ |  | $\begin{array}{ll} n \\ \stackrel{n}{n} \\ \stackrel{y}{0} \\ 0 \end{array}$ |  | $\stackrel{n}{\bullet} \stackrel{n}{\bullet} \stackrel{n}{0}$ |  | $\stackrel{0}{v}$ | $0$ | $\stackrel{0}{2} \stackrel{0}{2}$ | $? \stackrel{\leftrightarrow}{2}$ |  |  | $\stackrel{\sim}{\square}$ |
| O็ | $\bigcirc$ | $0 \frac{n}{0}$ | $\stackrel{6}{\circ}$ | － | N- | $\underset{\sim}{c} \stackrel{\sim}{\sim}$ | $\stackrel{\sim}{\circ}$ |  | ${ }^{3} 0$ |  | $\circ$ | $\stackrel{6}{\circ}$ | $\stackrel{\stackrel{\rightharpoonup}{\dot{r}}}{ }$ | N |  |  | $\stackrel{\square}{\square}$ |
| ¢ | $\frac{\varepsilon}{a}$ | － | レ | ソ | V | $\stackrel{1}{0}$ | จソ | ท | จ ท | ） 0 | V | ๒ | V | ソ | V | 0 | $\stackrel{\bullet}{\vee}$ |
| 毋 | 름 |  | $\underset{\sim}{2}$ | 0 | $\stackrel{\text { N}}{\stackrel{\circ}{v}}$ | $\stackrel{y}{v}$ | $\overbrace{0}$ | مٌ | $1$ | $\bigcirc$ | $\underset{0}{\star}$ | $\begin{gathered} \text { Ň } \\ \stackrel{\rightharpoonup}{v} \end{gathered}$ | $\begin{gathered} \text { Ň } \\ \text { V } \end{gathered}$ | $\underset{\sim}{v} \underset{\sim}{v}$ |  |  | N |
| ¢ ¢ |  | － | N | $\overline{0}$ | ～ | ＊ | ）通 | ת | 欠 | $\stackrel{\circ}{\circ}$ | $\infty$ | － | ̇ |  | ¢ |  | 8 |
|  | $\begin{aligned} & 0 \\ & \hline 10 \\ & \hline \end{aligned}$ | $\begin{array}{r\|r\|} \hline \end{array}$ | $\begin{aligned} & \bar{y} \\ & \underset{x}{x} \\ & \stackrel{y}{0} \\ & \dot{x} \end{aligned}$ | $\begin{aligned} & \mathbf{D}_{0}^{0} \\ & 0 \\ & 0 \end{aligned}$ |  |  |  |  |  |  |  |  |  |  |  |  |  |


| Tributary | Zn | Mercury | Total C | Inorganic C | Organic <br> C | LOI |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Units | ppm | ppm | \% | \% | \% | \% |
| Ravine River | 34 | 0.018 | 1 | 0 | 1 | 2.3 |
| Rock River | 73 | 0.086 | 9.9 | 0.2 | 9.7 | 19.3 |
| Sand River | 46 | 0.06 | 5 | 0 | 5 | 10 |
| Shelldrake River | 17 | 0.009 | 1.9 | 0.1 | 1.8 | 3.5 |
| Silver River | 56 | 0.049 | 5.2 | 0.1 | 5.1 | 11.2 |
| Silver River B | 45 | 0.033 | 3.3 | 0 | 3.3 | 6.8 |
| Sioux River | 36 | 0.024 | 3.4 | 0.1 | 3.3 | 7 |
| Splitrock Creek | 80 | 0.042 | 4.6 | 0 | 4.6 | 10.6 |
| St Louis River | 81 | 0.02 | 1.7 | 0.2 | 1.6 | 3.5 |
| Sturgeon River | 34 | 0.029 | 1.9 | 0 | 1.9 | 4.5 |
| Sturgeon River Slough | 15 | 0.011 | 1.1 | 0 | 1.1 | 2.6 |
| Sucker River | 28 | 0.028 | 6.2 | 0 | 6.2 |  |
| Tahquamenon River | 34 | 0.025 | 3 | 0 | 3 | 6.2 |
| Temperance River | 79 | 0.039 | 3.9 | 0 | 3.9 | 9 |
| Waiskey River | 34 | 0.02 | 2.5 | 0.1 | 2.4 | 5.5 |

