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THE INFLUENCE OF PIPELINE CONSTRUCTION UPON ASPECTS
OF WATER QUALITY OF THE OGILVIE AND SWIFT RIVERS,
YUKON TERRITORY

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TABLE OF CONTENTS

	<u>Page</u>
Abstract	1
Introduction	2
Materials and Methods	3
Results and Discussion	9
Literature Cited	20
List of Tables	24
Appendix 1	42
Appendix 2	54
Appendix 3	68
Appendix 4	72

ABSTRACT

The sub-Arctic Ogilvie and Swift Rivers are characterized by annual cycles of microalgal and bacterial biomasses and activities which are maximal in Spring following ice "break up" and decrease for the remainder of the year such that minimal values are noted in late Winter. Levels of algal biomasses and photosynthetic rates appear to be regulated by available light whereas biomasses and activities of heterotrophic bacteria are probably controlled by DOC content of the water.

Perturbation of these two lotic systems by streamside material (of concentrations of 100 ppm or greater) (1) decreases algal productivities (2) increases dark productivities (3) increases bacterial heterotrophic activities . . (4) increases biological oxygen demand and (5) overlayers and smothers benthic micro-organisms.

The seasonal order of sensitivity of the microflora of the Ogilvie and Swift Rivers to streamside and sediment additions appears to be Winter > Fall > Summer > Spring.

INTRODUCTION

Construction of a natural gas pipeline which generally parallels the route of the Alaska Highway across the Yukon Territory is due to commence in the early 1980's. Since this pipeline will cross numerous rivers, streams and creeks, the trenching and backfilling associated with its construction will probably stress each biological aquatic ecosystem through amongst other things (1) increased silt concentrations in waters and on stream bottoms and (2) increased organic and inorganic water loadings.

To adequately assess the impact of pipeline construction upon Northern Canadian rivers it is thus essential that regulatory agencies have a better understanding of (1) their biological, chemical and hydrological features and (2) the potential effect of pipeline construction on their water quality. Thus, the intent of this study was to (1) examine microbiological and related characteristics of two selected rivers (Ogilvie and Swift) in the northern and southern sections of the Yukon pipeline corridor and (2) to assess the impact of pipeline construction upon their microbial ecology as it relates to water quality.

MATERIALS AND METHODS

Periphytic algae were sampled by removing representative rocks (usually four from each site, each ca. 20 cm in diameter) from the stream bed of each river or creek, at locations noted (Figs. 1 and 2) and ca. 100 cm² areas were immediately scrubbed with a nylon toothbrush (when necessary a forceps or razor blade was used to remove encrusted forms). The detached microflora on the rock and brush were then sluiced into a container with the aid of a wash bottle containing distilled water; this entire procedure was repeated twice. See Sheenan *et al.* (1978) for a more detailed description of this technique.

The area of each rock sampled was determined by fitting aluminum foil to the scrubbed contour, removing the foil, pressing it flat and estimating its area with the aid of a polar planimeter.

Each periphyton suspension was wet filtered, within 24h, onto either 5.5 or 10 cm diameter Whatman GF/C glass fibre filters using a vacuum of -18 cm Hg. Immediately following this filtration each filter was bisected and the cells of one half of the filter washed into a sterile container, and preserved with acid Lugol's solution till assayed for algal species and numbers. The remaining half of each filter was treated with a MgCO₃ suspension, frozen and kept in the dark until used for chlorophyll a extraction and analysis.

Chlorophyll a content of each sample was assayed by extracting each GF/C filter (with filtered algae) in an acid free acetone: water (9:1) mixture using a High-Speed Polytron homogenizer followed by filtration (0.45 um Gelman Alpha-6 material filter) of the sample to remove debris and particulate matter. The residue was then re-homogenized and filtered again. Both filtrates were combined and made to 15 ml with 90% acetone; chlorophyll a content was measured spectrophotometrically (corrected for phaeophytin) using extinction values of Lorenzen (1967) as described by Strickland and Parsons (1968).

Diatom identifications and cell counts were made on subsamples which were first cleaned in H_2O_2 or nitric acid (Patrick and Reimer, 1966), evaporated onto cover slips (Battarbee, 1973) and then mounted on microscope slides with Hyrax medium. Microscope transects were then examined using phase contrast microscopy (1000 x magnification); all diatoms were identified and counted till there was a total count of 200 frustules. Suitable conversion factors were then used to transform counts to cells/cm².

The works of Patrick and Reimer (1966), Cleve-Euler (1951-1955), Hustedt (1930, 1931-1959), Huber-Pestalozzi (1942), Sreenivasa and Duthie (1973) and Weber (1966) were consulted for identification of the diatoms. The species classification outlined by Van Landingham (1967-1975) was followed except that *Cymbella caespitosa* was recognized as a distinct species. For genera not covered by Van Landingham (starting after the genus *Navicula*), the species taxonomy of Patrick and Reimer (1966), Cleve-Euler (1951-1955), Hustedt (1930), and Huber-Pestalozzi (1942) was followed.

The relative abundance of each algal phyla was measured in another subsample with an inverted microscope using methods detailed in Northcote, Ennis, and Anderson (1975). *Cyanophyta* and *Chlorophyta* species were qualitatively measured for relative abundance and identified using Prescott (1962), Hoek (1963), and Bourrelly (1966, 1968, 1970).

Samples for phytoplankton analyses were obtained by allowing water to flow into a 100 ml sterile container placed 20 cm. below the surface of each river or creek sampled. Two ml of acid Lugol's solution were then added to preserve each sample which were subsequently analyzed using the Utermohl (1958) technique which involved sample transfer to 5, 10, or 25 ml settling chambers (depending upon algal density) and enumeration with an inverted phase contrast microscope. Microscope transects were examined at 500 times magnification and all phytoplankters identified and counted until there was a total count of 200 cells (except; colonies composed of small cells were counted as individual colonies). Suitable conversion factors were used to transform counts to cells/ml.

Planktonic diatoms were identified using the reference works

described above as well as Patrick and Reimer (1975). Geitler (1932) and Prescott (1962) were used to identify planktonic algae from other classes.

Viable heterotrophic bacterial numbers were determined using the spread plate technique, an incubation temperature of 5°C and the following medium: Bacto-beef extract, 3g; Bacto-peptone, 5g; Bacto-agar, 15g, distilled water, 1 litre; Ph 7.2.

One-tenth ml samples of water were plated directly to assay planktonic bacterial numbers whereas epilithic bacterial counts were determined using the "scrub water" obtained as outlined above and before the addition of Lugol's solution.

Total planktonic and epilithic bacterial counts using water and "scrub water" respectively were done using epifluorescent microscopy, as described by Daley and Hobbie (1975).

Benthic invertebrates were sampled by placing a Surber sampler (1 mm mesh size) on a stream bed (water depth of ca. 20 cm) and picking up all larger rocks and washing them in front of the net. The remaining fine material was thoroughly stirred to dislodge organisms which were subsequently washed into the net. In this way stream bed material was sampled to a depth of ca. 10 cm. Each sample was then placed in a plastic bag, preserved with a 5% formaldehyde solution and transported to the laboratory for analysis by J. Keays, Powell River, British Columbia.

Laboratory analyses were done by placing each sample in a white enamel tray and removing organisms larger than ca. 1 cm. Each sample was then sorted in Petri plates and organisms removed. All organisms were identified with the aid of dissecting and compound microscopes. The texts used in identifications were (Johannsen, 1969; Usinger, 1956; Needham *et al.*, 1935; Smith, 1968; Ross, 1944; Ward and Whipple, 1959; Ricker, 1952 and Classen, 1931).

Dissolved organic carbon (DOC) was assayed by the wet oxidation technique of Menzel and Vaccaro (1964) whereas particulate carbon and nitrogen values were determined using a Beckman CHN analyzer.

Total inorganic carbon (TIC) values of sampled river waters were determined with the use of a dual channel carbon analyzer - Beckman model 915 equipped with a Beckman model 215B infrared analyzer. Aliquots of blended water were injected into a combustion tube heated to 175°C containing 85% H_3PO_4 on quartz chips. The CO_2 liberated was quantitatively assayed using an infrared analyzer and standard inorganic carbon solutions.

Glucose heterotrophic potentials (V_{max}), turnover times (T_t) and $K_t + S_n$ (K_t refers to the transport constant whereas S_n denotes the natural concentration of substrate) values were determined by aspirating water samples into ten-50 ml disposable plastic syringes (with attached needles replaced with Beckton-Dickinson two way valves). Each syringe contained an appropriate concentration of tritiated glucose (D-[6- 3H] glucose, specific activity 10 Ci/mmol; Amersham/Searle Corp.) and was rotated following water addition to mix the sample. Controls consisted of samples poisoned with 5% glutaraldehyde before addition of water. In this study tritiated glucose was added to yield a final concentration range in the various water samples of from 0.001 $\mu\text{Ci}/\text{ml}$ (10^{-10}M glucose) to $\mu\text{Ci}/\text{ml}$ (10^{-7}M glucose). Water samples were incubated *in situ* for periods of 30 - 240 min, depending upon water temperatures; care was taken to ensure substrate utilization was linear with time. The incubations were stopped by filtering the contents of each syringe through 0.45 μm pore size membrane filters, Millipore Corp. (25 mm in diameter contained within a Swinnex filter holder attached to the exit port of each two way valve. Each filter was then washed with a 50 ml volume of river or creek water by repeating the refilling and evacuation procedure. Care was taken to do all manipulations under water as exposure to ambient air temperatures (in some cases as low as -40°C) might immediately freeze the contents of each reaction vessel. Filters were placed in scintillation vials containing 15 ml of Aquasol scintillation cocktail (New England Nuclear Corp.) with 10% ethyl ether. All samples were counted in a Beckman LS-250 liquid scintillation spectrometer. Counts per minute (cpm) were corrected for quench (by the external standard method), machine efficiency and half-life decay and were reported as disintegrations per minute (dpm). The tritiated glucose was diluted in carbon-free water prepared by the method of Strickland

and Parsons (1962) and filtered through sterile membrane filters (0.22 μm pore size, 25 mm diameter, Millipore Corp.) prior to use. No attempt was made to quantitate glucose respiration. See Dietz *et al.* (1977) and Wright and Hobbie (1966) for a more complete description of the technique as well as equations for calculation of V_{\max} , $K_t + S_n$ and T_t . Bacterial specific activities were calculated as the glucose heterotrophic activity per viable heterotrophic bacterium.

Light (algal) and dark (mainly bacterial) productivities of the plankton were determined with the one of $^{14}\text{CO}_2$. Light (two), dark (two) and killed (two; 2 ml of 5% glutaraldehyde added to each) BOD bottles (300 ml capacity) were filled with river or creek water and 5 uCi of $^{14}\text{CO}_2$ (as bicarbonate - pre-membrane filtered (0.22 μm)) added to each. The bottles were stoppered, rotated to mix the contents and incubated *in situ* for 4 h (incubation times between 1000 and 1500 h were chosen). Following incubation, the reactions were stopped by adding 2 ml of 5% glutaraldehyde to the light and dark bottles and filtering the contents of each through 0.45 μm pore size membrane filters (Millipore Corp.) Following this, each filter was washed with 50 ml of pre-filtered river or creek water, placed in scintillation cocktail, as described above, and the dpm determined.

The light minus dark dpm values were used to calculate algal productivities whereas the dark minus killed values were used to determine heterotrophy productivities. See Romanenko *et al.* (1972) for a more complete description of this technique.

Ten-litre plexiglass containers similar to those described by Schindler *et al.* (1973) were used to assay benthic light and dark productivities. River or creek bottom rocks (4-5) of average volume of 400-700 ml each were gently placed into each of two light, two dark and two killed (with 30 ml of 5% glutaraldehyde) control boxes and the remaining space filled with river or creek water. Each box was then placed on the river bed for ca. 30 min. to allow a resettlement of the partially disturbed material. Following this 55 uCi of ^{14}C -bicarbonate (New England Nuclear Corp.) was injected into each box, the contents mixed, and the entire system was allowed to incubate for 4h (between the hours of 1000 to 1500). The

reactions in each light and dark box were then stopped with the addition of 30 ml of 5% glutaraldehyde. Known areas (see above for description of areal determination technique) of incubated rocks were scrubbed (see above) to obtain ¹⁴C-labelled microflora which was then membrane-filtered and dpm assayed as described above. The volumes of all rocks which occupied each box were also determined at this time. Algal (based upon light minus dark dpm) and dark (dark minus killed control dpm) productivities were subsequently calculated on a per cm² basis.

Biochemical oxygen demand (BOD) values of Ogilvie and Swift River waters were determined with the use of standard BOD bottles which were incubated in the dark for periods of up to 55 d. At times zero, 21 and 55 d dissolved oxygen contents were assayed by the Winkler technique (Amer. Pub. Health Assoc., Amer. Water Wks. Assoc., and Water Poll. Con. Fed. (1965)).

RESULTS AND DISCUSSION

A significant feature of both the Ogilvie and Swift Rivers was an annual cycle in the values of many of the microbial (biomasses and activities) as well as chemical and physical parameters assayed. These included microalgal, bacterial and invertebrate standing crops, algal and bacterial activities, as well as TIC, DO and DOC concentrations. Generally, maximum and minimum values of each parameter occurred in Spring and Winter respectively.

Relatively large concentrations of both planktonic and epilithic microalgae were present in these sub-Arctic Canadian rivers during the Spring, Summer and Fall of 1977 - 1978 (Tables 1 and 2). However, massive declines in standing stocks of these cells occurred in early Winter; generally the phytoplankton values decreased to approximately 1% of their Spring and Summer levels whereas the decline was somewhat less amongst the periphyton (Tables 1 and 2). The periphyton declines noted between October and December 1977 were ca. 93% and 98% for the Ogilvie and Swift Rivers respectively. However, due to difficulty in quantitatively sampling periphyton these values should be interpreted with caution (see Tett *et al.* (1978) for a discussion of adequately sampling benthic microalgae).

Karlstrom and Backlund (1977) noted similar fluctuations in numbers of planktonic diatoms of the river Rickleian (Sweden, lat. of ca. 64° 5' N.). Cell concentrations were greatest during Spring and declined quite abruptly during late Autumn and early Winter whereas a slight recovery in numbers was noted between January and March. In both the Ogilvie and Swift Rivers a modest recovery in numbers of both phytoplankton and periphyton occurred between early and late Winter.

The factors which cause the massive decline in phytoplankton numbers in late fall are not known with certainty. However, two significant ones were probably (1) low insolation and (2) low temperature. By late December daylight lengths were minimal at both rivers whereas water temperatures were near freezing (Table 7).

Several investigators have noted that many microalgae are able to withstand exposure to both low temperatures and darkness. Jansz and MacLean (1973) found that when the blue-green alga *Anacystis nidulans* was exposed to 0 - 5 C culture viability was reduced. However, residual numbers of viable cells remained, even at these low temperatures. Talling (1955) found that two diatom genera (*Asterionella* and *Fragilaria*) grew throughout the year in Lake Windermere with the division rates lowest in January. As Winter progressed, the division rates increased from January to March. This author concluded that "the mean relative growth rates of cells at 1 m. depth are primarily determined by daylength and temperature". Antia and Cheng (1970) showed that although 31 species of marine unicellular algae showed no significant growth in darkness at 20 C., several species were able to survive up to 24 w and resume normal growth rates upon transfer to light. It is possible that some algal cells treated in this fashion, including the microalgae of the Ogilvie and Swift Rivers, may have (1) decreased their endogenous metabolism and (2) shifted to a heterotrophic pattern of cellular maintenance. Many diatoms are able to survive in the absence of light using heterotrophic processes (Hellebust, 1968, Hellebust and Lewin, 1972 and Lewin and Hellebust 1975).

There is no facile explanation for the minor increase in numbers of both planktonic and epilithic microalgae between early and late winter noted in both this study (Tables 1 & 2) and by Karlstrom and Backlund (1977) in the river Ricklean. Water temperatures remained approximately the same. However, daylength increased during this time period and it is possible that photosynthetic rates increased which resulted in slow algal growth.

The algal species composition data showed similar trends in all rivers assayed, ie. the Spring - Summer - Fall populations were *Bacillario-phyceae* or *chlorophyceae* dominated whereas the bulk of the overwintering cells were diatoms (Table 3). The reason for this is not known.

The levels of both planktonic and epilithic chlorophyll a in these two lotic systems (Table 4) follow a pattern similar to that of the algal standing crops (Tables 1 and 2). That is, a marked decrease in early Winter followed by a partial recovery in late Winter and a rapid increase to Spring.

and Summer values.

The seasonal changes of bacterioplankton cell numbers in both rivers were remarkably similar to those of the phytoplankton (Table 5, cf. Table 1). That is, the viable bacteria numbers as determined by plate counts were minimal in Winter and increased to greater numbers in Spring and Summer. In both lotic systems, a slight increase in numbers was noted between early and late Winter.

Other investigators have noted numbers of both planktonic and epilithic bacteria which were similar to those observed in these two rivers (Table 5). Geesey *et al.* (in press) found that the epilithic and planktonic bacterial numbers (assayed by epifluorescent microscopy) of a pristine sub-alpine stream system in the Canadian Rocky Mountains ranged seasonally from 1×10^6 to 1×10^8 cells/cm² and from 2×10^3 to 2×10^5 cells/ml respectively. Both planktonic and epilithic bacterial numbers were minimal in the Winter and maximal in the Summer. The ratio of numbers of bacterioplankton to that of the sessile bacteria was approximately 2 orders of magnitude.

Data of this study of total bacterial numbers, as assayed by epifluorescent microscopy were not sufficient to interpret seasonal trends. However, the ratio between planktonic bacteria/ml and sessile bacteria/cm² was approximately 10^2 in all waters assayed by this technique (Table 5) and is the same as the value obtained by Geesey *et al.* (See above). Since the average depths of both the Ogilvie and Swift Rivers were ca. 2 m, the calculated ratios of planktonic to epilithic cells vary from 3.38 to 0.16. The ratio of phytoplankton to periphyton cells vary from 2.7 to 0.007 (based upon calculations of the data of Tables 1 and 2 and an average river depth of 2m.) Within the limits of these experimental data it would appear that similar proportions of the total bacterial and algal cell numbers in the river are suspended in the water as compared to the bottom.

The numbers of invertebrates /m² of river or stream bottom sampled using a Surber device ranged from 32 to 5208 with a mean value for all samples of 1284 (Table 6) which are equivalent to those which have been found by other investigators using similar streams and creeks in both the Yukon and North West Territories. Brunskill *et al.* (1973) found that "zoobenthos density in the Yukon and North Slope areas ranged from a few hundred to a few thousand

organisms /m². Hoos and Holman (1973) found the numbers of organisms to be ca. 3,600/m² at a site in Rose Creek which was not perturbed by mine tailings. Since so few samples within each lotic system studied in this investigation were taken it is difficult to adequately compare these data, although the numbers of benthic invertebrates appear to be greater in the Ogilvie River (March to August, 1978) as compared to the Swift River (Table 6). The *Chironomidae* were the most numerous macroinvertebrates in the majority of samples, an observation which has also been noted by other investigators (Brunskill *et al.*, 1973 and Hoos and Holman, 1973).

Generally, samples with large numbers of invertebrates also had more plant and filamentous algae present.

The average water temperatures of both the Ogilvie and Swift Rivers varied between Winter lows of ca. 0 C and Summer highs of 12 - 15 C during 1977 - 78. For at least 6 months of the year the temperatures were at or near freezing (Table 7). Hence, this physical parameter may have had a large influence upon the numbers of microflora of these rivers (see above), as well as microbial activities (see below).

DOC values increased in Spring and decreased as the seasons progressed through Summer, Fall and Winter (Table 7). These Spring increases were probably due to both allo- and autochthonous addition of organic matter to the two rivers, although a distinction between the relative importance of each to the total DOC levels cannot be made on the basis of these data. However, the contributions by both of these sources are probably decreased in Fall, lowest in Winter and highest in Spring and Summer since (1) freezing conditions would greatly slow tributary and land run-off into these rivers (allochthonous addition) and (2) phytoplankton and periphytic algal productivities were lowest in Fall and Winter (autochthonous addition) (see below).

TIC values of Ogilvie River water are almost always greater than those of the Swift River (at similar times of the year). Since the Ogilvie River system flows over extensive limestone substratum whereas the Swift River does not, there is probably a much greater non-biological contribution of CO₂, HCO₃⁻ and CO₃⁼ to the former river. However, both rivers displayed marked seasonal

variations in TIC. These values increased from Summer to their highest concentrations in late Winter (Table 7). A reasonable explanation is that this increase may be partially due to community respiration. This is, during the ca. 7 months of ice cover a significant portion of the planktonic and benthic DOC and POC were metabolized with the concomitant release of CO_2 which tended to collect under the ice as TIC. At Spring break-up a sudden drop in TIC values were noted in each river (Table 7). This may be due to the abrupt release of TIC to the atmosphere as CO_2 as well as its fixation by biological processes which accelerated (see below).

POC values were generally greater in the Ogilvie than in the Swift River water (table 7). Since these concentrations are a function of both microbial (cf. data of Tables 1,2 and 5) and detritus content of these waters, the concentrations of which in turn are regulated by many biotic and abiotic factors, it is difficult to interpret these data in a simple fashion other than to state that the Ogilvie River appears to be the richer of the two systems.

The planktonic microflora of both the Ogilvie and Swift Rivers displayed glucose heterotrophic activities for most of the year, the exception being late Winter of 1978. At that time glucose heterotrophic uptake versus time kinetics by the microorganisms of the water were linear with time, but Michaelis-Menten uptake kinetics were not observed (Table 8). This phenomenon has been noted previously in other aquatic ecosystems and it is the experience of this author that this occurs under at least three conditions, viz. (1) the microbial contents of the waters are extremely low (2) the concentrations of naturally occurring metabolites are minimal or (3) pollutants (e.g. mercury) and/or natural physico-chemical environmental conditions (e.g. low temperatures) are unduly stressing the microbial ecosystems. The bacterial content of these waters were not decreased excessively in Winter (Table 5), but, both DOC levels and temperatures did decrease appreciably as compared to Summer values. However, linear Michaelis-Menten curves were obtained during the Fall and early Winter when the temperatures were ca. 0°C which would tend to eliminate low temperatures as a cause of scattered uptake of glucose in late Winter. A more likely cause may be the quantity and quality of DOC present in late Winter (Table 7). As Winter progressed the DOC levels of both rivers decreased, probably due to

microbial utilization. Since allochthonous and autochthonous production of DOC is minimal in Winter, the DOC remaining in late Winter may be highly refractive, and not readily available for heterotrophic microbial utilization. Hence the bacterial cells may not have been able to readily metabolize this material and displayed non-linear heterotrophic activity plots. Following Spring "break-up" allo- and autochthonous production greatly increased the quantity and quality of DOC. This, coupled with increased water temperatures may have greatly influenced glucose heterotrophic activities in Spring and Summer (Table 8).

Both DOC concentrations and temperatures have been shown to influence heterotrophic bacterial activities in other aquatic ecosystems (Albright, 1977, Wright and Hobbie, 1965, Hamilton *et al.* 1966 and Dietz *et al.* 1977)

Values of glucose K_t and S_n decreased significantly from early Fall to late Winter (Table 8) which is indirect evidence that glucose concentrations may have been decreasing during this time frame. That is, the DOC may have become more refractive.

Since the streamside materials of the Swift River have a large organic matter content* addition of this to the Swift River would probably increase the heterotrophic activities of the microflora at all times of the year. A predicted result would be increased oxygen demands of these aquatic ecosystems.

The glucose heterotrophic potentials assayed in this study fall within the lower part of the range noted for several other freshwater rivers and lakes (Table 9). These values may indicate that the more Northerly the water body, the less the glucose heterotrophic potentials.

Phytoplankton photosynthesis occurred throughout the year in these two rivers with maximal activity in the Spring and Summer and minimal activity during the Fall and Winter (Table 10). These data are highly variable which is probably due to changing daily conditions within each watershed. These

* LFH and Btm contain 0.32 and 1.38 mg. DOC, 158 and 13 mg. organic carbon and 9 mg and trace (<1mg) organic nitrogen /g dry weight respectively.

include sunlight, silt load, temperature and water velocity (which may tend to suspend periphytic algae in the water column). An excellent example of the results of one of these natural perturbations is the minimal photosynthetic activity observed on 23 June, 1978 (Table 10) in the Ogilvie River. Heavy rainfall naturally perturbed this river with silt (0.117 g silt/litre water (Schreier, 1978)) within 2 d which resulted in heavy turbidity and little light penetration beyond a depth of ca. 0.1 m. Hence, photosynthesis did not appreciably occur (Table 10) although phytoplankton concentrations were high (Tables 1 and 4). This natural silt loading is probably in excess of that which would occur by streamside and streambed disturbances due to pipeline construction. Thus, in at least one of these two rivers construction disturbances in the Spring and Summer may result in little more than a pale imitation of natural processes which dramatically alter silt load.

The planktonic dark productivities were generally greater than light productivities for much of the year in the Ogilvie River whereas the opposite was noted for the Swift River (light productivities generally exceeded dark productivities) (Tables 10 and 11). In addition, the dark productivities of the Ogilvie River exceeded those of the Swift during the yearly cycle assayed (Table 11). The reasons for this are not known although there are several observations which may be pertinent: (1) DOC levels of the Ogilvie River water were greater than those of the Swift River during the annual cycle of Oct. 1977 - August 1978 (2) Ogilvie River bacterioplankton biomasses exceeded those of the Swift River throughout the year and (3) glucose $K_t + S_n$ values of the Ogilvie River were generally greater than those of Swift River water. Thus, bacterial productivities (which are a major portion of the total dark productivities of aquatic microbial ecosystems) should on the basis of these observations be greater in the Ogilvie as compared to the Swift River.

In summary, microbial (microalgal and bacterial) biomasses and activities of these two lotic ecosystems were greatest in Spring with decreasing values noted through Summer, Fall and Winter. The data support the hypothesis that light and DOC may be the major factors controlling standing crops and activities of microalgae and bacteria respectively.

The Ogilvie River appears to be more productive than the Swift River with regard to microorganisms since generally both standing crops and activities of microalgae and bacteria were greater in the former throughout the year (Tables 1 - 11). The reasons for this are not known although this difference may be a reflection of the greater levels of both TIC (microalgal substrate) and DOC (heterotrophic bacterial substrate) noted in the Ogilvie as compared to the Swift River at equivalent times of the year (Table 7).

The foregoing data and observations may be useful in predicting general influences of streamside and sediment addition to the river waters upon standing crops and activities of microalgae and bacteria. An increased silt and sediment load would probably lower microalgal photosynthetic rates which in turn would adversely influence algal biomasses. A natural example of this type of perturbation was noted in June in the Ogilvie River. However, this influence may not be significant if silt and/or sediment loading were of short duration.

Since streambank material contains relatively large amounts of DOC (see above) and stream sediments also have a large DOC component the results of these additions may be increased biomasses and activities of heterotrophic bacteria and increased oxygen demand. Since, the heterotrophic bacteria in these two lotic systems may be mainly DOC limited (see above, cf. Tables 5 and 7), DOC added at any time of year (including Winter) via silt and sediment would probably increase heterotrophic activities. However, DO levels of these two rivers tend to be lowest in late Winter. Hence these two lotic systems may be most sensitive to DOC addition at that time.

Most streambank and sediment material would settle to the stream bottoms within several km of its site of addition, overlaying and smothering much of the periphyton and epilithic bacteria, which would tend to lower their biomasses, productivities and activities. Recolonization of streambottom surfaces would eventually occur although the rates in Winter would be extremely slow (see above).

Thus, the least sensitive time for perturbation of these two lotic systems by streambank and sediment materials would probably be Spring following

"break up". The most sensitive time with regard to influences upon algal and bacterial activities (which in turn would influence DO levels) as well as recolonization rates would be Winter. The remainder of this manuscript deals with a quantitative evaluation of DO levels and BOD values of the Ogilvie and Swift rivers at various times of the year as influenced by streambank and sediment additions in the context of microalgal and bacterial biomasses and activities.

Data of Schreier (1978) and that of others (Schallock and Lotspeich, 1974) indicate that many Alaskan and Canadian Arctic and sub-Arctic Rivers have similar annual DO concentration trends. That is "the waters are near saturation during Spring "break up" and Fall "freeze up" when water temperatures are near 0 C, somewhat lower DO concentrations during warm Summer periods; and yearly minimum concentrations during the Winter (January - March) interval..... Data indicate that DO depression begins in October and continues into February" (Schallock and Lotspeich, 1974). A second important observation of these authors was that DO depletion usually becomes more severe as the river water flows from its headwaters to its mouth under an ice cover. This is probably due to continued biological and chemical utilization of DO of each water mass as it travels the length of the river. In several rivers these DO depletions may be severe, e.g. the Yukon River from the Alaska - Canadian border to its mouth displayed DO levels of ca. 10.5 mg/l and 1.9 mg/l respectively during March of 1971 (Schallock and Lotspeich, 1974).

Ogilvie and Swift River water DO values also showed marked variation in concentrations in both rivers over a yearly cycle (Schreier, 1978). In both cases the levels of the DO dropped significantly from early Fall to late Winter and remained at relatively high concentrations during the ice-free seasons of Spring and Summer. The drop in DO levels under ice cover was probably due to a variety of physical, chemical and biological influences such as dilution of river water by ground water, abiotic reduction by various types of organic and inorganic materials and respiration by the aquatic micro- and macroorganisms. When these data are used to calculate the *in situ* rate of DO depletion in the Swift River during the period of ice cover (October to March) a net respiration value of 0.041 mg O₂/l/d is obtained. This value is of the same order of magnitude of net respiration rate values reported by Welch (1974)

for three Canadian Arctic and sub-Arctic lakes during ice cover (Table 12).

In late Winter the experimentally determined BOD values of both the Ogilvie and Swift Rivers were in the range of ca. 0.5 to 1.0 mg O₂/l/21 d (ca. 0.023 to 0.048 mg O₂/l/d) (Table 13). These were probably maximum values of BOD since the treatments that the waters underwent would tend to increase BOD values. These are (1) an incubation temperature of 1°C (as compared to ca. 0°C at the time of sampling - this temperature increase would tend to increase biological activity due to a Q₁₀ effect) and (2) enclosure of natural waters in glass containers tends to accelerate biological activities, including oxygen utilization (bottle effect).

The *in situ* (0.041 mg O₂/l/d) and experimentally assayed (0.023 to 0.048 mg O₂/l/d) rates of DO depletion in the Swift River and its water respectively are approximately the same. However, these data comparisons must be interpreted with a great deal of caution since BOD values assayed *in vitro* are subject to experimental errors which tend to bias the results upwards (see above discussion). In addition, these *in vitro* BOD values do not take into consideration both biological and chemical oxygen demands of the stream bottom as well as oxygen utilization by planktonic macroorganisms (including fish). Benthic bacteria, fungi, insect larvae and other invertebrates may exert a considerable oxygen demand upon the overlying water. McDonnell and Hall (1969), for example, noted that ca. 50% of the benthic oxygen utilization by organisms was due to invertebrates in a river system.

Treatments of both Ogilvie and Swift River waters with several (LFH, Ae and Btm) soil horizon materials (from the Swift Riverbank) markedly influenced BOD values (Table 14). All treatments greater than 0.10 g streambank material/litre of stream water (100 ppm) significantly increased oxygen utilization by the native microflora of these two waters beyond that of the untreated control values.

Both algal and heterotrophic productivities of the Swift River waters were also influenced by these streambank materials additions (Table 15). Btm and LFH additions at 100 ppm (0.10 g/l) or greater and 1 ppt respectively significantly increased microbial productivities, particularly by heterotrophic

microorganisms. However, these algal productivity data may be misleading since all perturbant material was contained within BOD assay bottles which were then replaced in the clear stream water for incubation. If sediment and streambank material had been added to the entire river system light penetration into the water would have been markedly less with resultant decreases in algal productivities (see above discussion dealing with a natural perturbation by silt of the Ogilvie river in June). However, these data do indicate that in the presence of adequate light levels Btm material (at concentrations greater than 100 ppm) increased algal productivities (Table 15). Since heterotrophic productivities are not light dependent the greatly increased activities noted at Btm levels of 100 ppm or greater may be significant in the context of BOD by these rivers in Winter (see previous discussion).

In summary, additions of several Swift river streambank materials (LFH, Ae and Btm) to Swift River water increased BOD values, glucose heterotrophic activities, and planktonic and epilithic heterotrophic productivities. Planktonic (see above discussion) and benthic algal productivities were in the one case probably and in the other definitely inhibited by streambank materials additions. In all cases significant influences were noted at perturbant addition levels of 100 ppm or greater. Hence, additions of streambank materials and sediments should be regulated such that water silt levels do not exceed 100 ppm.

The time of year at which streamside and sediment additions to river and creek waters occurs is of great importance since these materials increase bacterial activities and BOD values which result in accelerated oxygen depletion rates. Therefore, preferential times for construction activity, in this context are Spring (following "break up"), Summer and Fall (before "freeze over").

Many swamps and bogs contain waters of relatively high organic matter content which may greatly increase BOD of receiving waters. Hence, pipeline crossings at these areas should be avoided.

As a water mass passes the length of a river which is ice covered, its DO level tends to become depleted due to micro- and macroorganism respiration. Therefore, construction activity (which increases streamside and sediment addition and hence enhances BOD values) should in general be done as close to a river's mouth as possible.

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LIST OF TABLES

- Table 1: Phytoplankton concentrations (cells/ml - 500 magnification) of several Yukon rivers.
- Table 2: Periphyton concentrations (cells/cm² - 1000 x magnification) of several Yukon rivers.
- Table 3: Percent composition of periphytic algae of several Yukon river waters.
- Table 4: Periphytic (mg/m²) and planktonic (mg/m³) chlorophyll a concentrations of several Yukon rivers.
- Table 5: Periphytic and planktonic bacterial counts of Ogilvie and Swift River water and stream bottom material, as determined by epifluorescent microscopy and nutrient agar plate counts.
- Table 6: Invertebrates (numbers/m²) of several Yukon Rivers.
- Table 7: Dissolved (DOC, mg C/l), particulate (POC, mg C/l) and total inorganic (TIC, mg C/l) carbon concentrations and water temperatures (C) of the Ogilvie and Swift Rivers.
- Table 8: Glucose turnover time (T_t), maximum velocity of uptake (V_{max}) and $K_t + S_n$ values of two Yukon Territory waters.
- Table 9: A comparison of glucose heterotrophic potentials (V_{max}) of several lakes and rivers.
- Table 10: Periphytic ($\mu\text{g C}/\text{m}^2/\text{day}$) and planktonic ($\text{mg C}/\text{m}^3/\text{day}$) algal productivities of two Yukon Territory rivers.
- Table 11: Periphytic ($\mu\text{g C}/\text{m}^2/\text{day}$ and planktonic ($\text{mg C}/\text{m}^3/\text{day}$) heterotrophic microbial productivities of two Yukon Territory rivers.
- Table 12: Respiration rates for several arctic and sub-Arctic Canadian water bodies.
- Table 13: Biologican oxygen demand* (mg O₂/l) of waters removed from the Ogilvie and Swift Rivers and incubated in the laboratory at 1 C.
- Table 14: The influences of three Swift River streamside soil horizons upon Biological Oxygen Demand of Swift and Ogilvie River waters. (Sampled 10 March, 1978). Incubation temerature = 1 C.
- Table 15: The influence of two streamside soil horizons upon planktonic algal ($\text{mg C}/\text{m}^3/\text{day}$) and heterotrophic ($\text{mg C}/\text{m}^3/\text{day}$) productivities of the Swift River.

Table 1. Phytoplankton concentrations (cells/ml - 500 x magnification) of several Yukon rivers.

River	Date of Sampling						
	4-10 Oct. 1977	10-20 Dec. 1977	22-30 Mar. 1978	15-26 May 1978	17-30 June 1978	1-8 Aug. 1978	
Ogilvie	250.5	1.4	18.7	267.1	266.9	116.0	
Swift	168.9	2.3	4.2	192.1	103.9	40.2	
Seagull	-	-	-	64.5	6.7	24.8	
Engineer	53.1	-	-	200.0	89.1	196.7	
Screw	-	-	-	36.5	66.4	46.9	
Partridge	-	-	-	21.9	99.1	49.4	
Logjam	-	-	-	68.7	102.8	184.8	
Smart	-	-	-	178.7	71.0	174.5	
Castle	-	-	-	-	-	111.1	

Table 2. Periphyton concentrations (cells/cm² - 1000 x magnification) of several Yukon rivers.

River	Date of Sampling						
	4-10 Oct. 1977	10-20 Dec. 1977	22-30 Mar. 1978	15-16 May 1978	17-30 June 1978	1-8 Aug. 1978	
Ogilvie	496,000	34,528	659,439	179,674	1,546,787		968,258
Swift	142,875	3,046	30,230	165,077	286,423		722,787
Seagull	1,316	-	25,599	21,728	14,212		133,307
Engineer	159,002	-	-	2,428	-		1,075,489
Screw	-	-	-	148,568	14,312		7,339,811
Partridge	-	-	-	30,214	90,200		64,858
Logjam	-	-	-	187,004	74,476		444,176
Smart	-	-	-	372,118	-		2,320,279
Castle	-	-	-	-	-		8,900,002

Table 3. Percent composition of periphytic algae of several Yukon river waters.

River	Date of Sampling						1-8 Aug. 1978
	4-10 Oct. 1977	10-20 Dec. 1977	22-30 Mar. 1978	15-26 May 1978	17-30 June 1978		
<u>Ogilvie</u>							
<i>Bacillariophyceae</i>	94	100	96	100	99	86	
<i>Chlorophyceae</i>	5	0	4	0	1	14	
<i>Cyanophyceae</i>	1	0	0	0	0	0	
<u>Swift</u>							
<i>Bacillariophyceae</i>	37	100	100	100	63	73	
<i>Chlorophyceae</i>	62	TR*	0	TR	37	27	
<i>Cyanophyceae</i>	1	0	0	0	TR	TR	
<u>Seagull</u>							
<i>Bacillario phyceae</i>	39	-	100	100	97	60	
<i>Chlorophyceae</i>	60	-	0	0	3	39	
<i>Cyanophyceae</i>	1	-	0	0	1	1	
<u>Engineer</u>							
<i>Bacillariophyceae</i>	95	-	-	100	-	98	
<i>Chlorophyceae</i>	5	-	-	0	-	2	
<i>Cyanophyceae</i>	0	-	-	0	-	0	
<u>Screw</u>							
<i>Bacillariophyceae</i>	-	-	-	100	100	100	
<i>Chlorophyceae</i>	-	-	-	0	0	0	
<i>Cyanophyceae</i>	-	-	-	0	0	0	

* Trace

Continued

Table 3. Percent composition of periphytic algae of several river waters. (continued)

River	Date of Sampling						17-30 June 1978	1-8 Aug. 1978
	4-10 Oct. 1977	10-20 Dec. 1977	22-30 Mar. 1978	15-26 May 1978	17-30 June 1978			
<u>Logjam</u>								
<i>Bacillariophyceae</i>	-	-	100	90	90	80	80	
<i>Chlorophyceae</i>	-	-	0	10	10	15	15	
<i>Cyanophyceae</i>	-	-	0	0	0	5	5	
<u>Smart</u>								
<i>Bacillariophyceae</i>	-	-	95	-	-	89**	89**	
<i>Chlorophyceae</i>	-	-	1	-	-	1	1	
<i>Cyanophyceae</i>	-	-	0	-	-	TR	TR	
<u>Partridge</u>								
<i>Bacillariophyceae</i>	-	-	-	-	-	99	99	
<i>Chlorophyceae</i>	-	-	-	-	-	0	0	
<i>Cyanophyceae</i>	-	-	-	-	-	1	1	
						60	60	
						0	0	
						40	40	

** 9% of sample was Hydrus foctidus

Table 4. Periphytic (mg/m^2) and planktonic (mg/m^3) chlorophyll a concentrations of several Yukon rivers.

River	Date of Sampling						
	4-10 Oct. 1977	10-20 Dec. 1977	22-30 Mar. 1978	15-26 May 1978	17-30 June 1978	1-8 Aug. 1978	
<u>Ogilvie</u>							
Periphytic	7.7	1.0	10.4	4.5	-	1.8	
Planktonic	0.6	0.1	0.1	0.7	1.4	0.1	
Periphytic phaeophytin to chl. a ratio	1.53	1.40	1.45	1.57	-	1.50	
<u>Swift</u>							
Periphytic	2.7	<0.5	0.6	2.9	2.7	13.8	
Planktonic	0.2	<0.5	0.1	0.3	0.1	0.1	
Periphytic phaeophytin to chl. a ratio	1.49	1.26	1.55	1.62	1.58	1.38	
<u>Seagull</u>							
Periphytic	1.5	-	0.5	0.9	0.8	-	
Periphytic phaeophytin to chl. a ratio	1.51	-	1.57	1.63	1.39	-	
<u>Engineer</u>							
Periphytic	2.6	-	-	0.1	-	3.2	
Periphytic phaeophytin to chl. a ratio	1.62	-	-	1.32	-	1.64	
<u>Screw</u>							
Periphytic	-	-	-	1.4	1.4	-	
Periphytic phaeophytin to chl. a ratio	-	-	-	1.59	1.57	-	

Continued

Table 4. (continued)

River	Date of Sampling						
	4-10 Oct. 1977	10-20 Dec. 1977	22-30 Mar. 1978	15-26 May 1978	17-30 June 1978	1-8 Aug. 1978	
<u>Partridge</u>	-	-	-	3.4	1.1	-	
Periphytic	-	-	-	-	-	-	
Periphytic phaeophytin	-	-	-	-	-	-	
to chl. a ratio	-	-	-	1.57	1.65	-	
<u>Logjam</u>	-	-	-	1.7	0.8	-	
Periphytic	-	-	-	-	-	-	
Periphytic phaeophytin	-	-	-	-	-	-	
to chl. a ratio	-	-	-	1.67	1.66	-	
<u>Smart</u>	-	-	-	-	-	-	
Periphytic	-	-	-	-	-	-	
Periphytic phaeophytin	-	-	-	-	-	-	
to chl. a ratio	-	-	-	5.6	-	-	
<u>Ogilvie tributary #1</u>	-	-	-	1.55	-	-	
Periphytic	-	-	-	-	-	-	
Periphytic phaeophytin	-	-	-	-	-	-	
to chl. a ratio	-	-	-	4.6	143.7	33.8	
<u>Ogilvie tributary #2</u>	-	-	-	-	-	-	
Periphytic	-	-	-	-	-	-	
Periphytic phaeophytin	-	-	-	-	-	-	
to chl. a ratio	-	-	-	<0.1	1.49	1.54	
				1.11	-	-	

Table 5. Periphytic and planktonic bacterial counts of Ogilvie and Swift River water and stream bottom material, as determined by epifluorescent microscopy and nutrient agar plate counts.

River	Date of Sampling					
	4-10 Oct. 1977	10-20 Dec. 1977	22-30 Mar. 1978	15-26 May 1978	17-30 June 1978	1-8 Aug. 1978
<u>Ogilvie</u>						
Periphytic (cells/cm ²)*	-	-	-	-	-	-
Planktonic (cells/ml)*	-	1.6 x 10 ⁴	-	1.6 x 10 ⁶	8.4 x 10 ⁵	2.1 x 10 ⁷
Periphytic (CFU/cm ²) **	-	3.6 x 10 ⁵	-	-	3.5 x 10 ⁶	3.8 x 10 ⁶
Planktonic (CFU/ml)**	2.4 x 10 ³	2.5 x 10 ²	9.0 x 10 ²	1.5 x 10 ⁴	7.0 x 10 ³	1.6 x 10 ³
<u>Swift</u>						
Periphytic (cells/cm ²)	-	-	-	-	8.0 x 10 ⁷	5.3 x 10 ⁷
Planktonic (cells/ml)	-	1.0 x 10 ⁴	-	-	8.4 x 10 ⁴	4.3 x 10 ⁴
Periphytic (CFU/cm ²)	-	1.4 x 10 ⁵	2.1 x 10 ⁵	5.7 x 10 ²	2.7 x 10 ⁴	9.1 x 10 ⁴
Planktonic (CFU/ml)	6.5 x 10 ²	3.2 x 10 ²	9.9 x 10 ²	1.0 x 10 ³	3.9 x 10 ²	1.9 x 10 ³

* epifluorescent counts

** plate counts, CFU = colony forming units

Table 6. Invertebrates (numbers/m²) of several Yukon Rivers.

River	Date of Sampling					
	4-10 Oct. 1977	28-29 Mar. 1978	22-30 May 1978	17-30 June 1978	1-8 Aug. 1978	
Ogilvie	-	979	3314	2432		1137
Swift	-	570	565	979		479
Seagull	-	-	118	366		-
Engineer	-	-	32	-		194
Screw	-	-	1302	2120		1463
Partridge	-	-	872	1442		968
Logjam	-	-	958	1679		1065
Castle	-	-	-	-		5208

Table 7. Dissolved (DOC, mg C/%) , particulate (POC, mg C/%) and total inorganic (TIC, mg C/%) carbon concentrations and water temperatures (°C) of the Ogilvie and Swift Rivers.

River	Date of Sampling				
	4-10 Oct. 1977	10-20 Dec. 1977	22-30 Mar. 1978	15-26 May 1978	17-30 June 1978
<u>Ogilvie</u>					
DOC	4.2	-	1.2	1.6	11.2
POC	-	-	0.6	0.6	6.5
TIC	35.1	-	38.8	47.0	8.2
Temperature	0.3	-1.0	-0.2	-0.2	0.8
<u>Swift</u>					
DOC	1.4	-	0.9	0.8	5.0
POC	-	-	1.2	0.6	0.3
TIC	12.6	-	13.8	17.0	10.1
Temperature	0.0	-0.5	-0.2	-0.2	5.0
					9.0
					12.0

Table 8. Glucose turnover time (T_t), maximum velocity of uptake (V_{max}) and $K_t + S_n$ values of two Yukon Territory Waters.

River		Date of Sampling					
		4-10 Oct. 1977	10-20 Dec. 1977	22-30 Mar. 1978	15-26 May 1978	17-30 June 1978	1-8 Aug. 1978
<u>Ogilvie</u>							
T_t	490.2 h	745.8 h					
$K_t + S_n$	$261.3 \times 10^{-6} \text{ mg/l}$	$30.5 \times 10^{-6} \text{ mg/l}$					
V_{max}	$533.1 \times 10^{-9} \text{ mg l/h}$	$40.9 \times 10^{-9} \text{ mg l/h}$					
Specific Activity	$22.2 \times 10^{-14} \text{ ng/cell/h}$	$16.4 \times 10^{-14} \text{ ng/cell/h}$					
<u>Swift</u>							
T_t	141.9 h	503.2 h					
$K_t + S_n$	$32.7 \times 10^{-6} \text{ mg/l}$	$19.6 \times 10^{-6} \text{ mg/l}$					
V_{max}	$230.7 \times 10^{-9} \text{ mg/l/h}$	$38.9 \times 10^{-9} \text{ mg/l/h}$					
Specific Activity	$35.5 \times 10^{-14} \text{ ng/cell/h}$	$12.2 \times 10^{-14} \text{ ng/cell/h}$					

Table 9. A comparison of glucose heterotrophic potentials (V_{max}) of several lakes and rivers.

Water Body	Range in glucose heterotrophic potentials (mg glucose/ $\% / h$) ($\times 10^{-4}$)	Source
Fraser River	0.27 - 9	Albright (1977)
Lake Erken	24 - 400	Hobbie & Wright (1968)
Lappland Lake	3.2*	Hobbie & Wright (1968)
Char Lake	0.01 - 0.08**	Morgan and Kalf (1972)
Chilliwack River	0.04	Albright & Wentworth (1973)
Capilano River	0.02	" "
Nicomekl River	7	" "
Serpentine River	1.8	" "
Ogilvie River	0.00041 -	This study
Swift River	0.00039 -	This study

* Result of one assay only.

** Assays from mid-November to mid-October were not reported.

Table 10. Periphytic ($\mu\text{g C/m}^2/\text{day}$) and planktonic ($\text{mg C/m}^3/\text{day}$) algal productivities of two Yukon Territory rivers.

River	Date of Sampling					
	4-10 Oct. 1977	10-20 Dec. 1977	22-30 Mar. 1978	15-26 May 1978	17-30 June 1978	1-8 Aug. 1978
<u>Ogilvie</u>						
Periphytic	-	-	0.80	0.15	17.4	ND*
Planktonic	-	-	-	-	-	12.8
<u>Swift</u>						
Periphytic	-	ND*	4.4	1.8	21.6	1.86
Planktonic	7.2	7.2	-	-	-	10.1

* No detectable activity.

Table 11. Periphytic ($\mu\text{g.C/m}^2/\text{day}$) and planktonic ($\text{mg.C/m}^3/\text{day}$) heterotrophic microbial productivities of two Yukon Territory Rivers.

River	Date of Sampling					
	4-10 Oct. 1977	10-20 Dec. 1977	22-30 Mar. 1978	15-26 May 1978	17-30 June 1978	1-8 Aug. 1978
<u>Ogilvie</u>						
Periphytic	-	-	-	1.79	1.51	25.5
Planktonic	-	-	-	-	70.2	1.6
<u>Swift</u>						
Periphytic	-	-	-	1.26	ND*	0.88
Planktonic	-	-	-	-	-	3.9

* No detectable activity.

Table 12. Respiration rates for several Arctic and sub-Arctic Canadian water bodies.

Water Body	Respiration Rate (mg O ₂ /l/day)	Source
Char Lake	0.011	Welch (1974)
Resolute Lake	0.021	Welch (1974)
Eleanor Lake	0.010	Welch (1974)
Swift River	0.041 * 0.023 **	This study This study
Ogilvie River	0.048 **	This study

* Assays *in situ*.

** BOD laboratory assays for 21 days at 1 C.

Table 13. Biological oxygen demand* (mg O₂/l) of waters removed from the Ogilvie and Swift Rivers and incubated in the laboratory at 1 C.

River	Date of Sampling		
	10-20 Dec. 1977	22-30 Mar. 1978	15-26 May 1978
Ogilvie	-	1.0	3.2
Swift	<0.1 (55 days)	0.5	0.7
Engineer	-	<0.1	-

* 21 day incubation unless otherwise noted.

Table 14. The influences of three Swift River streamsides soil horizons upon Biological Oxygen Demand of Swift and Oglivie River Waters. (Sampled 10 March 1978). Incubation temperature = 1 C.

Water Addition	Swift River Water			Ogilvie River Water		
	21	51	Days of incubation at 1 C.	21	45	
None	0.4	1.7*		0.0		0.4
0.01 g/l LFH**	0.5	1.3		-		-
0.10 g/l LFH	1.7	9.3		0.0		0.8
1.00 g/l LFH	4.5	10.3		1.2		4.4
10.00 g/l LFH	10.5	10.8		9.6		10.4
0.01 g/l Ae***	0.9	1.7		-		-
0.10 g/l Ae	1.0	2.6		0.0		0.6
1.00 g/l Ae	3.1	9.9		0.3		1.8
10.00 g/l Ae	9.4	10.7		9.4		10.1
0.01 g/l Btm****	0.7	1.1		-		-
0.10 g/l Btm	0.7	1.4		0.0		0.6
1.00 g/l Btm	1.4	5.9		0.3		1.0
10.00 g/l Btm	5.5	10.4		1.9		6.4

* all values expressed as mg O₂/l

** leaf, ferment and humus horizons

Table 15. The influence of two streamsides soil horizons upon planktonic algal ($\text{mg C/m}^3/\text{day}$) and heterotrophic ($\text{mg C/m}^3/\text{day}$) productivities of the Swift River.

Water Addition	Algal Productivity (23 June 1978)	Heterotrophic Productivity (7 Aug. 1978)
None	1.4	3.0
0.10 g/ L Btm	3.3	8.1
1.00 g/ L Btm	1.4	4.9
10.00 g/ L Btm	16.8	20.0
1.00 g/ L LFH	6.1	27.7

APPENDIX 1

Table 1a - Phytoplankton Concentrations (4 - 10 October, 1977) (cells/ml) of several Yukon Territory Rivers
 (analyzed using 500 X magnification). (1 of 2 pages)

SPECIES	SAMPLE LOCATION		
	Ogilvie River	Swift River	Engineer Creek
<i>Achnanthes flexella</i> (Kutz.) Grun.	3.6	0	1.3
<i>Achnanthes lanceolata</i> (Breb.) Grun.	0.7	11.0	9.7
<i>Achnanthes minutissima</i> Kutz.	18.1	14.3	0
<i>Amphipleura pellucida</i> Kutz.	0	1.1	0
<i>Amphora</i> sp.	2.5	0	4.4
<i>Anomoocensis uitrea</i>	51.5	20.9	0.9
<i>Cocconeis placenta</i> Ehr.	0	4.4	0
<i>Cyclotella ocellata</i> Pant.	0.2	2.2	0
<i>Cymatopleura solea</i> (Breb.) W. Sm.	0.4	0	0
<i>Cymbella caespitosa</i> (Kutz.) Brun.	0.7	7.7	0
<i>Cymbella sinuata</i>	0.7	0	1.3
<i>Diatoma hiemale</i> (Rotn) Heib.	2.2	7.7	23.0
<i>Diatoma tenue</i> var. <i>elongatum</i>	30.1	13.2	0
<i>Fragilaria construens</i> var. <i>binodis</i> (Ehr.) Grun.	14.8	0	0
<i>Fragilaria cotonensis</i> Kitton	0.4	0	0
<i>Fragilaria vaucheriae</i> (Kutz.) Peters	0.5	0	0
<i>Gomphonema olivaceum</i> (Lyngb.) Kutz.	3.7	0	0
<i>Navicula papula</i> Kutz.	0.4	0	0
<i>Navicula triponata</i> (O.F. Mull.) Bory	0.2	0	0
<i>Navicula</i> sp.	2.2	4.4	0
<i>Nitzschia acicularis</i> W. Sm.	0.5	0	0
<i>Nitzschia frustulum</i> (Kutz.) Grun.	0	2.2	0
<i>Nitzschia</i> sp.	61.5	46.8	10.7
<i>Synedra angustata</i>	0	0	1.8
<i>Tabellaria fenestrata</i> (Lyngb.) Kutz.	0	0	0

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Table 1a - Continued: (2 of 2 pages)

SPECIES	SAMPLE LOCATION		
	Ogilvie River	Swift River	Engineer Creek
<i>Tabellaria flocculosa</i> (Roth) Kutz.	1.5	0	0
<i>Ankistrodesmus falcatus</i>	3.0	6.6	0
<i>Chlamydomonas</i> sp.	0	20.9	0
<i>Chroomonas acuta</i>	44.3	5.5	0
<i>Chroomonas</i> sp.	1.1	0	0
<i>Cryptomonas</i> sp.	5.7	0	0
	<hr/>	<hr/>	<hr/>
TOTAL	250.5	168.9	53.1
	<hr/>	<hr/>	<hr/>

Table 1b - Phytoplankton Concentrations (December 10 - 20, 1977) (cells/ml) of several Yukon Territory Rivers (analyzed using 500 X magnification).

SPECIES	SAMPLE LOCATION		
	Ogilvie River	Swift River	
<i>Achnanthes flexella</i> (Kutz.) Grun.	.0084*	0.	
<i>Amphipleura pellucide</i> Kutz.	0	.0521	
<i>Cymatopleura solea</i> (Breb.) W. Sm.	0	.0076	
<i>Cymbella ventricosa</i> Kutz.	0	.0115	
<i>Cymbella</i> sp.	.0252	.0179	
<i>Cocconeis placentula</i> Ehr.	0	.0377	
<i>Diatoma niemae</i> (Roth) Heib.	0	.0276	
<i>Diatoma niemae</i> v <i>mesodon</i>	.0252	.0038	
<i>Diatoma tenue</i> v <i>elongatum</i>	.4452	.0076	
<i>Epithemia turgida</i> (Ehr.) Kutz.	0	.0101	
<i>Eunotia pectinalis</i> (Kutz.) Rabenhorst	0	.0069	
<i>Diatoma vulgare</i> Bory	.0168	0	
<i>Fragilaria capucina</i> Desm.	.0756	.0973	
<i>Fragilaria construens</i> v <i>binoëdis</i> (Ehr.) Grun.	.1512	.188	
<i>Frustulia rhomboides</i>	0	.031	
<i>Gomphonema geminatum</i>	0	.031	
<i>Gomphoneme olivaceum</i> (Lyngb.) Kutz.	0	.03	
<i>Harmaea arcuata</i> (Ehr.) Pair.	.0168	0	
<i>Melosira granulata</i> (Ehr.) Ralfs.	0	.176	
<i>Meridion circulare</i> (Grev.) Ag.	.0252	0	
<i>Navicula cryptocephala</i>	0	.0052	
<i>Navicula pupula</i> Kutz.	0	.0133	
<i>Navicula radiosa</i> Kutz.	0	.0032	
<i>Navicula tripunctata</i> (O.F. Mull.) Bory	0	.0152	

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Table 1b - Continued: (2 of 2 pages)

SPECIES	SAMPLE LOCATION	
	Ogilvie River	Swift River
<i>Navicula</i> sp.	.0252	.0305
<i>Neidium</i> sp.	0	.0076
<i>Mitzschia dissipata</i>	0	.0103
<i>Pinnularia</i> sp.	0	.0070
<i>Rhopalodia gibba</i>	0	.0310
<i>Stauroneis</i> sp.	0	.0038
<i>Suriacella</i> sp.	0	.0038
<i>Synechra ulna v oxyrranclus</i>	.0336	.0258
<i>Tabellaria fenestrata</i>	0	.0070
<i>Synechra acus</i>	.0168	0
<i>Synechra</i> sp.	.4956	1.3754
<i>Cymbella turgida</i>	0	.0412
<i>Cymbella caespitosa</i>	0	.0063
		.0063
TOTAL	1.3608	2.3290

* Only figures to the first decimal place are significant. The remainder are carried for calculation purposes only.

Table 1c - Phytoplankton Concentrations (22 - 30 March, 1978) (cells/ml) of several Yukon Territory Rivers (analyzed using 500 X magnification),

SPECIES	SAMPLE LOCATION	
	Ogilvie River	Swift River
<i>Achnanthes flexella</i>	0.367 *	0
<i>Achnanthes macrocephala</i>	0.458	0
<i>Achnanthes minutissima</i>	3.672	0.184
<i>Achnanthes</i> sp.	0.092	0.184
<i>Cymbella affinis</i>	0.092	0
<i>Cymbella sinuata</i>	0.000	0.061
<i>Cymbella ventricosa</i>	0.275	0.184
<i>Cymbella</i> sp. A	0.092	0
<i>Cymbella</i> sp.	0.000	0.061
<i>Diatoma hiemale</i>	4.312	0
<i>Diatoma tenuvelongatum</i>	1.376	.306
<i>Epithemia turgida</i>	0.000	.061
<i>Fragilaria capucina</i>	0.459	.122
<i>Fragilaria crotoneensis</i>	1.009	0
<i>Fragilaria vaucheriae</i>	2.683	.184
<i>Gomphonema acuminatum</i>	0.000	.061
<i>Gomphonema olivaceum</i>	0.367	.244
<i>Melosira granulata</i>	0.092	0
<i>Meridion circulans</i>	0.367	.061
<i>Navicula radiosa</i>	0.000	.122
<i>Navicula</i> sp.	0.092	0
<i>Nitzschia linearis</i>	0.184	.061
<i>Nitzschia palea</i>	0.459	.366
<i>Synechra acus</i>	1.927	0
<i>Synechra</i> sp.	0.092	.549

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Table 1c. Continued (2 of 2 pages)

SPECIES	SAMPLE LOCATION	
	Ogilvie River	Swift River
<i>Tabellaria fenestrata</i>	0.000	.122
<i>Cocconeis placentula</i>	0.000	.305
<i>Synechra ulna</i>	0.000	.366
<i>Chlorophyta</i>	0.000	0
<i>Chlamydomones</i> sp.	0.184	.549
<i>Chrysophyta</i>	0.000	0.
<i>Dinobryon sentularia</i>	0.000	.061
	—————	—————
TOTAL	18.651	4.214
	—————	—————

* Only figures to the first decimal place are significant. The remainder are carried for calculation purposes only.

**Table 1d - Phytoplankton Concentrations (15 - 26 May, 1978) (cell/ml) of several Yukon Territory Rivers.
(analyzed using 500X magnification).**

SPECIES	SAMPLE LOCATION					
	Ogilvie River	Swift River	Seagull Creek	Engineer Creek	Screw Creek	Partridge Creek
<i>Achnanthes minutissima</i>						26.2
<i>Achnanthes flexella</i>	65.7	47.8	0	27.5	9.6	8.2
<i>Achnanthes sp.</i>	3.0	0.5	0	1.6	0	0
<i>Amphipleura pellucida</i>	2.5	0.8	0	0.5	1.4	0
<i>Cocconeis placentula</i>	0.5	0.5	0	0	0.5	0
<i>Cymbella caespitosa</i>	0.3	2.2	0	9.6	0.9	4.1
<i>Cymbella cistula</i>	2.2	1.4	0	0	2.8	0
<i>Cymbella ventricosa</i>	2.4	0.3	0	0	0	0.9
<i>Cymbella sinuata</i>	7.5	4.4	0	2.3	0	0.5
<i>Cymbella sp.</i>	0	0.3	0	0.9	0	0.9
<i>Diatoma hiemale</i>	0.6	2.3	0	2.8	0	2.8
<i>Diatoma hiemale v mesodon</i>	4.9	5.8	5	4.4	0	9.2
<i>Diatoma tenuv v elongatum</i>	29.2	37.9	0	60.2	0	44.7
<i>Diatoma vulgare</i>	0.5	1.4	17.9	8.6	0	1.9
<i>Epithemia turgida</i>	0	0	0	0	0	0
<i>Eragilaria construens v binadis</i>	16.4	8.8	0	26.2	2.3	1.4
<i>Eragilaria construens v venten</i>	10.8	1.5	0	0.7	1.4	4.6
<i>Eragilaria capucina</i>	20.7	8.9	0	0	0	4.1
<i>Eragilaria vaucheriae</i>	6.0	5.8	0	13.9	0.5	0
<i>Gomphonema acuminatum</i>	0	1.4	0	0	0.5	11
<i>Gomphonema geminatum</i>	0.3	0	0	0	0.5	0
<i>Gomphema olivaceum</i>	15.1	6.9	0	11.6	4.1	7.4
<i>Gyrosigma sciontense</i>	0	0	0	0	0.5	0
<i>Hannaea arcus</i>	16.6	11.0	30.9	12.3	2.7	20.5
<i>Hannaea arcus v amphioxys</i>	4.5	2.5	6.4	0	0.9	4.6
<i>Melosira granulata</i>	0	10.4	0	0	0	0
<i>Melosira granulata v angustissima</i>	9.5	0.2	0	0	0.5	0
<i>Meridion circulare</i>	0	1.4	0	3.8	0	1.4
<i>Navicula cryptocephala</i>	0	1.2	0	0	0	0
<i>Navicula pupula</i>	0	0	0	0	0.5	0
<i>Navicula salinarum v intermedia</i>	2.7	2.1	0	2	0.5	1.8
<i>Navicula sp.</i>	7.7	4.1	0.5	6.8	0.5	1.9
<i>Neidium sp.</i>	0.3	0.8	0	2.7	0	0
<i>Nitzschia hantzschia</i>	0	0.9	0	0	0.5	7.4
<i>Nitzschia palea</i>	5.8	6.2	0.5	1.4	0.5	0
<i>Nitzschia sigma</i>	0	0.2	0	0	0.5	6.9
<i>Synechra ulna</i>	28.5	10.1	2.3	8.0	1.8	0.5
<i>Synechra ulna v oxyrhynchus</i>	0.4	0.2	0	0	0.5	0.5

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Table 1d - Continued (2 of 2 pages)

SPECIES	SAMPLE, LOCATION					
	Ogilvie River	Swift River	Seagull Creek	Engineer Creek	Screw Creek	Partridge Creek
<i>Tabellaria fenestrata</i>	0	0.3	0	1.4	0	0
<i>Nitzschia aciculans</i>	0	0.3	0	0	0	7.4
<i>Anomoconeis virtrella</i>	1.5	0	0	0	0	0
<i>Fragilaria leptostauron</i>	0.5	0	0	0	0	6.9
<i>Nitzschia linearis</i>	0.5	0	0	0	0	0.5
<i>Oscillatoriopsis sp.</i>	0	0	0	0	0	0.2
<i>Ankistrodesmus falcatus</i>	0	0.2	0	0	0	0
<i>Oedogonium sp. (mm)</i>	0	0.1	0	0	0	0
<i>Ulothrix sp. (mm)</i>	0	0	0.2	0	0	0
<i>Chlamydomonas sp.</i>	0	0.6	0	0	0	0.5
<i>Cryptomonas boralalis</i>	0	0.2	0	0	0	0
<i>Chroomonas acuta</i>	0	0.2	0.5	0	0.9	1.4
TOTAL	267.1	192.1	64.5	200.0	36.5	68.7
						177.9

Table 1e - Phytoplankton Concentrations (17 - 30 June, 1978) (cell/ml) of several Yukon Territory Rivers.
 (analyzed using 500 X magnification). (1 of 2 pages)

SPECIES	SAMPLE LOCATION					
	Ogilvie River	Swift River	Seagull Creek	Engineer Creek	Screw Creek	Partridge Creek
<i>Achnanthes flexella</i>	0.5	0.6	0.9	0.9	0	0.5
<i>Achnanthes minutissima Kutz</i>	41.0	31.5	0.9	26.7	20.3	16.6
<i>Achnanthes</i> sp.	6.9	0.5	0	3.7	0.5	2.3
<i>Amphipleura pellucida</i>	0	0.2	0	0	0	1.4
<i>Amphora</i> sp.	0	0	0	0	0	0
<i>Anomoconis vitrea</i>	1.0	0.8	0	0	0	0
<i>Cocconeis placentula</i>	0.2	0.8	0.5	0	0	0.5
<i>Cyclotella ocellata</i>	0	0.2	0	0	0	0
<i>Cymbella caespitosa</i>	2.2	1.3	0.5	0	2.8	1.4
<i>Cymbella cistula</i>	0.6	0.2	0	0	0	0.5
<i>Cymbella sinuata</i>	3.5	0.3	0	0	0.9	0
<i>Cymbella ventricosa</i>	8.3	2.8	0	0	2.8	1.4
<i>Cymbella</i> sp.	0	0	0	0	0.5	0
<i>Diatoma triemale v mesodon</i>	15.0	1.4	0	0	0.9	1.4
<i>Diatoma tenui velongatum</i>	57.4	21.4	0	23.9	4.1	7.4
<i>Diatoma vulgare</i>	0	0	0.9	0	0	0.5
<i>Diploneis decipiens</i>	0	0	0	0	0	0.5
<i>Fragilaria capucina</i>	19.9	2.8	0	7.4	3.2	4.6
<i>Fragilaria construens v construens</i>	11.1	0.9	0	0	0	0
<i>Fragilaria construens v binodis</i>	4.5	8.3	0	0	7.8	4.6
<i>Fragilaria construens v venter</i>	4.9	1.1	0	0	0	5.5
<i>Fragilaria vaucheriae</i>	12.2	1.9	0	10.1	1.8	11.5
<i>Gomphonema acuminatum</i>	0	0.7	0	0	0	2.8
<i>Gomphonema geminatum</i>	0	0.3	0	0	0.5	0
<i>Gomphonema olivaceum</i>	6.7	3.3	0.5	4.6	10.1	2.3
<i>Gomphonema parvulum</i>	0	2.2	0	0	0	0
<i>Hannaea arcus</i>	25.8	2.3	0.5	1.8	1.4	36.8
<i>Hannaea arcus v amphioxys</i>	0	0	0	0	0.9	9.2
<i>Melosira granulata</i>	0	1.7	0	0	1.8	0
<i>Meridion circulare</i>	8.7	0.2	0	0	0	0.5
<i>Navicula cryptocephala</i>	0.2	0.3	0	0	0	0.5
<i>Navicula radiosa</i>	0.2	0	0	0	0	0.5
<i>Navicula salinarum v intermedia</i>	0	0.2	0	0	0.5	8.3
<i>Navicula tripunctata</i>	0.3	0	0	0	0	0
<i>Navicula</i> sp.	4.1	2.2	0.9	0.5	0	0.9
<i>Neidium</i> sp.	0	0.2	0	0	0.5	1.4
<i>Nitzschia acicularia</i>	0.2	0.2	0	0	0	0
<i>Nitzschia hantzschia</i>	0.3	0	0	0	0	0

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Table 1e - Continued (2 of 2 pages)

SPECIES	SAMPLE LOCATION						Smart Creek
	Ogilvie River	Swift River	Seagull Creek	Engineer Creek	Screw Creek	Partridge Creek	
<i>Nitzschia linearis</i>	0.4	0	0	0	0	0	0.5
<i>Nitzschia palea</i>	0.2	0.6	0.5	0	0	0.5	0.9
<i>Nitzschia sigma</i>	0.4	0	0	0	0	0	0
<i>Rhopalodia gibba</i>	0	0.1	0	0	0	0	0.9
<i>Stephanodiscus astreæ</i>	0	0	0	0	0	0	0.5
<i>Synedra ulna</i>	26.3	10.4	0	5.5	0	1.8	3.2
<i>Synedra ulna v oxyriyehus</i>	12.7	1.5	0	6.4	0	0	0
<i>Tabellaria fenestrata</i>	0.5	0.3	0	0	1.4	0	0
<i>Tabellaria floeculosa</i>	0	0	0	0	0.5	0	0
<i>Cyanophyta:</i>							
<i>Anabaena</i> sp. (mm)	0	0	0	0	0	0	0
<i>Chlorophyta:</i>							
<i>Chlamydomonas</i> sp.	0	0	0.5	0	0	0.5	0
<i>Chrysophyta:</i>							
<i>Dinobryon sertularia</i>	0	0.2	0.5	0	0	0	0
<i>Cryptophyta:</i>							
<i>Chroomonas acuta</i>	0.6	0.1	0.5	0	0	0	0
<i>Cryptomonas borealis</i>	0	0.2	0	0	0	0	0
TOTAL	266.9	103.8	6.7	89.1	66.4	102.8	71.0

Table 1f - Phytoplankton Concentrations (1 - 8 Aug., 1978) (cells/ml) of several Yukon Territory Rivers.
 (analyzed using 500 X magnification). (1 of 2 pages)

SPECIES	SAMPLE LOCATION					
	Ogilvie River	Swift River	Seagull Creek	Engineer Creek	Screw Creek	Partridge Creek
<i>Achnanthes flexella</i>	4.6	0	0	1.1	0	0
<i>Achnanthes minutissima</i>	60.9	7.6	4.6	72.6	8.3	2.8
<i>Achnanthes</i> sp.	0.2	0	0	0	0	0
<i>Amphipleura pellucida</i>	0	0	0	0	0	0
<i>Amphora coffeiformis</i>	0.2	0	0	0	0	0
<i>Amphora</i> sp.	0.4	0	0	0	0	0
<i>Anomoconis vitreae</i>	1.5	0.5	0.9	0	0	0
<i>Cocconeis placentula</i>	0	0.5	0.9	0	0	0
<i>Cymbella affinis</i>	0	0.2	0	0	0	0
<i>Cymbella caespitosa</i>	1.7	0.2	0	0	1.8	0
<i>Cymbella sinuata</i>	0.2	0	0	0	0	0
<i>Cymbella ventricosa</i>	2.6	1.6	0.9	0	2.2	0
<i>Cymbella</i> sp.	0.2	0.5	0	0	0	0
<i>Diatoma hiemale v mesodan</i>	0	0	0.9	0	0	0
<i>Diatoma tenue v elongatum</i>	19.0	1.8	0.9	78.1	0	0
<i>Diatoma vulgare</i>	0.2	0.7	0	0	0	0
<i>Diploneis decipiens</i>	0	0	0	0	0	0
<i>Fragliaria capucina</i>	5.9	4.6	7.4	13.0	0.9	0
<i>Fragliaria construens v binodis</i>	0.4	1.8	0	0	0	0
<i>Fragliaria construens v venter</i>	0	1.0	0	0	0	0
<i>Fragliaria vaucheriae</i>	0.6	0.9	0	0	1.8	0
<i>Gomphonema geminatum</i>	0.4	0	0	0	0	0
<i>Gomphonema herculeana</i>	0	0	0	0	0	0
<i>Gomphonema olivaceum</i>	3.3	1.4	2.8	3.3	0.9	0
<i>Gomphonema parvulum</i>	0	0.7	0	0	0	0
<i>Hormaea arcus vanphioxyx</i>	0	0	0	0	1.8	0
<i>Meridion circulare</i>	0	0.2	0	0	0	0
<i>Navicula cryptocephala</i>	0.6	0.2	0	0	0	0
<i>Navicula tripunctata</i>	0	1.4	0	0	0	0
<i>Navicula viridula</i>	0	0	0	0	0	0
<i>Navicula</i> sp.	0.2	0.9	0	0	0	0
<i>Nitzschia dissipata</i>	0	0	0	0	0.9	0
<i>Nitzschia frustulum</i>	0.2	0	0	0	0	0
<i>Nitzschia hantzschia</i>	0.2	0	0	0	0	0
<i>Nitzschia linearis</i>	0.5	0.2	0	0	0	0
<i>Nitzschia palea</i>	1.5	1.2	0	0	0	0
<i>Nitzschia sigma</i>	0	0.2	0	0	0	0
<i>Stauroneis</i> sp.	0.2	0	0	0	0	0

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Table 1f - Continued: (2 of 2 pages)

SPECIES	SAMPLE LOCATION						
	Ogilvie River	Swift River	Seagull Creek	Engineer Creek	Screw Creek	Partridge Creek	Logjam Creek
<i>Surirella angustata</i>	0.2	0.5	0	0	0	0	0
<i>Synedra ulna</i>	6.2	6.4	0	4.4	2.8	13.8	6.9
<i>Synedra ulna v oxyrhynchus</i>	0	0.7	0.9	0	11.0	14.7	12.4
<i>Hannaea arcus</i>	0	0.7	0.9	0	0	1.8	0
<i>Mitrochria acicularis</i>	0	0.2	0	0	0	0	0
<i>Navicula pupula</i>	0	0.5	0	0	0	0	0
<i>Tabellaria flocculosa</i>	0	0	0	0	1.8	1.1	1.4
<i>Chlorophyta:</i>							
<i>Ankistrodesmus falcatus</i>	0	0.6	0	0	0	0	0
<i>Chlamydomonas sp.</i>	0.6	0.6	0	0	0	0	0
<i>Cosmarium sp.</i>	1.7	0.7	0.9	1.1	0	0	0.9
<i>Cladophora sp.</i>	0	0.2	0	0	0	1.8	0
<i>Cryptophyta:</i>							
<i>Dinobryon vertularia</i>	0	0.8	0	0	0	0	0
<i>Cryptophyta:</i>							
<i>Chroomonas acuta</i>	0.7	0	0	0	0	0	3.7
<i>Cryptomonas borealis</i>	0.2	0.2	0	0	0	0	0
<i>Cryptomonas sp.</i>	0.7	0.3	0	0	0	0	0
TOTAL	116.0	40.2	24.8	196.7	46.9	184.8	174.5

APPENDIX 2

Table 2a - Periphytic Diatom Concentrations (4 - 10 October, 1977) (cells/cm²) of several Yukon Territory Rivers (analyzed using 1000 X magnification). (1 of 4 pages)

SPECIES	SAMPLE LOCATION			
	Ogilvie River	Swift River	Seagull Creek	Engineer Creek
<i>Achnanthes clevei</i> Grun.	0	7734	30	2232
<i>Achnanthes inflata</i> (Kutz.) Grun.	0	256	0	0
<i>Achnanthes flexella</i> (Kutz.) Grun.	6654	174	0	1116
<i>Achnanthes lanceolata</i> (Breb.) Grun.	0	768	0	0
<i>Achnanthes microcephala</i> (Kutz.) Grun.	10295	5603	0	0
<i>Achnanthes minutissima</i> Kutz.	251379	37360	538	8201
<i>Achnanthes</i> sp.	884	297	6	0
<i>amphipleura pellucida</i> Kutz.	386	0	0	0
<i>Amphora</i> sp.	776	0	0	0
<i>Anomoconeis nitrea</i> (Grun.) Ross	27770	3059	0	1674
<i>Anomoconeis zetlensis</i> (Grun.) C1.	0	111	0	0
<i>Anomoceis</i> sp.	0	0	0	0
<i>Asterionella formosa</i> Hass	0	0	0	0
<i>Cocconeis placentula</i> Ehr.	4540	11101	0	0
<i>Cyclotella bodanica</i>	0	0	0	0
<i>Cyclotella glomerata</i> Bachm.	0	3184	0	0
<i>Cyclotella ocellata</i> Pant.	0	0	0	0
<i>Cymatopleura solea</i> (Breb.) W.Sm.	0	0	0	0
<i>Cymatopleura</i> sp.	0	0	0	0
<i>Cymbella caespitosa</i> (Kutz.) Grun.	3728	1818	30	3347
<i>Cymbella cistula</i> Hempr.	0	222	0	0
<i>Cymbella sinuata</i>	2855	881	0	0
<i>Cymbella turgida</i> Grun.	0	0	0	0
<i>Cymbella ventricosa</i> Kutz.	9466	2953	55	8369
<i>Cymbella</i> sp. "A"	202	0	0	0

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Table 2a - Continued: (2 of 4 pages)

SPECIES	SAMPLE LOCATION				
	Ogilvie River	Swift River	Seagull Creek	Engineer Creek	
<i>Cymbella</i> sp.	1282	0	0	0	0
<i>Diatoma niemalae</i> (Roth) Heib	2550	819	0	0	5021
<i>Diatoma niemalae</i>	0	1131	30	0	0
<i>Diatoma tenue</i> var. <i>elongatum</i>	43824	2867	24	0	27895
<i>Diatoma vulgare</i> Bory	2587	546	0	0	0
<i>Diploneis decipiens</i> A. CL.	0	767	0	0	0
<i>Denticula</i> sp.	386	0	0	0	0
<i>Epithemia turgida</i> (Ehr.) Kutz.	0	142	0	0	0
<i>Eunotia pectinalis</i> (Kutz.) Raben Horst	0	0	0	0	0
<i>Eunotia</i> sp.	0	0	1013	67	0
<i>Fragilaria capusina</i> Desm.	733	5051	36	0	0
<i>Fragilaria construens</i> var. <i>binaeis</i> (Ehr.) Grun	1031	5073	0	0	0
<i>Fragilaria construens</i> var. <i>construens</i> (Ehr.) Grun.	0	1598	0	0	0
<i>Fragilaria construens</i> var. <i>venter</i> (Ehr.) Grun	0	1056	0	0	0
<i>Fragilaria cotonensis</i> Kitton	91121	1574	0	0	0
<i>Fragilaria leptostauron</i> (Ehr.) Hust.	0	398	0	0	0
<i>Fragilaria vaucheriae</i> (Kutz.) Peters	1900	13217	212	0	9484
<i>Gomphonema geminatum</i> (Lyngb.) n. schm.	130	471	0	0	0
<i>Gomphonema herculeanum</i>	0	533	0	0	0
<i>Gomphonema intricatum</i>	0	0	0	0	0
<i>Gomphonema olivaceum</i> (Lyngb.) Kutz.	4508	8971	0	0	0
<i>Gomphonema parvulum</i> Kutz.	260	371	0	0	1116
<i>Gyrosigma sciotoense</i> (Sulliv. & Wormley) CL.	0	0	0	0	0
<i>Hannaea arcus</i> (Ehr.) PAIR.	1557	266	26	0	1116
<i>Hannaea arcus</i> var. <i>amphioxys</i> (Ehr.) Grun.	0	0	0	0	0

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Table 2a - Continued: (3 of 4 pages)

SPECIES	SAMPLE LOCATION			
	Ogilvie River	Swift River	Seagull Creek	Engineer Creek
<i>Meridion circulare</i> (Grev.) Ag.	0	192	65	0
<i>Melosira granulata</i> (Ehr.) Ralfs.	0	2219	13	0
<i>Navicula bicephala</i> Hust.	0	0	0	0
<i>Navicula pupula</i> Kutz.	0	534	0	0
<i>Navicula radiosa</i> Kutz.	0	111	0	0
<i>Navicula Salinarum</i> var. <i>intermedia</i> Grun.	1279	2768	65	1116
<i>Navicula scutellariae</i> W. Sm.	0	111	0	0
<i>Navicula tripunctata</i> (O.F. Mull.) Bory	1086	447	20	558
<i>Navicula</i> sp.	445	266	13	0
<i>Navisula</i> sp.	0	434	26	558
<i>Navicula</i> sp.	0	0	7	0
<i>Neidium</i> sp.	0	0	0	0
<i>Nitzschia acicularis</i> W. Sm.	367	224	0	0
<i>Nitzschia angustata</i> (W. Sm.) Grun	1442	0	0	0
<i>Nitzschia frustulem</i> (Kutz.) Grun	4479	1326	0	1116
<i>Nitzschia hantzschii</i> Rabh	0	1026	0	0
<i>Nitzschia linearis</i> W. Sm	666	0	0	0
<i>Nitzschia paeae</i>	7428	4395	33	3347
<i>Nitzschia</i> sp.	0	0	0	0
<i>Nitzschia dissipata</i> (Kutz.) Grun.	0	1708	0	0
<i>Pinnulari</i> sp.	0	0	0	0
<i>Rhopalodia gibba</i> (Ehr.) C. Mull.	0	557	0	0
<i>Stauronectis phoenicentron</i> (Nitz.) Ehr.	0	0	0	0
<i>Stauronectis anceps</i> Ehr.	130	0	0	0
<i>Stauronectis</i> sp.	367	59	0	0

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Table 2a - Continued:
(4 of 4 pages)

SPECIES	SAMPLE LOCATION			
	Ogilvie River	Swift River	Seagull Creek	Engineer Creek
<i>Surirella angustata</i> Kutz.	0	0	0	0
<i>Surirella ovata</i>	0	0	0	0
<i>Synedra angustata</i>	0	0	0	0
<i>Synedra delicatissima</i> W. Sm.	733	0	0	0
<i>Synedra radians</i> Kutz.	1135	0	0	0
<i>Synedra ulna</i> (Nitz.) Ehr.	5639	4137	0	8926
<i>Tabellaria fenestrata</i> (Lyngb.) Kutz.	0	237	0	0
<i>Tabellaria flacculosa</i> (Roth) Kutz.	0	739	20	0
TOTAL	496,000	142,875	1,316	159,002

Table 2b - Periphytic Diatom Concentrations (10 - 20 December, 1977) (cells/cm²) of several Yukon Territory Rivers (analyzed using 1,000 X magnification).

SPECIES	SAMPLE LOCATION	
	Ogilvie River	Swift River
<i>Achnanthes flexella</i>	2322	0
<i>Achnanthes lanceolata</i>	0	55
<i>Achnanthes microcephala</i>	791	32
<i>Achnanthes minutissima</i>	17563	407
<i>Achnanthes</i> sp.	144	233
<i>Amphipleura pellucida</i>	0	54
<i>Amphora coffeaeformis</i>	68	14
<i>Amphora</i> sp.	127	63
<i>Anomconcis vitrea</i>	8147	31
<i>Cocconeis placentula</i>	133	581
<i>Cyclotella ocellata</i>	273	131
<i>Cyclotella comta</i>	0	68
<i>Cymbella affinis</i>	76	0
<i>Cymbella caespitosa</i>	817	108
<i>Cymbella cistola</i>	468	0
<i>Cymbella sinuata</i>	0	79
<i>Cymbella ventricosa</i>	151	0
<i>Diatoma tenui v elongatum</i>	1323	30
<i>Denticula elegans</i>	51	0
<i>Fragilaria capucina</i>	68	0
<i>Fragilaria construens v binadis</i>	29	179
<i>Fragilaria construens v construens</i>	0	0
<i>Fragilaria construens v venter</i>	43	62
<i>Fragilaria leptostauron</i>	0	74
<i>Fragilaria vaucheriae</i>	67	48
<i>Frustulia rhomboides</i>	0	32
		33

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Table 2b - Continued: (2 of 2 Pages)

SPECIES	SAMPLE LOCATION	
	Ogilvie River	Swift River
<i>Gomphonema acuminatum</i>	0	21
<i>Gomphonema olivaceum</i>	388	161
<i>Gomphonema parvulum</i>	57	0
<i>Gomphonema</i> sp.	193	108
<i>Harnaea arcus</i>	51	0
<i>Melosira granulata</i>	0	8
<i>Navicula bicephala</i>	0	21
<i>Navicula salinarum v intermedia</i>	101	19
<i>Navicula tripunktata</i>	68	0
<i>Navicula</i> sp	0	30
<i>Neidium</i> sp.	0	17
<i>Nitzschia frustulum</i>	0	86
<i>Nitzschia palca</i>	0	18
<i>Nitzschia</i> sp.	205	70
<i>Rhopalodia gibba</i>	0	6
<i>Synechra acus</i>	671	167
<i>Nitzschia dissipata</i>	133	0
TOTAL	34,528	3,046

Table 2c - Periphytic Diatom concentrations (22 - 30 March, 1978) (cells/cm^2) of several Yukon Territory Rivers.
 (analyzed using 1000 x magnification).

SPECIES	SAMPLE LOCATION		
	Ogilvie River	Swift River	Seagull Creek
<i>Achnanthes flexella</i>	44,966	1,688	1,419
<i>Achnanthes lanceolata</i>	379	0	20
<i>Achnanthes microcephala</i>	0	0	141
<i>Achnanthes minutissima</i>	279,783	13,688	14,500
<i>Achnanthes</i> sp.	353	111	305
<i>Amphipleura pellucida</i>	0	0	0
<i>Amphora</i> sp.	6,590	22	492
<i>Anomooneis virtea</i>	5,143	416	742
<i>Cocconeis placentula</i>	3,655	955	661
<i>Cyclotella cincta</i>	0	228	0
<i>Cyclotella glomerata</i>	0	0	569
<i>Cyclotella ocellata</i>	1,589	376	0
<i>Cymbella affinis</i>	0	0	0
<i>Cymbella caespitosa</i>	949	76	0
<i>Cymbella cistula</i>	2,278	0	0
<i>Cymbella sinuata</i>	477	307	179
<i>Cymbella ventricosa</i>	8,690	702	120
<i>Cymbella</i> sp.	0	290	199
<i>Diatoma hiemale</i>	0	0	0
<i>Diatoma hiemale v mesadon</i>	54,089	216	0
<i>Diatoma tenue v elongatum</i>	23,351	1,856	1,482
<i>Diatoma vulgare</i>	8,891	1,188	0
<i>Fragilaria capucina</i>	12,787	0	0
<i>Fragilaria construens v construens</i>	477	687	13
<i>Fragilaria construens v binodis</i>	9,141	907	369
<i>Fragilaria construens v venter</i>	6,136	1,358	1,622
<i>Fragilaria crotonensis</i>	73,785	452	559
<i>Fragilaria leptostauron</i>	1,571	203	105
<i>Fragilaria vaucheriae</i>	6,122	171	171
<i>Frustulia rhomboides</i>	0	76	0
<i>Gomphonema lanceolatum</i>	5,550	15	0
<i>Gomphonema geminatum</i>	2,508	0	0
<i>Gomphonema herculeanum</i>	0	6	0
<i>Gomphonema olivaceum</i>	12,025	1,399	813
<i>Gomphonema parvulum</i>	0	82	6
<i>Gyrosigma sciotense</i>	0	0	67
<i>Hannaea arcus</i>	2,031	0	0
<i>Hannaea arcus v amphioxys</i>	0	15	0

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Table 2c - Continued (2 of 2 pages)

SPECIES	SAMPLE LOCATION		
	Ogilvie River	Swift River	Seagull Creek
<i>Melosira granulata</i>	0	3	0
<i>Meridion circulare</i>	9,479	82	0
<i>Navicula radiosa</i>	0	3	0
<i>Navicula salinarum v intermedia</i>	1,931	409	82
<i>Navicula viridula</i>	0	188	0
<i>Navicula</i> sp.	1,515	0	141
<i>Neidium</i> sp.	0	0	0
<i>Nitzschia dissipata</i>	2,741	640	13
<i>Nitzschia frustulum</i>	6,334	607	180
<i>Nitzschia hantzschiae</i>	176	0	0
<i>Nitzschia linearis</i>	10,415	7	0
<i>Nitzschia palea</i>	30,979	1,282	389
<i>Nitzschia sigma</i>	353	15	0
<i>Rhopolodia gibba</i>	0	236	45
<i>Synedra delicatissima</i>	379	44	0
<i>Synedra ulna</i>	21,821	198	156
<i>Synedra ulna v oxyrhynchus</i>	0	0	45
<i>Stephanodiscus astraea</i>	0	152	0
<i>Tabellaria fenestrata</i>	0	11	0
TOTAL	659,439	30,230	25,599

**Table 2d - Periphytic diatom concentrations (15 - 25 May, 1978) (cells/cm^2) of several Yukon Territory Rivers.
(analyzed using 1000 X magnification).**

SPECIES	SAMPLE LOCATION						
	Ogilvie River	Swift River	Seagull Creek	Engineer Creek	Screw Creek	Partridge Creek	Logjam Creek
<i>Achnanthes flexella</i>	937	1,293	129	77	0	0	0
<i>Achnanthes lanceolla</i>	483	102	0	0	0	0	4,686
<i>Achnanthes microcephala</i>	0	41,383	3,336	0	0	0	6,248
<i>Achnanthes minutissima</i>	117,691	7,782	54	1,017	120,382	14,801	110,454
<i>Achnanthes</i> sp.	1,688	0	774	0	4,607	429	0
<i>Amphipleura pellucida</i>	0	2,527	1,492	0	0	0	0
<i>Amphora</i> sp.	0	2,089	0	77	0	0	0
<i>Anomooneis vitrea</i>	0	2,385	4,412	0	8,548	881	37,218
<i>Cocconeis placentula</i>	2,453	471	129	0	0	0	384
<i>cymbella caespitosa</i>	681	157	0	0	0	156	0
<i>Cymbella sinuata</i>	2,471	1,876	345	0	750	466	9,405
<i>Cymbella ventricosa</i>	61	0	366	154	1,510	588	768
<i>Cymbella</i> sp.	179	157	0	0	0	1,876	3,124
<i>De-ticula</i> sp.	3,541	313	129	0	0	0	0
<i>Diatoma hiemale</i>	2,384	198	582	77	38	0	0
<i>Diatoma tenue</i> v <i>elongatum</i>	3,031	38,086	162	0	0	783	7,862
<i>Diatoma vulgare</i>	2,107	0	623	149	3,101	0	160,108
<i>Epithemia turgida</i>	0	469	0	0	979	0	0
<i>Eunotia</i> sp.	0	0	129	0	0	0	0
<i>Fragilaria capucina</i>	0	7,636	2,819	0	1,510	0	3,892
<i>Fragilaria construens</i> v <i>construens</i>	2,536	10,522	257	0	0	0	0
<i>Fragilaria construens</i> v <i>binodis</i>	5,568	2,429	216	0	3,020	2,056	3,839
<i>Fragilaria construens</i> v <i>venter</i>	4,172	6,839	0	308	0	0	2,687
<i>Fragilaria crotonensis</i>	300	1,586	0	0	0	0	0
<i>Fragilaria ioucheriae</i>	2,404	3,816	1,022	32	3,058	4,797	31,982
<i>Fragilaria leptostauran</i>	7,954	0	0	154	0	0	0
<i>Gomphonema germinatum</i>	0	0	0	755	0	2,280	2,280
<i>Gomphonema olivaceum</i>	263	1,450	129	154	307	1,482	9,730
<i>Comphonema parvulum</i>	3,151	5,143	386	32	38	392	768
<i>Hannaea arcus</i>	0	3,009	6,680	16	0	294	1,152
<i>Meiosira granulata</i>	0	0	1,850	0	0	0	0
<i>Meridion circulare</i>	316	0	0	129	0	0	0
<i>Navicula radiosa</i>	0	1,320	1,910	0	0	0	1,545
<i>Navicula salinarum</i> v <i>intermedia</i>	28	0	0	0	0	0	0
<i>Navicula tripunctata</i>	0	2,324	129	115	0	0	5,629
<i>Navicula</i> sp.	0	0	0	0	0	0	0

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Table 2d - Continued (2 of 2 pages)

Table 2e - Periphytic diatom concentrations (17 - 30 June, 1978) (cells/cm²) of several Yukon Territory Rivers.
(analyzed using 1000 X magnification).

SPECIES	SAMPLE LOCATION					Logjam Creek
	Ogilvie River	Swift River	Seagull Creek	Screw Creek	Partridge Creek	
<i>Achnanthes flexella</i>	0	2,029	0	0	0	0
<i>Achnanthes minutissima</i>	460,963	53,031	4,591	6,711	63,120	51,288
<i>Achnanthes</i> sp.	10,695	68,283	403	0	0	462
<i>Amphora</i> sp.	15,782	1,469	0	0	0	0
<i>Amphipleura pellucida</i>	0	1,706	0	0	0	66
<i>Anomoeneis vitrea</i>	0	1,259	564	0	0	0
<i>Cocconeis placentula</i>	0	6,851	60	1,249	0	9,591
<i>Cyclotella bodanica</i>	0	0	60	0	0	0
<i>Cyclotella comta</i>	0	988	0	0	0	0
<i>Cyclotella ocellata</i>	0	0	0	0	0	0
<i>Cymbella caespitosa</i>	0	0	0	0	0	0
<i>Cymbella cistula</i>	0	309	0	0	0	0
<i>Cymbella prostrata</i>	0	155	0	0	0	0
<i>Cymbella sinuata</i>	0	1,014	0	0	0	1,848
<i>Cymbella ventricosa</i>	131,740	3,865	310	0	0	1,583
<i>Cymbella</i> sp.	0	2,031	689	0	0	0
<i>Diatoma hiemale</i>	0	1,861	90	0	0	33
<i>Diatoma hiemale v mesodon</i>	344,765	194	1,629	0	0	1,583
<i>Diatoma tenui v elongatum</i>	0	42,071	207	0	0	0
<i>Diatoma vulgare</i>	7,766	126	251	0	0	0
<i>Denticula</i> sp.	3,946	0	0	0	0	0
<i>Epithemia sorex</i>	0	0	0	0	0	0
<i>Epithemia turgida</i>	0	107	275	0	0	0
<i>Eunotia</i> sp.	0	1,745	41	0	0	0
<i>Fragilaria capucina</i>	0	3,659	0	0	0	66
<i>Fragilaria construens v construens</i>	117,677	8,594	1,282	0	923	1,054
<i>Fragilaria construens v binadis</i>	106,426	16,891	126	0	121	0
<i>Fragilaria construens v venter</i>	0	931	0	0	0	0
<i>Fragilaria crotonensis</i>	28,510	1,464	0	0	0	0
<i>Fragilaria leptostauron</i>	9,924	7,950	1,861	402	3,628	1,643
<i>Fragilaria vaucheriae</i>	0	0	0	0	0	66
<i>Frustulia rhomboides</i>	0	0	0	0	882	0
<i>Gomphonema lanceolatum</i>	0	0	0	0	0	0
<i>Gomphonema geminatum</i>	0	0	0	0	0	99
<i>Gomphonema hemiculeatum</i>	0	0	0	0	0	2,571
<i>Gomphonema olivaceum</i>	49,328	8,571	156	3,476	14,436	0
<i>Gomphonema parvulum</i>	54,764	3,718	75	0	0	0
<i>Gyrosigma sciontense</i>	0	0	0	0	0	0

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Table 2e - Continued (2 of 2 pages)

SPECIES	SAMPLE LOCATION					Logjam Creek
	Ogilvie River	Swift River	Seagull Creek	Screw Creek	Partridge Creek	
<i>Hannaea arcus</i>	0	2,179	376	207	243	197
<i>Hannaea arcus v amphioxys</i>	7,766	0	0	0	1,764	0
<i>Melosira granulata</i>	0	0	0	0	0	0
<i>Melosira granulata v angustissima</i>	0	238	0	0	0	0
<i>Meridion circulare</i>	69,004	0	0	0	376	0
<i>Neidium sp.</i>	0	1,550	0	0	126	0
<i>Navicula convergens</i>	0	0	0	0	0	0
<i>Navicula cryptocephala</i>	0	0	0	0	0	0
<i>Navicula radiosa</i>	0	0	0	0	0	0
<i>Navicula sculletonoides</i>	0	0	0	0	0	0
<i>Navicula tripunctata</i>	0	0	0	0	0	0
<i>Navicula salinarum v intermedia</i>	4,962	2,549	0	0	0	0
<i>Navicula sp.</i>	41,428	2,622	0	0	0	0
<i>Nitzschia dissipata</i>	0	1,032	0	0	0	0
<i>Nitzschia frustulum</i>	19,728	155	0	0	0	0
<i>Nitzschia hantzschia</i>	3,946	2,622	0	0	0	0
<i>Nitzschia linearis</i>	0	44	0	0	0	0
<i>Nitzschia palea</i>	10,695	4,424	0	0	0	0
<i>Rhopalodia gibba</i>	0	226	0	0	0	0
<i>Surirella angusta</i>	0	0	0	0	0	0
<i>Stephanodiscus tenui</i>	0	232	0	0	0	0
<i>Synedra delicatissima</i>	0	1,464	0	0	0	0
<i>Synedra ulna</i>	46,972	19,571	0	0	41	66
<i>Synedra ulna v oxyrhynchus</i>	0	931	0	0	41	0
<i>Tabellaria flocculosa</i>	0	0	0	0	0	0
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	1,546,787	286,423	14,212	14,312	90,200	74,426

Table 2f - Periphytic Diatom concentrations (1 - 8 August, 1978) (cells/cm²) of several Yukon Territory Rivers.
 (analyzed using 1000 X magnification).

SPECIES	SAMPLE LOCATION						Castle Creek
	Ogilvie River	Swift River	Seagull Creek	Engineer Creek	Screw Creek	Partridge Creek	
<i>Achnanthes flexella</i>	29,269	15,649	0	0	0	0	0
<i>Achnanthes lancelota</i>	0	2,227	0	0	0	0	0
<i>Achnanthes minutissima</i>	646,293	260,519	50,784	293,315	5,897,160	33,534	37,802
<i>Achnanthes sp.</i>	10,559	30,765	0	0	0	0	0
<i>Amphipleura pellucida</i>	0	7,428	0	0	0	0	0
<i>Amphora sp.</i>	0	4,454	0	0	0	0	0
<i>Anomocercis vitrea</i>	26,921	267	1,270	0	0	0	0
<i>Cocconais placentula</i>	643	23,600	0	0	0	0	0
<i>Cyclorella glomerata</i>	0	12,561	0	0	0	0	0
<i>Cyclotella ocellata</i>	0	0	0	0	0	0	0
<i>Cymbella affinis</i>	0	1,467	0	0	0	0	0
<i>Cymbella caespitosa</i>	1,134	0	1,270	0	0	0	0
<i>Cymbella cistula</i>	0	3,217	0	0	0	0	0
<i>Cymbella prostrata</i>	0	3,217	0	0	0	0	0
<i>Cymbella sinuata</i>	2,657	1,467	0	0	50,619	0	14,176
<i>Cymbella ventricosa</i>	23,272	11,365	5,078	69,837	101,239	737	0
<i>Denticula sp.</i>	6,490	0	0	0	50,619	0	0
<i>Diatoma hiemale v mesodon</i>	0	3,295	2,539	0	0	737	0
<i>Diatoma tenua v elongatum</i>	71,632	43,364	0	479,547	0	0	0
<i>Epithemia turgida</i>	0	0	0	0	0	11,813	145,087
<i>Eunotia sp.</i>	52,271	32,917	27,931	0	101,239	0	0
<i>Fragilaria capucina</i>	0	5,945	0	0	1,843	0	22,321
<i>Fragilaria construens v construens</i>	26,614	62,064	3,174	0	253,097	3,685	4,725
<i>Fragilaria construens v binodis</i>	0	43,702	0	0	0	1,474	0
<i>Fragilaria construens v venter</i>	2,329	18,338	20,948	74,493	50,619	13,266	56,702
<i>Fragilaria vaucheriae</i>	0	1,114	0	0	0	0	0
<i>Fragilaria leptostauron</i>	0	0	0	0	50,619	1,474	0
<i>Gomphonema geminatum</i>	1,146	0	0	0	0	0	0
<i>Gomphonema intricatum</i>	0	0	0	0	151,858	1,474	54,346
<i>Gomphonema olivaceum</i>	18,281	35,648	0	0	0	0	9,450
<i>Gomphonema ventricosum</i>	0	0	0	0	0	0	0
<i>Gyrosigma sciotense</i>	0	4,684	0	0	0	0	0
<i>Harmaea arcus</i>	2,858	1,734	3,809	0	101,239	1,843	0
<i>Melosira granulata</i>	0	0	0	0	0	0	0
<i>Meridion circulare</i>	0	0	0	0	0	0	0
<i>Navicula cryptaceaephala</i>	0	0	0	0	0	0	0
<i>Navicula pupula</i>	0	0	0	0	0	0	0

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Table 2f - Continued (2 of 2 pages)

APPENDIX 3: Cellular volumes of algal species listed in Appendices 1 and 2.

Species	Volume: μm^3
<i>Achnanthes clevci</i> Grun.	150
<i>inflata</i> (Kutz.) Grun.	160
<i>flexella</i> (Kutz.) Grun.	1270
<i>lancelota</i> (Breb.) Grun.	120
<i>microcephala</i> (Kutz.) Cl.	100
<i>minutissima</i> Kutz.	60
<i>sp.</i>	210
<i>Amphipleura pellucida</i> (Kutz.) Kutz.	1920
<i>Amphora</i> <i>sp.</i>	50
<i>Anomoconeis vitrea</i> (Grun.) Ross	280
<i>zellensis</i> (Grun.) Cl.	370
<i>Asterionella formosa</i> Hass.	220
<i>Coccconeis placentula</i> Ehr.	700
<i>Cyclotella bodanica</i> Eulenst.	1520
<i>comta</i> (Ehr.) Kutz.	845
<i>glomerata</i> Bachm.	200
<i>ocellata</i> Pant.	900
<i>Cymatopleura solea</i> (Breb.) W. SM.	2970
<i>Cymbella affinis</i> Kutz.	4525
<i>caespitosa</i> (Kutz.) Grun.	2070
<i>cistula</i> (Ehr.) Kirchn.	16090
<i>prostrata</i> (Berk.) Cl.	3990
<i>sinuata</i> Greg.	100
<i>turgida</i> Greg.	1760
<i>ventricosa</i> Kutz.	260
<i>sp. "A"</i>	37470
<i>sp.</i>	530
<i>Diatoma hiemale</i> (Lyngb.) Heib.	250
<i>hiemale v mesodon</i> (Ehr.) Grun.	1210
<i>tenue v elongatum</i> Lyngb.	110
<i>vulgare</i> Bory	1460
<i>Diploncis decipiens</i> A. Cl.	630
<i>Denticula elegans</i> Kutz.	350
<i>sp.</i>	430
<i>Frustulia rhomboides</i> (Ehr.) DcT.	960

Species		Volume: μm^3
<i>Epithemia sorex</i>	Kutz.	1720
<i>turgida</i>	(Ehr.) Kutz.	25030
<i>Eunotia pectinalis</i>	(O.F. Mull.?) Rabh.	1830
	sp.	440
<i>Fragilaria capucina</i>	Desm.	1850
<i>construens v construens</i>	(Ehr.) Grun.	210
<i>construens v binodis</i>	(Ehr.) Grun.	480
<i>construens v venter</i>	(Ehr.) Grun.	250
<i>crotonensis</i>	Kitton	360
<i>leptostauron</i>	(Ehr.) Hust.	520
<i>vaucheriae</i>	(Kutz.) Peters	170
<i>Gomphonema acuminatum</i>	Ehr.	620
<i>intricatum</i>	Kutz.	690
<i>lanceolatum</i>	(Ag.) Ehr.	630
<i>olivaceum</i>	(Lyngb.) Kutz.	370
<i>parvulum</i>	Kutz.	980
<i>Gomphoneis herculeana</i>	(Ehr. cl.) (<i>Gomphonema hereuleanum</i>)	4750
<i>Didymosphenia geminata</i>	(Lyngb.) M. Schmidt.	21260
<i>Gomphonema</i> sp.	(<i>Gomphonema gemiactum</i>)	300
<i>Gyrosigma sciotense</i>	(Sulliv. & Wormley) Cl.	14190
<i>Hannaea arcus</i>	(Ehr.) Patr.	2520
	<i>arcus v amphioxys</i> (Rabh.) Patr.	1070
<i>Meridion circulare</i>	(Grev.) Ag.	1670
<i>Melosira granulata</i>	(Ehr.) Ralfs	1800
	<i>granulata v angustissima</i> O.F. Mull	750
<i>Navicula bicephala</i>	Hust.	360
	<i>convergens</i> Patr.	350
	<i>cryptocephala</i> Kutz.	780
	<i>pupula</i> Kutz.	420
	<i>radiosa</i> Kutz.	550
	<i>salinarum v intermedia</i> (Grun.) Cl.	830
	<i>scutelloides</i> W. SM.	970

Species	Volume: μm^3
<i>Navicula tripunctata</i> (O.F. Mull.) Bory	1480
<i>viridula</i> (Kutz.) Kutz.	1690
<i>sp. A</i>	540
<i>sp. B</i>	280
<i>sp. C</i>	480
<i>Neidium sp.</i>	1400
<i>Nitzschia acicularis</i> W.S.M.	280
<i>angustata</i> (W.S.M.) Grun.	920
<i>dissipata</i> (Kutz.) Grun.	410
<i>frustulum</i> (Kutz.) Grun.	170
<i>hantzschia</i> Rabf.	250
<i>linearis</i> W. SM.	3370
<i>palea</i> (Kutz.) W. SM.	645
<i>sigma</i> (Kutz.) W. SM.	500
<i>sp.</i>	660
<i>Pinnularia sp.</i>	940
<i>Rhoicosphenia curvata</i> (Kutz.) Grun.	510
<i>Rhopalodia gibba</i> (Ehr.) O. Mull	13470
<i>Stauroneis phoenicentron</i> (Nitz.) Ehr.	3020
<i>anceps</i> Ehr.	560
<i>sp.</i>	150
<i>Surirella angustata</i> Hust.	3030
<i>ovata</i> Kutz.	15350
<i>sp.</i>	1200
<i>Synedra delicatissima</i> W. SM.	4590
<i>ulna</i> (Nitz.) Ehr.	3460
<i>ulna voxrychhus</i> (Forti) Hust. (<i>Synedra angustata</i>)	3520
<i>radians</i> Kutz.	1240
<i>Stephanodiscus astraea</i> (Ehr.) Grun.	2280
<i>tenuis</i> Hust.	230
<i>Tabellaria fenestrata</i> (Lyngb.) Kutz.	840
<i>flocculosa</i> (Roth) Kutz.	520

Species	Volume: μm^3
<i>Chlorophyta</i>	260
<i>Ankistrodesmus falcatus</i> (Corda) Ralfs.	260
<i>Chlamydomonas</i> sp.	160
<i>Ulothrix</i> sp.	669,660 $\mu\text{m}^3/\text{mm}$ of algae
<i>Oedogonium</i> sp.	951,200 $\mu\text{m}^3/\text{mm}$ of algae
<i>Mougeotia</i> sp.	473,630 $\mu\text{m}^3/\text{mm}$ of algae
<i>Closterium</i> sp.	14380
<i>Cosmarium</i> sp.	5870
<i>Cryptophyta:</i>	
<i>Chroomonas acuta</i> Utermohl	100
sp.	440
<i>Cryptomonas</i> sp.	400
<i>Chrysophyta:</i>	
<i>Dinobryon sertularia</i> Ehr.	1140
<i>Cyanophyta:</i>	
<i>Oscillatoria</i> sp.	13,850 $\mu\text{m}^3/\text{mm}$
<i>Anabaena</i> sp.	10,180 $\mu\text{m}^3/\text{mm}$

Appendix 4. Invertebrates (number sampled in several years)

Appendix 4

Appendix 4 cont'd

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