# STUDY OF MERCURY MOBILIZATION FROM RESUSPENDED DREDGE SPOILS IN HOWE SOUND

bу

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#### ABSTRACT

The rate of release of mercury from suspended Howe Sound sediments under wide ranges of ionic strength and pH was studied. The effect of temperature, poison and storage time was also examined.

In all cases a period of rapid mercury release occurred, usually reaching a maximum value within several hours of sediment addition, followed by a slow decline of mercury levels to an equilibrim value after several days. The maximum mercury released in no case exceeded 30% of the total sediment mercury. Temperature, storage and biological effects were small. Salinity and pH had much greater effects; the maximum mercury release occurred for low pH and high salinity. The flux of mercury across the sedimentestuarine water interface under aerobic conditions was in the order of 100 ng/cm<sup>2</sup>/day.

The variation of disolved, particulate and interstitial mercury over a tidal cycle in the tidal flat area of the Squamish River Estuary suggest that the tidal oscillation is a major mechanism for mobilization of mercury from sediments into overlying Howe Sound waters.

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This report summarizes a study performed between December 6, 1976 and March 18, 1977 under DSS contract 08SS.KF832-6-1065. It is divided into two parts. PART A deals with

- (1) a laboratory investigation of the rate of release of mercury from resuspended Howe Sound dredge spoils in relation to temperature, salinity, pH, length of water storage time and biological effects; and
- (2) a laboratory investigation of the flux of mercury across the sediment water interface under aerobic conditions.

PART B deals with a field study conducted December 1 to December 2, 1976 on the Squamish River tidal flats and concerned with the measurement of dissolved, particulate and interstitial mercury over a tidal cycle.

Because of the large amount of data generated in this study, only a very general treatment of the data was possible. Note also that all mercury values reported represent the sum of inorganic and methyl mercury fractions.

#### 1. INTRODUCTION

Numerous laboratory experiments, for example those of Lee et al. and Lindberg et al., studying the release of potentially hazardous or undesirable components from resuspended dredge spoils have been conducted in recent years as a response to the anticipated large-scale ocean dumping of contaminated dredge spoils. In most cases, only qualitative results were obtained because insufficient care was exercised in experimental design. The common usage of small volume containers for reaction vessels in these studies has at least two drawbacks. Firstly, a large surface area to volume ratio increases the importance of wall effects especially at trace and ultra trace concentrations. Secondly, comparison of results from a large number of containers in time studies greatly increases the data scatter because of variations in the properties of the containers themselves. In this study, an attempt was made to minimize container effects by using specially cleaned 50-l glass vessels. Such large containers also permitted serial sampling for a given experimental condition in a single vessel.

An important factor in release experiments, seldom studied yet frequently used to excuse inconsistent results, is the presence of biological communities. There are two main reasons for anticipating that this effect is important especially in the case of mercury:

- (1) certain bacteria can reduce Hg(II) to Hg(0) which can then be lost through volatization. Baier et al. (1975) found that, at the sub-ppb level of dissolved mercury and at natural water pH, over 80% of the mercury in solution may be lost within days due to this effect; and
- (2) the suspension of dredge spoil in water results in the release of nutrients which encourages the proliferation of bacteria.

An estimate of the magnitude of biological effects was made in this study by comparing the Hg release rate for a biologically active and biologically inactive case. The biologically inactive case was created by the use of CuSO<sub>4</sub> as a poison. Copper sulphate was chosen over other biocides because it was effective, safe and supposedly free from undesirable properties such as surface activity and mercury complexing ability.

As in any research situation there were several unexpected developments. The most severe of these was an air contamination problem. Ambient levels of mercury in the air of our laboratory increased from < 1 ng/k to in excess of  $10^6 \text{ ng/k}$ . The sewer system was identified as the source of the contamination. This was verified by measuring high levels of airborne mercury in buildings in the same vicinity of the laboratory. The problem was eventually solved but unfortunately samples for the static release experiment (anaerobic case) were contaminated and the experiment was therefore lost.

## EXPERIMENTAL MATERIALS AND METHODS

## 2.1 Materials Used

## 2.1.1 Water

For the purposes of this study the salinity definitions for freshwater, estuarine water and seawater were  $<10^{\circ}/oo$ ,  $15-25^{\circ}/oo$  and  $>28^{\circ}/oo$  respectively. The water was collected at two different times (1) November 29 - December 2, 1976 (Cruise OC-76-IS-008) and (2) February 14, 1977 (Cruise OC-77-IS-002) at several stations in Howe Sound (for exact positions see appendix A.1)

Polyethylene carboys (220 1) with lids covered and sealed with double thickness 4 mil thick polyethylene sheeting, were used to store the water. Prior to use, each carboy was cleaned by leaching with a 0.02°/o (w/v) freshwater solution of K<sub>2</sub>Cr<sub>2</sub>O<sub>7</sub> in 1 N ACS grade HNO<sub>3</sub> for several days. Before filling with sample water, each carboy was pumped empty using a Jabsco in-line pump, rinsed twice with about 20 liters of sample water then pumped full through acid-cleaned polyethylene tubing during which time the salinity was monitored using a hand-held refractometer/salinometer. All manipulations were accomplished using plastic gloves.

# 2.1.2 Sediment Samples

The sediment used in the experiments was collected at a series of locations on the tidal flats in the vicinity of the FMC plant at Squamish November 18 and 19, 1976 by pounding an acrylic tube into the sediment and withdrawing it. The mercury contents of 2 cm subsamples of these cores were provided by Chemex Labs Ltd., North Vancouver, B.C. These subsamples were, in turn, subsampled after being homogenized by rolling and then used wet in the experiments.

#### 2.1.3 Box Cores

Box cores used in the static release experiments were collected during cruise OC-76-IS-008 using a Bruland-Soutar box corer and kept frozen until used.

# 2.1.4 Chemicals

Sodium hydroxide was purified by adding SnCl<sub>2</sub> (5% by weight) and purging with nitrogen. Reagent grade copper sulphate and mercury-determination-grade nitric acid were used without further purification.

# 2.2 Cleaning Procedures

#### Glassware:

New 500 ml glass-stoppered pyrex bottles were cleaned by the hot oxidation method described by Bothner (1974; also see 3.1 below). As many as three oxidations were necessary before new bottles provided an acceptable blank value. Fifty-litre pyrex containers used as reaction vessels for the release experiments were cleaned prior to initial use by filling with a 0.1% (w/v) solution of K<sub>2</sub>Cr<sub>2</sub>O<sub>7</sub> in 1 N ACS grade HNO<sub>3</sub> for three days then emptying and rinsing at least three times with forty liters of tapwater. Between experiments these containers were cleaned as above but only for one day. Other glassware and TFE teflon stirring rods were cleaned by heating (at 70°C for 12 h in ACS grade nitric acid (3 N) then rinsed three times with cold tap water and stored double-bagged in polyethylene until used.

#### Other Equipment:

All TFE teflon and polyethylene tubing, teflon pumps and peristaltic pump tubing was cleaned by recirculating concentrated ACS grade nitric acid for 1 h, then cold tap water five times for 15 minutes or until the recirculating wash water was no longer acidic.

Nuclepore filters were cleaned prior to use by leaching in a 0.01% (w/v)  ${\rm K_2Cr_2O_7}$  solution in 1 N ACS grade nitric acid

overnight then rinsing three times with tap water before use.

Plexiglass containers and lids were cleaned by soaking in 1 N ACS grade nitric acid at room temperature for several hours before rinsing with tap water.

## 2.3 Experimental Procedures

## 2.3.1 General Description of Laboratory Experiments

A series of nine experiments was conducted to measure the release rate of mercury from suspended sediment in water. In addition, an experiment was conducted to measure the flux of mercury across the sediment-water interface. The following is a description of these experiments.

## Experiment 1:

Measurements to determine experimental precision.

# Experiment 2:

A five-day experiment to measure the release rate of mercruy from mercury-rich sediment into freshwater, seawater and estuarine water, CuSO<sub>4</sub> poisoned estuarine water and the variation of dissolved mercury in estuarine water at pH 4 and pH 7.7.

## Experiment 3:

A five-day experiment - repeat of experiment 2 with blanks of seawater and estuarine water at natural pH. Combined with experiment 2, this experiment provides an estimation of the effects of water storage on the rate of mercury release.

## Experiment 4:

A one-day experiment to measure both dissolved and particulate phase mercury to determine whether or not the total mercury concentration changes with time.

# Experiment 5: Effect of Poison

A five-day experiment to determine the effect of  ${\rm CuSO}_4$  (as a biocide) on the rate of mercury release from mercury-rich and Hg-poor sediment at natural pH in estuarine water. Two concentrations of  ${\rm CuSO}_4$  were used - 0.2 g and 2.0 g

per 30 liters solution.

# Experiment 6: Effect of pH

A five-day experiment measuring the rate of mercury release from mercury-rich sediment into freshwater at pH 4,6,7,8 and 9. Experiment 7:

Continuation of experiment 6 using seawater at pH 4,6,7 and 9 and estuarine water at pH 4,6 and 9.

# Experiment 8: Effect of pH

A five-day experiment similar to experiments 6 and 7 but using Hg-poor sediment in the place of mercury-rich sediment. Due to a lack of high salinity water only the freshwater and estuarine water cases were examined.

# Experiment 9:

A 15-day experiment to measure the release rate of mercury into fresh, estuarine and sea waters from Hg-poor sediment at natural pH for poisoned an unpoisoned cases.

# Static Release Experiment:

A thirty-day experiment to measure the migration of mercury across the sediment-estuarine water interface and within the sediment for aerobic and anaerobic conditions using a sediment of known organic carbon content, moisture content and mineralogy.

Due to contamination problems the results of the anaerobic case were lost.

A separate run was not made for the effect of temperature since a combination of experiments 2, and 7 provide the rate of mercury release from sediment at  $\sim 17\,^{\circ}\text{C}$  and  $\sim 6\,^{\circ}\text{C}$ 

#### 2.3.2 Experimental Apparatus

#### 2.3.2.1 Release Experiments

The release experiments were conducted in a series of  $50-\ell$  rectangular Pyrex glass containers, as many as twelve being used at any one time. Each container was covered by a fitted 3/8" plexiglass lid containing two holes - one for a Teflon stirring

rod and a second for a pH electrode and sampling tube. Black and Decker ½" drills (Model #5718) mounted above each container by means of a lab rack and wired in series with dimmer switches provided variable, easily controlled stirring. The chance of drill-generated contamination was reduced by shielding the drills with 6 mil polyethylene sheeting. The entire configuration was enclosed in a wooden box with hinged sides providing easy access to the containers at sampling times. Mechanical noise was reduced by lining the inside surface of the box with ½" foam while radio frequency electrical noise was reduced by grounding the drills to chicken wire which had been stapled over the foam.

## 2.3.2.2 Static Release Experiment

The static release experiments were conducted in plexiglass containers as shown in Figure 1. The container dimensions were designed to exactly fit a 20 cm thick Bruland Soutar box core covered by approximately 10 liters of water. A gas inlet tube (A) allows the maintenance of a nitrogen atmosphere by the introduction of nitrogen gas. A slurry of HgCl<sub>2</sub> is introduced through the injection port (B).

# 2.3.3 Experimental Procedure and Sampling Technique

## 2.3.3.1 Release Experiments

Each cleaned container was filled with water of the desired salinity by pumping it directly from the storage carboys through 3/8" diameter polyethylene tubing using a Saturn teflon pump (Fluorocarbon Company). No attempt was made to resuspend the settled sediment because of the difficulty in producing an homogeneous suspension. Concentrated mercury-free nitric acid or 10 N mercury-free sodium hydroxide was then added to adjust the pH if necessary. After twelve hours of stirring, samples were drawn for dissolved mercury analysis (reported as "background" value) and salinity. Temperature and pH were measured in situ. A homogenized pre-weighed amount of wet sediment was then added

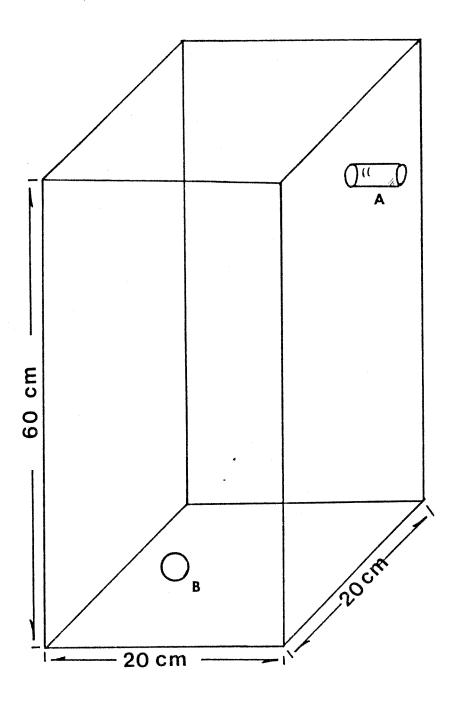


FIGURE 1. Plexiglass container used in Static Release Experiment

to each container and a sample drawn for particle size analysis and dissolved mercury. This sampling period was designated t = Oh. Routinely, the temperature and pH were measured in situ and the pH adjusted to the t = O value, if necessary, before drawing samples for dissolved mercury at t = 1h, 2h, 4h, 8h, 12h, 1d, 2d, 3d, 4d and 5d although these sampling times were modified on occasion to accommodate experimental problems.

At each sampling period, the technique for drawing the dissolved mercury sample was as follows: the teflon-weighted end of 1/8" i.d. teflon tubing was inserted into each container, in turn, to a position exactly six inches from the bottom and sample withdrawn by pumping through a Masterflex peristaltic pump (Horizon Ecology Company). Approximately six tubing dead volumes (150 cm<sup>3</sup>) were discarded ensuring that it was free of traces of the previous sample. Next, 50 ml were pumped into an all-glass Millipore filtration apparatus and filtered through a 0.4 µm Nuclepore filter under reduced pressure into a 1 liter glass filter flask. This 50 ml wash was also discarded. 500 ml was filtered and transferred to a 500 ml ground glass Ten milliliters of a 0.5% (w/v) stoppered sample bottle.  $K_2Cr_2O_7/HNO_3$  solution was then added to the filtrate to prevent adsorption of mercury onto the container walls and the filter stored in a Petri plate. The rate of filtration was variable. In some cases the rate was so slow that less than 500 ml had to be sampled in order to remain on the sampling schedule. The filtration apparatus was washed with  $1\ \mathrm{N}\ \mathrm{HNO}_3$  and tap water between samples and with 0.5% (w/v) K2Cr2O7/HNO3 solution between experiments.

Beginning with experiment three this sampling technique was modified to reduce possible contamination introduced by excessive handling. Accordingly, the 1 liter filter flask was replaced by a 500 ml sample bottle fitted with an adapter that reduced the Millipore holder to a 24/29 ground glass joint. To overcome the problem of losing mercury quickly to the glass walls of the sample bottle, the 10 ml of 0.5% K<sub>2</sub>Cr<sub>2</sub>O<sub>7</sub>/HNO<sub>3</sub> was added prior to the time of filtration.

At all sampling times it was assumed that the sediment was evenly distributed throughout the liquid phase. (No water was added to the containers to replace that pumped out during sampling). The validity of this assumption was tested by measuring the weight of particulates filtered from ten 500 ml samples. Turbulent mixing was provided by 18" glass propellors rotating at two speeds 7 one fast enough to keep all sediment in suspension, the second fast enough to keep only part of the sediment in suspension. Thirty liters of water and 0.600 g of sediment were used. Hence, if homogeneously dispersed, each 500 ml sample should contain 10 mg suspended solids. The results were as follows:

mg suspended solids per 500 ml

LOW SPEED				HIGH SPEED					
(	some sediment	visible	at bottom of	container)	(all	sediment	in	suspension)	

3.25		9.87
1.80		10.12
1.17		9.93
2.40	·	9.98
0.99		10.01
2.11		9.59
2.02		9.90
0.83		9.77
1.48	•	10.25
2.94	•	9.96

The particle size distribution was also measured for each of the above ten samples (see Appendix A.3). Only for high speed stirring were the distributions reproducible. Thus, only for the condition where all sediment is kept in suspension is the above assumption valid.

Unfortunately, the mechanical strength of the glass propellors was insufficient to withstand the strain of prolonged high speed stirring. Accordingly, the glass propellors were replaced by some constructed of 3/4" solid TFE teflon rod into the bottom of which were drilled two holes for the orthogonal placement of two, fourinch  $\frac{1}{2}$ " solid teflon rods. These propellors were then machined to

thread directly onto the rotating shaft of the drill and provided excellent turbulence at low revolutions which decreased wear of the drills substantially. In addition, a series of ten samples of estuarine water were drawn to assess the experimental sampling precision which should give a better estimate to the precision of the dissolved mercury concentrations than the precision obtained by measuring replicate mercury-spiked samples. The ten samples were drawn from the same glass container using the sampling technique began in Experiment 3, as outlined above. Between the drawing of each sample, estuarine water containing in excess of 100 ng/1 was pumped through the hosing to test the efficiency of 150 cm<sup>3</sup> of wash water in eliminating large scale mercury fluctuations. The results were as follows:

Sample	Dissolved Mercury (ng/1)
1	45.92
2	37.34
3	44.10
4	38.00
5	48.56
6	38.66
7	39.32
8	37.34
9	36.04
10	38.28
	$\overline{x} = 40.36$

Thus, the relative standard deviation at the 40 ng/l level is 10.01% or roughly four times the r.s.d. found for replicate analyses of spiked samples.

# 2.3.3.2 Static Release Experiment

Two approximately 20 cm long frozen box core sections were placed in two plexiglass containers (described above) and covered with ten liters of estuarine water (salinity  $= 16.6^{\circ}/\circ$ ). Anaerobic conditions were encouraged in one container by maintaining a nitrogen atmosphere over the core while aerobic conditions

were encouraged in the other by allowing the free movement of air over the core. After two days, background samples were drawn for dissolved mercury determinations and in situ pH values were measured. Then into each sediment, through the injection port 10 cm below the sediment-water interface, 0.23 g of mercury (as  $HgCl_2$ , 10 ml of a 0.023 gHg/ml solution) was injected using a 10 cm cannula. Sampling was then performed once a day for thirty days in the aerobic experiment while the anaerobic experiment was aborted after 15 days because the samples were lost due to a severe airborne mercury contamination problem (the experiments were not run concurrently). At each sampling period, 1000 ml of water were removed (500 ml for a sample and 500 ml for washing the tubing) and 1000 ml of estuarine water were added to maintain the 10 liter volume above the sediment. Control of the amount of water removed and added was critical because the amount of mercury in solution at any time had to be corrected for the fact that mercury rich water was being replaced by mercury poor water (estuarine) at each sampling period.

At the completion of aerobic experiment, nine 20 cm x 2 cm cores were taken in a symmetrical pattern around the original point of injection by pushing acrylic tubes into the sediment, freezing the sediment in a chest freezer then withdrawing the acrylic tubes containing the frozen cores. The cores were then extruded (frozen) and, after thawing, sampled at several depths. These sediment samples were then analyzed for mercury in order to obtain an impression of the three dimensional migration of mercury around the original point of injection.

#### 3. ANALYTICAL PROCEDURES

# 3.1 Mercury in Seawater by Flameless Atomic Absorption

A modified version of the wet chemical flameless atomic absorption technique developed by Bothner (1974) was used to determine total mercury in seawater. The sample is first strongly oxidized, then the mercury is reduced to the elemental state by reduction with stannous chloride and is purged from the sample with a stream of nitrogen gas. Water vapour is removed by passing the nitrogen carrier gas through a magnesium perchlorate drying tube. The dry gas is passed through a U.V. monitor with a 30 cm path length cell, and the absorption is measured at 253.7 nm. The absorbance is proportional to the mercury concentration in the sample.

The Bothner (1974) method was modified to increase analysis rate and precision by increasing the sample volume from 100 to 250 ml, oxidizing the sample directly in the collection flask, decreasing the purge gas flow rate, and damping the recorder signal to reduce instrument noise. By using a 47 µF capacitor across the recorder input, the instrument signal to noise ratio at the 20 ng/l level was ~ 50, compared to ~ 10 without dampling. Use of larger sample volumes (250 ml) gave a flat-top peak for 20-40 seconds as opposed to a fast (2-3 sec) peak for 100 ml sample volumes. The peak response time increased because mercury was purged from a larger volume using a slower (½) rate of nitrogen gas flow.

After incorporating these changes, the precision at the 20-40 ng/l level increased from  $\pm$  20% (2 $\sigma$ ) for 100 ml samples to  $\pm$  5% (2 $\sigma$ ) for 250 ml samples. The detection limit (twice noise) was 0.8 ng/l.

Analysis of a  $1.18 \pm 0.04$  ng/ml (National Bureau of Standards) mercury reference sample using our revised method gave a value of  $1.19 \pm 0.06$  ng/ml.

The recorder sensitivity (or recorder span factor) in ng Hg (free)/mm peak height was determined 2-3 times per day for 4-5 aliquots. Differences in sensitivity between fresh and saline

water, and free mercury vs total mercury (Berrang and Erickson, 1976) are within experimental error. Variations in sensitivity were ± 5% over several weeks. At the 20 ng Hg/l level, the sensitivity was 0.043 ng/nm (at 40 µv/cm recorder and 0.02 U.V. monitor absorbance ranges) with a baseline noise level of 2-3 mm and a long term (several hours) baseline drift of 10-20 mm. For 20 ng/l the signal/noise was ~ 50. Fluctuations in 60 Hz AC line voltages were apparently reduced during night-time hours and week-ends, with a corresponding reduction in baseline noise of 2-5. Subsequently, a voltage line regulator was used for daytime operation.

Some common problems encountered using this technique were high (oxidizing agent) blanks, sample flask contamination, "bubble-blanks", mercury absorption onto flask walls and AC input line fluctuations. Since specially purified  $KMnO_4$  and  $K_2S_2O_8$ were not available, ACS reagent grade purity was used to oxidize total merucyr samples. Blanks from 5 ml of 5% (w/v) solutions of  $\mathrm{KMnO_4}$  and  $\mathrm{K_2S_2O_8}$  per 250 ml sample were 1.0 ng (i.e. a 4 ng/1 correction). Cold tap water, containing 15 ng Hg/1, was used to clean, rinse, prepare solutions and standards etc., and gave a blank of  $\tilde{}$  0.15 ng (or  $\tilde{}$  10% of the total reagent blank). Whereas the SnCl  $_2$  and NH  $_2\cdot$  OH HCl solutions could be purged with N<sub>2</sub> gas to remove mercury, and clean nitric acid purchased, KMnO<sub>4</sub> and K2S208 are not readily purified, except by perhaps subliming  $\mathrm{KMnO}_4$  and recrystalizing  $\mathrm{K_2S_2O_8}$ . Use of lesser amounts of both  $\mathrm{KMnO}_4$  and  $\mathrm{K_2S_2O_8}$  may be feasible, although time did not permit an investigation of this possibility. New sample flasks require very rigorous cleaning (at least two hot oxidizing steps of 2-3 hours each) to eliminate bottle blanks. The "bubble-blank", i.e. the difference in the baseline between sample by-pass and sample purging was 2-3 mm, apparently due to a 25% increase in N<sub>2</sub> line pressure in the by-pass mode. Initially, a very large ( $^{\sim}$  50 mm) "bubble-blank" was encountered, and subsequently reduced by oxidizing the polyethylene drying tube and packing the Mg  $(C10_4)_2$ .

Since few mercury values analysed during this study were below  $\sim 30 \text{ ng/k}$ , a reagent blank of  $\sim 4 \text{ ng/k}$  and a baseline noise level equivalent to 0.4 ng/k were acceptable.

#### 3.1.1 Procedure

## 3.1.1.1 Preparation of Mercury Standard Solutions

A 1,000 ppm mercury solution (Fisher Scientific Co. Ltd.) was diluted first to 1 ppm, then to 10 ppb (ng/ml) with a tap water solution\* containing 2% (v/v) nitric acid (specially purified for mercury determination) ++ and 0.01% (w/v) potassium dichromate (ACS grade). The 10 ppb working standardwas prepared daily.

## 3.1.1.2 Preparation and Purification of Reagents

# Nitric acid/potassium dichromate solution

A solution containing 0.5% (w/v) potassium dichromate (ACS grade) in mercury free nitric acid<sup>++</sup> was added to a 1 liter digidispensing pipettor. The potassium dichromate was dissolved in a small volume of tap water prior to the addition of nitric acid.

## Stannous chloride/sulphuric acid solution

A 20% (w/v) stannous chloride (ACS grade) solution and a 20% (v/v) sulphuric acid (ACS grade) solution were prepared with tap water and purged with  $N_2$  gas for 4-6 hours to remove mercury. The purified solution should have < 1 ng/ $\ell$  mercury.

# Hydroxylamine hydrochloride solution

Five ml of the stannous chloride solution were added to approximately 2 liters of a 12% (w/v) hydroxylamine hydrochloride (ACS grade) solution prepared with tap water. This solution was purged for 4-6 hours with N $_2$  gas to remove mercury. The mercury blank should be < 1 ng/ $\ell$ .

<sup>\*</sup> Mercury contamination from this solution represents ~ 0.2% error in the 10 ppb standard.

<sup>++</sup> Available from Canadian Laboratories Co. Ltd.

# Potassium permangante solution

Two liters of a 5% (w/v) potassium permanganate (ACS grade) solution was prepared with tap water. Aliquots of this solution were measured for mercury blank before additions to samples and discarded if > 0.15 ng/ml solution.

# Potassium persulphate solution

Prepared in the same way as the potassium permanganate solution above.

## 3.1.1.3 Sample Analysis

The 500 ml sample and 10 ml nitric acid/dichromate solution were aliquoted into two 255 ml portions. Five milliliters of persulphate and 5 ml of permanganate solution were added to each aliquot prior to heating at 80°C for two hours in a water bath. Two and one-half milliliters of hydroxylamine hydrochloride solution were added to the cooled samples and the permanganates dissolved with gentle swirling. The air space above the sample solution was purged with N<sub>2</sub> gas for one minute to remove traces of chlorine gas, since chlorine absorbs at 253.7 nm. Just prior to analysis, 10 ml of stannous chloride solution were added, the diffuser inserted, and the sample shaken for 30 seconds, let stand for 30 seconds and purged with N<sub>2</sub> gas at a flow rate of 0.4 1/min for approximately 1 minute and the peak height measured in mm. Peak heights from two 250 ml aliquots were averaged for each sample.

The instrument settings were:

U.V. Monitor (Laboratory Data Control, Riviera Beach, Florida - 30 cm path length cell)
Range - 0.02 Absorbance

Recorder (Fisher Recordall - Series 5000) Range - 1 mv Full Scale (25 cm) Chart speed - 2 inches/minute

Nitrogen gas (Grade G) flow rate - 0.4 l/minute.

The system waspurged between samples using tap water. The 6 cm (length) x 2 cm (diameter) polyethylene drying tube was packed with fresh ACS grade magnesium perchlorate after analysis of  $^{\sim}50$  aliquots. Glass wool is used at each end of the drying tube to prevent Mg(ClO<sub>4</sub>) from entering the U.V. gas cell.

# 3.1.2 Blanks and Calibration Samples

Total reagent blanks were determined as follows: To a 500 ml flask containing 250 ml tap water were added 5 ml of nitric acid/dichromate, 2.5 ml of hydroxylamine hydrochloride, 5 ml of persulphate, and 5 ml of permanganate solutions. After gentle swirling, 10 ml of stannous chloride solution were added and the mercury purged with  $N_2$  gas. Precision of peak heights should be  $\pm$  5-10%. Peak heights should be less than 4 ng/1.

The recorder span factor (ng Hg/mm peak height) was determined by spiking each of 3-5 aliquots of 250 ml of tap water, containing 5 ml nitric acid/dichromate solution, with 5 ng Hg. The precision (20) of (free) mercury peaks should be 5%. Standard spiked samples were done prior to every run (9 samples).

# 3.1.3 Calculation of Results

Mercury concentrations in ng/liter are calculated as follows:
C (ng/l) = [Average of two sample peak heights (mm) - Total Reagent Blank (mm)].

3.2 Mercury in Sediments

About 2 g of wet sediment was placed into a teflon decomposition bomb and 15 ml of freshly prepared aqua-regia added. After standing for 1 h at room temperature, the solution was brought to boiling for an additional hour. The digest was then cooled and transferred to a 500 ml glass sample bottle and analyzed in duplicate as per the method of section 2.4.1 omitting the oxidation step.

3.3 Mercury in Particulate Phase

The Nuclepore filters and filtered particulate matter were transferred directly into 500 ml sample bottles then 10 ml of 0.5% (w/v)  $\rm K_2 \rm Cr_2 \rm O_7$  in mercury-free nitric acid and 500 ml tap water added. This volume was then digested by the procedure of 3.1 , aliquotted and mercury determined in duplicate.

3.4 Organic Carbon in Sediments

The Walkley-Black titrimetric procedure as outlined in Black et al (1972) was used to determine the organic matter in sediments. Oxidizable matter in the sample is first oxidized by K<sub>2</sub>Cr<sub>2</sub>O<sub>7</sub> and the reaction is facilitated by the heat generated when two volumes of H<sub>2</sub>SO<sub>4</sub> are mixed with one of 1N K<sub>2</sub>Cr<sub>2</sub>O<sub>7</sub>. The excess Cr<sub>2</sub>O<sub>7</sub> is determined by titration with standard FeSO<sub>4</sub> solution and the quantity of substances oxidized is calculated from the amount of Cr<sub>2</sub>O<sub>7</sub> reduced. Ferroin (o-phenanothroline ferrous sulphate) is used as the indicator. Since Cl interferes by producing a positive error, Ag<sub>2</sub>SO<sub>4</sub> is added to precipitate out the Cl. The oxidizable organic matter determined by this method differs from the total carbon content by a correction factor, f equal to 1.33. Then,

% Organic Carbon =  $\frac{(\text{meq K}_2\text{Cr}_2\text{O}_7 - \text{meq FeSO}_4)}{\text{g water-free sediment}} \times 1.33 \times 0.003 \frac{\text{gC}}{\text{meq C}} \times 100\%$ 

The reagent blank was less than 0.01 ml and precision at 2% organic carbon content was  $\pm$  5% (2 $\sigma$ ).

# Reagents

(1) 1 N  $K_2 Cr_2 O_7$ :

49.00 g ACS grade K<sub>2</sub>Cr<sub>2</sub>O<sub>7</sub> (pulverized and dried at 105°C for 3-4 h) dissolved in Milli-Q water and made up to 1 liter in a volumetric flask.

(2)  $H_2SO_4/Ag_2SO_4$ :

15.0 g  $Ag_2SO_4$  added to 1 liter of  $H_2SO_4$ .

(3)  $0.5 \text{ N FeSO}_4$ :

278 g ACS FeSO<sub>4</sub>-7H<sub>2</sub>O dissolved in Milli-Q water. Eighty ml conc H<sub>2</sub>SO<sub>4</sub> added solution allowed to cool and made up to one liter.

(4) Indicator:

o-phenanthroline ferrous sulphate prepared by the method outlined by
Kolthoff and Sandell. Stoichiometric
amount of o-phenanthroline monohydrate
dissolved in a 0.025 M solution of
ferrous sulphate in Milli-Q water.

#### Procedure

The sediment was ground to pass a 0.42 mm nylon sieve. Five ml of 1 N  $\rm K_2Cr_2O_7$  were added to 0.1 g of sediment in an Erlenmeyer flask and the soil dispersed into the solution with vigorous swirling. Next, 10 ml of conc.  $\rm H_2SO_4$  were added and, after swirling for about one minute, the flask placed on an asbestos sheet for 30 minutes. The solution was then diluted with about 200 ml Milli-Q water, 3-5 drops of ferroin were added and the solution titrated to a red brown end point with 0.05 N  $\rm FeSO_4$ . The  $\rm FeSO_4$  was standardized daily.

3.5 Dissolved Organic Carbon

The procedure was similar to the one used for sediment except that the concentrations of the oxidizing and reducing agents were decreased by an order of magnitude. The size of sample was 25 ml.

3.6 pH

The pH  $(\pm~0.05~\text{unit})$  was measured in situ using a Horizon Ecology Digi Sense Model 5985-40 portable pH meter with combination electrode.

3.7 Salinity

The salinity  $(\pm 0.1^{\circ}/o_{\circ})$  was measured with an Environmental Devices Corporation Type 102 Handheld Refractometer/Salinometer.

3.8 Temperature

The temperature  $(\pm 0.05 \text{ C}^{\circ})$  was measured using a thermometer (range  $0^{\circ}$ C to  $50^{\circ}$ C) calibrated against an NBS thermometer.

3.9 Particle Size Spectra

Particle size distributions were measured using a Coulter Counter Model TA II adjusted to the following instrumental parameters:

Operation Mode: Concentration Index

Scope Display: Differential

Auto/Man: Auto
Print/Plot: Print
Sampling volume: 2 ml
Active Channels: 2-15

Size Calibration Channels: 10-11

Aperture tube: 200 µm

The instrument was calibrated according to the procedure outlined in the Operator's Manual for Coulter Counter Model TA II using 27.3  $\mu m$  pollen grains dispersed in ethanol. Samples were run by the method as outlined in the manual.

# 4. RESULTS AND DISCUSSION

## 4.1 Release Experiments

The data for the release experiments are tabulated in Appendix A.3. In all the experiments, the relationship between released mercury and time was similar. After the addition of sediment, there was a rapid increase in dissolved mercury, usually within the first hour, followed by a decrease to an apparent equilibrium value after several days. This behaviour is similar to that observed in similar studies by Lindberg et al (1975), Lee et al (1975), and Thomas (1976). Two essential characteristics of the release-time curves (Figures 2 to 29), (a) the maximum mercury released per gram sediment and (b) the equilibrium value of released mercury per gram sediment will be used for comparing the experimental results. In no case was the amount of mercury released greater than 30% of total available in the sediment.

The effect of water storage on the rate of mercury release from suspended sediment can be assessed by comparing the approximate maximum and equilibrium mercury concentrations for seawater, estuarine water and freshwater in Experiments 2 and 3. The following summarizes the results obtained at natural pH using mercury rich sediment:

	Experiment	Experiment 3		
	Max. Released Hg (ng/g)	Eq <sup>m</sup> Hg (ng/g)	Max. Released Hg(ng/g)	Eq <sup>m</sup> Hg (ng/g)
Seawater	1100	120	1123	400
Estuarine Water	650	125	600	150
Freshwater	650	-400	650	100

The generally good agreement of maximum and equilibrium values in the two separate experiments indicates that storage of water has no significant effect on the release of mercury from suspended sediment.

That mercury was not being stripped out of solution by adsorption onto the walls of the glass reaction vessels was confirmed by measuring the dissolved mercury concentration in these containers at pH 3.95 and 7.80 over a period of five days (Figures 2 and 3). A similar result was obtained for all other blank runs. The validity of concluding that changes in the dissolved mercury concentrations at much higher levels than background were the result of sediment water interactions alone was tested by measuring the variation of total mercury in a glass reaction vessel over a 24 hour period. Figures 8 and 9 show the results for seawater and estuarine water respectively. In both cases total mercury is approximately constant with changes in dissolved mercury being compensated by equal and opposite changes in particulate mercury.

The preliminary measurement of the effect of poison (2.0 g CuSO<sub>4</sub>) on the rate of mercury release from mercury-rich sediment in estuarine water (Figure 6) indicated that, compared to the unpoisoned case (Figure 5), there was little difference between release maxima but a large difference between equilibrium values (poisoned lower). Experiment 5 examined the effect of using CuSO<sub>4</sub> at 2.0 g and 0.2 g per 30 litres of estuarine water on the release rates from mercury-rich and relatively mercury poor sediments. Although the concentration of CuSO<sub>4</sub> had little effect on the mercury released, the values in Experiment 5 were much higher than those of Experiment 2. Because mercury values for the CuSO<sub>4</sub> blank were similar to those for non-poisoned blanks, it is reasonable to conclude that rather than a biological effect, the high mercury concentrations in Experiment 5 resulted from a sediment preference for Cu over Hg.

The release of mercury from both mercury-rich and mercurypoor sediment showed similar behaviour for seawater, estuarine water and freshwater (Figures 10-29). The highest release occurred at lowest pH (4) and the lowest release at highest pH (9). At high pH, the equilibrium values may be negative (Figures 16, 17, 20, 21). There is ample evidence to suggest that this could be the result of the high pH flocculation of humic substances which are very effective mercury scavengers (Eckert and Sholkovitz, 1976; Sholkovitz, 1976). For each pH, the maximum release depended on ionic strength (largest for seawater, smallest for freshwater).

The composition of the sediment used in Experiment 9 (Chemex.Core #1, 0-10 cm) was as follows:

(a)	X-ray ar	nalysis	Vermiculite 10A Mica Kaolinite Chlorite les	ss than		40% 20-4 10-2	
			Amphiboles, and quart		•	trac	e
(b)	Organic	carbon				0.15	5%
(c)	Sand Silt Clay	2 mm - 0.0 0.05 mm - less than	0.002 mm		,	91.2 8.3 0.5	%

The same salinity dependence occurred in this 15-day experiment as has already been noted. The difference between the poisoned and unpoisoned cases was quantitatively unimportant for all salinities (Figures 22-29).

The effect of temperature on the release of mercury was small. For example the difference between the release maximum for the mercury-rich sediment-seawater system at  $6^{\circ}$ C (Figure 16) and  $17^{\circ}$ C (Figure 7) was smaller than experimental error.

Several authors (Kurbatov et al (1951); Posselt et al (1968), Loganathan and Burau (1973)) have, in the past, used mass-law relationships to study the adsorption and desorption of trace metals on various adsorbents. This approach can be applied to the exchange reaction,

 $Hg-Sed + n S^{O}/oo \stackrel{K}{\rightarrow} Hg^{++} + n Sed-S^{O}/oo$ 

whence,

$$K = \frac{(Hg^{++}) (Sed-S^{o}/oo)^{n}}{(Hg-Sed) (S^{o}/oo)^{n}}$$

Rearranging and taking logs

$$log (\frac{Hg \ released}{Hg \ sorbed}) = log K + n log (S^{o}/oo) + n log (Sed-S^{o}/oo)$$

Assuming constant sorbent activity which is strictly true only in the case of low exchange, and that  $Hg_{total} = Hg_{dissolved} + Hg_{particulate}$ 

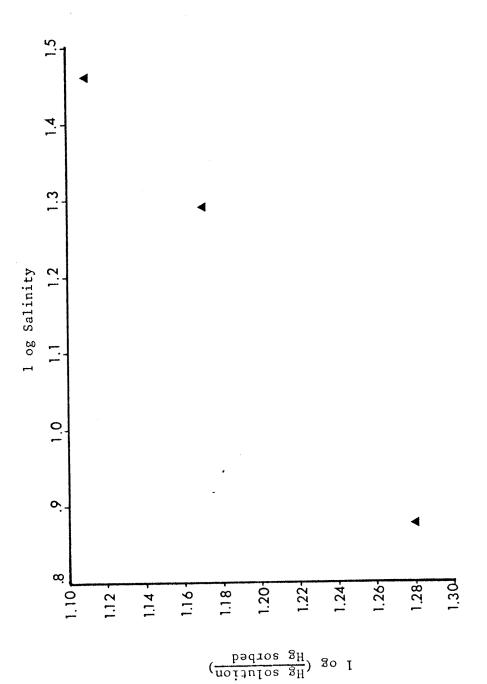
$$\frac{\log \left(\frac{\text{Hg released}}{\text{Hg sorbed}}\right) = \log K' + n \log (S^{0}/oo)$$

Figure 30 is a plot of the logarithm of the solid-solution phase distribution coefficient vs.  $\log (S^{O}/oo)$ , is given for the unpoisoned sea, fresh, and estuarine water cases of Experiment 9. This relationship does not imply that the interaction between mercury and water is controlled by salinity alone but the linearity does suggest the importance of ionic strength to the overall equilibrium. (It is difficult to assign any theoretical significance to the curve because the assumption of constant sorbent activity during sorption is not true here).

# 4.2 Static Release Experiment

The data for the 30-day aerobic static release experiment appear in Appendix A, 4. At t = 0D, the concentration of dissolved mercury was about 400 ng/l increasing steadily to over 40,000 ng/l after 19 days and then decreasing to about 3,000 ng/l by 30 days. These data, once corrected for dilution effects, were used to calculate an average daily flux of mercury across the sediment-water interface by dividing the total change in dissolved mercury by  $400 \text{ cm}^2$ . This gave fluxes as high as  $470 \text{ ng/cm}^2/\text{day}$  but most were less than  $100 \text{ ng/cm}^2/\text{day}$ .

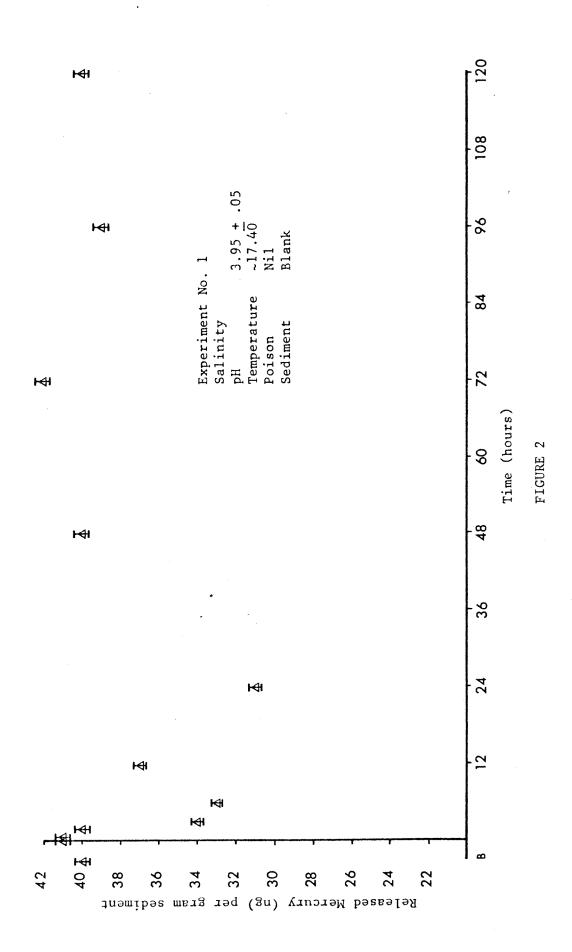
A theoretical flux considering only diffusion transport of mercury through the interstitial waters can be calculated from

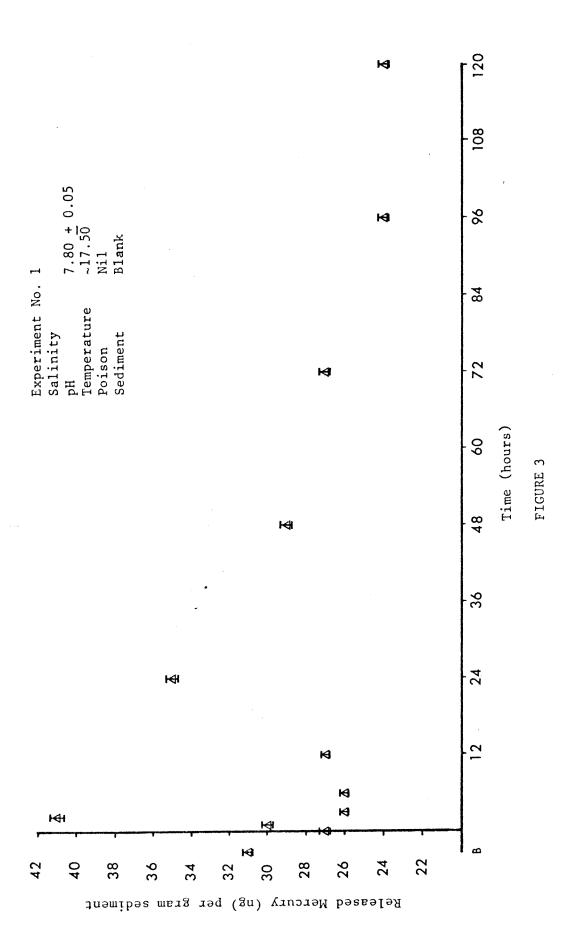


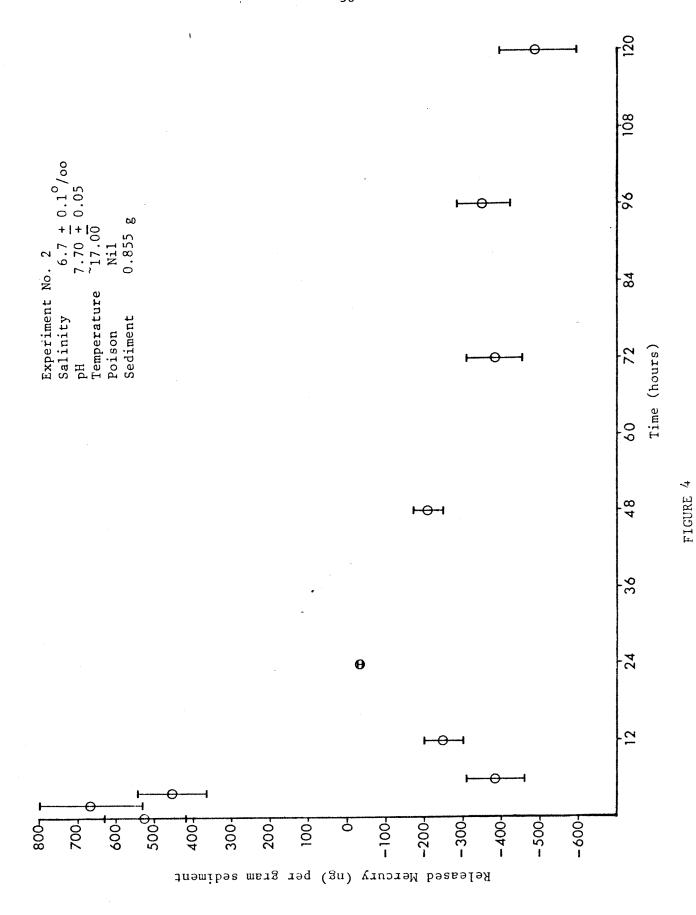
Logarithm of the distribution coefficient as a function of salinity for Experiment 9, unpoisoned case. FIGURE 30.

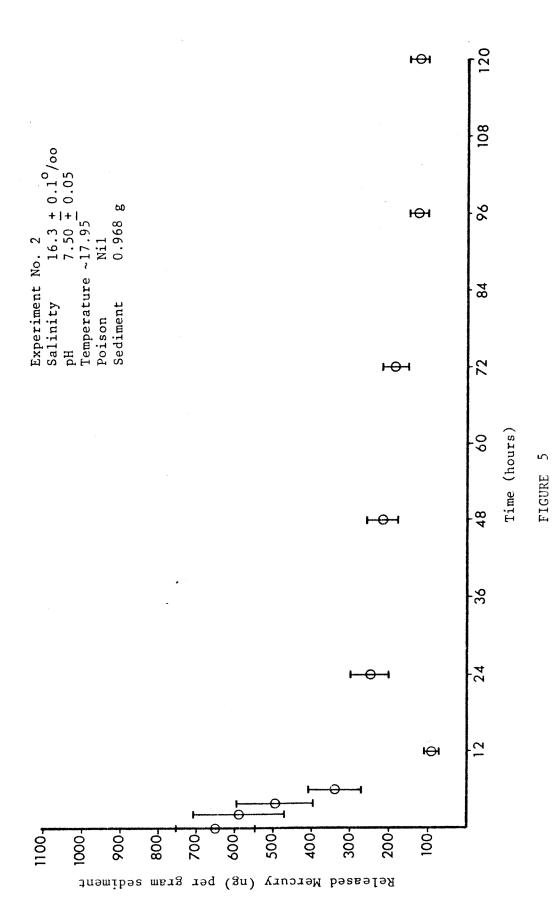
Fick's first law. Assuming a maximum molecular diffusion coefficient of  $10^{-5}$  cm<sup>2</sup>/sec and using a value of 300 ng/l-7 cm for the concentration gradient (based on interstitial water analyses at t = 30 d, 6.0 ppb Hg and 6.3 ppb Hg at 10 cm and 3 cm respectively), the calculated flux is 37 ng/cm<sup>2</sup>/sec which compares favourably with the observed fluxes. In contrast, Lu and Chen (1977) reported that for polluted mercury sediments under aerobic conditions, there was no change in interfacial mercury concentrations over a period of five months.

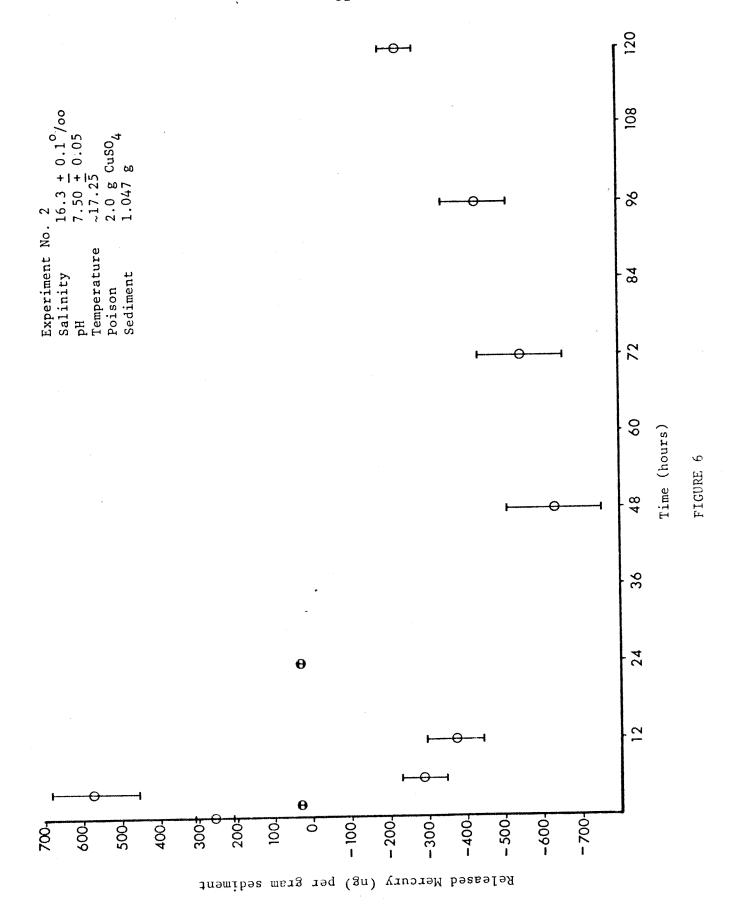
Analyses of sediment at the completion of the experiment indicated that most of the added mercury had been lost. Enough mercury was added to yield an average concentration of 30 ppm in the sediment and interstitial water but no value was found to exceed 1 ppm.

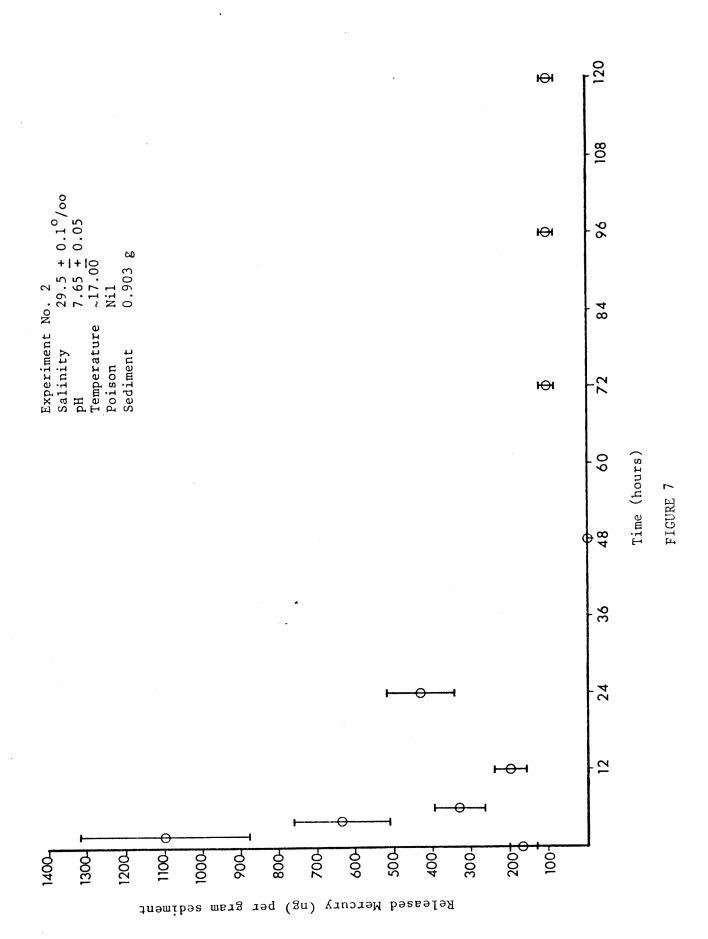


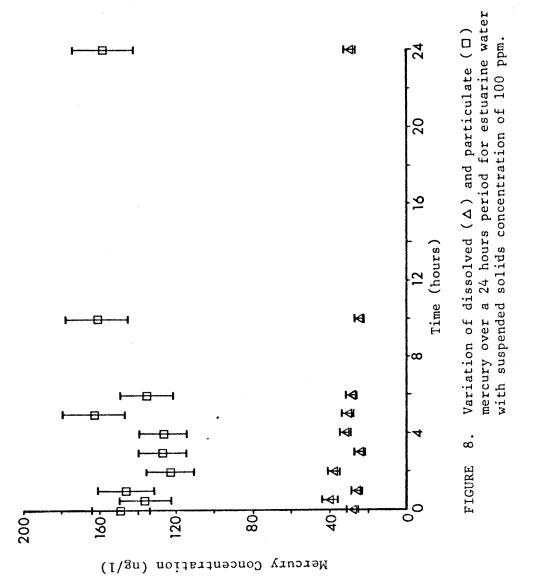


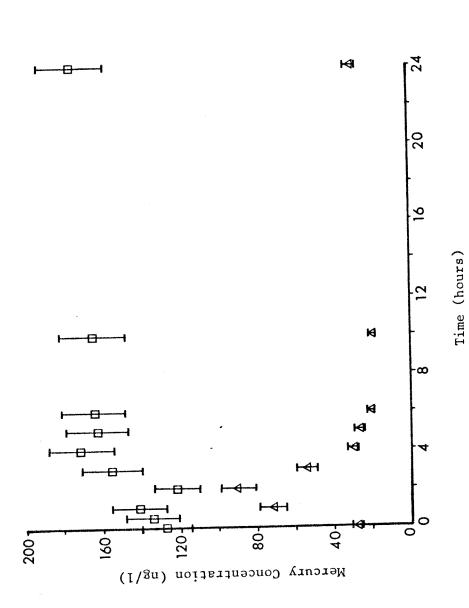




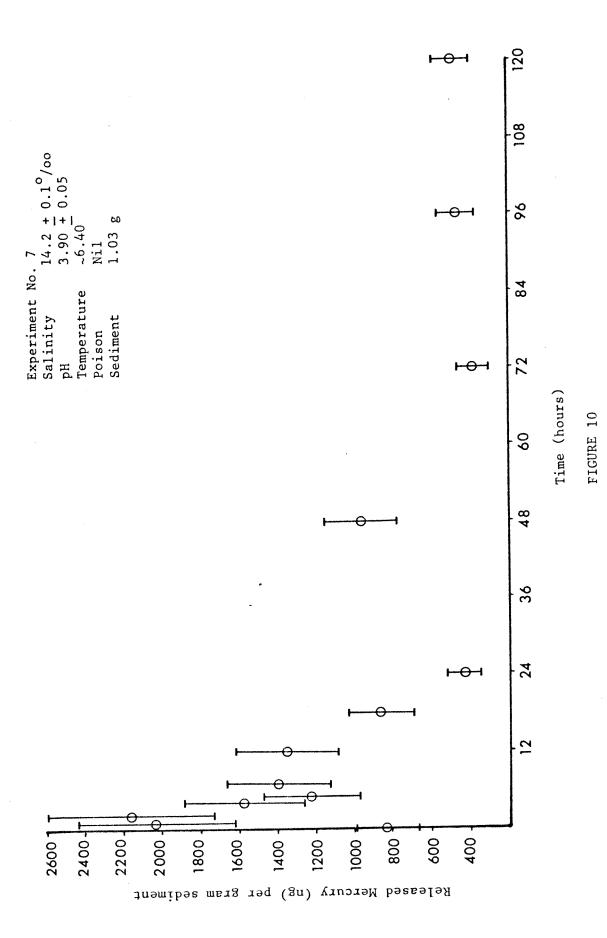


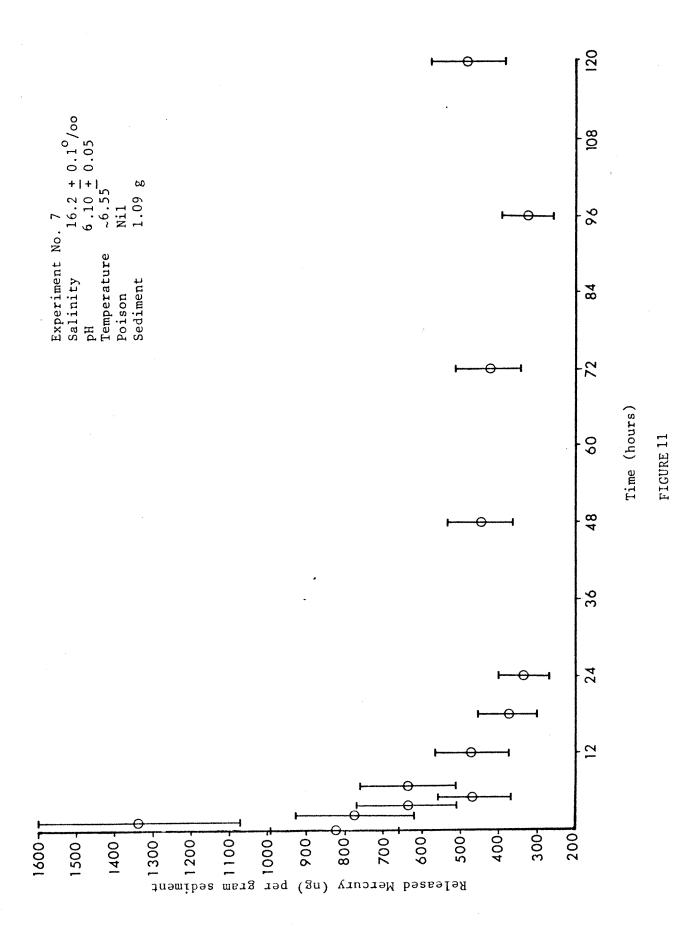


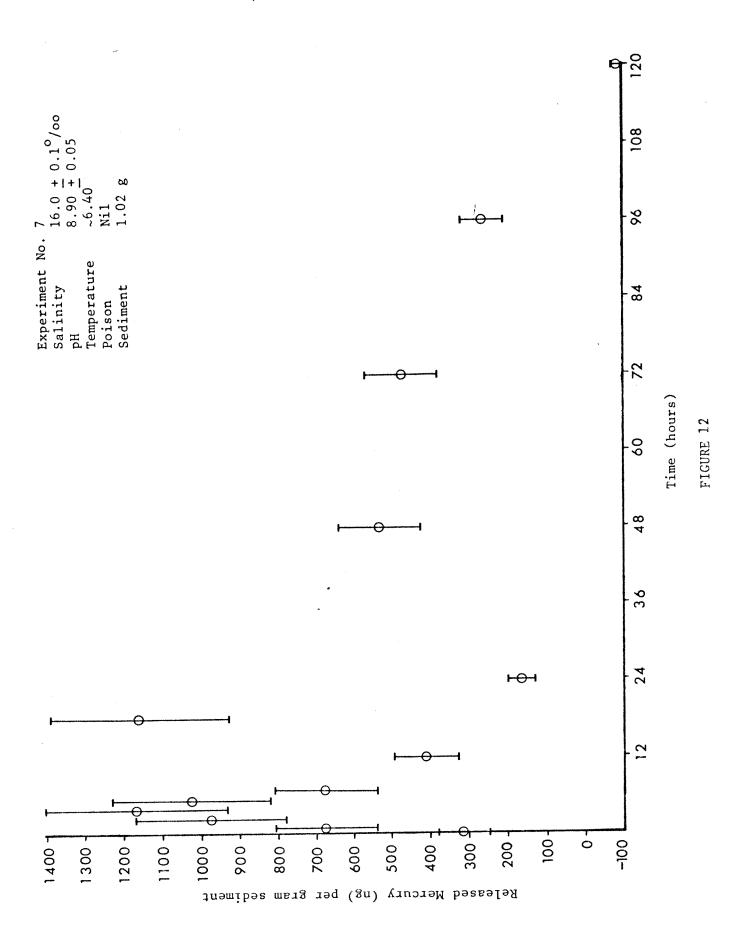


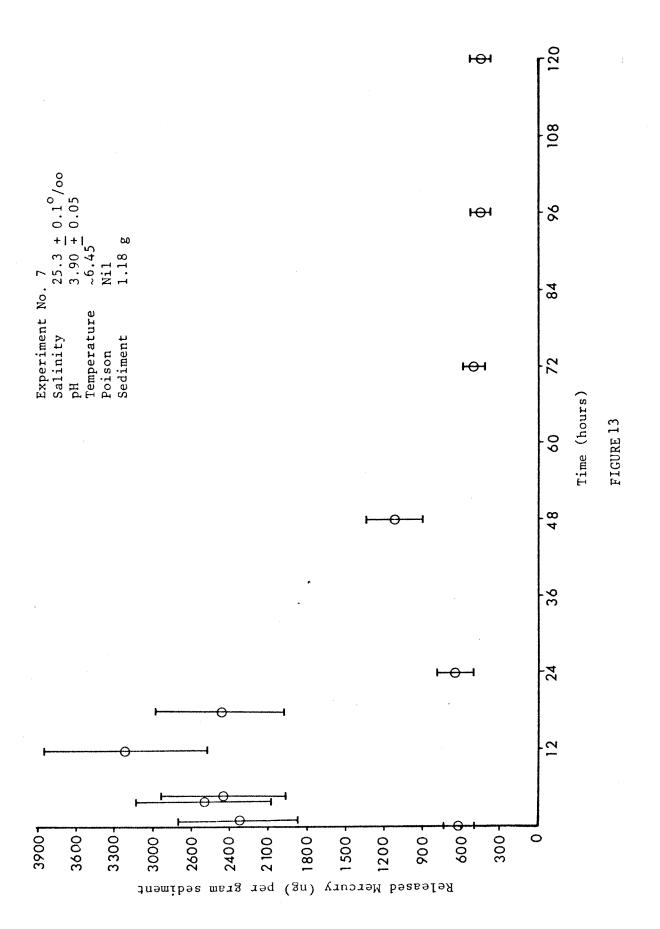


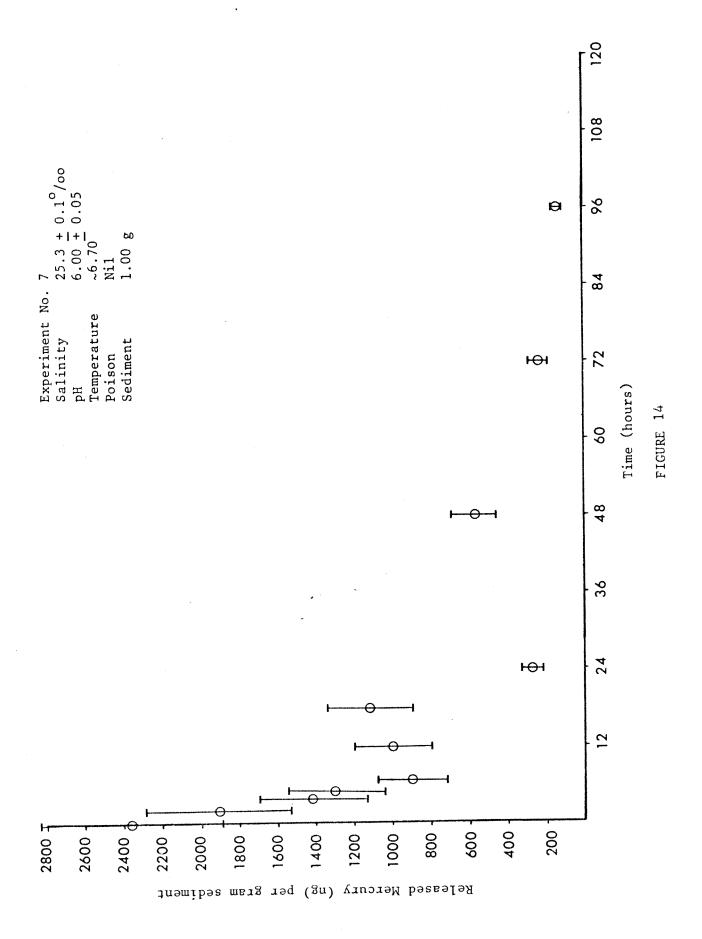
Variation of dissolved  $(\Delta)$  and particulate  $(\Box)$  mercury over a 24 hour period for seawater with suspended solids concentration of 100 ppm. FIGURE 9.

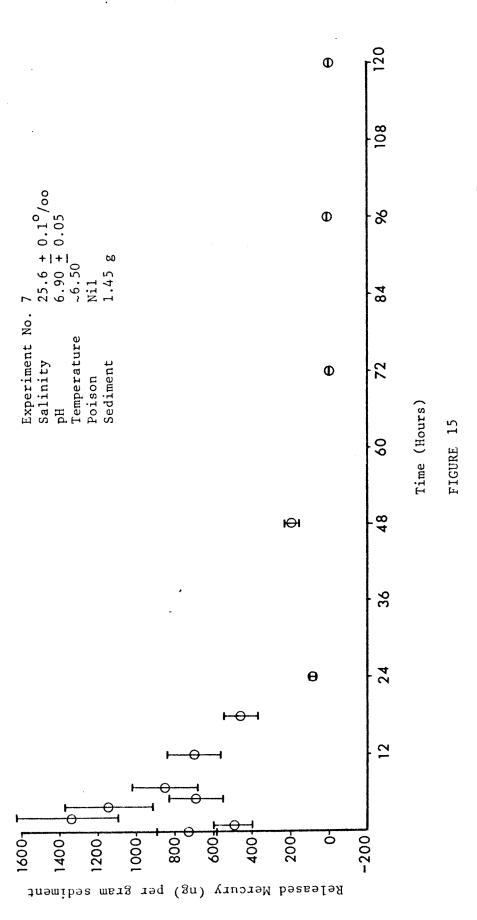


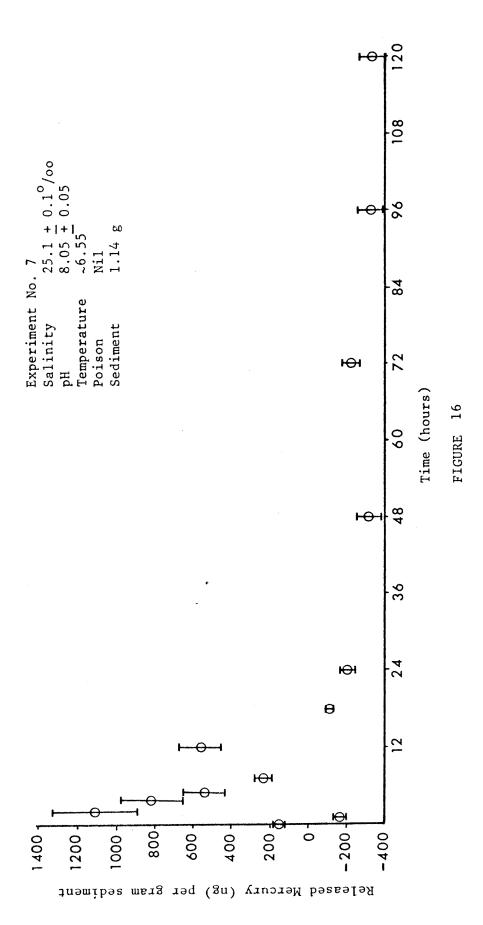


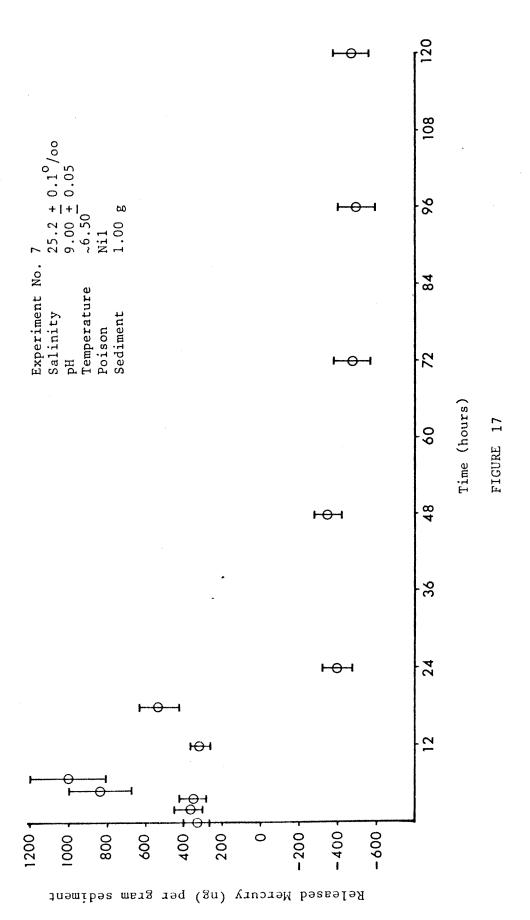


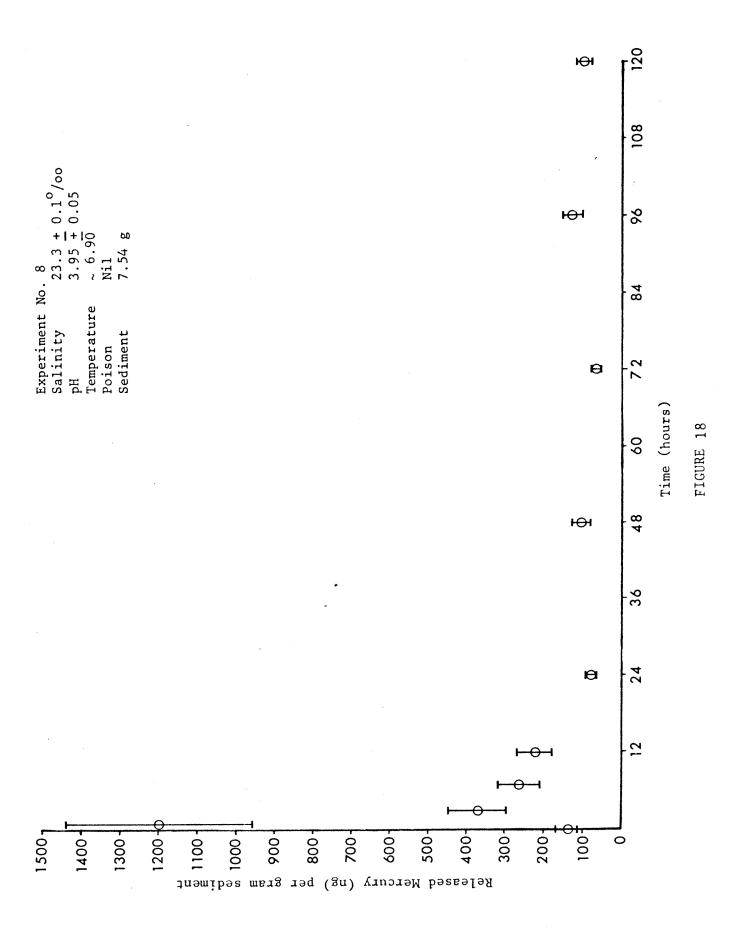


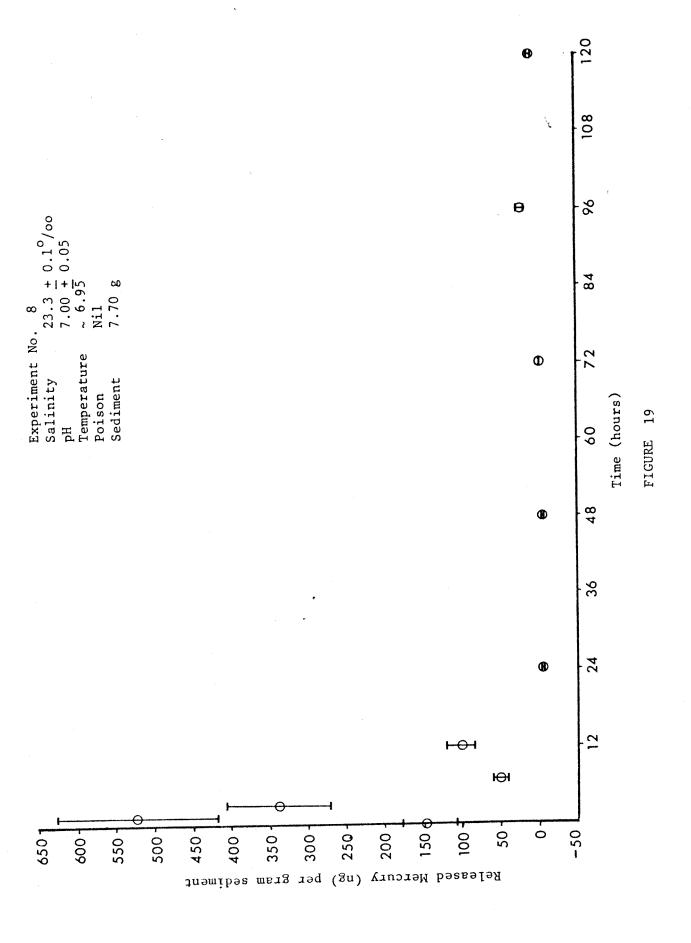


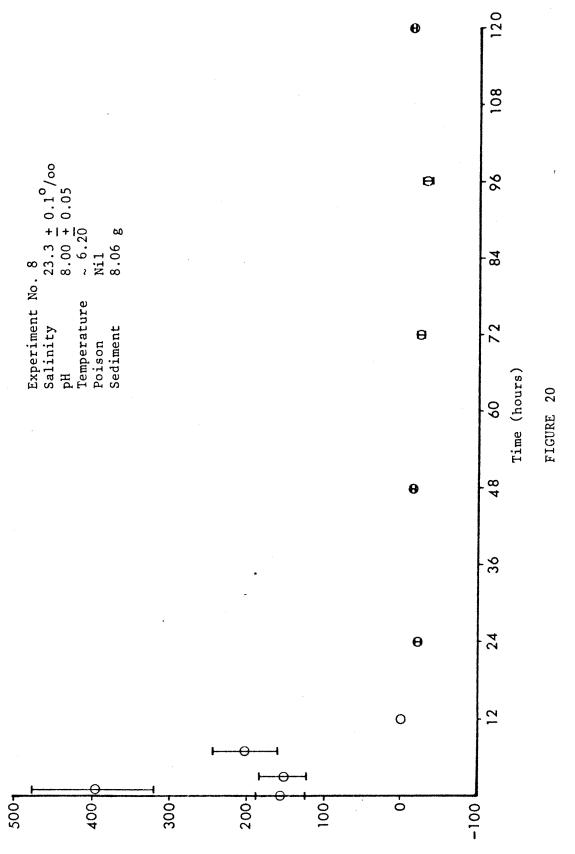




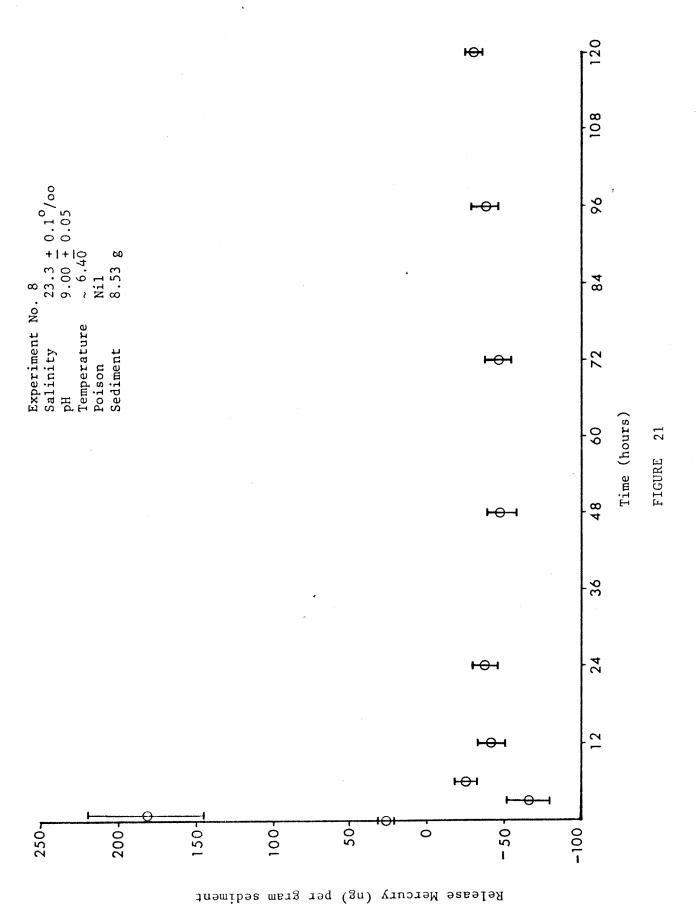


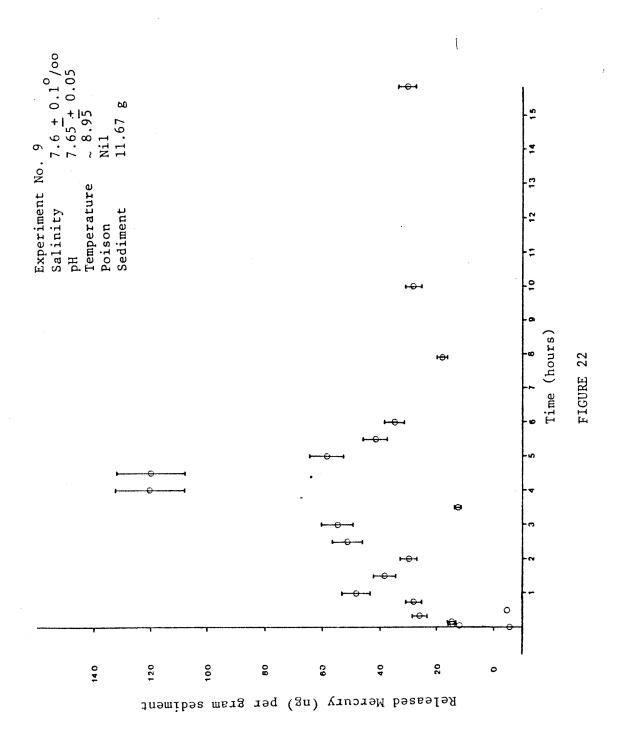


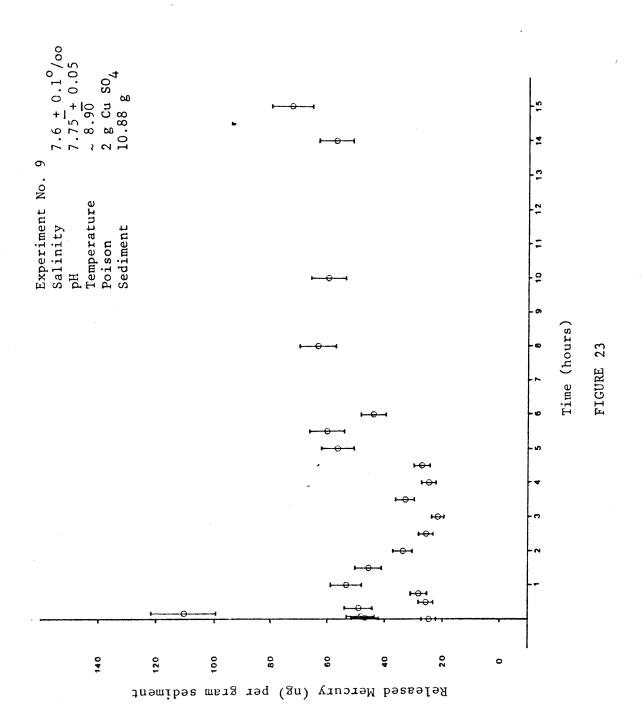


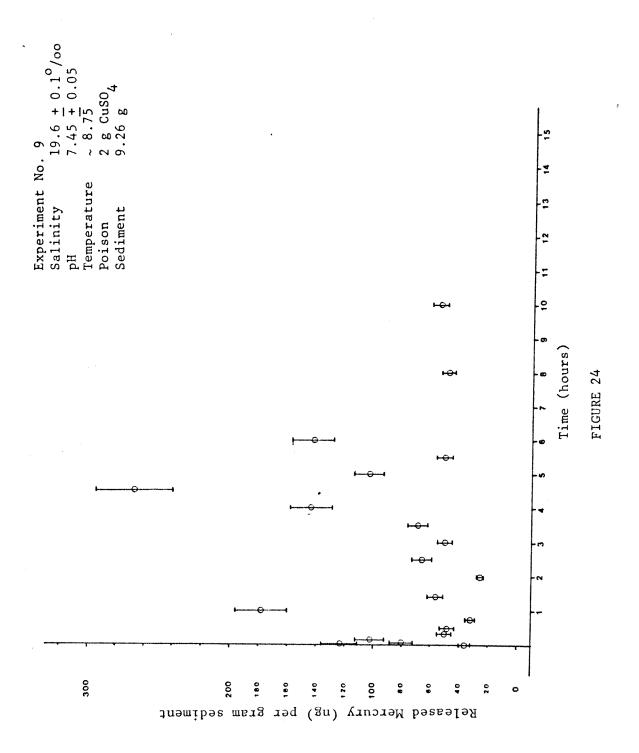


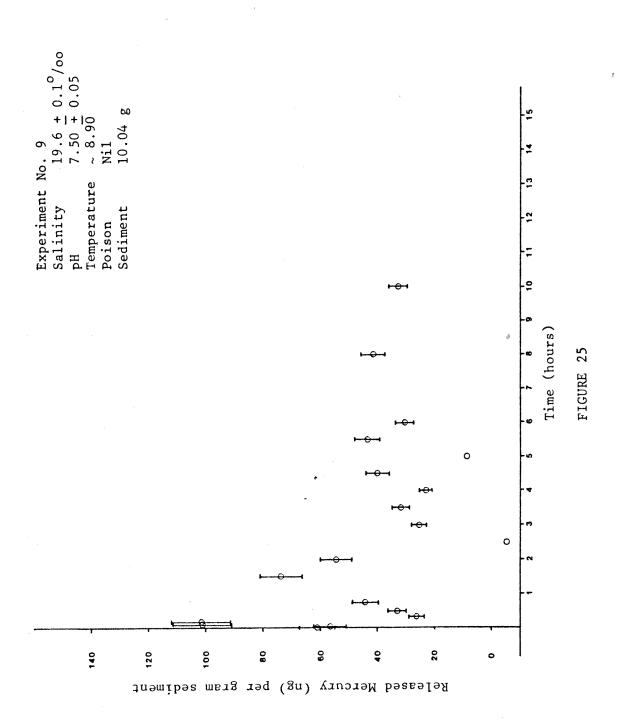
Released Mercury (ng) per gram sediment

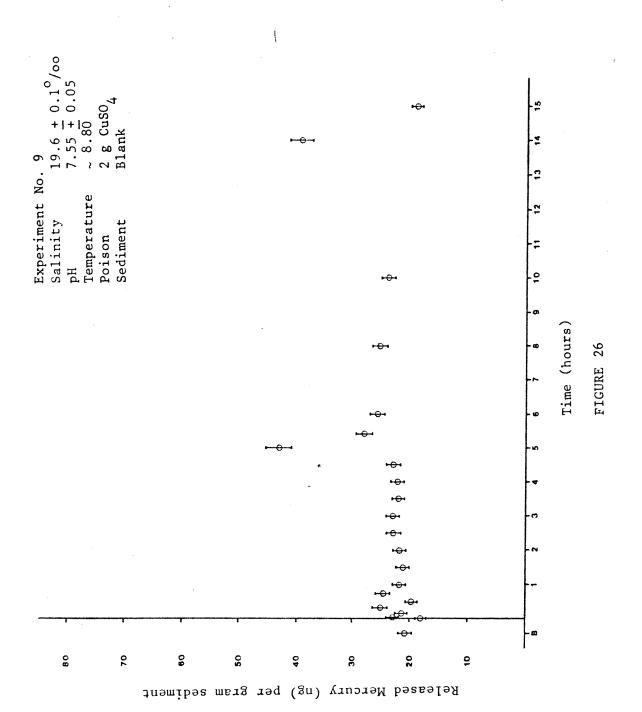


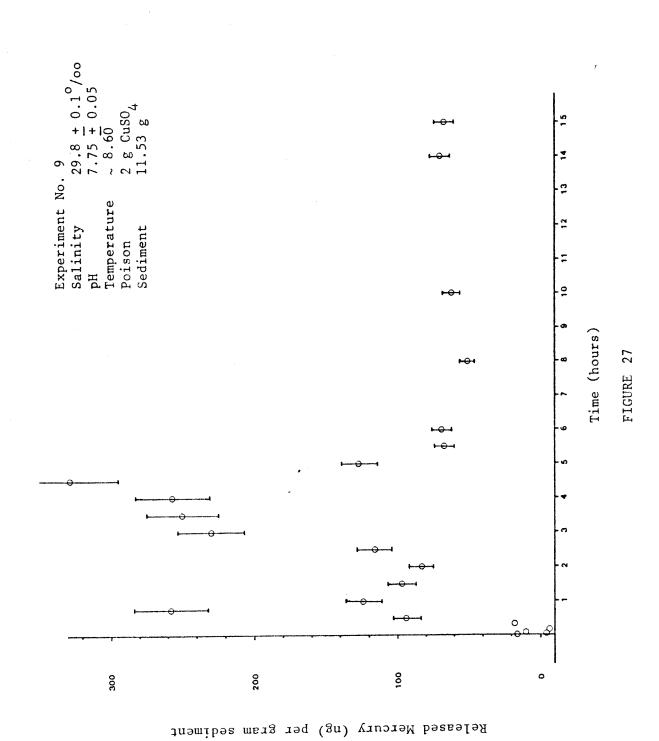


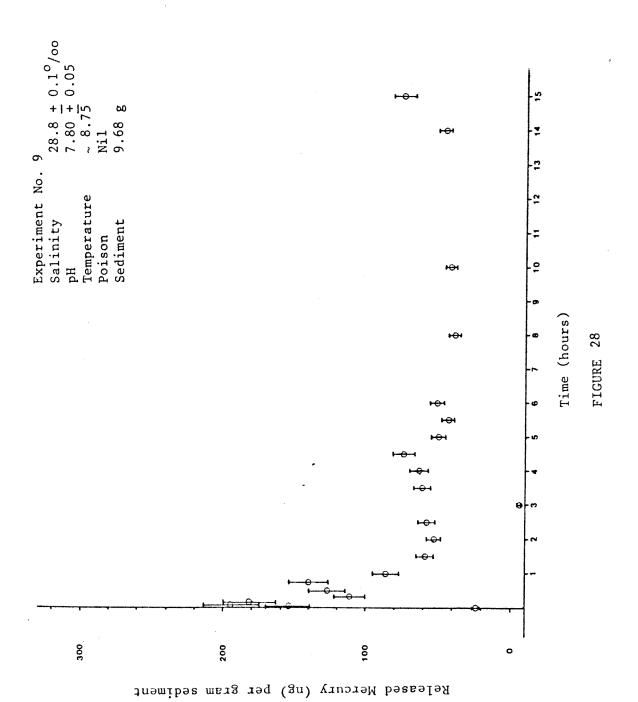


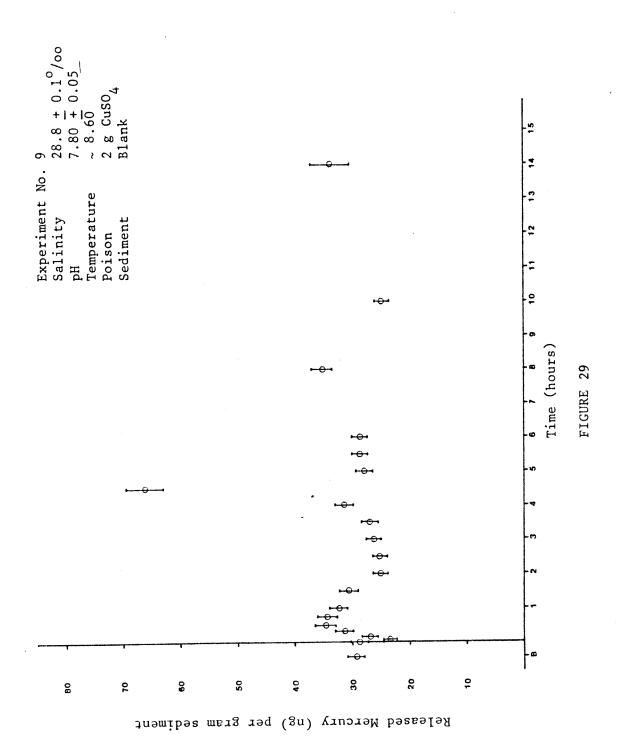












#### PART B: IN SITU MERCURY MOBILIZATION STUDY

#### 1. INTRODUCTION

Studies involving trace metal distributions in the Dutch Wadden Sea (Duinker, 1974) have indicated that, in inter-tidal waters, the observed mobilization of trace metal from the sediment to the overlying waters may be related to the tidal oscillation. A conception of the process, though likely oversimplified, is that during a flooding tide, the sediment is disturbed causing solid particles to be re-suspended and interstitial waters (usually higher in dissolved constituents than overlying waters) to be mixed upward. The net result may be a sharp and significant increase in dissolved metal concentration.

Other mechanisms of metal mobilization have been proposed, for example, biological reduction followed by removal through evolution of gas bubbles (Alberts et al, 1974) biological resuspension effects (Horne, 1969), simple diffusion (Presley, et al, 1967), desorption (Rohatgi and Chen, 1975), methylation (Jensen and Jernelov, 1969) and complex formation (Linberg and Harris, 1974). However, these mechanisms, with the possible exception of complex formation involving humic acids (Nissenbaum and Swaine, 1976) are not considered capable of accounting for the major part of the mercury mobilized from sediment (Lindberg et al, 1975).

Therefore, emphasis in this study was placed on investigating in situ, the variation of dissolved, particulate and pore water mercury over a tidal cycle. An area adjoining the wastewater outfall of the FMC chlor-alkali plant on the Squamish River mudflats at Squamish, B.C. was chosen as the study site for several reasons. Firstly, because of past pollution, the sediments in this area are reported to contain high levels of mercury (Thompson and McComas, 1973). Secondly, the mercury-rich sediments of the tidal flats near the FMC outfall are easily accessible and enclosed on three sides by groins forming a relatively small and well sheltered embayment.

Thirdly, there is a continuing source of mercury rich waters at the present time.

# SAMPLING PROGRAMME

On December 1 - December 2, 1976 during an 8.8' tide, a series of five sampling sites spaced about equally along a line extending 150 m seaward from the outfall area was occupied (Figure 31). Site 1 was located directly in the effluent stream approximately 5 metres from the outfall. Site 2 was located in a weed bed growing upon a great deal of putrifying organic debris and fine silt. Site 3 was in the main effluent stream (between a large well-sorted san bar and the weed bed. The sediment was poorly sorted sands, silts and organic debris. Site 4 was located in the centre of a large sandy depression. Site 5 was about 150 metres from the outfall in an area of highly compacted and well sorted sand.

The path of the effluent stream varied with tidal conditions. At high tide the effluent from the FMC holding pond veered left on entering the embayment and described the path shown in Figure 31. As the water level fell with the ebbing tide the path of the water became distributed between two paths (see Figure 31) until at low tide the water effluent was carried away exclusively by apath to the right of the outfall.

Using a 12 foot Canova boat as a working platform, eightfoot wooden markers were driven into the sediment for site identification during high slack water. Water, salinity, particle size distribution and sediment core samples were collected as conditions permitted at the times shown on Figure 32 . Water samples were collected as follows: acid-cleaned teflon tubing, taped two inches from the end of a ten-foot wooden pole was placed very carefully to the water-sediment interface. This ensured that a water sample unaffected by disturbed sediment was being obtained. The water was then pumped into 500 ml glass sampling bottles using a portable peristaltic pump. To each sample for total mercury, ten millilitres of a 0.5% (w/v)  $\frac{C^2}{C^2} \frac{C^2}{C^2}$  solution in mercury-free ACS nitric acid was added. Sediment cores were obtained in duplicate by driving 12 inch long, 3 cm in diameter acrylic tubes into the sediment and

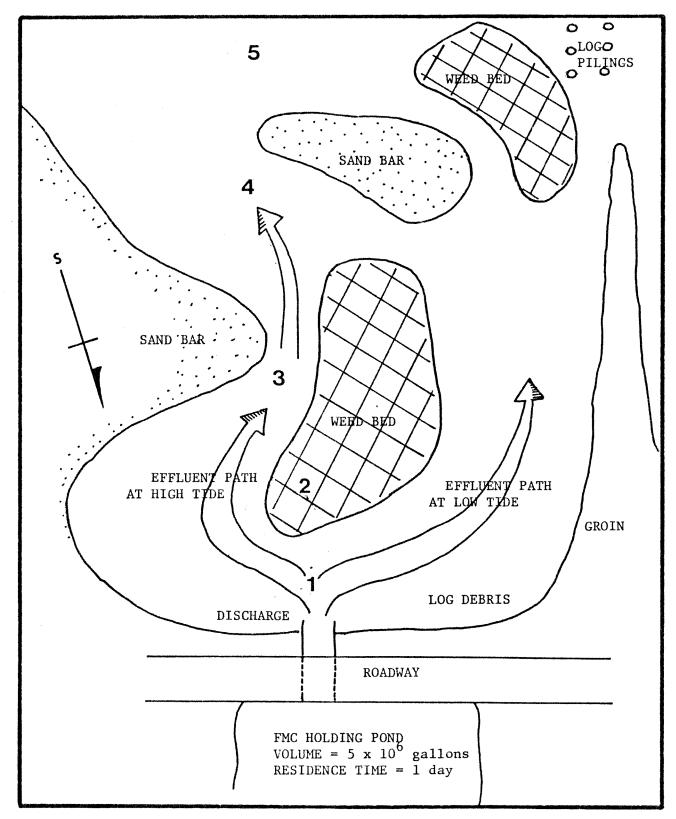
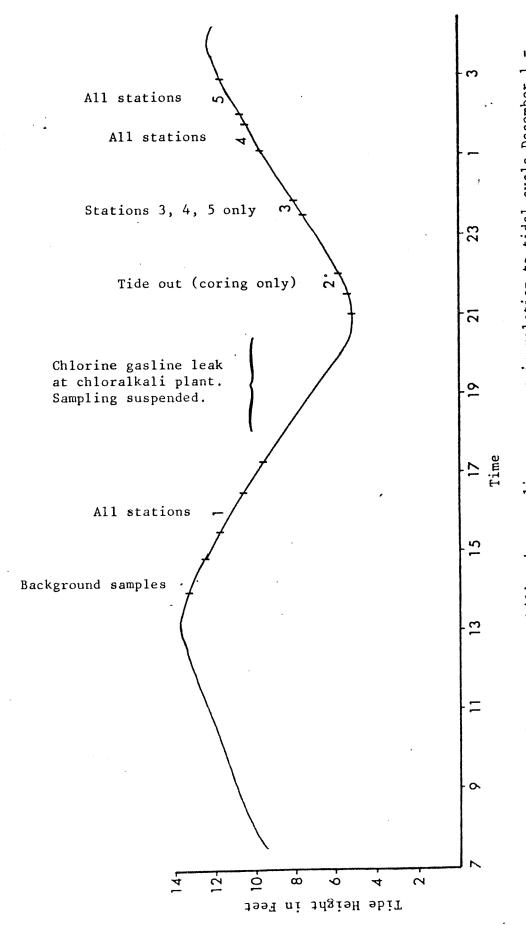


FIGURE 31. Sketch map (not to scale) showing the study area and main topographical features.



In situ mercury mobilization sampling programme in relation to tidal cycle December 1 December 2, 1976. Numbers on curve refer to sampling periods. FIGURE 32.

withdrawing. One core was frozen immediately, the other was used for the extraction of interstitial waters.

Water samples were filtered through 0.4  $\mu m$  pre-weighed Nuclepore membrane filters which had been prewashed with 0.1 N HNO3. To the filtrate (500 ml) was added 10 ml 0.5% (w/v)  $K_2 Cr_2 O_7 / HNO_3$  solution.

Interstitial waters were extracted under a nitrogen pressure of 2 to 5 atmospheres using a nylon squeezer similar to the one described by Reeburgh (1967). About 20 minutes of squeezing yielded 50-100 ml of pore water which was stored by adding 5 ml  $\rm K_2 Cr_2 O_7/HNO_3$  solution to the 150 ml collection flask.

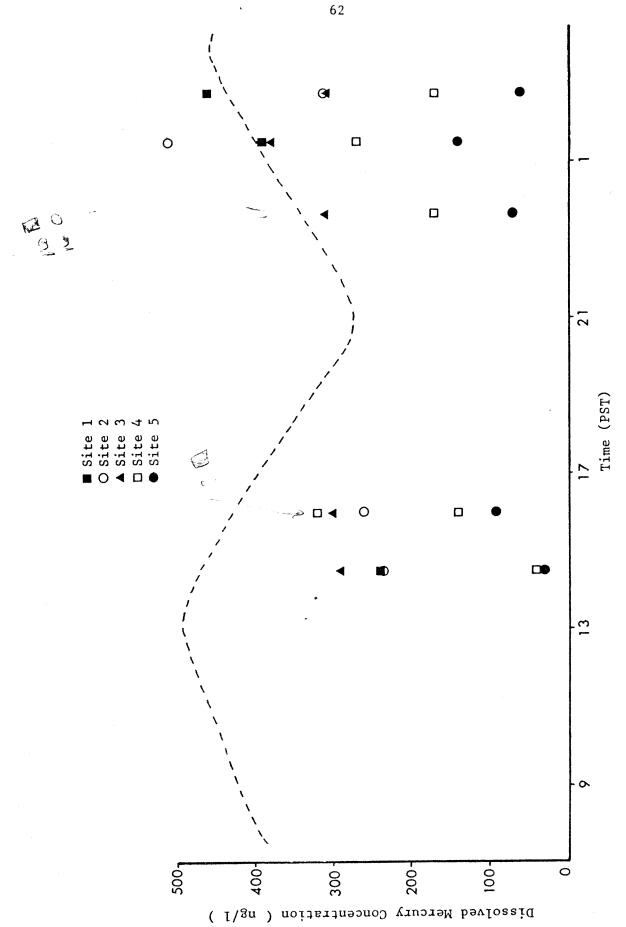
## 3. ANALYTICAL METHODS

Mercury in water, sediment and particulates was determined by methods described in Part A, 3.

## 4. RESULTS AND DISCUSSION

The data for the in situ mercury mobilization study appear in Appendix B1. Figure 33 shows the variation of dissolved mercury over a tidal cycle. At each site, the dissolved mercury levels were generally lowest at low water and, as the tide flooded there was a substantial (about 40%) short-lived increase. This trend, although based on limited data, is similar to the one observed by Duinker (1976). The interstitial waters, by contrast, behave oppositely being at a maximum during low tide and minimum during high tide which is a consequence of dilution. These data, however, may be in error because of possible mercury scavenging that occurred when iron hydroxide precipitated during the squeezing of several cores.

The dissolved mercury levels around the outfall at FMC were, at slack water, approximately ten times Howe Sound background values ( ~ 15 ng/l). During a flood tide this increased to about twenty times Howe Sound background. Assuming that (a) most of the mercury mobilized out of the tidal flat area is dissolved mercury (b) the sediments of the entire tidal flat



Dissolved Mercury in waters of Squamish River Tidal Flats off FMC Plant over a tidal cycle December 1 - December 2, 1976. FIGURE 33.

area contain about 1.0 ppm Hg in the top 10 cm (c) all the sediments in the tidal flats behave similarly to those for the five sites studied here and (d) on the average the sediments were covered by about 2 meters of water during each tidal cycle, then the observed released (~ 100 ng/1) corresponds to a half-life of mercury in the sediments of about 500 days. Although this is an extremely crude estimate, it does compare favourably with the half-life for mercury release from (a) Ottawa River sediments about one year (Waslenchuk, 1975) and (b) Bellingham Bay sediments of about 1.5 years (Carpenter, 1973). It is, therefore, possible that the mobilization of mercury out of the tidal flat area is driven mainly by the tidal oscillation.

## REFERENCES

- Alberts, J.J., Schindler, J.E. and Miller, R.W., 1974. Elemental mercury evolution mediated by humic acid. Science, 184, 895.
- 2. Baier, R.W. et al., 1975. Mercury loss from culture media. Anal. Chem., 47, 14, 2464.
- 3. Berrang, P.G. and Erickson, P.E., 1976. Ocean Dumping Project (Chemical Component). Data report for cruise OC-76-IS-001. Prepared by Seakem Oceanography Ltd., Victoria, B.C. under contract to Ocean Chemistry Division, D.O.E., Victoria, B.C. Contract No. SS08-KF832-5-SP009B.
- 4. Black, C.M. et al., 1972. Methods of soil analysis, Prentice Hall, New York.
- 5. Bothner, M.H., 1974. Mercury: Some aspects of its marine geochemistry in Puget Sound, Washington, Ph.D. Thesis, University of Washington, Seattle, Washington.
- 6. Carpenter, R. and Bothner, M.H., 1973. The rate of mercury loss from contaminated estuarine sediments in Bellingham Bay, Washington. Paper presented at the First Annual NSF Trace Contaminants Conference, Oak Ridge National Laboratory, Oak Ridge, Tennessee, August 8-10.
- Duinker, J.C., 1974. On the behaviour of copper, zinc, iron and manganese and evidence for mobilization processes in the Dutch Wadden Sea, Neth. J. Sea. Res., 8, 2, 214.
- 8. Eckert, J.M. and Sholkovitz, E.R., 1976. The flocculation of iron, aluminium and humates from river water by electrolytes. Geochim et Cosmochim. Acta, 40, 847.
- 9. Horne, R.A., 1969. <u>Marine Chemistry</u>. Wiley-Interscience, New York, N.Y.
- 10. Jensen, S. and Jernelov, A., 1967. Biological methylation of mercury in aquatic organisms, Nature, 223, 753.
- 11. Kolthoff, I.M. and Sandell, E.B., 1968. Quantitative Chemical Analysis, MacMillan and Co., New York, N.Y.
- Kurbatov, M.H., Wood, G.B. and Kurbatov, J.D., 1951. Isothermal adsorption of cobalt from dilute solutions, J. Phys. Chem., <u>55</u>, 1170.
- 13. Lee, G.F. et al., 1975. Research study for the development of dredged material disposal criteria. Dredged Material Research Program Report D-75-4.

- 14. Lindberg, S.E., Andren, A.W. and Harriss, R.C., 1974. Geochemistry of mercury in the estuarine environment in Estuarine Research, Vol. I, L.E. Cronin (Ed). Academic Press Inc., New York.
- 15. Lindberg, S.E. and Harriss, R.C., 1974. Mercury-organic matter associations in estuarine sediments and interstitial water, Environ. Sci. Tech., 8, 459.
- Loganathan, P. and Burau, R.G., 1973. Geochem. et Cosmochim. Acta, 37, 1277.
- 17. Lu, J.C.S. and Chen, K.Y., 1977. Migration of trace metals in interfaces of seawater and polluted surficial sediments, Envir. Sci. Tech., <u>11</u>, 2, 174.
- 18. Nissenbaum, A. and Swaine, D.J., 1976. Organic matter-metal interactions in recent sediments: The role of humic substances, Geochimica et Cosmochimica Acta, 40, 809.
- 19. Posselt, H.S., Anderson, F.J. and Weber, W.J., 1968. Catron sorption on colloidal hydrous manganese dioxide, Envir. Sci. Tech., 2, 1087.
- 20. Presley, B.J., Brooks, R.R., Kaplan, I.R., 1967. Science, 158, 906.
- 21. Rohatgi, N. and Chen, K.Y., 1975. J. Water Pollution Control Federation, 47, 9, 2299.
- 22. Sholkovitz, E.R., 1976. Flocculation of dissolved organic and inorganic matter during the mixing of river water and seawater, Beochim et Cosmochim Acta, 40, 831.
- 23. Thomas, D.J., 1976. Laboratory Study of the Release of Trace Metals from Dump Material in Seawater. Unpublished manuscript. Prepared by Seakem Oceanography, Ltd., Victoria, B.C. under contract (SS08-KF832-5-1568) to Ocean Chemistry Division, Victoria, B.C.
- 24. Thompson, J.J. and McComas, F.T., 1973. Distribution of Mercury in the Sediments and Waters of Howe Sound, B.C. F.R.B.C. Technical Report No. 396.
- 25. Waslenchuk, D.G., 1975. Mercury in fluvial bed sediments subsequent to contamination, Envir. Geol., <u>1</u>, 131.

APPENDICES

Appendix A·1

#### POSITIONS OF WATER COLLECTION STATIONS

CRUISE	LAT (°N)	LONG (ON)	WATER COLLECTED
OC-76-IS-008	49 <sup>°</sup> 22.3'	123 <sup>o</sup> 30.6'	Seawater
	49 <sup>°</sup> 40.9'	123 <sup>o</sup> 10.8'	Freshwater, Estuarine water
OC-77-IS-002	49 <sup>°</sup> 26.5'	123°59.6'	Seawater
	49 <sup>°</sup> 40.9'	123°10.6'	Estuarine water

Appendix A.2

# Dissolved Organic Carbon Content of Water's Used in Experiments

Salinity (°/oo)	D.O.C. (mg/1)
5.2	0.7
6.7	0.9
7.6	0.9
14.9	1.5
16.0	1.8
16.3	1.9
16.4	2.0
16.9	2.3
9.6	1.6
23.3	1.2
25.2	1.0
28.8	0.6
29.5	0.5

### Appendix A.3

Data for Release Experiments

Initial Conditions	1.	Water:	Salinity	$16.0 \pm 0.1^{\circ}/\circ \circ$
			рН	$3.95 \pm .05$
			Volume used	30 ± 0.5 1
			Poison	Nil

2. Sediment: type

> Nil amount added

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	3.85	17.40	40	
Oh	3.80	17.35	41	
1h -	3.80	17.75	41	
2h	3.90	17.40	40	
3h	3.95	17.65	34	
6h	4.05	17.80	33	
12h	4.10	17.80	37	
1d	3.85	17.85	31	
2 d	3.90	17.95	40	
3d	4.00	18.10	42	
4d	4.00	18.30	39	
5d	3.90	18.05	40	

1. Water:	Salinity	$16.5 \pm 0.1^{\circ}/00$
	рН	7.80 <u>+</u> 0.05
	Volume used	30 <u>+</u> 0.5 1
	Poison	Ni1
	1. Water:	pH Volume used

2. Sediment: type

amount added Nil

Released Hg (ng) Corrected Hg  $T(^{o}C)$ per g sediment conc (ng/1) Sampling time pН 31 7.85 17.50 Background 27 7.80 17.55 Oh 30 7.75 17.80 1h 41 7.75 17.55 2h 26 7.75 17.80 3h 26 7.85 17.85 6h 27 17.90 7.90 12h 35 18.10 7.75 1d29 18.15 7.80 2d27 7.80 18.40 3d24 18.00 7.85 4d 24 7.75 18.10 5d

Initial Conditions	1. Water:	Salinity	$6.7 \pm 0.1^{\circ}/00$
		pН	7.70 <u>+</u> 0.05
		Volume used	30 ± 0.5 1
		Poison	Nil
	2. Sediment:	type	Core #8, 0-2 cm
•		amount added	0.855 g

Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.60	17.00	38	
Oh	7.65	17.00	53	. 526
2h	7.75	17.10	57	667
4h	7.50	17.10	51	456
6h	7.50	17.25	27	-386
12h	7.55	17.10	31	-246
1d	7.60	17.20	37	<b>-</b> 35
2d	7.60	17.25	32	-211
3d	7.65	17.10	27	-386
. 4d	7.50	17.90	28	-351
40 5d	7.60	18.25	24	-491

Experiment No. 2

Initial	Conditions	1.	Water:	Salinity	$16.3 \pm 0.1^{\circ}/00$
				рН	7.50 <u>+</u> 0.05
				Volume used	30 <u>+</u> 0.5 1
				Poison	Nil
		2.	Sediment:	type	Core #8, 0-2 cm
				amount added	0.968 g

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.45	17.95	45	
Oh	7.40	18.00	66	651
	7.45	18.00	64	589
2h	7.45	18.05	61	496
4h	7.55	18.00	56	341
6h		18.30	48	93
12h	7.60		53	248
1d	7.65	18.40		217
2d	7.50	18.20	52	186
3d	7.50	18.00	51	
4d	7.55	18.80	49	124
5d ·	7.50	18.95	49	124

Expe	rim	ent	No	. 2
				•

Initial	Conditions	1.	Water:	Salinity	$16.3 \pm 0.1^{\circ}/\circ \circ$
				pН	7.50 <u>+</u> 0.05
				Volume used	30 <u>+</u> 0.5 1
				Poison	2.0 g CuSO <sub>4</sub>
		2.	Sediment:	type	Core #8, 0-2 cm
				amount added	1.047 g

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.50	17.25	47	
Oh	7.55	17.25	56	258
2h	7.55	17.40	48	29
4h	7.65	17.75	67	573
6h	7.40	17.70	37	-287
12h	7.40	17.50	34	-372
1d	7.45	17.80	48	29
2d	7.55	17.80	25	-630
3d	7.60	18.15	28	-544
4 d	7.50	18.00	32	-430
5d	7.50	18.05	39	-229

Initial Conditions	1.	Water:	Salinity	$29.5 \pm 0.1^{\circ}/_{\circ}$
			рН	7.65 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Ni1
	2.	Sediment:	type	Core #8, 0-2 cm

amount added 0.903 g

Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.55	17.00	24	
Oh	7.60	17.05	39	166
2h	7.60	17.00	57	1096
4h	7.65	17.10	43	631
6h	7.70	17.15	34	332
12h	7.70	17.10	30	199
1d	7.70	17.20	37	432
2d	7.50	17.05	24	0
3d	7.55	17.00	27	100
4d	7.60	18.10	27	100
5d .	7.70	18.55	27	100

Experiment No. 3

				67.010/
Initial Conditions	1.	Water:	Salinity	$6.7 \pm 0.1^{\circ}/\circ \circ$
			pН	7.65 <u>+</u> 0.05
			Volume used	30 ± 0.5 1
			Poison	Nil
	2.	Sediment:	type	Core #8, 0-2 cm
			amount added	0.909 g
		0	Corrected Hg	Released Hg (ng

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.65	17.45	25	
Oh	7.75	17.50	23	-44
1h	7.50	17.50	33	267
3h	<b>7.</b> 50	17.45	45	<b>6</b> 67
6h	7.55	17.30	29	133
12h	7.60	17.90	28	100
1 d	7.65	18.35	28	100
2d	7.70	18.40	26	33
3d	7.70	17.95	29	133
4d	7.65	18.00	41*	
5d	7.65	17.80	22	-100

<sup>\*</sup>Sample contaminated

Initial	Conditions	1.	Water:	Salinity	$16.3 \pm 0.1^{\circ}/\circ \circ$
				ηН	7.65 <u>+</u> 0.05
				Volume used	30 <u>+</u> 0.5 1
				Poison	Nil
		2.	Sediment:	type	Core #8, 0-2 cm
				amount added	0.951

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.65	17.80	22	
Oh	7.70	17.85	22	0
1h	7.70	17.85	38	505
3h	7.75	17.85	41	599
6h	7.55	17.80	26	126
12h	7.55	17.95	22	0
1 d	7.60	18.00	25	95
2d	7.60	18.05	24	63
3d	7.60	18.50	26	126
4d	7.60	18.00	37	473
5d ·	7.60	18.05	27	158

			٠,
Exper	ımen	t No.	

Initial	Conditions	1.	Water:	Salinity	$16.3 \pm 0.1^{\circ}/00$
				рН	7.50 <u>+</u> 0.05
				Volume used	30 ± 0.5 1
				Poison	2.0 g CuSO <sub>4</sub>
		2.	Sediment:	type	Core #8, 0-2 cm
				amount added	1.00 g

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.50	17.20	25	
Oh	7.50	17.30	31	180
1h	7.50	17.40	47	660
3h	7.50	17.45	41	480
6h	7.60	17.45	34	270
12h	7.65	17.70	31	180
1d	7.45	17.80	34	270
2d	7.50	17.95	40	450
3d	7.55	18.20	32	210
4d	7.55	18.35	38	390
5d -	<b>7.</b> 50	18.25	42	510

Initial Conditions	1.	Water:	Salinity	16.4 ± 0.1°/00
			рН	7.45 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Nil

2. Sediment: type

amount added Nil

Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.45	17.80	21	
Oh	7.55	17.80	32	
1h ,	7.60	17.85	30	
3h	7.40	17.95	34	
6h	7.40	18.05	29	
12h	7.45	18.00	36	
1 d	7.50	18.35	35	
2d	7.50	18.45	31	
3d	7.40	18.35	38	
4d	7.45	18.30	38	
5d	7.45	18.25	36	

Experiment No. 3

Initial	Conditions	1.	Water:	Salinity	29.5 <sup>0</sup> /00
				рН	7.60 <u>+</u> 0.05
				Volume used	30 ± 0.5 1
				Poison	Ni1
		2.	Sediment:	type	Core #8, 0-2 cm
				amount added	0.910 g

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.70	17.15	24	
Oh	7.70	17.25	27	99
<b>1</b> h	7.55	17.20	47	758
3h	7.60	17.20	58	1123
6h	7.60	17.15	35	363
12h	7.60	17.00	37	429
1d	7.60	18.10	35	363
2d	7.60	16.65	36	396
3d	7.60	17.80	76*	
4d	7.65	17.75	37	429
5d <sup>-</sup>	7.60	17.90	36	396

<sup>\*</sup>Contaminated sample

Initial Conditions	1.	Water:	Salinity	$29.5 \pm 0.1^{\circ}/00$
			pH	7.70 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Ni1
	2.	Sediment:	type	
			amount added	Ni1
			Corrected Ho	Released Ho (no

Sampling	time	pH	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Backgrou	ınd	7.70	17.80	37	
Oh		7.70	17.40	37	
1 h		7.70	17.45	34	
3h		7.80	17.55	34	
6h		7.65	17.85	29	
12h		7.65	17.95	35	
1d		7.70	17.80	31	
2d		7.70	17.90	38	
3d		7.70	18.20	99*	
4 d		7.70	18.15	38	
5d	*	7.70	18.15	34	

<sup>\*</sup> Contaminated sample

Experiment No. 4

17 ng/1	2 ng/1	$29.4 \pm 0.1^{\circ}/00$	$14.9 \pm 0.1^{\circ}/00$
Tap Water Blank	Nuclepore Filter Blank	Seawater Salinity	Estuarine Salinity
Initial Conditions:			

IR I NE INK	Particulate Hg (ng/1)		31	32	26	21	17	32	20	26	19	21	26	20
ESTUARINE BLANK	Dissolved Hg (ng/1)	38	23	23	29	36	26	26	23	28	26	34	26	29
RINE	Particulate Hg (ng/1)		31	19	149	136	146	123	127	126	163	135	161	158
ESTUARINE	Dissolved Hg (ng/l)	33	23	21	28	70	26	38	24	32	31	29	25	30
TER	Particulate Hg (ng/1)		33	17	128	135	142	122	156	171	163	165	165	175
SEAWATER	Dissolved Hg (ng/1)		26	20	28	Lost	72	06	54	30	27	21	20	30
	Time	Total Background	Filtered Background	Background No Sed.	Oh	0.5h	Ih	2h	3h	4h	5h	6h	10h	24h

Experiment No. 5

Initial Conditions	1. Water:	Salinity	$16.8 \pm 0.1^{\circ}/\circ \circ$
	y.	Н	7.45 <u>+</u> 0.05
		Volume used	30 ± 0.5 1
		Poison	0.2 g Cu SO <sub>4</sub>
	2. Sedimer	nt: type	Core #6, 6-8 cm
		amount added	6.45 g

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.45	17.45	30.2	
Oh	7.35	17.85	68.6	179
1h	7.35	17.85	62.1	148
2h	7.40	18.10	87.7	267
4h	7.40	17.85	86.7	263
8h	7.40	17.60	68.1	176
12h	7.55	17.95	84.2	251
18h	7.45	17.95	73.4	201
24h	7.45	18.10	77.9	222
1.5d	7.40	18.80	72.6	197
2d ·	7.50	18.75	73.8	203

Experiment No. 5

Initial Conditions	1.	Water:	Salinity	$16.5 \pm 0.1 / 00$
		•	pН	7.55 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	2.0 g CuSO <sub>4</sub>
	2.	Sediment:	type	Core #6, 6-8 cm
			amount added	7.02
				,

рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
7.55	18.10	38.1	
7.60	18.35	75.8	238
7.45	18.20	61.3	99
7.45	18.45	Lost	
7.50	18.30	101.8	271
<b>7.</b> 50	17.95	77.2	167
7.60	17.80	74.5	156
7.50	18.10	82.9*	191
7.55	18.25	161.8*	529
7.60	18.40	124.6*	370
7.60	18.05	126.6*	378
	7.55 7.60 7.45 7.45 7.50 7.50 7.60 7.55 7.60	7.55 18.10  7.60 18.35  7.45 18.20  7.45 18.45  7.50 18.30  7.50 17.95  7.60 17.80  7.50 18.10  7.55 18.25  7.60 18.40	PH T(°C) conc (ng/1)  7.55 18.10 38.1  7.60 18.35 75.8  7.45 18.20 61.3  7.45 18.45 Lost  7.50 18.30 101.8  7.50 17.95 77.2  7.60 17.80 74.5  7.50 18.10 82.9*  7.55 18.25 161.8*  7.60 18.40 124.6*

<sup>\*</sup>Brown precipitate in samples

Experiment No. 5

Initial Condi	itions 1.	Water:	Salinity	$16.7 \pm 0.1^{\circ} \rho \circ$
			рН	7.45 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	0.2 g CuSO <sub>4</sub>
	2.	Sediment:	type	Core #8, 0-2 cm
			amount added	1.00 g
				•

Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.45	17.80	43.8	·
Oh	7.40	18.45	54.0	306
1h	7.45	18.40	81.3	1125
2h	7.50	18.30	95.7	1557
8h	7.55	18.35	65.9	663
18h	7.65	18.15	60.2	492
24h	7.40	18.20	46.7	87
1.5d	7.40	18.00	66.5	681
2.5d	7.40	17.65	52.5	261
4d	7.45	17.90	64.0	606
5d ·	7.40	18.10	66.2	672

Experiment No. 5

Initial	Conditions	1.	Water:	Salinity	$16.6 \pm 0.1^{\circ}/00$
				pН	7.60 <u>+</u> 0.05
				Volume used	30 + 0.5 1
				Poison	2.0 g CuSO <sub>4</sub>
		2.	Sediment:	type	Core #8, 0-2 cm
				amount added	1.05 g

Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.60	17.90	30.4	
Oh	7.55	18.35	41.5	317
1h	7.45	18.30	75.7	1294
2h	7.55	18.20	92.9	1786
8h	7.55	18.25	91.5	1746
18h	7.60	18.45	60.9	871
1 d	7.65	17.95	62.9	929
1.5d	7.55	18.10	61.1	877
2.5d	7.45	18.05	313.3*	
4d	7.60	18.35	288.8*	
5d ·	7.60	18.45	268.6*	

<sup>\*</sup> Contaminated sample

Initial Conditions	1.	Water:	Salinity	$16.9 \pm 0.1^{\circ}/\circ o$
			рН	7.60 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	2.0 g CuSO <sub>4</sub>
	2.	Sediment:	type	
			amount added	Ni 1

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.60	17.65	40.7	
Oh	7.55	17.65	27.8	
1h	7.55	17.95	24.2	
2h	7.60	17.70	29.3	
8h	7.65	17.80	28.4	
18h	7.70	17.80	95.1*	
1d	7.55	17.90	24.1	
1.5d	7.55	17.90	26.5	
2.5d	7.60	18.15	33.3	
4d	7.60	18.40	27.1	
5d	7.60	18.90	28.9	

<sup>\*</sup> Contaminated sample

Initial Conditions	1. Water:	Salinity	$16.9 \pm 0.1^{\circ}/\circ \circ$
		рН	7.65 <u>+</u> 0.05
		Volume used	30 <u>+</u> 0.5 1
		Poison	0.2 g CuSO <sub>4</sub>

2. Sediment: type

amount added Nil

Released Hg (ng) Corrected Hg T(°C) Sampling time  $\mathbf{p}\mathbf{H}$ conc (ng/1)per g sediment Background 7.65 17.10 22.5 Oh 17.80 23.5 7.75 7.80 17.85 27.3 1h 2h 7.50 17.85 30.2 8h 24.1 7.50 17.60 18h 7.55 17.70 28.0 1d7.60 17.80 30.6 1.5d 17.95 27.6 7.60 2.5d 7.65 17.65 37.7 18.10 4d 7.55 30.5 5d 7.65 18.60 27.5

Experiment No. 6

Initial Conditions	1.	Water:	Salinity	5.2 <u>+</u> 0.1°/00
			рН	4.00 <u>+</u> 0.05
			Volume used	30 ± 0.5 1
			Poison	Ni1
	2.	Sediment:	type	Core #8 0-2 cm

amount added 1.39 g

Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	4.05	6.65	23.0	
Oh	4.00	6.20	49.0	561
1h	4.00	6.00	64.8	902
2h	3.95	6.00	53.0	648
5h	4.00	5.80	81.5	1759
8h	4.05	5.75	49.0	561
12h	4.00	5.30	36.5	291
1d	4.00	5.75	33.5	723
2d	4.05	5.65	33.1	218
3d	4.00	6.00	30.6	164
4d	4.05	5.80	29.3	136
5d	4.00	5.80	30.9	171

8h

1 d

2d

3d

4d

5d

12h

Initial Conditions	1.	Water:	Salinity	5.2 <u>+</u> 0.1°/oo
			рН	5.90 <u>+</u> 0.05
•			Volume used	30 <u>+</u> 0.5 1
			Poison	Ni1
	2.	Sediment:	type	Core #8 0-2 cm
			amount added	1.00 g
Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	5.90	19.30	17.0	
Oh 1h	5.80 5.95	19.90 18.95	32.6 33.4	468 492
2h	5.90	20.00	28.9	357
5h	5.90	20.25	37.3	609
J11	5.70	20.23	31.3	003

21.45

20.65

20.10

20.15

20.45

20.70

20.10

32.3

32.3

27.8

18.8

17.5

18.4

18.0

459

459

324

54

15

42

30

5.95

5.80

5.95

6.00

5.90

5.90

5.90

Experiment No. 6

Initial Conditions	1.	Water:	Salinity	$5.2 \pm 0.1^{\circ}/00$
			рН	7.00 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Nil
	2.	Sediment:	type	Core #8 0-2 cm
			amount added	1.05 g

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	6.90	19.55	11.5	
Oh	6.95	19.75	25.1	389
1h	6.80	19.20	28.9	497
2h	6.95	18.90	26.9	440
5h	7.05	20.60	30.6	546
8h	7.00	21.50	28.6	489
12h	6.90	20.90	22.3	309
1d	7.00	20.05	28.8	494
2d	7.00	19.00	22.4	311
3d	7.05	19.85	13.6	· 60
4 d	7.10	19.55	13.3	51
5d	7.00	19.15	13.8	66

Initial (	Conditions	1.	Water:	Salinity	$5.2 \pm 0.1^{\circ}/_{\circ \circ}$
				pН	8.00 <u>+</u> 0.05
•				Volume used	30 ± 0.5 1
				Poison	Nil
		2.	Sediment:	type	Core #8, 0-2 cm
				amount added	1.15 g
Sampling	time	pН	т( <sup>о</sup> с)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment

Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	8.05	19.30	17.5	
Oh	8.00	19.15	28.3	281
1h	8.00	19.05	41.3	621
2h	7.95	19.30	41.3	621
5h	8.10	21.20	32.5	391
8h	8.05	22.00	34.5	444
12h	8.10	20.15	24.5	183
24h	8.00	19.85	25.3	203
2d	8.05	18.95	28.1	277
3d ·	8.00	19.75	26.0	222
4d	8.00	20.15	25.7	214
5d	8.00	19.60	24.9	193

Experiment No. 6

Initial Conditions	1.	Water:	Salinity	5.2 <u>+</u> 1°/oo
			pН	8.95 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Nil
	2.	Sediment:	type	Core #8 0-2 cm
			amount added	1.19 g
Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	8.95	19.5	15.0	
Oh	8.80	19.0	20.5	139
1h	9.00	19.0	17.9	73
2h	8.85	19.5	16.7	43
5h	8.90	19.9	16.5	38 <b>-</b> 50
8h	8.90	20.1	13.0 113.0*	-50
12h 1d	8.90 8.90	20.9 20.5	20.7	144
2d	8.90	20.5	11.1	-98
2d 3d	8.90	19.4	9.9	<b>-129</b>
3d 4d	8.90	20.0	10.6	-111
5d	8.90	20.0	11.0	-101

<sup>\*</sup>sample contaminated

Experiment No. 6

Initial Conditions	1.	Water:	Salinity	5.2 ± 0.1°/00
			pН	7.70 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Ni1
	2.	Sediment:	type	
			amount added	Ni1
	**	m/ <sup>0</sup> a >	Corrected Hg	Released Hg

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.65	18.90	17.0	
Oh	7.70	19.45	21.7	
1h	7.70	19.60	19.1	
2h	7.70	19.55	22.6	
5h	7.65	20.95	16.1	
8h	7.75	21.10	16.1	
12h	7.75	19.90	19.0	
1d	7.70	19.90	13.7	
2d	7.65	19.65	15.8	
3d	7.65	19.50	15.3	
4d	7.70	19.95	16.0	
5d	7.70	20.05	16.9	

Experiment No. 7

•				
Initial Conditions	1.	Water:	Salinity	16.2 <u>+</u> 0.1°/oo
			рН	3.90 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Ni 1
	2.	Sediment:	type	Core #8, 0-2 cm
			amount added	1.03 g
Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	3.95	6.40	25.8	
Oh	3.90	6.35	54.2	827
1h	4.00	6.25	95.5	2030
2h	4.00	6.00	100.0	2161
4h	3.85	5.90	80.0	1579
5h	3.90	5.85	67.6	1217
7h	3.90	6.15	73.7	1395
12h	3.95	6.50	72.3	1354
18h	4.00	6.45	55.4	862
1d	4.00	6.45	40.7	434
2d	4.00	6.45	58.8	961
<b>3</b> d	4.00	6.45	38.9	382
4d	4.05	6.30	41.8	466
5d	3.95	€.05	42.5	486

Experiment No. 7

Initial Conditions	1.	Water:	Salinity	$16.2 \pm 0.1^{\circ}/00$
			рН	4.10 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Ni1
	2.	Sediment:	type	Core #8, 0-2 cm
			amount added	1.09 g

Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	4.05	6.55	27.1	
Oh	4.10	6.50	57.1	826
1h	4.10	6.75	75.7	1338
2h	4.05	6.40	55.2	773
4h	4.05	6.30	50.4	641
5h	4.10	6.15	44.0	465
7h	4.00	6.00	50.2	636
12h	4.00	5.70	44.2	471
18h	4.00	5.95	40.7	374
1d	4.00	6.40	39.3	336
2d	4.00	6.35	43.3	446
3d	4.00	6.25	42.5	424
4d	4.00	6.10	38.9	325
5d	4.00	6.40	44.5	479

Experiment No. 7

Initial	Conditions	1.	Water:	Salinity	$16.0 \pm 0.1^{\circ}/00$
				pH	8.90 <u>+</u> 0.05
				Volume used	30 ± 0.5 1
				Poison	Ni1
		2.	Sediment:	type	Core #8, 0-2 cm
				amount added	1.02 g

Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	8.95	6.40	25.2	
Oh	8.90	6.35	35.9	315
1h	9.00	6.10	48.1	674
2h	9.10	5.95	58.3	974
4h	8.85	5.95	64.9	1168
5h	8.90	5.80	60.0	1024
7h	8.90	6.40	48.1	674
12h	9.00	6.50	39.2	412
18h	9.00	6.85	64.6	1159
1d	9.10	5.90	30.8	165
2d*	8.95	6.15	43.3	532
3d	8.95	6.30	41.3	474
4d	8.90	6.40	34.2	265
5d	9.10	6.35	22.3	<del>-</del> 85

<sup>\*</sup>stir rod and drill replaced

Experiment No. 7

Initial Conditions	1.	Water:	Salinity	$25.3 \pm 0.1^{6}/00$
			pН	3.90 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Ni1
	2.	Sediment:	type	Core #8, 0-2 cm

amount added 1.18 g

Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	3.85	6.45	38.2	
Oh	3.95	6.75	62.5	618
1h	3.80	6.80	129.9	2331
2h	3.90	6.40	Lost	
4h	4.00	6.30	140.8	2608
5h	4.00	6.15	134.6	2450
7h	3.90	5.90	Lost	
12h	3.85	5.80	164.9	3221
18h	3.85	6.05	135.8	2481
1d	3.95	6.35	63.6	646
2d	3.90	6.00	82.3	1121
3d	4.00	6.20	57.1	481
4d	4.00	6.10	55.7	445
5d	3.95	6.15	55.7	445

Experiment No. 7

Initial Conditions	1. Wat	er: Sal	linity	25.3 <u>+</u>	0.10/00	
		рН		6.00 <u>+</u>	0.05	
		Vo1	lume used	30 <u>+</u> 0	.5 1	
		Poi	ison	Nil		
	2. Sed	liment: typ	oe .	Core #	8, 0-2 cm	
		amo	ount added	1.00 0		

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	6.00	6.70	39.6	
Oh	6.00	6.80	118.2	2358
1h	6.00	6.80	Lost	
2h	6.00	6.70	103.3	1911
4h	6.05	6.50	86.9	1419
5h	6.00	6.20	82.8	1296
7h	6.00	6.10	69.6	900
12h	5.95	6.30	72.8	996
18h	5.90	6.75	76.9	1119
1d	6.00	5.95	48.7	273
2d	6.00	6.10	58.8	576
<b>3</b> d	6.00	6.20	47.7	243
4d	6.00	6.45	44.2	138

Experiment No. 7

Initial Conditions	1. Water:	Salinity	$25.6 \pm 0.1^{\circ}/00$
		рН	6.90 <u>+</u> 0.05
		Volume used	30 <u>+</u> 0.5 1
		Poison	Nil
	2. Sediment:	type	Core #8, 0-2 cm
		amount added	1.45 g

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.00	6.50	44.4	
Oh	6.90	6.60	79.8	732
1h	6.95	6.75	68.5	499
2h	7.00	6.20	109.9	1355
4h	7.00	6.10	99.5	1140
5h	7.00	6.25	77.7	689
7h	7.00	6.40	85.4	848
12h	6.95	6.80	78.2	699
18h	7.00	6.35	66.7	461
1d	7.10	6.45	48.5	85
2d	7.00	6.30	53.4	186
3d	7.05	6.20	44.2	-4.
4d	7.00	6.35	44.6	4
5d	6.90	6.00	44.1	-6

Experiment No. 7

Initial Conditions	1.	Water:	Salinity	$25.1 \pm 0.1^{\circ}/00$
			рН	8.05 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Ni1
	2.	Sediment:	type	Core #8, 0-2 cm

amount added 1.14 g

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	8.10	6.55	45.5	
Oh	8.00	6.40	51.3	153
1h	8.00	6.30	39.2	-166
2h	8.05	6.15	87.6	1108
4h	8.00	5.90	76.5	816
5h	7.90	6.40	66.0	539
7h	7.95	6.70	54.4	234
12h	7.95	6.90	66.7	558
18h	7.95	6.75	41.2	-113
1d	7.95	6.85	37.4	-213
2d	8.15	6.30	33.3	-321
3d	8.10	6.45	36.8	-229
4d	8.10	6.10	33.0	-329
5d	8.00	6.25	32.6	-339.

Experiment No. 7

Initial Conditions	1. Water	: Salinity	$25.2 \pm 0.1^{\circ}/00$
		рН	9.00 <u>+</u> 0.05
		Volume used	30 <u>+</u> 0.5 1
		Poison	Ni1
	2. Sedim	ent: type	Core #8, 0-2 cm
		amount added	1.00 g

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	8.85	6.50	40.5	
Oh	8.80	6.30	51.6	333
1h	8.95	6.15	Lost	
2h	9.05	5.95	53.0	375
4h	9.00	6.10	52.3	354
5h	9.05	5.80	68.4	837
7h	9.10	6.15	73.8	999
12h	9.10	6.20	51.2	321
18h	9.10	6.30	58.1	528
1d	8.95	6.35	27.3	<del>-</del> 396
2d	8.90	6.40	28.7	-354
3d	8.95	6.55	24.7	-474
4d	9.00	6.60	23.5	-510
5d	9.10	6.35	24.8	-471

Experiment No. 8

Initial Conditions	1.	Water:	Salinity	$4.5 \pm 0.1^{\circ}/\circ \circ$
			pН	4.00 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Nil
	2.	Sediment:	type	#2 0-2 cm
			amount added	7.97 g

Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Amount Hg released/g sed.
Background	4.00	6.40	30.8	
Oh	4.00	6.30	93.0	234.1
1h	4.00	6.25	193.4	612.0
3h	4.00	6.35	79.7	184.1
7h	4.00	6.70	74.9	166.0
12h	4.00	6.60	64.0	125.0
1d	4.00	6.75	47.7	63.6
2d	4.00	6.80	45.1	53.8
3d	4.00	6.40	42.2	42.9
4d	4.00	6.50	44.6	51.9
5 <b>d</b>	4.00	6.30	48.3	65.9

Experiment No. 8

Initial Conditions	1.	Water:	Salinity	4.5 <u>+</u> 0.1°/00
	•		рН	6.00 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Nil
	2.	Sediment:	type	#2 0-2 cm

amount added 8.77 g

Sampling time	рН	T(°C)	Corrected Hg	Amount Hg
nambiting cime	pii	1( ()	conc (ng/1)	released/g sed.
Background	6.05	6.90	30.7	
Oh	6.00	6.85	43.9	45.2
1h	6.15	6.40	97.2	227.5
3h	6.00	6.20	137.5	365.3
7h	5.95	6.60	160.9	445.4
1 2h	5.90	6.70	104.6	252.8
1 d	5.95	6.30	62.3	108.1
2d	6.00	6.80	26.9	-13.0
3d	6.00	6.70	26.7	-13.7
4d	6.00	6.50	26.6	-14.0
5 <b>d</b>	6.00	6.40	27.1	-12.3

Initial Conditions	1.	Water:	Salinity	$4.5 \pm 0.1^{\circ}/00$
			pН	6.80 <u>+</u> 0.05
		•	Volume used	$30 \pm 0.5 1$
			Poison	Ni1
	2.	Sediment:	type	#2 0-2 cm

amount added 8.47 g

Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Amount Hg released/g sed.
Background	6.90	6.85	28.7	
Oh	6.85	6.15	88.2	210.7
1h	6.95	5.95	98.3	246.5
3h	6.90	6.80	65.2	129.3
7h	6.90	6.70	46.4	62.7
12h	6.95	6.50	48.1	68.7
1d	6.90	6.75	23.2	-19.5
2d	6.95	6.70	32.4	13.1
3d	7.10	6.40	31.5	9.9
4d	6.95	6.60	29.0	1.1
5 <b>d</b>	6.85	6.30	30.0	4.6

Initial Conditions	1.	Water:	Salinity	$4.5 \pm 0.1^{\circ}/\circ \circ$
			pН	7.95 <u>+</u> 0.05
			Volume used	31 <u>+</u> 0.5 1
			Poison	Ni1
	2.	Sediment:	type	#2 0-2 cm
			amount added	7.75 g

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Amount Hg released/g sed.
Background	7.85	6.85	28.2	
Oh	7.95	6.80	33.3	20.4
1h	8.00	6.40	58.8	122.4
3h	8.05	6.90	76.2	192.0
7h	7.95	6.85	50.9	90.8
12h	7.90	6.80	50.9	90.8
1d	7.95	6.70	27.5	-2.8
2d	7.90	6.90	32.4	16.8
3d	7.90	6.95	33.4	20.8
4d	7.95	6.40	27.7	-2.0
5d	7.90	6.60	79.0*	

<sup>\*</sup> Contaminated sample

Experiment No. 8

Initial Conditions	1. Water	: Salinity	$4.5 \pm 0.1^{\circ}/\circ \circ$
		рН	$9.00 \pm 0.05$
		Volume used	30 <u>+</u> 0.5 1
		Poison	Nil
•	2. Sedim	ent: type	#2 0-2 cm

amount added 7.89 g

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Amount Hg released/g sed.
Background	9.10	6.90	29.9	
Oh	9.00	6.70	81.3	195.4
1h	8.80	6.70	85.1	210.0
3h	8.95	6.75	50.2	77.2
7h	9.10	6.70	75.9	174.9
12h	9.15	6.80	50.9	79.8
1d	9.00	6.40	47.0	79.8
2d	9.15	6.30	40.8	65.0
3d	9.00	6.50	33.2	12.5
4d .	9.05	6.80	30.6	2.7
5d	9.10	6.20	30.9	3.8

Experiment No. 8

4d

5d

Initial Conditions	1. Water:	Salinity	23.3 ± 0.1°/oo
		pН	3.95 <u>+</u> 0.05
	v.	Volume used	31 <u>+</u> 0.5 1
		Poison	Ni1
	2. Sediment:	type	#2 0-2 cm

Corrected Hg Amount Hg  $T(^{o}C)$ Sampling time conc (ng/1)released/g sed. pН 6.90 32.4 Background 4.05 4.00 6.85 66.3 139.4 Oh 4.00 6.50 323.5 1196.8 1h 123.1 4.15 6.55 372.9 3h 3.90 6.50 96.5 263.5 7h 3.95 6.8086.7 223.2 12h 77.7 3.95 7.00 51.3 1d104.0 2d3.95 6.40 57.7 6.90 48.3 65.4 3d4.00

6.80

6.85

4.00

3.90

amount added

63.4

56.9

7.54 g

127.5

97.5

				_		^
Exp	er	1m	en	t	No	. 8

Initial Conditions	1. Water:	Salinity	$23.3 \pm 0.1^{\circ}/\circ o$
		рН	6.00 <u>+</u> 0.05
		Volume used	30 <u>+</u> 0.5 1
		Poison	Ni1
	2. Sediment	: type	#2 0-2 cm
		amount added	8.64 g

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Amount Hg released/g sed.
Background	6.00	6.05	37.8	
Oh	6.05	6.15	99.0	212.5
1h	6.00	6.20	157.1	414.2
3h	6.00	6.30	78.5	141.3
7h	6.00	6.00	79.4	144.4
12h	6.05	6.90	65.5	96.2
1d	5.95	6.85	50.2	43.1
2d	5.90	6.90	45.3	26.0
3d	6.05	6.80	37.8	0
4d	6.00	6.40	41.4	12.5
5d	6.00	6.45	49.1	39.2

Experiment No. 8

Initial Conditions	1.	Water:	Salinity	$23.3 \pm 0.1^{\circ}/00$
			рН	7.00 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Ni 1
	2.	Sediment:	type	#2 0-2 cm

amount added 7.70 g

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Amount Hg released/g sed.
Background	7.00	6.95	28.5	
Oh	7.00	6.85	66.4	147.7
1h	7.00	6.40	162.9	523.6
3h	7.00	6.30	115.4	338.6
7h	7.00	6.80	41.6	51.0
12h	6.90	6.80	54.3	100.5
1d	6.90	6.20	27.0	-5.8
2d	7.00	6.10	27.1	-5.5
3d	7.00	6.50	28.0	-1.9
4 d	7.00	6.90	34.1	21.8
5d	6.95	6.60	31.1	10.1

Experiment No. 8

Initial Conditions	1.	Water:	Salinity	$23.3 \pm 0.1^{\circ}/00$
			pH	8.00 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Nil
	2.	Sediment:	type	#2 0-2 cm

amount added 8.06 g

Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Amount Hg released/g sed.
Background	8.00	6.20	31.6	
Oh	8.00	6.30	73.6	156.3
1h	8.10	6.40	138.5	397.9
3h	8.05	6.45	72.9	153.7
7h	8.00	6.40	86.2	203.2
12h	8.00	6.10	31.6	0
1d	8.00	6.00	25.6	-22.3
2d	8.00	6.05	27.4	-15.6
3d	7.95	6.35	24.8	-25.3
4d	7.90	6.90	22.6	-33.5
5d	8.05	6.55	26.5	-15.3

Experiment No. 8

Initial Conditions	1.	Water:	Salinity	23.3 ± 0.1°/oo
			pН	$9.00 \pm 0.05$
			Volume used	30 <u>+</u> 0.5 1
			Poison	Nil
	_			

2. Sediment: type #2 0-2 cm

amount added 8.53 g

Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Amount Hg released/g sed.
Background	9.10	6.40	34.6	
Oh	9.00	6.40	42.2	26.7
1h	8.85	6.30	86.4	182.2
3h	8.95	6.30	16.2	-64.7
7h	8.95	6.35	27.3	-25.7
12h	8.95	5.90	22.6	-42.2
1 d	8.95	5.85	23.9	-37.6
2d	8.95	6.80	20.9	-48.2
3d	8.90	7.00	21.4	-46.4
4d	8.95	7.10	23.5	-39.0
5d	8.95	6.75	25.7	-31.3

Experiment No. 8

Initial Conditions	1.	Water:	Salinity	$23.3 \pm 0.1^{\circ}/\circ \circ$
			pН	6.90 <u>+</u> 0.05
			Volume used	30 <u>+</u> 0.5 1
			Poison	Nil
	2.	Sediment:	type	
			amount added	Nil
Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	
Background	6.95	6.40	38.0	
Oh 1h 3h 7h 12h 1d 2d 3d 4d	7.00 7.05 7.00 7.00 7.15 6.90 6.95 7.05 7.00	6.30 6.30 6.80 6.10 6.40 5.95 6.80 6.70 6.80	124.7* 203.0* 39.6 52.5 37.1 37.8 38.7 35.2 27.3	
5d	6.95	6.40	39.2	

<sup>\*</sup>Contaminated sample

					••	^
Exp	er	1 m	en	t	No	y

Initial Conditions	1.	Water:	Salinity	$7.6 \pm 0.1^{\circ}/_{\circ}$
			pН	7.65 <u>+</u> 0.05
			Volume used	40 <u>+</u> 0.5
			Poison	Ni1
	2.	Sediment:	type	Core #1, 2-4 cm
			amount added	11.67 g

Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.60	8.95	19.2	
Oh	7.60	8.95	17.6	- 5.5
1h	7.65	8.90	22.7	12.0
2h	7.65	8.90	23.4	14.4
4h	7.65	8.85	23.5	14.7
8h	7.60	8.70	26.8	26.0
12h	7.65	8.30	17.8	- 4.8
18h	7.70	7.90	27.4	28.1
24h	7.70	8.95.	33.3	48.3
1.5d	7.65	8.90	30.4	38.4
2d	7.65	8.80	29.8	36.3
2.5d	7.60	8.70	34.2	51.4
3d ·	7.65	8.90	35.2	54.8
3.5d	7.60	8.40	22.9	12.7
. 4d	7.60	8.10	54.3	120.3
4.5d	7.65	8.95	54.2	120.0
5d	7.70	8.70	36.3	58.6
5.5d	7.70	8.90	31.4	41.8
6d	7.65	8.30	29.4	35.0
8d	7.70	8.45	24.5	18.2
10d	7.65	8.95	27.5	28.4
11d	7.65	8.90	1415.1**	
12d	7.60	8.55	392.9**	

Experiment No. 9 Cont'd

13d

14d

15d

Initial Conditions	1.	Water:	Salinity	$7.6 \pm 0.1 / 00$
			pН	7.65 <u>+</u> 0.05
			Volume used	40 <u>+</u> 0.5
			Poison	Nil
	2.	Sediment:	<b>ty</b> pe	Core #1, 2-4 cm
			amount added	11.67 g
Sampling time	pН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment

92.3\*\*

69.2\*\*

30.8

28.2

7.55

7.60

7.65

8.65

8.60

8.20

<sup>\*\*</sup>Possible air contamination from broken Hg thermometer

Salinity

pН

 $7.6 \pm 0.1 \, ^{\rm O}/{\rm oo}$ 

 $7.75 \pm 0.05$ 

1. Water:

Experiment No. 9

			hu	7.75 ± 0.05
	. •		Volume used	40 ± 0.5 l
			Poison	2 g CuSO <sub>4</sub>
		2. Sediment:	type	core #1, 2-4 cm ,
			amount added	10.88 g
Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.75	8.90	20.9	-
0h	7.70	8.85	27.5	24.3
1h	7.70	8.85	33.6	46.7
2h	7.70	8.85	34.0	48.2
4h	7.80	8.85	50.9	110.3
8h	7.80	8.60	34.2	48.9
12h	7.75	8.10	27.8	25.4
18h	7.75	8.00	28.5	27.9
24h	7.70	8.80	35.4	53.3
1.5d	7.80	8.80	33.3	45.6
2d	7.85	8.70	30.0	33.5
2.5d	7.85	8.70	27.8	25.4
3d	7.80	8.85	26.7	21.3
3.5d	7.75	8.85	29.8	32.7
4d	7.80	8.75	27.5	24.3
4.5d	7.80	8.80	28.2	26.8
5d .	7.75	8.90	36.2	56.3
5.5d	7.75	8.90	37.2	60.0
6d	7.80	8.90	32.8	43.8
8d	7.70	8.50	38.1	63.2
10d	7.70	8.80	37.1	59.6
11d	7.70	8.80	1464.7**	
<b>1</b> 2d	7.70	8.65	249.5**	-
13d	7.70	8.75	66.1**	-
14d	7.75	8.55	36.4	57.0
15d	7.75	8.25	40.6	72.4

<sup>\*</sup> hole in nuclepore filter

<sup>\*</sup> possible air contamination from broken Hg thermometer

Initial Conditions		1.	Water:	Salinity pH	7.6 ± 0.1 °/oo 7.65 ± 0.05
				Volume used	45 ± 0.5 l
	• •			Poison	2 g CuSO <sub>4</sub>
		_			2 8 00004
		2.	Sediment:		₹ , , , , , , , , , , , , , , , , , , ,
				amount added	Ni1
Sampling time	рН		T(°C)	Corrected Hyconc (ng/1)	
Background	7.65		8.75	22.7	
0h	7.60		8.70	112.4*	
<b>1</b> h	7.60		8.70	27.9	
2h	7.65		8.95	18.5	
4h	7.65		8.85	18.2	•
8h	7.65		8.85	18.6	
12h	7.75		8.40	18.5	
18h	7.65		8.35	20.0	
24h	7.65		8.05	17.2	
1.5d	7.70		8.80	17.7	•
2d	7.65		8.85	18.9	
2.5d	7.65		8.75	19.8	
3d	7.70		8.75	19.4	
3.5d	7.70		8.90	23.8	•
4d	7.75		8.55	17.5	
4.5d	7.80		8.75	65.2*	
5d .	7.65		8.15	22.2	
5.5d .	7.65		8.80	23.8	•
6d	7.65		8.80	19.9	
8d	7.65		8.95	18.3	
10d	7.65		8.40	20.7	
11d	7.65		8.55	756.7**	
12d	7.70		8.80	153.3**-	
13d	7.65		8.70	112.6**	
14d	7.65		8.60	55.8**	
15d	7.65	,	8.20	28.7	

hole in nuclepore filter

possible air contamination from broken Hg thermometer

Salinity

pН

 $19.6 \pm 0.1 \, ^{\rm O/oo}$ 

 $7.50 \pm 0.05$ 

Experiment No. 9

					Volume used	40 ±	0.5 l
					Poison	Nil	
			2.	Sediment:	type		#1, 2-4 cm
					amount added	10.04	g
Sampling	time	pН		T(°C)	Corrected Hg conc (ng/1)	;	Released Hg (ng) per g sediment
Backgrou	ınd	7.50		8.90	36.4		<del>-</del>
Oh		7.50		8.80	51.7		61.0
1h		7.55		8.70	50.6		56.6
2h		7.55		8.80	61.8		101.2
4h		7.60		8.80	61.9		101.6
8h		7.55		8.40	43.0		26.3
12h		7.50		7.90	44.7		33.1
18h		7.50		7.90	47.5		44.2
24h		7.50		8.75	787.8*		-
1.5d		7.50		8.80	54.9		73.7
2d		7.50		8.80	50.1		54.6
2.5d		7.50		8.85	35.1		-5.2
3đ		7.50		8.60	42.8		25.5
3.5d		7.50		8.55	44.4		31.9
4d		7.65		8.40	42.2		23.1
4.5d		7.55		8.85	46.5		40.2
5d		7.55		8.85	38.6		8.8
5.5d		7.50		8.45	47.4		43.8
6d		7.55		8.35	44.1		30.7
8d		7.55		8.85	46.9		41.8
10d		7.55		8.85	44.7		33.1
11d		7.55		8.85	129.0**		-
12d		7.55		8.90	90.7**		<del>-</del>
13d		7.60		8.85	535.6**		<del>-</del>
14d		7.60		8.50	452.0**		-
15d		7.55		8.65	67.0**		-

<sup>\*</sup> hole in nuclepore filter

<sup>\*\*</sup> possible air contamination from broken Hg thermometer

Salinity

pН

 $19.6 \pm 0.1$   $^{\rm o}/_{\rm oo}$ 

 $7.45 \pm 0.05$ 

#### Experiment No. 9

	•		P.	
	. •		Volume used	40 ± 0.5
			Poison	2 g CuSO <sub>4</sub>
		2. Sediment:	type	core #1, 2-4 cm
			amount added	9.26 g
Sampling time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.55	8.75	31.7	_
0h	7.45	8.65	40.0	35.9
1h	7.45	8.70	60.2	123.1
2h	7.50	8.70	50.3	80.3
4h	7.50	8.65	55.3	101.9
8h	7.40	8.40	43.2	49.7
12h	7.50	8.00	42.8	47.9
18h	7.55	7.90	39.2	32.4
24h	7.55	8.15	72.9	178.0
1.5d	7.55	8.80	44.7	56.2
2d	7.55	8.80	37.5	25.1
2.5d	7.55	8.75	47.0	66.1
3d	7.55	8.75	43.3	50.1
3.5d	7.50	8.80	47.7	69.1
4d	7.55	8.75	65.0	143.8
4.5d	7.55	8.85	93.5	267.0
5d	7.55	8.75	55.6	103.2
5.5d	7.55	8.75	43.4	50.5
6d	7.55	8.70	64.6	142.1
8d	7.55	8.60	42.9	48.4
10d	7.55	8.65	44.2	54.0
11d	7.55	8.55	1401.9**	· <del>-</del>
12d	7.55	8.80	233.5**	-
13d	<b>7.</b> 55	8.80	249.7**	_
14d	7.55	8.70	127.8**	-
15d	7.55	8.70	79.4**	-

<sup>\*</sup> hole in nuclepore filter

<sup>\*\*</sup> possible air contamination from broken Hg thermometer

#### Experiment No. 9

Initial Conditio	ns	1.	Water:	Salinity pH Volume used Poison	7.45	± .1 °/oo ± 0.05 0.5 l
		2.	Sediment:	type amount added	- Nil	
Sampling time	pН		T(°C)	Corrected Hg conc (ng/1)	3	Released Hg (ng) per g sediment
Background	7.50		8.85	19.0		
0h	7.45	•	8.80	19.9		
1h	7.50		8.70	15.6		
2h	7.55		8.90	90.6*		
4h	7.45		8.80	23.5		•
8h	7.50		8.50	20.5		
12h	7.50		8.45	21.2		•
18h	7.50		8.40	17.4		
24h	7.45		8.35	30.0		
1.5d	7.50		8.90	19.6		
2d	7.50		8.85	18.3		
2.5d	7.45		8.70	17.6		
3d	7.55		8.85	19.0		
3.5d	7.45		8.80	17.2		
4d	7.50		8.75	25.2		
4.5d	7.45		8.80	159.2*		
5d .	7.50		8.85	21.0		
5.5d .	7.50		8.85	25.4		
6d	7.50		8.90	19.4		•
8d	7.50		8.75	20.9		
10d	7.45		8.80	18.5		
11d	<b>7.</b> 50		8.80	580.9**		
12d	7.50		8.80	216.9**		
13d	<b>7.</b> 50		8.40	75.6**		
14d	7.45		8.20	43.5**		
15d	7.50		8.45	40.4		

<sup>\*</sup> hole in nuclepore filter

<sup>\*\*</sup> possible air contamination from broken Hg thermometer

						•
Initial Conditi	Lons	1.	Water:	Salinity	19.6	± 0.1 °/oo
	-			pН	7.55	± 0.05
	. •			Volume used	40 ±	0.5 l
				Poison	2 g (	CuSO <sub>4</sub>
		2.	Sediment:	type	_	i
				amount added	Nil	
Sampling time	pН		T(°C)	Corrected Hagonic (ng/1)	3	Released Hg (ng) per g sediment
Background	7.50		8.80	20.8		
0h	7.55		8.85	18.1		
1h	7.60		8.85	23.0		
2h	7.60		8.85	97.0*		
4h	7.60		8.80	21.4		
8h	7.60		8.40	25.1		
12h	7.50		8.30	19.7		
18h	7.60		8.80	24.6		
24h	7.55		8.90	21.8		
1.5d	7.60		8.95	21.2		
2d	7.60		8.40	21.8		
2.5d	7.55		8.80	22.9		
3d	7.55		8.55	23.0		
3.5d	7.55		8.80	22.0		
4d	7.55		8.30	22.2		
4.5d	7.55		8.55	22.9		
5d .	7.55		8.80	43.1		
5.5d .	7.60		8.85	28.0		
6d	7.55		8.90	25.7		
<b>8</b> d	7.55		8.90	25.3		
10d	7.40		8.95	23.9		
11d	7.55		8.75	991.2**		

7.55

7.60

7.60

7.55

12d

13d

14d

15d

8.85

8.55

8.55

8.55

288.8\*\*

105.7\*\*

39.3

19.0

<sup>\*</sup> hole in nuclepore filter

<sup>\*\*</sup> possible air contamination from broken Hg thermometer

Salinity

28.8 ± 0.1 °/oo

1. Water:

#### Experiment No. 9

			pН	7.80 ± 0.05
	. •		Volume used	40 ± 0.5 l
			Poison	Nil
	,	2. Sediment:	type	core #1, 2-4 cm ,
			amount added	9.68 g
Sampling t	time pH	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Backgroun	nd 7.75	8.75	19.2	-
0h	7.85	8.95	24.9	23.6
1h	7.80	8.95	56.5	154.1
2h	7.70	8.90	66.3	194.6
4h	7.70	8.85	63.2	181.8
8h	7.70	8.80	46.0	110.7
12h	7.75	8.35	50.0	127.3
18h	7.80	8.05	53.0	139.7
24h	7.90	8.80	40.0	86.0
1.5d	7.85	8.75	33.4	58.7
2d	7.85	8.80	32.0	52.9
2.5d	7.80	8.85	33.2	57.9
3d	7.80	8.85	17.8	-5.8
3.5d	7.80	8.20	33.9	60.7
4d	7.80	8.05	34.5	63.2
4.5d	7.80	8.80	37.0	73.6
5d	. 7.75	8.70	31.3	50.0
5.5d	7.80	8.85	29.7	43.4
6d	7.75	8.10	31.6	51.2
8d	7.75	8.15	28.7	39.3
10d	7.80	8.20	29.3	41.7
11d	7.80	8.95	702.1**	- -
12d	7.85	8.80	88.6**	-
13d	7.80	8.75	65.9**	
14d	7.80	8.70	30.4	46.3
15d	7.80	8.45	37.3	74.8

<sup>\*</sup> hole in nuclepore filter

<sup>\*\*</sup> possible air contamination from broken Hg thermometer

Salinity

28.8  $\pm$  0.1  $^{\rm o}/{\rm oo}$ 

Experiment No. 9

				рН	$7.75 \pm 0.05$
				Volume used	40 ± 0.5 ℓ
				Poison	2 g CuSO <sub>4</sub>
		2.	Sediment:	type	core #1, 2-4 cm
				amount added	11.53 g
Sampling	time	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Backgrou	und	7.80	8.60	25.0	-
0h		7.80	8.60	29.6	16.0
1h		7.80	8.65	23.8	-4.2
2h		7.70	8.60	27.9	10.1
4h		7.75	8.60	23.3	-5.9
8h		7.75	8.50	30.2	18.0
12h		7.75	8.00	52.0	93.7
18h		7.80	8.60	99.3	257.8
24h		7.80	8.65	60.7	123.9
1.5d		7.85	8.65	53.0	97.1
2d		7.80	8.65	49.0	83.3
2.5d		7.80	8.70	58.4	115.9
3d		7.80	8.65	91.2	230.0
3.5d		7.80	8.65	97.1	250.1
4d		7.85	8.35	99.0	256.7
4.5d		7.75	8.30	119.7	328.5
5d		7.85	8.70	61.5	126.6
5.5d		7.75	8.70	44.3	67.0
6d		7.80	8.65	44.9	69.0
<b>8</b> d		7.80	8.60	39.6	50.7
10d		7.75	8.40	42.8	61.8
11d		7.75	8.45	534.9**	-
12d		7.85	8.70	121.6**	-
13d		7.75	8.75	83.9**	-
14d		7.80	8.60	45.2	70.1
15d		7.75	8.40	44.3	67.0

<sup>\*</sup> hole in nuclepore filter

possible air contamination from broken Hg thermometer

Initial Conditions	1.	Water:	Salinity	$28.8 \pm 0.1$ °/00
			pН	$7.80 \pm 0.05$
.•			Volume used	40 ± 0.5 l
			Poison	Nil

2. Sediment: type

amount added Nil

Sampling tim	рН	T(°C)	Corrected Hg conc (ng/1)	Released Hg (ng) per g sediment
Background	7.80	8.80	28.1	
0h	7.80	8.80	23.7	
1h	7.80	8.75	38.7	
2h	7.80	8.75	23.1	•
4h	7.75	8.75	22.9	
8h	7.75	8.60	21.4	
12h	7.80	8.25	22.2	
18h	7.80	7.80	20.8	
24h	7.85	8.65	19.0	•
1.5d	7.85	8.70	24.7	
2d	7.80	8.70	21.8	
2.5d	7.80	8.65	24.9	
3d	7.80	8.80	20.1	
3.5d	7.80	8.20	23.0	
4d	7.80	8.00	22.8	r
4.5d	7.80	8.80	28.7	
5d	. 7.85	8.70	20.0	
5.5d	7.85	8.75	26.9	
6d	<b>7.</b> 75	8.45	24.4	
8d	7.80	8.55	22.0	
10d	7.80	8.75	25.1	
11d	7.80	8.85	703.1**	
12d	7.80	8.35	172.6**	
13d	7.80	8.35	68.1**	
14d	7.85	8.80	23.4	
15d	7.80	8.30	89.8**	

<sup>\*</sup> hole in nuclepore filter

<sup>\*\*</sup> possible air contamination from broken Hg thermometer

Experiment No. 9

Initial Conditions		1.	Water:	Salinity pH Volume used Poison	7.80	± 0.1 °/00 ± 0.05 0.05 l CuSO <sub>4</sub>
		2.	Sediment:	type amount added	Blank Nil	t .
Sampling time	рН		T(°C)	Corrected Hg		Released Hg (ng) per g sediment
Background	7.80		8.60	29.4		
Oh	7.80		8.80	28.8		
<b>1</b> h	7.85		8.85	223.4*		
2h	7.90		8.80	23.5		
4h	7.80		8.80	27.0		
8h	7.80		8.40	31.4		
12h	7.85		8.20	34.7		•
18h	7.75		7.90	34.4		
24h	7.80		8.45	32.4		
1.5d	7.80		8.80	30.7		
2d	7.80		8.85	25.2		
2.5d	7.85		8.80	25.3		
3d	7.80		8.80	26.4		
3.5d	7.80		8.85.	27.0		
4d	7.85		8.90	31.4		
4.5d	7.85		8.85	66.2		
5d .	7.80		8.80	27.9		
5.5d	7.80		8.90	28.7		
6d	7.90		8.90	28.7		
8d	7.80		8.65	35.1		
10d	7.75		8.70	24.9		
11d	7.75		8.80	228.1**		
12d	7.75		8.80	201.5**		
13d	7.75		8.65	132.4**		
14d	7.75		8.40	33.7		
15d	7.75		8.40	72.5**		

<sup>\*</sup> hole in nuclepore filter

possible air contamination from broken Hg thermometer

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			$\cdot$									-								)			•					•		

# Appendix A.4

Data for Static Release Experiments

# Aerobic Static Release Experiment Daily Averaged Mercury Flux

Time (days)	Mercury Concentration in Solution (ng/1)	n Total ∆ Hg (ng)	Hg Flux (ng/cm²/day)
0	398	•	
1	697	3348	8 . 4
	782	1507	3.8
2 3	2204	14962	37.4
4	3200	12124	30.3
	3860	9660	24.2
5 6	4308	8390	21.0
7	5328	14468	36.2
8	8098	32988	82.5
9	6030	-12622	- 31.6
10	7720	22890	57.2
11	8050	10980	27.5
12	10170	29210	73.0
13	10926	17690	44.2
14	11860	20226	50.6
15	8550	-21280	- 53.2
16 .	9520	18210	45.5
17	17726	91540	228.9
18	23798	78406	196.0
19	40250	188278	470.7
20	23400	128290	320.7
21	9855	-122090	-280.2
22	7330	- 15434	- 38.6
23	7078	4770	11.9
24	2173	- 42012	-105.0
25	6665	47053	117.6
26	6288 ·	2855	7.1
27	6320	6568	16.4
28	6028	3360	8.4
29	12358	69228	5.4 5.71.3 5.71.3
30	3388	- 28519	71.3

Concentration of Mercury in Sediment (ppb) at Completion of Aerobic Static Release Experiment, t = 30 d

Depth Along Core (cm)	0-1	4-5	9-10	14-15	19-20
П	471 (2.0)*	374 (1.1)	543 (0.8)	347 (0.3)	125
2	109 (2.0)	339	91 (1.1)	460 (0,2)	, 952 ( .1)
е	508 (1.8)	428 (1.6)	511 (1.4)	888 (1.3)	440
4	566 (1.9)	441 (1.7)	440 (0.5)	894 (1.5)	
₽	423 (1.8)	360 (1.5)	361 (0.9)	402 (0.6)	368
9	421 (2.0)	199	428 (1.1)	233 (0.4)	210
7	385 (1.6)	$^{217}_{(1.2)}$	332 (1.0)	501 (1.0)	278
œ	324 (2.1)	576 (1.1)	466 (0.6)	570 (0.8)	944
6	602 (2.0)	559 (1.1)	453 (0.3)	328 (0.1)	281

\* Bracketted numbers refer to % organic carbon content of sediment

Plan view of box core in container Sediment injected into 5 through 2 at a depth of  $10\ \mathrm{cm}.$ 

1 2 3

9

4 5

Appendix B.1

Data for In Situ Study

### Dissolved Mercury in Water and Interstitial Waters Squamish River Estuary and Mud Flats

FMC Outfall

 $1.0 \mu g/1$ 

Station	Sampling Period	Dissolved Mercury (ng/l)	Interstitial Mercury (µg/1)
1	Background	240	
	1	320	2.0
	1 2 3 4		
	3		
	4	390	1.5
	5	460	
2	Background	235	
	1	260	14.4
	2 3		12.1
\$			
	4	510	2.0
	5	310	
3	Background	290	
	1	300	14.1
	2 3		16.0
		310	5.5
	4	380	1.4
	5	310	
4	Background	42	
	1	140	16.7
	2		14.6
	3	• 170	4.9
	4	_ 270	1.5
	5	170	
5	Background	33	
	1	92	1.2
	2 3 4 5		2.1
	3	69	1.9
	4	91	1.8
	5	60	,

# Particulate Mercury in Waters of Squamish River Estuary

Station	Sampling Period	Suspended Solids	Particulate Hg (ng/1)
1	1	30.2	280
	2		
	1 2 3		
	4	680.5	860
	5	860.3	78000
2	1	41.6	520, 480
	2		
	3		
	4	785.4	3200, 3400
	4 5	880.2	1400
3	1	95.1	3800
	2 3		
	3	65.3	3200
	4	159.8	2800
	5 1	585.5	1440
4		17.8	780
	2 3		
		173.4	320
	4	496.7	300
	5	182.8	220
5	1	1.0	40
	2		
	3	1.8	44
	4 5	560.3	360
	5	· 77.7	1340

7. 2 . 6 . 6 . 6 . 7 . 6 . 7 . 6 . 7 . 6 . 7 . 6 . 7 . 6 . 7 . 6 . 7 . 6 . 7 . 6 . 7 . 6 . 7 . 6 . 7 . 6 . 7 . 6 . 7 . 6 . 7 . 7
7. 2
100   100
1/124   14/43   3141/2   3/4
17.2
100   100
2007   2007
2003   2005
7/1/2 27/
7/1/5 2/1/5
7/1/25 7/1/25 7/1/27 7/2/2/2/2/2/2/2/2/2/2/2/2/2/2/2/2/2/
2007   2007
2000 2000 2000 2000 2000 2000 2000 200
7/15 27/75 27/77 27/75 27/77 27/75 27/77 27/75 27/77 27/75 2
2005   20
2008 21/15 27/27 2008 7/1946 22/29 2008 7/1946 22/29 2007 2008 7/1946 2007 2008 7/1947 2007 2008 7/1947 2007 2008 7/1947 2008 7/1948 7/1948 2008 7/1948 2008 7/1948 2008 7/1948 2008 7/1948 2008 7/1948 2008 7/1948 2008 7/1948 2008 7/1948 2008 7/1948 2008 7/1948 2008 7/1
7648 7774 7777 7588 77976 7229 76814 78976 7574 7 7757 7788 77976 7 7757 7788 7788 7787 7757 7788 7788 7788 7758 7788 7788 7788 7758 7788 7788 7788 7788 7788 778
2012 2114 2214 2214 2214 2214 2214 2214
44/2 76 12 47/2 57/2 57/2 57/2 57/2 57/2 57/2 57/2 5
10 20 10 10 10 10 10 10 10 10 10 10 10 10 10
12 5 12 12 12 12 12 12 12 12 12 12 12 12 12
7/14 27 12/15 27/17 28/15 28/1
- 3775 3775 - 3776 - 37
76/16 5/1/2 3/20