



FEDERAL CONTAMINATED SITE
RISK ASSESSMENT IN CANADA:
**Toxicological Reference
Values (TRVs)**

VERSION 4.0



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Également disponible en français sous le titre :

L'évaluation des risques pour les sites contaminés fédéraux au Canada : Valeurs toxicologiques de référence (VTR), version 4.0

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Publication date: June 2025

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Cat.: H129-108/2025E-PDF

ISBN: 978-0-660-78211-9

Pub.: 250150

TABLE OF CONTENTS

| | |
|--|----|
| PREFACE | 01 |
| SUMMARY OF REVISIONS | 02 |
| ACRONYMS AND ABBREVIATIONS | 03 |
| UNITS | 06 |
| 1.0 INTRODUCTION | 07 |
| 1.1 GENDER-BASED ANALYSIS PLUS | 08 |
| 2.0 TRVs RECOMMENDED FOR CONTAMINATED SITES | 09 |
| 2.1 TRVs FOR ENVIRONMENTAL CONTAMINANTS | 09 |
| 2.2 TRVs FOR ESSENTIAL TRACE ELEMENTS | 10 |
| 2.3 RELATIVE POTENCY FACTORS/TOXIC EQUIVALENCY FACTORS | 11 |
| 3.0 RELATIVE ABSORPTION FACTORS FOR DERMAL EXPOSURE | 12 |
| 4.0 SUMMARY TABLES | 13 |
| 5.0 REFERENCES | 52 |

LIST OF TABLES

| | |
|---|----|
| TABLE 1: TRVs Recommended for Use in Human Health Risk Assessments of Federal Contaminated Sites..... | 14 |
| TABLE 2: Recommended RPFs for Carcinogenic PAHs..... | 49 |
| TABLE 3: Provisional RPFs for Carcinogenic PAHs..... | 49 |
| TABLE 4: TEFs for PCDDs, PCDFs, and Dioxin-Like PCBs..... | 50 |
| TABLE 5: Recommended Dermal Relative Absorption Factors (RAF_{Derm})..... | 51 |

PREFACE

The Federal Contaminated Sites Action Plan (FCSAP) was established in 2005 as a 15-year horizontal program with funding of \$4.54 billion from the Government of Canada. In 2019, the program was renewed for another 15 years, from 2020 until 2035.

The primary objective of FCSAP is to reduce environmental and human health risks from known federal contaminated sites in Canada and their associated federal financial liabilities. To achieve this objective, FCSAP funds federal departments, agencies and consolidated Crown corporations (collectively referred to as “custodians”) to assess, remediate and risk manage the federal contaminated sites for which they are responsible. FCSAP also provides guidance, tools and resources to custodians to ensure that federal contaminated sites are managed in a scientifically sound and a nationally consistent manner. The *Federal Approach to Contaminated Sites* and the *FCSAP Decision-Making Framework (DMF)* provide a 10-step roadmap that outlines the specific activities, requirements and key decisions to effectively address federal contaminated sites in Canada. The *DMF* along with other FCSAP-related resources can be found on the [FCSAP website](#).

This guidance document supplements Health Canada’s (HC’s) preliminary and detailed quantitative risk assessment guidance and assists federal custodial departments with the consistent assessment of human health risks posed by federal contaminated sites across Canada.

Guidance documents on human health risk assessment (HHRA) prepared by HC in support of FCSAP may be obtained by contacting HC at cs-sc@hc-sc.gc.ca or from our website at: www.canada.ca/en/health-canada/services/environmental-workplace-health/contaminated-sites.html.

As is common with any national guidance, this document will not satisfy all requirements presented by federal contaminated sites, custodial departments or risk assessors. As the practice of HHRA advances and as FCSAP proceeds, new and updated information on various aspects of HHRA will be published. As a result, it is anticipated that revisions and/or addenda to this document will be necessary from time to time to reflect this new information. Please consult the HC website above to confirm that the version of the document in your possession is the most recent.

HC requests that any questions, comments, suggested additions or revisions to this document be directed to HC at the email address identified above.

SUMMARY OF REVISIONS

Federal Contaminated Site Risk Assessment in Canada: Toxicological Reference Values (TRVs), Version 4.0 reflects revisions to text and tables, relative to the previous version, *Federal Contaminated Site Risk Assessment in Canada, Toxicological Reference Values (TRVs), Version 3.0* (HC, 2021).

The report was updated to include the addition of text for consistency with Canada's Gender-Based Analysis Plus (GBA Plus) policy, and [Table 1](#) (was previously placed in Appendix A).

The following changes to the listed TRVs were included:

- › barium: updated tolerable daily intake (TDI)
- › benzo[a]pyrene: updated TDI, added note to oral slope factor (SF) regarding application of age-dependent adjustment factors (ADAFs) for early-life exposures
- › chromium, trivalent: updated TDI, removed inhalation tolerable concentration (TC)
- › dichloroethylene, 1,1: new inhalation unit risk (UR)
- › manganese: new inhalation TC
- › perfluorooctanoic acid (PFOA): removed oral TDI (currently under review)
- › perfluorooctane sulfonate (PFOS): removed oral TDI (currently under review)
- › tetrachloroethylene: new inhalation UR
- › trichloroethylene: added note to oral SF regarding application ADAFs for early-life exposures

The table of WHO's dioxin toxic equivalency factors (TEFs) that was presented in [Table 4](#) has been updated (DeVito et al., 2024).

ACRONYMS AND ABBREVIATIONS

| | |
|--------------------------------|---|
| ADAF | age-dependent adjustment factor |
| AROI | acceptable range of oral intake |
| ATSDR | Agency for Toxic Substances and Disease Registry (United States) |
| B[a]P | benzo[a]pyrene |
| BCL | Battelle Columbus Laboratories (now Battelle) |
| BMC | benchmark concentration |
| BMCL | lower limit of a one-sided 95% confidence interval on the BMC |
| BMCL_{01/05/10} | lower 95% confidence limit on a benchmark concentration associated with a 1%, 5%, or 10% response |
| BMD | benchmark dose |
| BMDL | lower limit of a one-sided 95% confidence interval on the BMD |
| BMDL_{01/05/10} | lower 95% confidence limit on a benchmark dose associated with a 1%, 5%, or 10% response |
| BW | body weight |
| CalEPA | California Environmental Protection Agency |
| CEPA | <i>Canadian Environmental Protection Act</i> |
| CSTEE | Scientific Committee on Toxicity, Ecotoxicity and the Environment |
| CF | conversion factor |
| CMP | Chemicals Management Plan |
| CCME | Canadian Council of Ministers of the Environment |
| CDHS | California Health and Human Services Agency's Department of Health Services |
| DRI | dietary reference intake |
| DQRA | detailed quantitative risk assessment |
| DWEL | drinking water equivalent level |
| EC | Environment Canada |
| ECB | European Chemicals Bureau |
| ECCC | Environment and Climate Change Canada |
| EFSA | European Food Safety Authority |
| EHMI | equivalent human monthly intake |
| ESOD | erythrocyte superoxide dismutase |
| ETE | essential trace element |
| FAO | Food and Agriculture Organization of the United Nations |
| FCSAP | Federal Contaminated Sites Action Plan |

| | |
|-----------------|--|
| GBA Plus | Gender-Based Analysis Plus |
| HC | Health Canada |
| HEC | human equivalent concentration |
| HEQ | human equivalent |
| HHRA | human health risk assessment |
| HQ | hazard quotient |
| IARC | International Agency for Research on Cancer |
| ILCR | incremental lifetime cancer risk |
| IOM | Institute of Medicine of the National Academies (renamed the National Academy of Medicine in 2015) |
| IPCS | International Programme on Chemical Safety |
| IRIS | Integrated Risk Information System (US EPA) |
| IRw | water ingestion rate |
| JISA | Japan Industrial Safety Association |
| JBRC | Japan Bioassay Research Centre |
| LEC | lowest effect concentration |
| LOAEL | lowest observable adverse effect level |
| MMAD | mean mass aerodynamic diameter |
| MOE | Ontario Ministry of the Environment (currently known as the Ministry of Environment, Conservation and Parks) |
| MF | modifying factor |
| MRL | minimal risk level |
| NIOSH | National Institute for Occupational Safety and Health |
| NOAEL | no observable adverse effect level |
| NTP | National Toxicology Program (US Department of Health and Human Services) |
| OEHHA | Office of Environmental Health Hazard Assessment (California Environmental Protection Agency) |
| PAH | polycyclic aromatic hydrocarbon |
| PBPK | physiologically based pharmacokinetic (model) |
| PCB | polychlorinated biphenyl |
| PCDD | polychlorinated dibenzodioxin |
| PCDF | polychlorinated dibenzofuran |
| PCE | perchloroethylene (tetrachloroethylene) |
| PFOA | perfluorooctanoic acid |
| PFOS | perfluorooctane sulfonate |

| | |
|---------------------------|---|
| PND | postnatal day |
| POD | point of departure |
| PQRA | preliminary quantitative risk assessment |
| PSL | Priority Substance List (Health Canada) |
| pTDI | provisional tolerable daily intake |
| pTMI | provisional tolerable monthly intake |
| pTWI | provisional tolerable weekly intake |
| RAF | relative absorption factor |
| RAF_{Derm} | dermal relative absorption factor |
| RAIS | Risk Assessment Information System |
| RDA | recommended dietary allowance |
| RDDR | regional deposited-dose ratio |
| RPF | relative potency factor |
| RfC | reference concentration |
| RfD | reference dose |
| SF | slope factor |
| TC | tolerable concentration |
| TC₀₅ | tumourigenic concentration found to induce a 5% increase in the incidence of, or deaths due to, tumours considered to be associated with exposure |
| TD₀₅ | tumourigenic dose found to induce a 5% increase in the incidence of, or deaths due to, tumours considered to be associated with exposure |
| TCDD | 2,3,7,8-tetrachlorodibenzo-p-dioxin |
| TCE | trichloroethylene |
| TDI | tolerable daily intake |
| TEF | toxic equivalency factor |
| TEQ | toxic equivalent |
| TRV | toxicological reference value |
| UF | uncertainty factor |
| UL | tolerable upper intake level |
| UR | unit risk |
| US EPA | United States Environmental Protection Agency |
| VOC | volatile organic compound |
| WHO | World Health Organization |

UNITS

| | |
|--|---|
| g/kg_{BW} | grams per kilogram of body weight |
| kg | kilograms |
| L/day | litres per day |
| mg/day | milligrams per day |
| mg/L | milligrams per litre |
| mg/kg_{BW} | milligrams per kilogram of body weight |
| mg/kg_{BW}-day | milligrams per kilogram of body weight per day |
| (mg/kg_{BW}-day)⁻¹ | per milligram per kilogram of body weight per day |
| mg/m³ | milligrams per cubic metre |
| (mg/m³)⁻¹ | per milligram per cubic metre |
| µg/day | micrograms per day |
| µg/g | micrograms per gram |
| µg/kg_{BW} | micrograms per kilogram of body weight |
| µg/kg_{BW}-day | micrograms per kilogram of body weight per day |
| µg/L | micrograms per litre |
| (µg/L)⁻¹ | per microgram per litre |
| µg/m³ | micrograms per cubic metre |
| ng/kg_{BW} | nanograms per kilogram of body weight |
| ng/kg_{BW}-day | nanograms per kilogram of body weight per day |
| pg/kg_{BW} | picograms per kilogram of body weight |
| pg/kg_{BW}-day | picograms per kilogram of body weight per day |
| pg/kg_{BW}-month | picograms per kilogram of body weight per month |
| ppm | parts per million |
| (ppm)⁻¹ | per part per million |

1.0 INTRODUCTION

This document is part of a series of guidance documents published by Health Canada (HC) for use in assessment of human health risks at federal contaminated sites in Canada. This document is designed for use with HC's Preliminary Quantitative Risk Assessment (PQRA) guidance (HC, 2024a) and Detailed Quantitative Risk Assessment (DQRA) guidance (HC, 2010a) to assist with standardization of human health risk assessment (HHRA) at federal contaminated sites funded under the Federal Contaminated Sites Action Plan (FCSAP). Application of the values presented in this document is discussed in the guidance referred to above.

Toxicological reference values (TRVs) are used to quantitatively assess human health risks associated with exposure to environmental contaminants and are published by a variety of national and international agencies. HC and other government agencies derive TRVs for chemical substances with a threshold mode of action and/or a non-threshold mode of action. Some substances may have both threshold and non-threshold endpoints.

- For substances with a threshold mode of action, the TRV is provided as a **tolerable daily intake (TDI)** for oral exposures, or **tolerable concentration (TC)** for inhalation exposures. The TDI and TC are typically derived based on a dose or exposure concentration at or below which no toxic effects are assumed to occur. The TRVs for threshold response chemicals represent the amount of exposure that is considered to be unlikely to cause adverse health effects in the general population, including sensitive individuals, but excluding those with allergy or other hypersensitivity (HC, 2010a), unless otherwise indicated.
- For substances without a threshold mode of action, such as certain carcinogens and germ cell mutagens, it is assumed that any level of exposure may result in an adverse effect. Typically, the TRV is derived after fitting the dose-response data for the observable/detectable range of effects to a mathematical model and then extrapolating to the low-dose region for which there are no data. This extrapolation allows for the estimation of oral **slope factors (SFs)** and inhalation **unit risks (URs)**.

Sources of TRVs for use in assessment of potential human health risks at Canadian federal contaminated sites include, but are not limited to, the following:

- HC:
 - › Contaminated Sites Reports and Publications - Federal Contaminated Site Risk Assessment in Canada: www.canada.ca/en/health-canada/services/environmental-workplace-health/reports-publications/contaminated-sites.html
 - › Environmental Contaminants: www.canada.ca/en/health-canada/services/environmental-workplace-health/reports-publications/environmental-contaminants.html
 - › Chemicals Management Plan: www.canada.ca/en/health-canada/services/chemical-substances/chemicals-management-plan.html
 - › Water Quality - Reports and Publications: www.canada.ca/en/health-canada/services/environmental-workplace-health/reports-publications/water-quality.html
 - › Air Quality and Health: www.canada.ca/en/health-canada/services/air-quality.html
- United States Environmental Protection Agency (US EPA)
 - › Integrated Risk Information System (IRIS): www.epa.gov/iris. TRVs are generally identified by the US EPA as oral reference doses (RfDs), inhalation reference concentrations (RfCs), oral slope factors (SFs), and inhalation unit risks (URs).

- California Environmental Protection Agency (CalEPA)
 - › Chemicals Database: <https://oehha.ca.gov/chemicals>
 - › The CalEPA employs the same general terminology as the US EPA.
- World Health Organization (WHO) and the International Programme on Chemical Safety (IPCS) - various sources including:
 - › Chemical Safety Information from Intergovernmental Organizations: www.inchem.org
 - › International Programme on Chemical Safety: www.inchem.org
 - › Air Quality: www.who.int/news-room/feature-stories/detail/what-are-the-who-air-quality-guidelines
 - › TRVs are generally identified by the WHO and the IPCS as tolerable daily intakes (TDIs).
- United States Agency for Toxic Substances and Disease Registry (ATSDR)
 - › Toxicological Profiles: <https://www.atsdr.cdc.gov/toxicological-profiles/about/index.html>
 - › The ATSDR generally identifies TRVs as minimal risk levels (MRLs).

1.1 GENDER-BASED ANALYSIS PLUS

In the context of contaminated site HHRA, gender-based analysis plus (GBA Plus) identifies and analyzes the differential impacts of contaminated sites on diverse population groups. The “plus” in GBA Plus acknowledges that GBA goes beyond biological (sex¹) and socio-cultural (gender²) differences. It highlights the pathways on which those differences develop and how they intersect with other determinants to shape health and well-being. It guides the consideration of sex and gender in framing, planning for, and conducting HHRA for contaminated sites. GBA Plus includes other individual and social identity factors such as race, religion, social position, income, age, ability, and education; this is called intersectionality³. The basic steps to applying GBA Plus include gathering appropriate data, understanding context, and asking analytical questions to determine whether there may be disproportionate effects on diverse populations. By working through a GBA Plus analysis, experts can better understand the possible differential effects on distinct groups of people, including on disproportionately affected or impacted populations and populations identified by sex and gender. Considering how a program, policy, plan or product related to assessment and management of a contaminated site might impact groups differently provides an opportunity for all those involved, to help address potential pitfalls before these become a problem, or to identify opportunities that would not have been otherwise considered.

¹ Sex refers to physical and physiological features including chromosomes, gene expression, hormone levels and function, and reproductive/sexual anatomy. <https://cihr-irsc.gc.ca/e/48642.html>

² Gender refers to the socially constructed roles, behaviors, expressions and identities of girls, women, boys, men, and gender diverse people. <https://cihr-irsc.gc.ca/e/48642.html>

³ Government of Canada’s Approach Gender Based Analysis Plus. <https://women-gender-equality.canada.ca/en/gender-based-analysis-plus/government-approach.html>

Key GBA Plus considerations in contaminated site HHRAs:

- Does the proposed assessment and management of the contaminated site identify the diverse communities who may be directly and indirectly affected by exposure to contamination at the site?
- Are data about potential impacts disaggregated by sex, gender, age, language and other social identities relevant to the local communities?
- Have perspectives and characteristics of affected diverse communities and disproportionately impacted groups been included in the HHRA?
- Has the toxicity assessment considered all age groups and is it protective of the entire population (i.e., protective of diverse communities and age groups)?
- What potential health impacts on the well-being of all people, including Indigenous peoples and disproportionately affected populations, may result from the contaminated site? What types of measures are needed to ensure equitable representation during engagement processes and subsequent stages of contaminated site assessment and management?
- What measures are needed to mitigate any adverse effects of the contaminated site on all people that may be affected by contamination, including Indigenous Peoples and disproportionately affected groups?

Identifying the range of concerns and interests of, and impacts on, diverse groups based on social characteristics like gender, age, ethnicity, occupation, and length of residency, for example, can help foster the development of more comprehensive mitigation and enhancement strategies.

2.0 TRVs RECOMMENDED FOR CONTAMINATED SITES

2.1 TRVs FOR ENVIRONMENTAL CONTAMINANTS

For the assessment of human health risks posed by substances found at federal contaminated sites in Canada, the TRVs in this document are recommended. If a published TRV other than those presented in [Table 1](#) is used in an HHRA (e.g., more recent data have been used by a different agency), these TRVs may be applied if supported by adequate scientific rationale. For substances that lack a TRV from regulatory or advisory agencies, please contact HC.

A list of TRVs recommended for environmental contaminants is presented in [Table 1](#); however, not all of the TRVs were derived by HC. In order to enable standardization of HHRAs for federal contaminated sites, where HC did not have a published TRV, TRVs were identified from other regulatory agencies. The basis, method of derivation, level of protection, uncertainty or confidence level of the TRVs, and any modifications made were considered in the identification of TRVs for use in HHRAs of federal contaminated sites.

It is not recommended to rely on occupational standards for the assessment of human health risks posed by substances found at federal contaminated sites. If occupational health and safety values are applied, an occupational health and safety specialist should be consulted.

The TRVs presented in [Table 1](#) are recommended for chronic exposures at contaminated sites. At this time, HC does not prescribe TRVs for exposures of lesser duration (i.e., acute, subchronic). Short-duration TRVs from other regulatory agencies may be used in risk assessments of federal contaminated sites, with scientific rationale. Refer to Appendix D of HC (2024a) PQRA guidance for additional information on human health risk assessment of short-duration exposures.

2.2 TRVs FOR ESSENTIAL TRACE ELEMENTS

Recommended TRVs for essential trace elements (ETEs) are presented in [Table 1](#).

The approach for establishing TRVs for ETEs considers the benefits and risks associated with these substances and reflects their characterization as essential elements. The Institute of Medicine of the National Academies (IOM, 2000, 2001, 2006) has published **tolerable upper intake levels** (UL) for ETEs. Health Canada recommends these ULs be interpreted and applied as TDIs for oral exposure for human health risk assessment when assessing risks associated with exposure to contaminants considered to be ETEs at federal contaminated sites in Canada. The use of ULs to assess the non-carcinogenic risks of an ETE does not preclude the need to quantify cancer risks for ETEs that may also have carcinogenic endpoints.

Some contaminants found at federal contaminated sites can also be ETEs. For example, the WHO considers the following trace elements to be essential in human nutrition: iron (FAO/WHO, 2001), cobalt, copper, iodine, molybdenum, selenium, zinc (WHO, 1996, 2002a) and manganese (IOM, 2001). There is a growing body of evidence that suggests that silicon, boron, nickel, and vanadium play essential metabolic roles in some species, and these elements have been considered to be **probable ETEs** by the WHO since 1996; however, given that human data on ULs for probable ETEs are limited, HHRAs of federal contaminated sites should address exposure to such ETEs based on the TRVs presented in [Table 1](#).

Due to their essentiality, ETE deficiency in the diet can result in biochemical, functional or structural effects. These effects may be reversed by adequate supplementation of the ETE (e.g., Mertz, 1980; WHO, 1996). For this reason, unlike RfD and TDI derivation, it is inappropriate to assume that a zero intake of ETEs is without risk (WHO, 2002a). Conversely, excess intake of an ETE may result in toxic effects that must be considered when establishing TDIs or RfDs. However, some TDIs or RfDs for ETEs derived using typical chemical risk assessment methods may be overly conservative when compared to dietary reference intakes (DRIs) established by the IOM's Food and Nutrition Board (IOM, 2000, 2001, 2006).

The Expert Advisory Committee on Dietary Reference Intakes (DRI Committee) developed a framework for development of dietary allowances and recommendations (IOM, 2000, 2001, 2006). The DRIs apply to healthy Canadian populations and consider bioavailability as well as nutrient and dietary interactions (Mertz, 1995; IOM, 2000; 2001; 2006; WHO, 2002a). Such DRIs are normally developed for the general population, according to age and gender (IOM, 2000; 2001, 2006). These DRIs consider physiological state to protect sensitive subpopulations (Mertz, 1998; Munro, 1999).

For each ETE, a safe range of oral intakes has been established to avoid both deficiency and toxicity ('acceptable range of oral intake' or AROI [WHO, 2005]). Each ETE has a homeostatic mechanism which involves regulation of absorption, excretion, and tissue retention. This mechanism allows adaptation to varying nutrient intakes for optimal systemic supply in order for essential functions to be carried out (WHO, 2002a). AROIs, including intake from food and water, are maintained under homeostasis in healthy populations (IOM, 2000, 2001, 2006). As nutrient needs vary considerably among individuals, deficiency and toxicity are not necessarily encountered at the lower and upper bounds of the AROIs, respectively (Becking, 1998). For DRIs within AROI limits, IOM (2000, 2001, 2006) defined the following:

- **Recommended Dietary Allowance (RDA):** average daily nutrient intake level sufficient to meet the nutrient requirement of nearly all (97% to 98%) healthy individuals in a particular life stage and gender group;
- **Adequate Intake:** recommended average daily intake level based on observed or experimentally determined approximations or estimates of nutrient intake by a group (or groups) of apparently healthy people that are assumed to be adequate - used when an RDA cannot be determined;
- **Estimated Average Requirement:** average daily nutrient intake level estimated to meet the requirement of half the healthy individuals in a particular age, life stage and gender group; and,
- **Tolerable Upper Intake Level (UL):** highest average daily nutrient intake level that is likely to pose no risk of adverse health effects to almost all individuals in the general population - as intake increases above the UL, potential risk of adverse effects may increase. The ULs are not specific data points from any particular dose-response relationship but are derived using well-established principles of risk assessment (WHO, 2002a). Various data sources, such as epidemiological studies, clinical trials, and experimental studies, can be used to estimate ULs (WHO, 1996, 2002a, IOM, 2000, 2001, 2006). Most ULs are derived from **no observable adverse effect levels** (NOAELs) and/or **lowest observable adverse effect levels** (LOAELs) (IOM, 2000, 2001, 2006). Uncertainty factors (UFs) are applied to NOAELs or LOAELs in the calculation of ULs (WHO, 2002a). However, these UFs tend to be lower than those traditionally used to establish TDIs or RfDs, while fully protecting human health (Mertz, 1995). The UFs used to establish ULs are generally less than 10, owing to the quality of available human data (Becking, 1998; Munro, 1999; Dourson et al., 2001). The ULs consider risks from nutrient deficiencies and toxicity, as well as inter-individual variability (WHO, 2002a).

2.3 RELATIVE POTENCY FACTORS/TOXIC EQUIVALENCY FACTORS

Some substances such as polycyclic aromatic hydrocarbons (PAHs), polychlorinated dibenzo-*p*-dioxins (PCDDs), polychlorinated dibenzofurans (PCDFs) and dioxin-like polychlorinated biphenyls (PCBs) are typically present as complex mixtures in the environment; however, for many individual PAHs and PCDDs/PCDFs/dioxin-like PCBs, toxicological data are insufficient to establish TRVs.

Mixtures of carcinogenic PAHs are assessed using relative potency factors (RPFs), also referred to as potency equivalence factors. An RPF is the ratio of carcinogenic potential of an individual PAH relative to benzo[a]pyrene (B[a]P). For a given mixture, the concentration of each carcinogenic PAH is multiplied by its RPF, and the resulting concentrations are summed to estimate a B[a]P equivalent concentration.

The RPFs for carcinogenic PAHs provided in [Table 2](#) are those recommended by the Canadian Council of Ministers of the Environment (CCME, 2010). [Table 3](#) presents provisional RPFs for a number of additional PAHs that may be present at Canadian federal contaminated sites, but for which there are limited relative potency data and RPFs have therefore not been formally derived by the CCME. The provisional RPFs were based on an analysis of RPFs published by other regulatory agencies and in the scientific literature (Equilibrium Environmental Inc. [EEI], 2006): Risk assessors may use RPFs based on more recent data and/or authoritative reviews, provided their use is accompanied by an appropriate supporting rationale

Exposures to mixtures of PCDDs/PCDFs and dioxin-like polychlorinated biphenyls (PCBs) are assessed using the WHO's toxic equivalency factors (TEFs) (DeVito et al., 2024). For a given mixture, the concentration of each PCDD, PCDF and PCB is multiplied by its respective TEF, and the resulting concentrations are summed to estimate a 2,3,7,8-tetrachlorodibenzo-p-dioxin (TCDD) toxic equivalent (TEQ) concentration. TEFs for PCDDs, PCDFs and certain carcinogenic PCBs are provided in [Table 4](#).

3.0 RELATIVE ABSORPTION FACTORS FOR DERMAL EXPOSURE

The degree of absorption of a substance into the systemic circulation depends on the route of exposure (oral, inhalation or dermal), the medium of exposure (e.g., soil, drinking water, food), as well as other factors, such as the physico-chemical properties of the substance, duration and frequency of exposure.

Ideally, health risks from an environmental exposure would be evaluated using a TRV derived from a study using the same route of exposure and the same medium of exposure. When this is not possible, relative absorption factors (RAFs) may be used to account for differences in absorption under environmental exposure conditions vs. conditions in the TRV study. As dermal TRVs are rarely available, dermal exposure associated with a contaminant is typically assessed in relation to an oral TRV, by incorporating a relative dermal absorption factor (RAF_{Derm}).

An RAF_{Derm} is calculated as follows:

$$RAF_{Derm} = \frac{\text{fraction of chemical absorbed through the skin from environmental medium}}{\text{fraction of chemical absorbed in principal oral TRV study}}$$

The denominator represents the chemical absorption efficiency in the principal study used to derive the oral TRV. For example, if dermal absorption is 10% and oral absorption in the principal TRV study is 100%, the RAF_{Derm} would be $10\% \div 100\% = 10\%$. Similarly, if oral absorption in the principal TRV study is only 50%, then the RAF_{Derm} would be $10\% \div 50\% = 20\%$. As such, an RAF_{Derm} of 1 (i.e., 100%) does not indicate that absorption is complete; rather, absorption from environmental exposure is considered equivalent to the absorption observed in the principal study upon which the TRV is based.

Recommended RAF_{Derm} values are provided in [Table 5](#). Unless otherwise indicated, these values were obtained from the Ontario Ministry of the Environment (MOE, 2011; currently known as the Ministry of the Environment, Conservation and Parks [MECP]). For substances not listed in [Table 5](#), RAF_{Derm} may be obtained from the sources listed at the beginning of Section 1, as well as from the Risk Assessment Information System (RAIS; rais.ornl.gov) or other recognized sources. Where alternate data sources are used, rationale with references should be provided in the report.

Dermal absorption of contaminants from contact with water during activities such as swimming, bathing, and showering can be estimated by employing dermal permeability constants (P_{Derm} , available from US EPA, 2004) and using methods described by the US EPA (1992, 2007a). HC uses a 'multiroute assessment approach' to determine the relative contribution of inhalation and dermal exposure associated with bathing and showering in relation to the total dose from exposure to a contaminant in drinking water (Krishnan and Carrier, 2008).

4.0 SUMMARY TABLES

The following tables provide a summary of the following values recommended by Health Canada:

- [Table 1](#): TRVs Recommended for Use in HHRA of federal contaminated sites, with a summary of the basis of each;
- [Table 2](#): Recommended RPFs for carcinogenic PAHs;
- [Table 3](#): Provisional RPFs for carcinogenic PAHs;
- [Table 4](#): TEFs for PCDDs, PCDFs, and dioxin-like PCBs; and,
- [Table 5](#): Recommended dermal RAFs.

It is recommended that selected TRVs and associated key health effects be described and summarised in the risk assessment report, with a discussion of threshold and non-threshold effects by exposure route (i.e., oral, dermal, inhalation), as appropriate.

TABLE 1: TRVs Recommended for Use in Human Health Risk Assessments of Federal Contaminated Sites

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|---------------------|----------------------|---|--|--|---|--------------------------------|--|--|
| Arsenic (inorganic) | Oral SF ¹ | 1.8E+00 (mg/kg _{BW} -day) ⁻¹ | <p>Study Type: epidemiological</p> <p>Species: humans</p> <p>Mode of Exposure: oral (drinking water)</p> <p>Exposure Concentrations: concentration of arsenic in drinking water varied from less than 10 to greater than 600 µg/L (groundwater arsenic concentrations)</p> <p>Duration: chronic</p> <p>Uncertainty Factors: N/A</p> | <p>Range of unit risks associated with ingesting 1 µg/L of arsenic in drinking water:</p> <p>3.06E-06 to 3.85E-05 (µg/L)⁻¹ (based on a 1% increase in risk)</p> | <p>TRV based on upper end of range of unit risks (URs) in drinking water: 3.85E-05 (µg/L)⁻¹</p> <p>Conversion to oral SF in (mg/kg_{BW}-day)⁻¹: Oral SF = UR × BW_{adult} × CF/IR_w [where BW_{adult} = 70.7 kg, IR_w = 1.5 L/day, and CF (conversion factor) = 1000 µg/mg]</p> | Cancer (bladder, lung, liver) | <p>CEPA: Group I carcinogenic to humans (EC and HC, 1993a)</p> <p>IARC: Group 1 carcinogenic to humans (IARC, 2012a)</p> | HC, 2006 (based on Morales et al., 2000; Chen et al., 1985; Wu et al., 1989) |
| | Inhalation UR | 6.4E+00 (mg/m ³) ⁻¹ | <p>Study Type: epidemiological (occupational)</p> <p>Species: humans</p> <p>Mode of Exposure: inhalation</p> <p>Exposure Concentrations: N/A</p> <p>Duration: chronic</p> <p>Uncertainty Factors: N/A</p> | <p>TC₀₅ (5% tumourigenic concentration) = 7.83 µg/m³</p> | <p>Relative risk model</p> <p>Inhalation UR = 0.05/TC₀₅ where 0.05 = 5% extra cancer risk</p> | Cancer (lung) | US EPA IRIS: Group A carcinogenic to humans (US EPA, 1995a) | EC and HC, 1993a (based on Higgins et al., 1986) |
| Barium | Oral TDI | 1.9E-01 mg/kg _{BW} -day | <p>Study Type: chronic</p> <p>Species: male and female B6C3F1 mice</p> <p>Mode of Administration: oral (drinking water)</p> <p>Exposure Regime: 0, 500, 1250, and 2500 ppm barium chloride dihydrate in drinking water (daily doses estimated to be 0, 30, 75, and 160 mg barium/kg_{BW}-day for males, and 0, 40, 90, and 200 mg barium/kg_{BW}-day for females)</p> <p>Duration: 2 years</p> <p>Uncertainty Factors: 300 (10 for intraspecies variability, 10 for interspecies variability, and 3 for database deficiencies)</p> | BMDL ₀₅ = 58 mg/kg _{BW} -day | TDI = BMDL ₀₅ /UF | Toxic nephropathy in male mice | <p>CEPA: Group VA inadequate data for evaluation (HC, 1990)</p> <p>IARC: not classified</p> <p>US EPA IRIS: inhalation route - carcinogenic potential cannot be determined; oral route - not likely to be carcinogenic to humans (US EPA, 1998a)</p> | HC, 2020 (based on NTP, 1994) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|-----------|---------------|---|---|---|---|--|---|---|
| Benzene | Oral SF | 8.3E-02 (mg/kg _{BW} -day) ⁻¹ | Study Type: chronic Species: rats and mice Mode of Administration: gavage, corn oil Exposure Regime: 0, 50, 100, and 200 mg/kg _{BW} -day (male rats); 0, 25, 50, and 100 mg/kg _{BW} -day (female rats, male and female mice), 5 days/week Duration: 103 weeks Uncertainty Factors: N/A | Range of unit risks associated with ingesting 1 µg/L of benzene in water: 2.03E-06 to 4.17E-06 (µg/L) ⁻¹ | Linearized multistage model and allometric scaling TRV based on upper bound estimate of unit risks (URs) in drinking water: 4.17E-06 (µg/L) ⁻¹ Conversion to oral SF in (mg/kg _{BW} -day) ⁻¹ : Oral SF = UR × BW _{adult} × CF/IR _W [where BW _{adult} = 70.0 kg, IR _W = 3.5 L eq/day, and CF = 1000 µg/mg] | Cancer (malignant lymphomas) and Bone marrow hematopoietic hyperplasia | CEPA: Group I carcinogenic to humans (EC and HC, 1993b) IARC: Group 1 carcinogenic to humans (IARC, 2012b) | HC, 2009 (based on NTP, 1986a) |
| | Inhalation UR | 1.6E-02 (mg/m ³) ⁻¹ | Study Type: epidemiological (occupational) Species: human Mode of Exposure: inhalation Exposure Concentrations: N/A Duration: chronic Uncertainty Factors: N/A | Unit lifetime leukemia risk to the general population, derived from these studies: Ohio Pliofilm cohort: 0.044 (ppm) ⁻¹ [0.014 (mg/m ³) ⁻¹] Chinese cohorts: 0.056 (ppm) ⁻¹ [0.018 (mg/m ³) ⁻¹] | Poisson regression and linear relative risk models Inhalation UR for lifetime inhalation exposures of the general population (based on the geometric mean of upper bound estimates of leukemia risk from these studies) Inhalation UR = [0.044 (ppm) ⁻¹ × 0.056 (ppm) ⁻¹] ^{1/2} = 0.050 (ppm) ⁻¹ [0.016 (mg/m ³) ⁻¹] | Cancer (leukemia) | US EPA IRIS: Group A carcinogenic to humans (US EPA, 2000a) | HC, 2013a and OEHHA, 2001 (based on Rinsky et al., 1987; Paxton et al., 1994; Hayes et al., 1997) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|----------------------|--|---|--|--|---|---|--|---|
| Benzo[a]pyrene (BaP) | Oral TDI | 3.0E-04 mg/kg _{bw} -day | <p>Study Type: developmental</p> <p>Species: neonate Sprague-Dawley rat pups (10 males and 10 females)</p> <p>Mode of Administration: gavage</p> <p>Exposure Regime: 0 (peanut oil only), 0.02, 0.2, or 2 mg/kg_{bw} administered daily from postnatal day (PND) 5 until PND 11</p> <p>Duration: until PND 71</p> <p>Uncertainty Factors: 300 (10 for intraspecies variability, 10 for interspecies variability, and 3 for database deficiencies)</p> | BMDL _{1SD} = 0.0921 mg/kg _{bw} -day | <p>BMDL_{1SD} derived from exponential polynomial model corresponding to an extra risk of 10%</p> <p>TDI = BMDL_{1SD}/UF (TDI rounded to 3.0E-04 mg/kg_{bw}-day)</p> | Neuro-developmental toxicity | | US EPA, 2017 (based on Chen et al., 2012) |
| | Inhalation TC | 2.0E-06 mg/m ³ | <p>Study Type: developmental</p> <p>Species: F344 rats (pregnant females)</p> <p>Mode of Administration: inhalation (nose only)</p> <p>Exposure Regime: 0 ("sham" carbon black or unexposed), 25, 75, and 100 µg/m³, 4 hours per day for 10 days (gestation days 11 to 20)</p> <p>Duration: 10 days (gestation days 11–20)</p> <p>Uncertainty Factors: 3000 (3 for toxicodynamic differences, 10 for intraspecies variability, 10 for LOAEL to NOAEL extrapolation, and 10 for database deficiencies)</p> | LOAEL = 0.025 mg/m ³ | <p>LOAEL adjusted for continuous daily exposure and converted to a human equivalent concentration based on a regional deposited dose ratio for extra respiratory effects</p> <p>LOAEL_{HEC} = 0.0046 mg/m³</p> <p>RfC = LOAEL_{HEC}/UF</p> | Developmental toxicity (decreased embryo/foetal survival) | <p>CEPA: Group II probably carcinogenic to humans (EC and HC, 1994a)</p> <p>IARC: Group 1 carcinogenic to humans (IARC, 2012b)</p> <p>US EPA carcinogenic to humans (US EPA, 2017)</p> | US EPA, 2017 (based on Archibong et al., 2002) |
| | Oral SF (As per HC [2013b, 2024a], given that BaP is known to have a mutagenic mode of action, HC recommends applying ADAFs ² to the oral SF when assessing risk associated with early life exposures at federal contaminated sites) | 1.289E+00 (mg/kg _{bw} -day) ⁻¹ | <p>Study Type: chronic</p> <p>Species: B6C3F1 female mice</p> <p>Mode of Administration: diet</p> <p>Exposure Regime: 0, 5, 25, and 100 ppm (corresponding to approximately 0, 0.7, 3.3, and 13.0 mg/kg_{bw}-day, as per HC [2016])</p> <p>Duration: 2 years</p> <p>Uncertainty Factors: N/A</p> | BMDL ₁₀ = 0.5389 mg/kg _{bw} -day | <p>Allometric scaling of the BMDL₁₀ (to account for interspecies variability and derive a human equivalent value)</p> <p>BMDL₁₀, HEC = 0.07758 mg/kg_{bw}-day</p> <p>Oral SF = 0.1/BMDL₁₀, HEC where 0.1 = 10% extra cancer risk</p> | Digestive tract toxicity (tumours of the forestomach) | | HC, 2016 (based on Culp et al., 1998 and Moffat et al., 2015) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|----------------------|---------------|---|---|--|--|--|---|---|
| Benzo[a]pyrene (BaP) | Inhalation UR | 6.0E-01 (mg/m ³) ⁻¹ | <p>Study Type: chronic</p> <p>Species: Syrian golden male hamsters</p> <p>Mode of Administration: inhalation (nose only) of benzo[a]pyrene condensed onto sodium chloride aerosols</p> <p>Exposure Regime: 2.2, 9.5, and 46.5 mg BaP/m³ (time-weighted average concentrations of 0, 0.25, 1.01, and 4.29 mg/m³, corresponding to 0, 2, 10, and 50 mg/m³ nominal study concentrations), 4.5 hours/day for the first 10 weeks, then 3 hours/day for the remainder of the study</p> <p>Duration: minimum of 10 weeks up to 130 weeks</p> <p>Uncertainty Factors: N/A</p> | BMDL ₁₀ = 0.16 mg/m ³ | <p>Multistage Weibull time-to-tumour dose-response model + linear extrapolation from the POD associated with 10% extra cancer risk</p> <p>Inhalation UR = 0.1/BMDL₁₀ where 0.1 is 10% extra cancer risk</p> | Cancer (tumours of the upper gastrointestinal tract and upper respiratory tract [squamous cell neoplasia in the larynx, pharynx, trachea, nasal cavity, esophagus, and forestomach]) | <p>CEPA: Group II probably carcinogenic to humans (EC and HC, 1994a)</p> <p>IARC: Group 1 carcinogenic to humans (IARC, 2012b)</p> <p>US EPA carcinogenic to humans (US EPA, 2017)</p> | US EPA, 2017 (based on Thyssen et al., 1981) |
| Beryllium | Oral TDI | 2.0E-03 mg/kg _{BW} -day | <p>Study Type: chronic</p> <p>Species: dogs (5 male and 5 female beagles)</p> <p>Mode of Administration: diet</p> <p>Exposure Regime: 0, 1, 5, 50, or 500 ppm beryllium as beryllium sulfate tetrahydrate (diets fed for 1 hour per day), corresponding to doses of 0.023, 0.12, 1.1, and 12.2 mg/kg_{BW}-day for male dogs and 0.029, 0.15, 1.3, and 17.4 mg/kg_{BW}-day for female dogs (using estimated time-weighted average body weights and a reported average food intake of 300 g/day)</p> <p>Duration: 0, 5, and 50 ppm group exposed for 172 weeks; 500 ppm dose group terminated at 33 weeks because of overt signs of toxicity; 1 ppm group exposed for 143 weeks.</p> <p>Uncertainty Factors: 300 (10 for intraspecies variability, 10 for interspecies variability, and 3 for database deficiencies)</p> | BMDL ₁₀ = 4.6E-01 mg/kg _{BW} -day | <p>BMDL₁₀ derived from exponential polynomial model corresponding to an extra risk of 10%</p> <p>TDI = BMDL₁₀/UF (TDI rounded to 2.0E-03 mg/kg_{BW}-day)</p> | Gastrointestinal toxicity (lesions of the small intestine) | <p>CEPA: see 2020 Chemicals Management Plan (CMP) assessment (ECCC and HC, 2020)</p> <p>IARC: Group 1 carcinogenic to humans (IARC, 2012a)</p> <p>US EPA IRIS: oral route – carcinogenic potential cannot be determined; inhalation route – known/likely human carcinogen (US EPA, 1998b)</p> | US EPA, 1998b (based on Morgareidge et al., 1976) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|-----------|---------------|--|--|---|--|--|---|--|
| Beryllium | Inhalation TC | 2.0E-05 mg/m ³ | <p>Study Type epidemiological (occupational study)</p> <p>Species: human</p> <p>Mode of Exposure: inhalation</p> <p>Exposure Concentrations: individual estimated average exposures for six chronic beryllium disease cases and two sensitized cases ranged from 0.2 to 1.1 µg/m³, and the median of estimated average beryllium exposure for these eight workers was 0.55 µg/m³. Cumulative exposure ranged from 92.6 to 1945 µg/m³-day.</p> <p>Duration: chronic</p> <p>Uncertainty Factors: 10 (3 for the sensitive nature of the subclinical endpoint [beryllium sensitization], and 3 for poor quality of exposure monitoring) [total uncertainty factor rounded to 10]</p> | LOAEL = 0.55 µg/m ³ | <p>LOAEL adjusted for occupational inhalation rate and for an intermittent working week schedule</p> <p>LOAEL_{HEC} = 0.20 µg/m³</p> <p>TC = LOAEL_{HEC}/UF</p> | Immunotoxicity and Respiratory toxicity (beryllium sensitization and progression to chronic beryllium disease [chronic inflammatory lung lesions]) | <p>CEPA: see 2020 Chemicals Management Plan (CMP) assessment (ECCC and HC, 2020)</p> <p>IARC: Group 1 carcinogenic to humans (IARC, 2012a)</p> <p>US EPA IRIS: oral route – carcinogenic potential cannot be determined; inhalation route – known/likely human carcinogen (US EPA, 1998b)</p> | US EPA, 1998b (based on Kreiss et al., 1996) |
| | Inhalation UR | 2.4E+00 (mg/m ³) ⁻¹ | <p>Study Type: epidemiological (occupational)</p> <p>Species: humans (male)</p> <p>Mode of Exposure: inhalation</p> <p>Exposure Concentrations: range of median exposure levels inside plants (100–1000 µg/m³) estimated in NIOSH's industrial hygiene reviews</p> <p>Duration: The cohort employed between 1942 and 1967 was followed through 1975. The subcohort upon which the inhalation UR is based was followed for at least 25 years.</p> <p>Uncertainty Factors: N/A</p> | Range of upper bound unit risks: 1.6E-04 (µg/m ³) ⁻¹ to 7.2E-03 (µg/m ³) ⁻¹ | <p>Linear relative risk model</p> <p>Geometric mean of upper bound unit risks = 2.4E-03 (µg/m³)⁻¹</p> | Cancer (lung) | US EPA, 1998b (based on Wagoner et al., 1980 and NIOSH, 1972) | |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|-----------|---------------------------|---|--|---|---|--|--|---|
| Cadmium | Oral TDI (provisional) | 8.0E-04 mg/kg _{BW} -day | <p>Study Type: epidemiological (meta-analysis)</p> <p>Species: humans</p> <p>Mode of Exposure: environmental (primarily through food)</p> <p>Exposure Concentrations: N/A</p> <p>Duration: chronic</p> <p>Uncertainty Factors: toxicodynamic and toxicokinetic variability incorporated (using Monte Carlo simulation) into the toxicokinetic model relating cadmium concentration in urine to dietary intake</p> | <p>NOAEL = 5.24 µg Cd/g creatinine in urine (corresponds to a dietary cadmium exposure of 1.2 µg Cd/kg_{BW}-day)</p> <p>[5th-95th percentiles: 0.8–1.8 µg Cd/kg_{BW}-day]</p> | <p>Lower bound of 0.8 µg/kg_{BW}-day retained as oral TDI to account for particularly susceptible individuals</p> <p>This oral TDI is reported by WHO (2011) as a provisional monthly tolerable intake of 25 µg/kg_{BW}</p> | Nephrotoxicity (renal tubular dysfunction) | <p>CEPA: probably carcinogenic to humans (inhalation pathway) (EC and HC, 1994b)</p> <p>IARC: Group 1 carcinogenic to humans (IARC, 2012a)</p> | WHO, 2011 |
| | Inhalation UR | 4.2E+00 (mg/m ³) ⁻¹ | <p>Study Type: epidemiological (occupational)</p> <p>Species: humans</p> <p>Mode of Exposure: inhalation of dusts of cadmium oxide and cadmium sulfide, and cadmium fumes</p> <p>Exposure Concentrations: Equivalent lifetime exposure in µg/m³ = 2, 11.8, and 41 µg Cd/m³ (based on 24 hour/day exposure and an estimated average lifetime of 61.5 years)</p> <p>Duration: at least two years</p> <p>Uncertainty Factors: N/A</p> | <p>Range of excess cancer risk for the exposed population: 2.0E-03 (µg/m³)⁻¹ to 1.2E-02 (µg/m³)⁻¹</p> | <p>Poisson regression model fitted to occupational mortality data + extrapolation to ambient levels in California</p> | Cancer (lung) | <p>US EPA IRIS: Group B1 probably carcinogenic to humans (US EPA, 1987a)</p> | OEHHA, 2011 (based on Thun et al., 1985, 1986 and CDHS, 1986, 1990) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|----------------------|---------------|---|---|--|---|--|--|---|
| Carbon tetrachloride | Oral TDI | 7.1E-04 mg/kg _{BW} -day | <p>Study Type: subchronic</p> <p>Species: male Sprague-Dawley rats</p> <p>Mode of Administration: gavage (corn oil)</p> <p>Exposure Regime: 0, 1, 10, or 33 mg/kg_{BW}-day, administered as a single oral bolus, 5 days/week</p> <p>Duration: 12 weeks</p> <p>Uncertainty Factors: 1000 (10 for intraspecies variability, 10 for interspecies variability, and 10 for major database deficiencies including lack of adequate chronic studies and evidence regarding carcinogenic mode of action in animals)</p> | NOAEL = 1 mg/kg _{BW} -day | <p>NOAEL adjusted for weekly continuous exposure</p> <p>NOAEL_{adj} = 0.71 mg/kg_{BW}-day</p> <p>TDI = NOAEL_{adj}/UF</p> | Hepatotoxicity (increased serum sorbitol dehydrogenase levels and mild centrilobular vacuolization in the liver) | <p>CEPA: not assessed</p> <p>IARC: Group 2B possibly carcinogenic to humans (IARC, 1999a)</p> <p>US EPA IRIS: likely to be carcinogenic to humans (US EPA, 2010)</p> | HC, 2010b (based on Bruckner et al., 1986) |
| | Inhalation UR | 6.0E-03 (mg/m ³) ⁻¹ [US EPA (2010): do not use with exposures >18 mg/m ³] | <p>Study Type: chronic</p> <p>Species: male BDF1 mice</p> <p>Mode of Administration: inhalation</p> <p>Dosing Regime: 0, 5, 25, or 125 ppm carbon tetrachloride vapour (0, 31, 157, or 786 mg/m³), 6 hours/day, 5 days/week</p> <p>Duration: 104 weeks</p> <p>Uncertainty Factors: N/A</p> | LEC ₁₀ = 18 mg/m ³ (lowest effective concentration) | <p>Internal mouse doses determined using PBPK model + BMD modelling using a log-probit model and an extra 10% cancer risk + Linear extrapolation from the POD (LEC₁₀) converted to a human equivalent concentration using a human PBPK model</p> <p>Inhalation UR = 0.1/LEC₁₀ [UR rounded to 6.0E-03 (mg/m³)⁻¹]</p> | Cancer (adrenal gland [pheochromocytomas]) | <p>US EPA IRIS: likely to be carcinogenic to humans (US EPA, 2010)</p> | HC, 2024b (based on US EPA, 2010 [derived from Nagano et al., 2007 and JBRC, 1998]) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|---------------|-----------------------------|-------------------------------------|---|---|--|---|---|---|
| Chlorobenzene | Oral TDI | 4.3E-01 mg/kg _{BW} -day | <p>Study Type: chronic</p> <p>Species: F344/N rats and B6C3F1 mice</p> <p>Mode of Administration: gavage, corn oil</p> <p>Exposure Regime: 0, 60, or 120 mg/kg_{BW}-day (male and female rats, and female mice); 0, 30, or 60 mg/kg_{BW}-day (male mice), 5 days/week, as monochlorobenzene</p> <p>Duration: 103 weeks</p> <p>Uncertainty Factors: 100 (10 for intraspecies variability and 10 for interspecies variability)</p> | NOAEL = 60 mg/kg _{BW} -day | <p>NOAEL adjusted for continuous exposure</p> <p>NOAEL_{adj} = 43 mg/kg_{BW}-day</p> <p>TDI = NOAEL_{adj}/UF</p> | Hepatotoxicity (neoplastic nodules in the liver) | CEPA: Group III possibly carcinogenic to humans (EC and HC, 1992a) | HC, 1996 (based on NTP, 1985a and Kluwe et al., 1985) |
| | Inhalation TC (provisional) | 1.0E-02 mg/m ³ | <p>Study Type: subchronic</p> <p>Species: Sprague-Dawley male rats, and male rabbits</p> <p>Mode of Administration: inhalation (whole body exposure chambers)</p> <p>Exposure Regime: 0, 75, or 250 ppm (0, 341, or 1138 mg/m³) chlorobenzene vapours, 7 hours/day, 5 days per week, for up to 120 exposure days</p> <p>Duration: 24 weeks</p> <p>Uncertainty Factors: 5000 (10 for intraspecies variability, 10 for interspecies variability, 10 for a less than chronic study, and 5 for use of a LOAEL rather than NOAEL)</p> | LOAEL = 341 mg/m ³ | <p>LOAEL adjusted for continuous exposure and differences in volume inhaled and body weight between rats and the human child</p> <p>LOAEL_{HEC} = 50.2 mg/m³</p> <p>TC = LOAEL_{HEC}/UF</p> | Nephrotoxicity (increased kidney weight, kidney lesions) and Endocrine system toxicity (lesions in the adrenal cortex) and Changes in red cell parameters | IARC: not classified US EPA IRIS: Group D not classifiable as to human carcinogenicity (US EPA, 1990a) | HC, 1996 and EC and HC, 1992a (based on Dilley, 1977) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|----------------------------------|---------------------------|-------------------------------------|--|---|--------------------------|---------------------------------------|---|---------------------------------|
| Chromium, trivalent ³ | Oral TDI (provisional) | 3.0E-01 mg/kg _{BW} -day | <p>Study Type: chronic</p> <p>Species: male and female F344/N rats and B6C3F1 mice</p> <p>Mode of Administration: diet</p> <p>Exposure Regime: chromium picolinate monohydrate with feed [0, 2000 mg/kg diet (10.7 mg Cr(III)/kg_{BW}-day), 10,000 mg/kg diet (55 mg Cr(III)/kg_{BW}-day), and 50,000 mg/kg diet (286 mg Cr(III)/kg_{BW}-day)]</p> <p>Duration: 105 weeks</p> <p>Uncertainty Factors: 1000 (10 for intraspecies variability and 10 for interspecies variability, and 10 for absence of adequate data on reproductive and developmental toxicity)</p> | NOAEL = 286 mg Cr(III)/kg bw-day | TDI = NOAEL Cr(III)/(UF) | No effects observed at any dose level | <p>CEPA: Group VI unclassifiable with respect to carcinogenicity to humans (EC and HC, 1994c)</p> <p>IARC: Group 3 not classifiable with respect to carcinogenicity to humans (IARC, 1990)</p> <p>US EPA IRIS: Group D not classifiable with respect to human carcinogenicity (US EPA, 1998c)</p> | EFSA, 2014 (based on NTP, 2010) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|----------------------|---------------|--|---|---|---|---|--|--|
| Chromium, hexavalent | Oral TDI | 2.2E-03 mg/kg _{BW} -day | <p>Study Type: chronic</p> <p>Species: male and female B6C3F1 mice</p> <p>Mode of Administration: oral (drinking water)</p> <p>Exposure Regime: Male mice received 0, 14.3, 28.6, 85.7, or 257.4 mg sodium dichromate dehydrate (SSD)/L (equivalent to 0, 5, 10, 30, and 90 mg Cr (VI)/L or 0, 0.4, 0.9, 2.4, and 5.9 mg Cr (VI)/kg_{BW}-day respectively). Female mice received 0, 14.3, 57.3, 172, or 516 mg sodium dichromate dihydrate/L (equivalent to 0, 5, 20, 60, and 180 mg Cr (VI)/L or 0, 0.4, 1.4, 3.1, and 8.7 mg Cr (VI)/kg_{BW}-day).</p> <p>Duration: 2 years</p> <p>Uncertainty Factors: 25 (10 for intraspecies variability and 2.5 for pharmacodynamic interspecies differences)</p> | BMDL ₀₁ = 0.67 mg Cr(VI)/kg _{BW} -day | <p>PBPK model used to convert mouse BMDL₀₁ into a human equivalent dose of 0.054 mg Cr (VI)/kg_{BW}-day</p> <p>TDI = human equivalent dose/UF</p> | Gastrointestinal toxicity (diffuse epithelial hyperplasia of the small intestine) | <p>CEPA: Group I carcinogenic to humans (EC and HC, 1994c)</p> <p>IARC: Group 1 carcinogenic to humans (IARC, 2012a)</p> | <p>HC, 2018 (based on NTP, 2008; Stout et al., 2009; Thompson et al., 2014; Summit Toxicology, 2014)</p> |
| | Inhalation TC | 1.0E-04 mg/m ³ | <p>Study Type: subchronic</p> <p>Species: male Wistar rats</p> <p>Mode of Administration: inhalation (whole body exposure chambers)</p> <p>Exposure Regime: sodium dichromate (chromium particulates), 0.05–0.4 mg Cr (VI)/m³ for 22 hours/day, 7 days/week</p> <p>Duration: 30 to 90 days</p> <p>Uncertainty Factors: 300 (10 for use of a subchronic study, 10 for intraspecies variability, and 3 for pharmacodynamic interspecies differences)</p> | BMC = 0.016 mg Cr (VI)/m ³ | <p>BMC₁₀ selected as POD and adjusted for continuous exposure + animal to human conversion based on regional deposited dose ratio (RDDR) of 2.1576 for particulates</p> <p>TC for Cr (VI) = BMC x RDDR /UF</p> | Respiratory tract toxicity (increased albumin and lactate dehydrogenase in bronchioalveolar lavage fluid, which may be indicative of initial injury and chronic inflammation) | <p>US EPA IRIS: Group A inhalation route: carcinogenic to humans (US EPA, 1998d); Group D oral route: not classifiable as to human carcinogenicity (US EPA, 1998d)</p> | <p>US EPA, 1998d (based on Glaser et al., 1990 and Malsch et al., 1994)</p> |
| | Inhalation UR | 7.6E+01 (mg/m ³) ⁻¹ | <p>Study Type: epidemiological (occupational)</p> <p>Species: humans (adult men)</p> <p>Mode of Exposure: inhalation</p> <p>Exposure Concentrations: N/A</p> <p>Duration: at least 1 year, up to 8 years</p> <p>Uncertainty Factors: N/A</p> | TC ₀₅ (5% tumourigenic concentration) = 0.66 µg/m ³ | Inhalation UR = 0.05/TC ₀₅ where 0.05 = 5% extra cancer risk | Cancer (lung) | | <p>HC, 1996 (based on Mancuso, 1975)</p> |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|-----------------------|-------------|--------------------------------------|---|---|--|---|--|--|
| Copper | Oral TDI | 4.26E-01 mg/kg _{BW} -day | <p>Study Type: epidemiological (prospective)</p> <p>Species: humans</p> <p>Mode of Exposure: oral (drinking water)</p> <p>Exposure Concentrations: healthy infants 3 to 12 months of age (n = 128) were given drinking water with < 0.1 mg copper/L (n = 48) or 2 mg copper/L (n = 80) (copper sulfate solution added to drinking water ingested by formula-fed or cow's milk-fed infants, or ingested by mothers for breast-fed infants, and/or added to water used for meal preparation)</p> <p>Duration: nine months</p> <p>Uncertainty Factors: none (considered conservative since no symptoms of copper toxicity were observed at this dose over the entire duration of the study)</p> | NOAEL = 2 mg/L (corresponding to an estimated mean \pm SD daily intake of 0.318 \pm 0.107 mg/kg _{BW} -day) | TDI = NOAEL + SD = 0.318 + 0.107 = 0.426 mg/kg _{BW} -day | Gastrointestinal toxicity and Hepatotoxicity (liver function) | <p>CEPA: see 2019 (draft) CMP assessment (ECCC and HC, 2019a)</p> <p>IARC: not classified</p> <p>US EPA IRIS: Group D not classifiable as to human carcinogenicity (US EPA, 1988a)</p> | HC, 2019a (based on Olivares et al., 1998) |
| Dichlorobenzene, 1,2- | Oral TDI | 4.3E-01 mg/kg _{BW} -day | <p>Study Type: chronic</p> <p>Species: rats and mice</p> <p>Mode of Administration: gavage, corn oil</p> <p>Exposure Regime: 0, 60, or 120 mg/kg_{BW}-day (male and female rats, female mice), 30 and 60 mg/kg_{BW}-day (male mice), 5 days per week</p> <p>Duration: 103 weeks</p> <p>Uncertainty Factors: 100 (10 for intraspecies variability and 10 for interspecies variability)</p> | NOAEL = 60 mg/kg _{BW} -day | <p>NOAEL adjusted for continuous exposure</p> <p>NOAEL_{adj} = 43 mg/kg_{BW}-day</p> <p>TDI = NOAEL_{adj}/UF</p> | Nephrotoxicity (increase in tubular regeneration in the kidney) | <p>CEPA: Group V probably not carcinogenic to humans (EC and HC, 1993c)</p> <p>IARC: Group 3 not classifiable as to its carcinogenicity to humans (IARC, 1999b)</p> <p>US EPA IRIS: Group D not classifiable as to human carcinogenicity (US EPA, 1990b)</p> | HC, 1996 (based on NTP, 1985b) |

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|----------------------------------|---------------|--|--|--|--|--|--|--|
| Dichlorobenzene, 1,4- | Oral TDI | 1.1E-01 mg/kg _{BW} -day | <p>Study Type: chronic</p> <p>Species: rats and mice</p> <p>Mode of Administration: gavage, corn oil</p> <p>Exposure Regime: male rats: 0, 150, or 300 mg/kg_{BW}-day, 5 days/week; female rats and male and female mice: 0, 300, or 600 mg/kg_{BW}-day, 5 days/week</p> <p>Duration: 103 weeks</p> <p>Uncertainty Factors: 1000 (10 for intraspecies variability, 10 for interspecies variability, and 10 for use of LOAEL vs NOAEL)</p> | LOAEL = 150 mg/kg _{BW} -day | <p>LOAEL adjusted for continuous exposure</p> <p>LOAEL_{adj} = 107 mg/kg_{BW}-day</p> <p>TDI = LOAEL_{adj} /UF</p> | Nephrotoxicity (renal tubular degeneration and atrophy) | CEPA: Group III possibly carcinogenic to humans (EC and HC, 1993d) | HC, 1996 (based on NTP, 1987) |
| | Inhalation TC | 6.0E-02 mg/m ³ | <p>Study Type: chronic</p> <p>Species: male and female F344 rats and BDF1 mice</p> <p>Mode of Administration: inhalation (whole body exposure chambers)</p> <p>Exposure Regime: 1,4-dichlorobenzene vapour at concentrations of 0, 20, 75, or 300 ppm (equivalent to 0, 120, 451, and 1804 mg/m³) for 6 hours/day, 5 days/week</p> <p>Duration: 104 weeks</p> <p>Uncertainty Factors: 30 (3 for intraspecies variability and 10 for interspecies variability)</p> | BMCL ₁₀ = 9.51 ppm (57.2 mg/m ³) | <p>BMCL₁₀ adjusted for duration (1.70 ppm [10.2 mg/m³]) + Conversion to a BMCL_{10,HEC} using a regional gas deposition ratio</p> <p>BMCL_{10,HEC} = 0.27 ppm (1.6 mg/m³) TC = BMCL_{10,HEC}/UF</p> | Respiratory tract toxicity (nasal lesions [eosinophilic changes in the nasal olfactory epithelium]) | IARC: Group 2B possibly carcinogenic to humans (IARC, 1999b) US EPA IRIS: not assessed | HC, 2024b (based on ATSDR, 2006 [derived from Aiso et al., 2005 and JBRC, 1995]) |
| Dichloroethane, 1,2- (DCA, 1,2-) | Oral SF | 3.3E-03 (mg/kg _{BW} -day) ⁻¹ | <p>Study Type: chronic</p> <p>Species: male and female F344 rats and BDF1 mice</p> <p>Mode of Administration: inhalation (whole body exposure chambers)</p> <p>Exposure Regime: rats: 0, 10, 40, or 160 ppm 1,2-dichloroethane vapour (0, 202, 809 and 2024 mg/m³ or 0, 12, 50, 200 mg/kg_{BW}), 6 hours/day, 5 days/week; mice: 0, 10, 30, or 90 ppm 1,2-dichloroethane vapour (0, 40, 121, 364 mg/m³ or 0, 54, 162, 486 mg/kg_{BW}), 6 hours/day, 5 days/week</p> <p>Duration: 104 weeks</p> <p>Uncertainty Factors: N/A</p> | BMD of the 1,2-DCA concentration rat blood based on an excess lifetime risk of 10 ⁻⁵ = 0.00027 mg/L | <p>Rat PBPK model used to extrapolate between exposure routes (inhalation to oral) and to estimate the lifetime average daily concentration in rat blood + Multistage modeling to determine rat BMD corresponding to an excess lifetime risk of 10⁻⁵ + PBPK model to extrapolate from internal animal dose to external dose in humans (0.003 mg/kg_{BW}-day)</p> <p>Oral SF = 10–5/0.003 mg/kg_{BW}-day</p> | Cancer (combined mammary gland tumours [adenoma, fibroadenoma, and adenocarcinoma of the mammary gland]) | CEPA: Group II probably carcinogenic to humans (EC and HC, 1994d) IARC: Group 2B possibly carcinogenic to humans (IARC, 1999a) US EPA IRIS: Group B2 probably carcinogenic to humans (US EPA, 1987b) | HC, 2014a (based on Nagano et al., 2006) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|-----------------------|--|-------------------------------------|--|--|--|---|---|--|
| Dichloroethylene, 1,1 | Oral TDI (referred to as an 'acceptable daily intake' in HC, 1994) | 3.0E-03 mg/kg _{BW} -day | <p>Study Type: chronic</p> <p>Species: Sprague-Dawley rats</p> <p>Mode of Administration: oral (drinking water)</p> <p>Exposure Regime: time weighted average daily doses: 0, 7, 10, and 20 mg/kg_{BW}-day (males); 0, 9, 14, and 30 mg/kg_{BW}-day (females)</p> <p>Duration: 2 years</p> <p>Uncertainty Factors: 3000 (10 for intraspecies variability, 10 for interspecies variability, 10 for use of a LOAEL, and 3 for limited evidence of carcinogenicity)</p> | LOAEL = 9 mg/kg _{BW} -day | TDI = LOAEL/UF | Hepatotoxicity (hepatocellular swelling with mid-zonal fatty changes) | CEPA: not assessed IARC: Group 2B possibly carcinogenic to humans (IARC, 2019) | HC, 1994 (based on Quast et al., 1983) |
| | Inhalation UR | 1.7E-07 (mg/m ³)-1 | <p>Study Type: chronic</p> <p>Species: male B6C3F₁/N mice</p> <p>Mode of Exposure: inhalation (whole body exposure chambers)</p> <p>Exposure Concentrations: 0, 6.25, 12.5, or 25 ppm (approximately 0, 5.1, 10, 20 mg/kg-day); 6.17 hours/day, 5 days/week, for 104 weeks</p> <p>Duration: chronic</p> <p>Uncertainty Factors: N/A</p> | Based on the lower bound on the dose associated with a 5% tumour response in male mice, multistage model | <p>Concentrations adjusted for continuous exposure; average BW used to estimate mouse inhalation rate (IR) to give average daily dose</p> <p>BMD modelling using 5% tumour response, multistage model</p> <p>CSF determined by linear extrapolation and converted to human exposure using BW scaling (BW^{3/4})</p> <p>IUR calculated from CSF using human IR of 15.1 m³/day</p> <p>Inhalation UR = 10⁻⁵/0.06 µg/m³</p> | Cancer (kidney [renal tubule adenoma or carcinoma]) | US EPA IRIS: Group C possible human carcinogen (US EPA, 2002) | HC, 2024b (based on OEHHHA, 2017 [derived from NTP, 2015]) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|---|--|---|---|---|---|--|---|---|
| Dichloromethane (methylene chloride) | Oral TDI | 1.4E-02 mg/kg _{BW} -day | <p>Study Type: chronic</p> <p>Species: male and female F344 rats</p> <p>Mode of Administration: oral (drinking water)</p> <p>Exposure Regime: 0, 5, 50, 125, and 250 mg/kg_{BW}-day; additional group at 250 mg/kg_{BW}-day</p> <p>Duration: 104 weeks; additional group was exposed for 78 weeks + 26 week recovery period</p> <p>Uncertainty Factors: 300 (10 for intraspecies variability, 10 for interspecies variability, and 3 for database deficiencies)</p> | BMDL ₁₀ = 4.2 mg/kg _{BW} -day | TDI = BMDL ₁₀ /UF | Hepatotoxicity (increased incidences of foci and areas of cellular alterations in liver) | | HC, 2011a (based on Serota et al., 1986a) |
| | Inhalation TC | 6.0E-01 mg/m ³ | <p>Study Type: chronic</p> <p>Species: Sprague-Dawley rats</p> <p>Mode of Administration: inhalation (whole body exposure chambers)</p> <p>Exposure Regime: 0, 50, 200, or 500 ppm (equivalent to 0, 174, 695, or 1737 mg/m³) dichloromethane (> 99.5% pure) for 6 hours/day, 5 days/week</p> <p>Duration: 2 years</p> <p>Uncertainty Factors: 30 (3.16 for intraspecies variability, 3.16 for interspecies variability, and 3 for database deficiencies)</p> | 1 st percentile HEC = 17.2 mg/m ³ | Rat PBPK model to estimate rat internal dose (BMDL ₁₀) + Adjustment to a human equivalent internal BMDL ₁₀ + Conversion to an HEC using a human PBPK model TC = 1 st percentile HEC/UF | Hepatotoxicity (hepatic vacuolation) | CEPA: Group II probably carcinogenic to humans (EC and HC, 1993e) IARC: Group 2A probably carcinogenic to humans (IARC, 2017) US EPA IRIS: carcinogenic by a mutagenic mode of action (US EPA, 2011a) | HC, 2024b (based on US EPA, 2011a [derived from Nitschke et al., 1988]) |
| | Oral SF [US EPA (2011a): do not use with exposures >60 mg/kg _{BW} -day; apply ADAFs ² to the oral SF for early life exposures] | 2.0E-03 (mg/kg _{BW} -day) ⁻¹ | <p>Study Type: chronic</p> <p>Species: male and female B6C3F1 mice</p> <p>Mode of Administration: oral (drinking water)</p> <p>Exposure Regime: 0, 60, 125, 185, or 250 mg/kg_{BW}-day (in deionized drinking water)</p> <p>Duration: 104 weeks</p> <p>Uncertainty Factors: N/A</p> | BMDL ₁₀ = 60 mg/kg _{BW} -day | BMDL ₁₀ estimated using a linearized multistage model Converted to human exposure using BW scaling (bw ^{3/4}) and Oral SF determined by linear extrapolation Oral SF calculated from adult exposure data and does not reflect presumed early-life susceptibility | Cancer (liver [hepatocellular carcinomas or adenomas]) | | US EPA, 2011a (based on Serota et al., 1986b and Hazleton Laboratories, 1983) |

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|---|---|---|--|---|--|---|--|---|
| Dichloromethane (methylene chloride) | Inhalation UR [US EPA (2011a): do not use with exposures exceeding >7700 mg/m ³ ; apply ADAFs ² for early life exposures] | 1.0E-05 (mg/m ³) ⁻¹ | <p>Study Type: chronic</p> <p>Species: male B6C3F1 mice</p> <p>Mode of Administration: inhalation (whole body exposure chambers)</p> <p>Exposure Regime: 0, 2000 or 4000 ppm (approximately 0, 7000, 14 000 mg/m³); 6 hours/day, 5 days/week</p> <p>Duration: 2 years</p> <p>Uncertainty Factors: N/A</p> | <p>BMDL₁₀ (mouse liver tumours) = 544.4 mg/m³</p> <p>BMDL₁₀ (mouse lung tumours) = 48.6 mg/m³</p> | <p>PBPK model to estimate internal mouse dose + Multistage dose-response model to determine mouse BMDL₁₀ values for combined liver and lung tumours + Allometric scaling to convert mouse BMDL₁₀ values to human equivalent BMDL₁₀ values + Probabilistic human PBPK model to determine distribution of internal human doses + corresponding inhalation URs expressed as external concentrations</p> <p>Inhalation UR based on combined risk for liver and lung tumours</p> | Cancer (liver and lung) [hepatocellular and bronchoalveolar carcinomas or adenomas] | <p>CEPA: Group II probably carcinogenic to humans (EC and HC, 1993e)</p> <p>IARC: Group 2A probably carcinogenic to humans (IARC, 2017)</p> <p>US EPA IRIS: carcinogenic by a mutagenic mode of action (US EPA, 2011a)</p> | US EPA, 2011a (based on NTP, 1986b and Mennear et al., 1988) |
| | Ethylbenzene | Oral TDI | 2.2E-02 mg/kg _{bw} -day | <p>Study Type: chronic</p> <p>Species: B6C3F1 mice</p> <p>Mode of Administration: inhalation (whole body exposure chambers)</p> <p>Exposure Regime: 0, 75, 250, or 750 ppm (0, 330, 1100, or 3300 mg/m³) for 6 hours/day, 5 days/week</p> <p>Duration: 103 weeks</p> <p>Uncertainty Factors: 25 (10 for intraspecies variability and 2.5 for interspecies variability)</p> | NOAEL = 330 mg/m ³ (75 ppm) | <p>Mouse PBPK model to estimate mouse internal liver concentration corresponding to NOAEL = 0.08 mg/L; human PBPK model used to estimate corresponding external oral dose in humans assuming 1.5L/day drinking water consumption = 0.54 mg/kg_{bw}-day</p> <p>TDI = human external oral dose/UF</p> | Pituitary gland toxicity (hyperplasia) and Hepatototoxicity (cellular alterations of the liver) | <p>CEPA: see 2016 CMP assessment (ECCC and HC, 2016)</p> <p>IARC: Group 2B possibly carcinogenic to humans (IARC, 2000)</p> |
| Inhalation TC | | 2.0E+00 mg/m ³ | <p>Study Type: chronic</p> <p>Species: male and female F344/N rats and B6C3F1 mice</p> <p>Mode of Administration: inhalation</p> <p>Exposure Regime: 0, 75, 250, or 750 ppm (0, 330, 1100, or 3300 mg/m³), 6 hours/day, 5 days/week</p> <p>Duration: 104 weeks (rats); 103 weeks (mice)</p> <p>Uncertainty Factors: 30 (10 for intraspecies variability and 3 for interspecies variability)</p> | NOAEL = 75 ppm (330 mg/m ³) | <p>NOAEL adjusted for continuous exposure</p> <p>NOAEL_{adj} = 57 mg/m³ TC = NOAEL_{adj}/UF</p> | Pituitary gland toxicity (hyperplasia) and Hepatototoxicity (liver cellular alterations and necrosis) | <p>US EPA IRIS: Group D not classifiable as to human carcinogenicity (US EPA, 1998e)</p> | HC, 2024b (based on OEHHHA, 2000 [derived from NTP, 1999 and Chan et al., 1998]) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|-----------|--|-------------------------------------|---|---|---|---|--|---|
| n-Hexane | Oral TDI (provisional) | 1.0E-01 mg/kg _{BW} -day | Study Type: subchronic Species: rats Mode of Administration: gavage Exposure Regime: 0, 66, 132, or 264 mg/day, 7 days/week Duration: 4 weeks Uncertainty Factors: 90 (10 for intraspecies variability, 3 for interspecies variability, and 3 for deficiencies in the database) | POD = 8 mg/ kg _{BW} -day | TDI = POD/UF | Neurotoxicity (motor nerve conduction velocity, mixed nerve conduction velocity) | CEPA: not assessed IARC: not classified | CCME, 2011 (based on EEI, 2008 [derived from Ono et al., 1979, 1981]) |
| | Inhalation TC (provisional) | 7.0E-01 mg/m ³ | Study Type: subchronic Species: Wistar male rats Mode of Administration: inhalation (whole body exposure chambers) Exposure Regime: 0, 500, 1200, or 3000 ppm n-hexane vapour (0, 1762, 4230, or 10 574 mg/m ³), 12 hours/day, 7 days/week Duration: 16 weeks Uncertainty Factors: 300 (10 for intraspecies variability, 3 for interspecies variability, 3 for use of a subchronic study, and 3 for database deficiencies) | BMCL = 122 ppm (430 mg/m ³) | BMCL adjusted for continuous exposure BMCL _{HEC} = 215 mg/m ³ TC = BMCL _{HEC} /UF | Neurotoxicity (peripheral neuropathy – decreased motor nerve conduction velocity) | US EPA IRIS: inadequate information to assess carcinogenic potential (US EPA, 2005b) | US EPA, 2005b (based on Huang et al., 1989) |
| Lead | Risk-specific dose ⁴ (provisional) | 5.0E-04 mg/kg _{BW} -day | Study Type: epidemiological (meta-analysis) Species: humans Mode of Exposure: N/A Exposure Concentrations: N/A Duration: from birth or infancy until 5 to 10 years of age Uncertainty Factors: none | BMDL ₀₁ = 0.5 µg/ kg _{BW} -day | 95 th lower confidence limit of the BMD associated with a 1 IQ point decrement (intake rate associated with a drop of 1 IQ point in a population of children) Risk-specific dose = BMDL ₀₁ (no adjustment) | Neuro-developmental toxicity (cognitive function) | CEPA: not classified IARC: Group 2A probably carcinogenic to humans (IARC, 2006) US EPA IRIS: Group B2 probable human carcinogen (US EPA, 1988b) | EFSA, 2013 (based on Lanphear et al., 2005) |

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|-----------|-------------|-------------------------------------|---|---|-----------------------|-------------------------------------|---|---|
| Manganese | Oral TDI | 2.5E-02 mg/kg _{BW} -day | <p>Kern et al., 2010</p> <p>Study Type: neonatal exposure</p> <p>Species: Sprague-Dawley rats</p> <p>Mode of Administration: oral</p> <p>Exposure Regime: 0, 25, or 50 mg manganese/kg_{BW}-day in a sucrose solution for 21 days following birth (postnatal days [PND] 1–21)</p> <p>Duration: follow-up through PND 46</p> | LOAEL = 25 mg/ kg _{BW} -day | TDI = LOAEL/UF | Neuro- developmental toxicity | <p>CEPA: not assessed</p> <p>IARC: not assessed</p> <p>US EPA: Group D not classifiable as to human carcinogenicity (US EPA, 1988c)</p> | <p>HC, 2019b (based on Kern et al., 2010; Kern and Smith, 2011; Beaudin et al., 2013)</p> |
| | | | <p>Kern and Smith, 2011</p> <p>Study Type: neonatal exposure</p> <p>Species: Sprague-Dawley rats</p> <p>Mode of Administration: oral</p> <p>Exposure Regime: 0, 25, or 50 mg manganese/kg_{BW}-day in a sucrose solution for 21 days following birth (PND 1–21)</p> <p>Duration: sacrificed on PND 24 or observed to PND 107</p> | | | | | |
| | | | <p>Beaudin et al., 2013</p> <p>Study Type: adult and neonatal exposure</p> <p>Species: Long-Evans rats</p> <p>Mode of Administration: oral</p> <p>Exposure Regime: 0, 25, or 50 mg manganese/kg_{BW}-day in a stevia for 21 days following birth (PND 1–21) or through adulthood; oral manganese exposure post-weaning (PND 22 to end of study) via drinking water</p> <p>Duration: exposure during PND 1–21 or through adulthood</p> | | | | | |
| | | | <p>Uncertainty Factors: (for the three studies): 1000 (10 for intraspecies variability, 10 for interspecies variability, and 10 for the use of a LOAEL rather than a NOAEL)</p> | | | | | |

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|-----------|---------------|--|--|---|--|-----------------------------------|---|--|
| Manganese | Inhalation TC | 5.0E-05 mg/m ³ (in PM _{3.5}) | <p>Study Type: epidemiological (occupational)</p> <p>Species: humans</p> <p>Mode of Exposure: inhalation</p> <p>Exposure Concentrations: Average respiratory exposure ranged from 1.21–285.16 µg/m³, mean (±SD) and median values were 70.92±66.88 µg/m³ and 51.42 µg/m³, respectively</p> <p>Duration: 15 years (geometric mean)</p> <p>Uncertainty Factors: 100 (10 to account for interindividual variability, including possible enhanced susceptibility, 10 to account for database limitations)</p> | BMCL ₀₅ = 0.020 mg/m ³ (in PM _{3.5}) | <p>BMCL adjusted for continuous exposure</p> <p>BMCL₀₅ = 0.020 mg/m³</p> <p>TC = BMCL₀₅ × 0.238/ UF</p> | Neurofunctional outcome variables | <p>CEPA: not assessed</p> <p>IARC: not assessed</p> <p>US EPA: Group D not classifiable as to human carcinogenicity (US EPA, 1988c)</p> | HC, 2010c (based on Lucchini et al., 1999) |

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|---------------------|--|-------------------------------------|---|---|--|--|---|--|
| Mercury (inorganic) | Oral TDI [For exposure to mercury through consumption of food, use the TRV for methylmercury, the predominant form of mercury in these foods.] | 3.0E-04 mg/kg _{BW} -day | <p>Druet et al., 1978</p> <p>Study Type: subchronic</p> <p>Species: Brown Norway rats</p> <p>Mode of Administration: subcutaneous injection</p> <p>Exposure Regime: 0, 0.1, 0.25, 0.5, 1, and 2 mg/kg_{BW}, 3 times a week for 8 weeks; additional group at 0.05 mg/kg_{BW} for 12 weeks</p> <p>Duration: 8 or 12 weeks</p> | LOAEL = 0.226 mg/kg _{BW} -day (after conversion from subcutaneous to oral route) | <p>US EPA selected a drinking water equivalent level (DWEL) of 0.010 mg/L based on the three studies.</p> <p>US EPA used the DWEL to derive an RfD:</p> <p>Oral RfD = DWEL × IR_W / BW_{adult} [where IR_W = 2 L/day and BW_{adult} = 70 kg]</p> | Immunotoxicity (autoimmune glomerulonephritis) | <p>CEPA: not classified</p> <p>IARC: Group 3 not classifiable as to its carcinogenicity to humans (IARC, 1993)</p> <p>US EPA IRIS: Group D not classifiable as to human carcinogenicity (US EPA, 1995b)</p> | <p>CCME, 1999a,b and US EPA, 1995b (based on Druet et al., 1978; Bernaudin et al., 1981; Andres, 1984)</p> |
| | | | <p>Bernaudin et al., 1981</p> <p>Study Type: subchronic</p> <p>Species: Brown Norway rats</p> <p>Mode of Administration: gavage (food)</p> <p>Exposure Regime: 0 or 3 mg HgCl₂ (equivalent to 2.22 mg Hg)/kg_{BW} per week</p> <p>Duration: 60 days</p> | LOAEL = 0.317 mg/kg _{BW} -day | | | | |
| | | | <p>Andres, 1984</p> <p>Study Type: subchronic</p> <p>Species: Brown Norway rats and Lewis rats</p> <p>Mode of Administration: gavage (water)</p> <p>Exposure Regime: 3 mg HgCl₂ (equivalent to 2.22 mg Hg)/kg_{BW}, 2 times per week</p> <p>Duration: 60 days</p> | LOAEL = 0.633 mg/kg _{BW} -day | | | | |
| | | | <p>Uncertainty Factors: 1000 (10 for use of subchronic studies, 10 for intraspecies and interspecies variability, and 10 for LOAEL to NOAEL conversion)</p> | | | | | |

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|-----------------------|---------------------------|--|--|--|---|---|---|--|
| Methylmercury | Oral TDI (provisional) | 2.0E-04 mg/kg _{BW} -day (people of child-bearing age, infants, and children < 12 years) | Study Type: epidemiological Species: humans (children) Mode of Exposure: diet Estimated Exposure: daily intake estimated at 0.001 mg/kg _{BW} -day Duration: chronic (maternal exposure) Uncertainty Factors: 5 (see HC [2007] for details) | Approximate threshold of 10 µg/g mercury in maternal hair, corresponding to a dietary methylmercury intake level of 0.001 mg/kg _{BW} -day | TDI = dietary methylmercury intake level of 0.001 mg/kg _{BW} -day/UF | Neuro- developmental toxicity | CEPA: not assessed IARC: Group 2B possibly carcinogenic to humans (IARC, 1993) | HC, 2007 (based on Grandjean et al., 1997) |
| | | 4.7E-04 mg/kg _{BW} -day (non-sensitive adults of the general population) | Study Type: epidemiological Species: humans (children) Mode of Exposure: diet Exposure Concentrations: daily intake estimated at 0.0015 mg/ kg _{BW} -day Duration: chronic (maternal exposure) Uncertainty Factors: 6.4 (2 for interindividual variability in the hair: blood mercury ratio, and 3.16 [10 ^{0.5}] for inter-individual variability in the rate of elimination) | Average mercury concentration of 14 µg/g in maternal hair, corresponding to an estimated dietary methylmercury daily intake of 0.0015 mg/ kg _{BW} -day | FAO/WHO (2007) pTWI = provisional tolerable dietary methylmercury weekly intake (daily intake × 7 days/week)/UF= 0.0016 mg/kg _{BW} -week Provisional TDI = pTWI × 2 for non-sensitive adults of the general population/7 days in a week | Neuro- developmental toxicity | US EPA IRIS: Group C possibly carcinogenic to humans (US EPA, 1995c) | FAO/WHO, 2007 |
| Methylnaphthalene, 2- | Oral TDI | 4.0E-03 mg/kg _{BW} -day | Study Type: chronic Species: male and female B6C3F1 mice Mode of Administration: diet Exposure Regime: 0, 54.3, or 113.8 mg/kg _{BW} -day (males); 0, 50.3, or 107.6 mg/kg _{BW} -day (females) Duration: 81 weeks Uncertainty Factors: 1000 (10 for intraspecies variability, 10 for interspecies variability, and 10 for database deficiencies) | BMDL ₀₅ = 3.5 mg/ kg _{BW} -day | TDI = BMDL ₀₅ /UF (rounded to 4.0E-03 mg/kg _{BW} -day) | Respiratory tract toxicity (pulmonary alveolar proteinosis) | CEPA: not assessed IARC: not assessed US EPA IRIS: inadequate information to assess human carcinogenic potential (US EPA, 2003a) | US EPA, 2003a (based on Murata et al., 1997) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|-----------------|---------------|-------------------------------------|---|--|--|---|---|--|
| Naphthalene | Oral TDI | 2.0E-02 mg/kg _{bw} -day | <p>Study Type: subchronic</p> <p>Species: male and female F344 rats</p> <p>Mode of Administration: gavage (corn oil)</p> <p>Exposure Regime: 0, 25, 50, 100, 200, or 400 mg/kg_{bw}, 5 days/week</p> <p>Duration: 13 weeks</p> <p>Uncertainty Factors: 3000 (10 for intraspecies variability, 10 for interspecies variability, 10 for use of a subchronic study, and 3 for database deficiencies)</p> | NOAEL = 100 mg/kg _{bw} -day | NOAEL adjusted for continuous exposure NOAEL _{adj} = 71 mg/kg _{bw} -day TDI = NOAEL _{adj} /UF (rounded to 2.0E-02 mg/kg _{bw} -day) | Decreased body weight | CEPA: not assessed IARC: Group 2B possibly carcinogenic to humans (IARC, 2002) | US EPA, 1998f (based on BCL, 1980) |
| | Inhalation TC | 1.0E-02 mg/m ³ | <p>Study Type: chronic</p> <p>Species: F344 rats</p> <p>Mode of Administration: inhalation (whole body exposure chambers)</p> <p>Exposure Regime: 0, 10, 30, or 60 ppm (0, 52, 157, or 315 mg/m³) for 6 hours per day plus T₉₀ (12 minutes for the time to achieve 90% of the target concentration after vapour generation), 5 days per week</p> <p>Duration: 105 weeks</p> <p>Uncertainty Factors: 1000 (10 for intraspecies variability, 10 for interspecies variability, and 10 for database deficiencies)</p> | LOAEL = 52 mg/m ³ (10 ppm) | LOAEL adjusted for continuous exposure LOAEL _{adj} = 9.3 mg/m ³ (1.8 ppm) TC = LOAEL _{adj} /UF | Respiratory tract toxicity (nasal lesions [neuroblastoma of the olfactory epithelium, and adenoma of the respiratory epithelium of the nose]) | US EPA IRIS: Group C possibly carcinogenic to humans (US EPA, 1998f) | HC, 2013c (based on NTP, 2000) |
| Nickel chloride | Oral TDI | 1.3E-03 mg/kg _{bw} -day | <p>Study Type: reproductive</p> <p>Species: female Long-Evans rats</p> <p>Mode of Administration: oral (drinking water)</p> <p>Exposure Regime: 0, 10, 50, and 250 ppm Ni²⁺ (equivalent to 0, 1.3, 6.7, and 31.6 mg Ni²⁺/kg_{bw}-day, respectively)</p> <p>Duration: 11 weeks prior to mating (with unexposed males). Nickel administration continued through two successive gestation and lactation periods.</p> <p>Uncertainty Factors: 1000 (10 for intraspecies variability, 10 for interspecies variability, and 10 for use of a LOAEL instead of a NOAEL)</p> | LOAEL = 1.3 mg Ni ²⁺ /kg _{bw} -day | TDI = LOAEL/UF | Reproductive toxicity (perinatal death) | See nickel, mixture of oxidic, sulfidic and soluble inorganic nickel compounds | HC, 1996 (based on Smith et al., 1993) |

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|-------------------------------------|---------------|---------------------------|---|---|--|--|---|---|
| Nickel oxide | Inhalation TC | 2.5E-05 mg/m ³ | <p>Study Type: subchronic</p> <p>Species: Wistar rats</p> <p>Mode of Administration: inhalation (whole body exposure chambers)</p> <p>Exposure Regime: 0, 0.025 and 0.150 mg nickel/m³ as NiO aerosols, 24 hours/day, 7 days/week</p> <p>Duration: 4 months</p> <p>Uncertainty Factors: 1000 (10 for intraspecies variability, 10 for interspecies variability, and 10 for less than chronic study)</p> | LOAEL = 0.025 mg/m ³ | TC = LOAEL/UF | Respiratory tract toxicity (increase in the number of alveolar macrophages, increase in the size and number of macrophages with more than one nucleus, and an increase in phagocytic activity) | <p>CEPA: Group I carcinogenic to humans (EC and HC, 1994e)</p> <p>IARC: see nickel, mixture of oxidic, sulfidic and soluble inorganic nickel compounds</p> <p>US EPA IRIS: not classified</p> | HC, 1996 (based on Spiegelberg et al., 1984) |
| Nickel subsulfide (sulfidic nickel) | Inhalation TC | 1.8E-05 mg/m ³ | <p>Study Type: subchronic</p> <p>Species: F344/N rats and B6C3F1 mice</p> <p>Mode of Administration: inhalation (whole body exposure chambers)</p> <p>Exposure Regime: 0, 0.11, 0.22, 0.44, 0.88, and 1.8 mg nickel/m³, 6 hours/day, 5 days/week</p> <p>Duration: 13 weeks</p> <p>Uncertainty Factors: 1000 (10 for intraspecies variability, 10 for interspecies variability, and 10 for less than chronic study)</p> | LOAEL = 0.1 mg/m ³ | LOAEL adjusted for continuous exposure LOAEL _{adj} = 0.018 mg/m ³ TC = LOAEL _{adj} /UF | Respiratory tract toxicity (increase in number of alveolar macrophages, hyperplasia of alveolar macrophages) | <p>CEPA: Group I carcinogenic to humans (EC and HC, 1994e)</p> <p>IARC: see nickel, mixture of oxidic, sulfidic and soluble inorganic nickel compounds</p> <p>US EPA IRIS: Group A carcinogenic to humans (US EPA, 1987c)</p> | EC and HC, 1994e and HC, 1996 (based on Benson et al., 1990 and Dunnick et al., 1989) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|----------------|---------------|-------------------------------------|--|---|---|--|---|--|
| Nickel sulfate | Oral TDI | 1.2E-02 mg/kg _{BW} -day | <p>Study Type: epidemiological (human controlled studies)</p> <p>Species: humans (1st study [men] = 8 non-allergic volunteers; 2nd study [women] = 20 nickel-sensitive subjects and 20 non-allergic age-matched controls, both groups having existing vesicular hand eczema of the pompholyx type)</p> <p>Mode of Exposure: oral (drinking water) Exposure Concentrations: 12 µg nickel/kg_{BW} in drinking water (exposed subjects in both studies), followed by a 72-hour observation period</p> <p>Duration: N/A (single administration)</p> <p>Uncertainty Factors: none (LOAEL was based on a highly sensitive human population [WHO, 2007])</p> | LOAEL = 12 µg Ni/kg _{BW} -day | TDI = LOAEL | Dermal toxicity (exacerbation of eczema in nickel-sensitive subjects) | CEPA and IARC: see nickel, mixture of oxidic, sulfidic and soluble inorganic nickel compounds | CCME, 2015 (based on WHO, 2007 [derived from Nielsen et al., 1999]) |
| | Inhalation TC | 2.0E-05 mg/m ³ | <p>Study Type: chronic</p> <p>Species: male and female F344/N rats and B6C3F1 mice</p> <p>Mode of Administration: inhalation (whole body exposure chambers)</p> <p>Exposure Regime: rats: 0, 0.12, 0.25, or 0.5 mg nickel sulfate hexahydrate/m³ (equivalent to 0, 0.03, 0.06, or 0.11 mg nickel/m³); mice: 0, 0.25, 0.5, or 1 mg nickel sulfate hexahydrate/m³ (equivalent to 0, 0.06, 0.11, or 0.22 mg nickel/m³); about 6 hours/day, 5 days/week</p> <p>Duration: 104 weeks</p> <p>Uncertainty Factors: 1000 (10 for intraspecies variability, 10 for interspecies variability, and 10 for use of a LOAEL)</p> | LOAEL = 0.06 mg/m ³ | <p>LOAEL adjusted for continuous exposure</p> <p>LOAEL_{adj} = 0.011 mg/m³</p> <p>Intermediate TC = LOAEL_{adj}/UF = 1.1E-05 mg/m³</p> <p>Value of 0.02 µg/m³ was recommended as the European air quality standard based on soluble nickel compounds constituting <50% of total nickel compounds in ambient air</p> | Respiratory tract toxicity (lung inflammation [chronic active inflammation, macrophage and lymphoid hyperplasia, alveolar proteinosis, fibrosis, lung lesions, and atrophy of the olfactory epithelium]) | US EPA IRIS: not assessed | CCME, 2015 (based on ECB, 2008 and CSTE, 2001, derived from NTP, 1996) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|--|-----------------------------|---|---|---|---|---|---|---|
| Nickel, mixture of oxidic, sulfidic and soluble inorganic nickel compounds | Inhalation UR | 1.3E+00 (mg/m ³) ⁻¹ | <p>Study Type: epidemiological (occupational)</p> <p>Species: humans</p> <p>Mode of Exposure: inhalation</p> <p>Exposure Concentrations: N/A</p> <p>Duration: > 6 months</p> <p>Uncertainty Factors: N/A</p> | TC ₀₅ (5% tumourigenic concentration) = 0.04 mg/m ³ | Inhalation UR = 0.05/ TC ₀₅ where 0.05 = 5% extra cancer risk | Cancer (lung, nasal, kidney, prostate, buccal cavity) | <p>CEPA: classified as Group I carcinogenic to humans (EC and HC, 1994e)</p> <p>IARC: Group 1 classified as carcinogenic to humans (IARC, 2012a)</p> <p>US EPA: see individual nickel substances</p> | EC and HC, 1994e and HC, 1996 (based on Doll et al., 1990) |
| Nickel, metallic | Inhalation TC (provisional) | 1.8E-05 mg/m ³ | <p>Study Type: subchronic</p> <p>Species: rabbits</p> <p>Mode of Administration: inhalation</p> <p>Exposure Regime: 0, 0.13 mg/m³ metallic nickel dust, 6 hours/day, 5 days/week</p> <p>Duration: 4 and 8 months</p> <p>Uncertainty Factors: 1000 (10 for intraspecies variability, 10 for interspecies variability, and 10 for database deficiencies and a less than chronic study)</p> | LOAEL = 0.13 mg/m ³ | <p>LOAEL (rounded to 0.1 mg/m³) adjusted for continuous exposure</p> <p>LOAEL_{adj} = 0.018 mg/m³</p> <p>TC = LOAEL_{adj} /UF</p> | Respiratory tract toxicity (morphological and biological effects on alveolar cells) | <p>CEPA: Group VI unclassifiable with respect to carcinogenicity to humans (EC and HC, 1994e)</p> <p>IARC: Group 2B possibly carcinogenic to humans (IARC, 1990)</p> <p>US EPA IRIS: not classified</p> | EC and HC, 1994e and HC, 1996 (based on Johansson et al., 1983) |
| Polychlorinated biphenyls (PCBs) (non dioxin-like i.e., non-coplanar) ⁵ | Oral TDI (provisional) | 1.0E-05 mg/kg _{BW} -day (based on an Aroclor 1254 mixture) | <p>Study Type: chronic</p> <p>Species: female rhesus monkeys</p> <p>Mode of Administration: oral (ingestion of capsules containing Aroclor 1254 in a 1:1 glycerol/corn oil mixture)</p> <p>Exposure Regime: 0, 0.005, 0.02, 0.04, or 0.08 mg/kg_{BW}-day</p> <p>Duration: 23 months and 55 months (same group)</p> <p>Uncertainty Factors: 300 (10 for intraspecies variability, 3 for interspecies variability, and 10 to extrapolate from a LOAEL to a NOAEL)</p> | LOAEL for Aroclor 1254 = 0.005 mg/kg _{BW} -day | <p>As per Baars et al. (2001), TDI for mixture of non dioxin-like (i.e., non-coplanar PCBs) = 50% of the TDI of Aroclor 1254 (based on chemical analysis of seven "indicator PCBs" [PCB # 28, 52, 101, 118, 138, 153, and 180])</p> <p>LOAEL Aroclor 1254/UF = 1.7E-05 mg/kg_{BW}-day (rounded to 2.0E-05 mg/kg_{BW}-day)</p> <p>TDI Aroclor 1254 = 2.0E-05 mg/kg_{BW}-day × 50% = 1.0E-05 mg/kg_{BW}-day</p> | Immunotoxicity (decreased antibody response) | <p>CEPA: not classified</p> <p>IARC: not classified</p> <p>US EPA IRIS: not classified</p> | WHO, 2003 (based on Tryphonas et al., 1989, 1991), and Baars et al., 2001 |

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|---|------------------------|--------------------------------------|---|--|--|--|---|--|---|
| Polychlorinated biphenyls (PCBs)⁶ (dioxin-like, i.e. coplanar) | Oral TDI (provisional) | 2.3E-09 mg TEQ/kg _{bw} -day | See PCDDs/PCDFs for study details. Dioxin-like (i.e., coplanar) PCBs should be evaluated with PCDDs/PCDFs, using appropriate TEFs (see Table 4). See PCDDs/ PCDFs for source information. | | | | | CEPA: not classified IARC: Group 1; dioxin-like (i.e., coplanar) PCBs classified as carcinogenic to humans (IARC, 2016) US EPA IRIS: Group B2 probably carcinogenic to humans (US EPA, 1996) | See PCDDs/PCDFs for source information. |
| Polychlorinated dibenzo-p-dioxins/ polychlorinated dibenzofurans⁶ (PCDDs/PCDFs) | Oral TDI (provisional) | 2.3E-09 mg TEQ/kg _{bw} -day | Faqi and Chahoud, 1998 Study Type: subchronic (developmental) Species: Wistar rats Mode of Administration: subcutaneous Exposure Regime and Duration: initial doses of 0, 25, 60, or 300 ng tetrachlorodibenzo-p-dioxin (TCDD)/kg _{bw} , followed by weekly maintenance doses at 0, 5, 12, or 60 ng TCDD/kg _{bw} beginning 2 weeks prior to mating and continuing through mating, gestation, and lactation Uncertainty Factors: 9.6 (3 for use of a LOAEL rather than a NOAEL, and 3.2 for intraspecies variability) | LOAEL (maternal body burden) = 25 ng/kg _{bw} -day | BMD modeling to extrapolate NOAEL and LOAEL based on maternal body burden, to estimate equivalent monthly human intakes (EHMIs) pTMI = EHMI/UF Range of pTMIs = 40–100 pg/kg _{bw} -month mid-point of pTMI range = 70 pg/kg _{bw} -month pTDI = pTMI/30 days per month | Developmental toxicity (decreased sperm production and altered sexual behaviour in male offspring) | CEPA: not assessed IARC: Group 3 not classifiable as to carcinogenicity to humans for PCDDs (other than 2,3,7,8-TCDD and 2,3,4,7,8-PCDF) (IARC, 1997) IARC: Group 1, carcinogenic to humans for 2,3,7,8-TCDD and 2,3,4,7,8-PCDF (IARC, 2012b) | WHO, 2002b (based on Faqi and Chahoud, 1998 and Ohsako et al., 2001) | |
| | | | Ohsako et al., 2001 Study Type: subchronic (developmental) Species: pregnant Holtzman rats Mode of Administration: single oral bolus dose by gavage on day 15 of gestation Exposure Regime and Duration: single bolus dose (0, 12.5, 50, 200, or 800 ng 2,3,7,8-TCDD/ kg _{bw}) on day 15 of gestation Uncertainty Factors: 3.2 for intraspecies variability | NOAEL (maternal body burden) = 13 ng/kg _{bw} -day | | Developmental toxicity (decrease of ventral prostate weight and anogenital distance in male offspring) | US EPA: Group B2 probable human carcinogen for HxCDD; other PCDDs/ PCDFs not assessed (US EPA, 1987d) | | |

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|-----------|---|---|---|--|---|--|---|--|
| Pyrene | Oral TDI | 3.0E-02 mg/kg _{BW} -day | <p>Study Type: subchronic</p> <p>Species: male and female CD-1 mice</p> <p>Mode of Administration: gavage (corn oil)</p> <p>Exposure Regime: 0, 75, 125, or 250 mg/kg_{BW}-day</p> <p>Duration: 13 weeks</p> <p>Uncertainty Factors: 3000 (10 for intraspecies variability, 10 for interspecies variability, 10 for a less than chronic study, and 3 for database deficiencies)</p> | NOAEL = 75 mg/kg _{BW} -day | TDI = NOAEL/UF (TDI rounded to 3.0E-02 mg/kg _{BW} -day) | Nephrotoxicity (renal tubular pathology [lesions], decreased kidney weights) | <p>CEPA: not assessed</p> <p>IARC: Group 3 not classifiable as to its carcinogenicity to humans (IARC, 2010)</p> <p>US EPA: Group D not classifiable as to human carcinogenicity (US EPA, 1990c)</p> | US EPA, 1990c (based on US EPA, 1989) |
| Selenium | UL (HC) 0 to <6 mo 6 mo to <5 yrs 5 to <12 yrs 12 to <20 yrs ≥20 yrs | mg/kg _{BW} -day 5.5E-03 6.0E-03 6.3E-03 6.2E-03 5.7E-03 | <p>Yang and Zhou, 1994 (adults)</p> <p>Study Type: epidemiological</p> <p>Species: humans (adults)</p> <p>Mode of Exposure: dietary intake</p> <p>Exposure Concentrations: initial estimated range of intake: 913 to 1907 µg/day; range of intake during re-examination (8 years later): 654 to 952 µg/day</p> <p>Duration: chronic</p> <p>Uncertainty Factors: 2 (to protect sensitive individuals)</p> | NOAEL = 800 µg/day (mean selenium intake upon re-examination) (adults) | UL (IOM) = NOAEL/UF IOM adult ULs were adjusted to account for differences in HC's adult age group (HC, 2010a) | Hair and nail brittleness and loss (signs and symptoms of chronic selenosis) | CEPA: see 2017 CMP assessment (ECCC and HC, 2017) | IOM, 2000 (based on Yang and Zhou, 1994 and Shearer and Hadjimarkos, 1975) |
| | | | <p>Shearer and Hadjimarkos, 1975 (infants, children, and adolescents)</p> <p>Study Type: epidemiological</p> <p>Species: humans (infants, 0–6 months of age)</p> <p>Mode of Exposure: diet (human milk)</p> <p>Exposure Concentrations: selenium concentration of human milk of unsupplemented women ranged from 7 to 60 µg/L (average of 18 µg/L)</p> <p>Duration: N/A (stage of lactation ranged from 17 to 869 days)</p> <p>Uncertainty Factors: 1 (because of a lack of evidence that maternal intake associated with a human milk level of 60 µg selenium/L results in infant or maternal toxicity)</p> | NOAEL = 60 µg/L (infants) | NOAEL adjusted for estimated average human milk intake of 0.78 L/day NOAEL _{adj} = 47 µg/day (rounded to 45 µg/day) Infant UL (IOM) = NOAEL _{adj} /UF IOM derived ULs for older infants, children, and adolescents based on the infant UL and relative body weight TRVs were calculated in mg/kg _{BW} -day for age groups in HC (2010a) guidance | No evidence of selenium toxicity | <p>CEPA: see 2017 CMP assessment (ECCC and HC, 2017)</p> <p>IARC: Group 3 not classifiable as to human carcinogenicity (IARC, 1987)</p> <p>US EPA IRIS: Group D not classifiable as to human carcinogenicity (US EPA, 1991)</p> | |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|------------------------------|-------------|-------------------------------------|--|--|--|----------------------------------|---|--|
| Tetrachloroethylene (PCE) | Oral TDI | 4.7E-03 mg/kg _{BW} -day | <p>Study Type: epidemiological (occupational)</p> <p>Species: humans</p> <p>Mode of Exposure: inhalation</p> <p>Exposure Concentrations: Two exposure groups. High exposure group (dry cleaners): exposure range = 0.38–31.19 ppm (2.6 to 211 mg/m³); mean 8-hour time-weighted average exposure = 7.27 ppm (49 mg/m³). Moderate exposure group (ironers): exposure range = 0.52–11.28 ppm (3.5 to 77 mg/m³); mean of 8-hour time-weighted average exposure = 4.8 ppm (33 mg/m³)</p> <p>Duration: 8.8 years (average)</p> <p>Uncertainty Factors: 1000 (10 for intraspecies variability, 10 to extrapolate from a less than lifetime exposure, and 10 for database deficiencies)</p> | <p>NOAEL = 4.8 ppm (33 mg/m³)</p> <p>BMD₁₀ = 7.2 ppm (49 mg/m³)</p> | <p>BMD power model BMD₁₀ = 6.6 ppm (45 mg/m³)</p> <p>PBPK model used to extrapolate from inhalation exposures to equivalent oral doses</p> <p>Peak kidney PCE concentrations used to estimate brain concentrations</p> <p>External dose associated with BMD₁₀ = 4.7 mg/kg_{BW}-day</p> <p>TDI = external dose associated with the BMD₁₀/UF</p> | Neurotoxicity (colour confusion) | <p>CEPA: Group III possibly carcinogenic to humans (EC and HC, 1993f)</p> <p>IARC: Group 2A probably carcinogenic to humans (IARC, 2014)</p> <p>US EPA IRIS: likely to be carcinogenic to humans (US EPA, 2012)</p> | HC, 2015 (based on Cavalleri et al., 1994) |

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|------------------------------|---------------|---------------------------|--|--|---|--|---|---|
| Tetrachloroethylene (PCE) | Inhalation TC | 4.0E-02 mg/m ³ | <p>Cavalleri et al., 1994 Study Type: epidemiological (occupational) Species: humans Mode of Exposure: inhalation Exposure Concentrations: Two exposure groups. High exposure group (dry cleaners): exposure range = 0.38–31.19 ppm (2.6 to 211 mg/m³); mean 8-hour time-weighted average exposure = 7.27 ppm (49 mg/m³). Moderate exposure group (ironers): range = 0.52–11.28 ppm (3.5 to 77 mg/m³); mean of 8-hour time-weighted average exposure level = 4.8 ppm (33 mg/m³). Duration: 8.8 years (average) Uncertainty Factors: 1000 (10 for intraspecies variability, 10 for uncertainties in extrapolating from a LOAEL to a NOAEL, and 10 for database uncertainties)</p> <p>Echeverria et al., 1995 Study Type: epidemiological (occupational) Species: humans Mode of Exposure: inhalation Exposure Concentrations: Three exposure zones identified for counter clerks, pressers, and operators, corresponding to air levels of 11.2, 23.2, and 40.8 ppm respectively (76, 156, and 277 mg/m³) Duration: chronic Uncertainty Factors: 1000 (10 for intraspecies variability, 10 for uncertainties in extrapolating from a LOAEL to a NOAEL, and 10 for database uncertainties)</p> | <p>LOAEL (Cavalleri et al., 1994) = 42 mg/m³ (time-weighted average mean concentration of both exposure groups)</p> <p>LOAEL (Echeverria et al., 1995) = 156 mg/m³</p> | <p>LOAELs adjusted for continuous exposure and breathing rate</p> <p>LOAEL_{adj} (Cavalleri et al., 1994) = 15 mg/m³</p> <p>LOAEL_{adj} (Echeverria et al., 1995) = 56 mg/m³</p> <p>TC = midpoint of the range of LOAELs/UF = 0.04 mg/m³</p> | Neurotoxicity (alterations in reaction times, cognitive function, and colour vision) | <p>CEPA: Group III possibly carcinogenic to humans (EC and HC, 1993f)</p> <p>IARC: Group 2A probably carcinogenic to humans (IARC, 2014)</p> <p>US EPA IRIS: likely to be carcinogenic to humans (US EPA, 2012)</p> | HC, 2024b (based on US EPA, 2012 [derived from Cavalleri et al., 1994 and Echeverria et al., 1995]) |

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|------------------------------|---------------|---|---|---|---|---|--|---|
| Tetrachloroethylene (PCE) | Inhalation UR | 2.6E-04 (mg/m ³) ⁻¹ | <p>Study Type: chronic</p> <p>Species: male Crj:BDF1 mice</p> <p>Mode of Exposure: inhalation (whole body exposure chambers)</p> <p>Exposure Concentrations: 0, 10, 50, or 250 ppm; 6 hours/day, 5 days/week, for 104 weeks</p> <p>Duration: chronic</p> | BMDL ₁₀ = 57 ppm or 390 mg/m ³ | BMD modelling based on the lower bound on the dose associated with a 10% extra cancer risk in male mice, multistage model Unit risk determined by linear extrapolation and converted to human exposure using a harmonized PBPK model Inhalation UR = 0.01/BMDL ₁₀ Where 0.01 = 10% extra cancer risk | Cancer (liver [hepatocellular adenomas or carcinomas]) | <p>CEPA: Group III possibly carcinogenic to humans (EC and HC, 1993f)</p> <p>IARC: Group 2A probably carcinogenic to humans (IARC, 2014)</p> <p>US EPA IRIS: likely to be carcinogenic to humans (US EPA, 2012)</p> | HC, 2024b (based on US EPA, 2012 [derived from JISA, 1993]) |
| Toluene | Oral TDI | 9.7E-03 mg/kg _{BW} -day | <p>Study Type: epidemiological (occupational)</p> <p>Species: humans (printing shop workers)</p> <p>Mode of Exposure: inhalation</p> <p>Exposure Concentrations: high exposure group (106 subjects) = 26 ppm (98 mg/m³); low exposure group (86 subjects) = 3 ppm (11 mg/m³)</p> <p>Duration: long duration (21 years) and shorter duration (6 years)</p> <p>Uncertainty Factors: 10 for intraspecies variability</p> | NOAEL = 26 ppm (98 mg/m ³) | <p>PBPK modeling to estimate an internal toluene blood concentration following inhalation exposure = 0.0075 mg/L + Conversion to external oral human dose assuming ingestion of 1.5 L drinking water/day:</p> <p>NOAEL_{HEC} = 0.097 mg/kg_{BW}-day</p> <p>TDI = NOAEL_{HEC}/UF</p> | Neurotoxicity (cognitive function: attention, memory, and psychomotor function) | <p>CEPA: Group IV unlikely to be carcinogenic to humans (EC and HC, 1992b)</p> <p>IARC: Group 3 not classifiable as to its carcinogenicity to humans (IARC, 1999a)</p> <p>US EPA IRIS: inadequate information to assess carcinogenic potential (US EPA, 2005c)</p> | HC, 2014b (based on Seeber et al., 2004, 2005) |
| | Inhalation TC | 2.3E+00 mg/m ³ | <p>Study Type: epidemiological (occupational)</p> <p>Species: humans (printing shop workers)</p> <p>Mode of Exposure: inhalation</p> <p>Exposure Concentrations: high exposure group (106 subjects) = 26 ppm (98 mg/m³); low exposure group (86 subjects) = 3 ppm (11 mg/m³)</p> <p>Duration: long duration (21 years) and shorter duration (6 years)</p> <p>Uncertainty Factors: 10 (3.16 for pharmacokinetic variability and 3.16 for pharmacodynamics variability)</p> | NOAEL = 26 ppm (98 mg/m ³) | NOAEL adjusted for continuous exposure (assuming 8 hours/day, 5 days/week to 24 hours/day, 7 days/week) NOAEL _{adj} = 23 mg/m ³ TC = NOAEL _{adj} /UF | Neurotoxicity (cognitive function: attention, memory and psychomotor function) | US EPA IRIS: inadequate information to assess carcinogenic potential (US EPA, 2005c) | HC, 2011b (based on Seeber et al., 2004, 2005) |

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|----------------------------|---------------|--------------------------------------|--|--|---|--|--|--|
| Trichloroethylene (TCE) | Oral TDI | 1.46E-03 mg/kg _{BW} -day | <p>Study Type: subchronic (developmental)</p> <p>Species: Sprague-Dawley rats</p> <p>Mode of Administration: oral (drinking water)</p> <p>Exposure Regime: 0, 1.5, and 1100 ppm (equivalent to 0, 0.18, and 132 mg/kg_{BW}-day); 3 dosing regimes: 1) 3 months before pregnancy, 2) 2 months before and 21 days during pregnancy, or 3) 21 days during pregnancy only</p> <p>Duration: variable (see Exposure Regime above)</p> | BMDL ₁₀ = 0.146 mg/kg _{BW} -day | BMD model TDI = BMDL ₁₀ /UF | Developmental toxicity (fetal heart defects) | | HC, 2005 (based on Dawson et al., 1993) |
| | Inhalation TC | 2.0E-03 mg/m ³ | <p>Keil et al., 2009</p> <p>Study Type: chronic</p> <p>Species: female B6C3F1 mice</p> <p>Mode of Administration: oral (drinking water)</p> <p>Exposure Regime: 0, 1.4, and 14 ppm (0, 0.35, 3.5 mg/kg_{BW}-day)</p> <p>Duration: 30 weeks</p> <p>Uncertainty Factors: 100 (10 for extrapolating from a LOAEL rather than a NOAEL, 3 for intraspecies variability, and 3 for interspecies variability) [total uncertainty factor rounded to 100]</p> <p>Johnson et al., 2003</p> <p>Study Type: developmental</p> <p>Species: pregnant Sprague-Dawley rats</p> <p>Mode of Administration: oral (drinking water)</p> <p>Exposure Regime: 0, 0.0025, 0.25, 1.5, and 1100 ppm (0, 0.00045, 0.048, 0.218 or 129 mg/kg_{BW}-day) on gestational days 1 to 22</p> <p>Duration: 3 weeks during pregnancy</p> <p>Uncertainty Factors: 10 (3 for intraspecies variability, and 3 for interspecies variability) [total uncertainty factor rounded to 10]</p> | <p>LOAEL = 0.35 mg/kg_{BW}-day</p> <p>POD^{internal dose} (LOAEL) = 0.139 mg TCE metabolized / kg³/day</p> <p>HEC₉₉ LOAEL = 0.033 ppm (0.19 mg/m³)</p> <p>POD^{internal dose} = BMDL₀₁ = 0.0142 mg TCE metabolized by oxidation/kg³ / day</p> <p>HEC₉₉ BMDL₀₁ = 0.0037 ppm (0.021 mg/m³)</p> | <p>Candidate RfCs derived using a PBPK model integrating combined intraspecies, interspecies, and route-to-route extrapolation, and dividing by a UF.</p> <p>Candidate RfC (Keil et al., 2009) = 0.0019 mg/m³</p> <p>Candidate RfC (Johnson et al., 2003) = 0.0021 mg/m³</p> <p>Selected RfC = midpoint between the candidate RfCs = 0.002 mg/m³</p> | <p>Developmental toxicity (fetal heart malformations) and Immunotoxicity (decreased thymus weight)</p> | <p>CEPA: Group II probably carcinogenic to humans (EC and HC, 1993g)</p> <p>IARC: Group 1 carcinogenic to humans (IARC, 2014)</p> <p>US EPA IRIS: carcinogenic to humans (US EPA, 2011b)</p> | <p>US EPA, 2011b (based on Keil et al., 2009 and Johnson et al., 2003)</p> |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|----------------------------|--|--|--|--|--|---|--|---|
| Trichloroethylene (TCE) | Oral SF (As per HC [2013b, 2024a], given that TCE is known to have a mutagenic mode of action, HC recommends applying ADAFs ² to the oral SF when assessing risk associated with early life exposures at federal contaminated sites) | 8.11E-04 (mg/kg _{bw} -day) ⁻¹ | Study Type: chronic Species: male and female rats Mode of Administration: gavage Exposure Regime: 0, 500, and 1000 mg/kg _{bw} -day, 5 days/week Duration: 103 weeks Uncertainty Factors: N/A | Range of oral SFs: 5.82E-04 to 8.11E-04 (mg/kg _{bw} -day) ⁻¹ | Linearized multistage model and allometric scaling Most conservative oral SF: 8.11E-04 (mg/kg _{bw} -day) ⁻¹ | Cancer (kidney [combined tubular cell adenomas and adenocarcinomas]) | CEPA: Group II probably carcinogenic to humans (EC and HC, 1993g) IARC: Group 1 carcinogenic to humans (IARC, 2014) | HC, 2005 (based on NTP, 1988 and NTP, 1990) |
| | Inhalation UR (As per HC [2013b, 2024a], given that TCE is known to have a mutagenic mode of action, HC recommends applying ADAFs ² to the oral SF when assessing risk associated with early life exposures at federal contaminated sites) | 4.1E-03 (mg/m ³) ⁻¹ | Study Type: epidemiological (occupational) Species: humans Mode of Exposure: inhalation Exposure Concentrations: N/A Duration: chronic Uncertainty Factors: N/A | LEC ₀₁ = 2.4 mg/m ³ (lowest effective concentration) | Linear low-dose extrapolation from the LEC ₀₁ (95% lower bound on the exposure associated with a 1% extra cancer risk) + Application of a factor of 4 to include non-Hodgkin's lymphoma and liver cancer risks Inhalation UR = 0.01 / LEC ₀₁ where 0.01 = 1% extra cancer risk | Cancer (liver, kidney [renal cell carcinoma], non-Hodgkin's lymphoma) | US EPA IRIS: carcinogenic to humans (US EPA, 2011b) | US EPA, 2011b (based on Charbotel et al., 2006 and Raaschou-Nielsen et al., 2003) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|------------------------------|-------------|-------------------------------------|--|---|-----------------------|--------------------------------|---|--|
| Uranium (non-radioactive) | Oral TDI | 6.0E-04 mg/kg _{BW} -day | <p>Study Type: subchronic</p> <p>Species: male and female Sprague-Dawley rats</p> <p>Mode of Administration: oral (drinking water)</p> <p>Exposure Regime: 0, 0.96, 4.8, 24, 120, or 600 mg/L uranyl nitrate hexahydrate (equivalent to uranium doses of 0, 0.06, 0.31, 1.52, 7.54, and 36.73 mg/kg_{BW}-day in male rats, and 0, 0.09, 0.42, 2.01, 9.98, and 53.56 mg/kg_{BW}-day in female rats)</p> <p>Duration: 91 days</p> <p>Uncertainty Factors: 100 (10 for intraspecies variability and 10 for interspecies variability).</p> <p>Other points of departure (including a NOAEL) were higher than the LOAEL identified by Gilman et al. (1998) and therefore HC (2019c) did not consider an uncertainty factor for use of a LOAEL instead of a NOAEL necessary. HC (2019c) did not apply an uncertainty factor for use of a subchronic study because the study was considered adequately sensitive since the observed kidney effects were similar to the minimal effects seen in a number of longer-term studies and since human absorption values were found to be independent of exposure duration (HC, 2019c).</p> | LOAEL = 0.06 mg/kg _{BW} -day (males) | TDI = LOAEL/UF | Nephrotoxicity (renal lesions) | CEPA: not assessed IARC: not assessed US EPA IRIS: not assessed | HC, 2019c (based on Gilman et al., 1998) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|----------------|--|--|---|--|---|---|---|---|
| Vinyl chloride | Oral SF | 2.4E-01 (mg/kg _{bw} -day) ⁻¹ for continuous lifetime exposure during adulthood 4.8E-01 (mg/kg _{bw} -day) ⁻¹ for continuous lifetime exposure from birth | Study Type: chronic Species: male and female rats Mode of Administration: diet (mixture of vinyl chloride monomer [VCM] and polyvinyl chloride [PVC] powder) Exposure Regime: 0, 1.7, 5.0, or 14.1 mg/kg _{bw} -day, 4 hour feeding period/day; a positive control group was administered 300 mg/ kg _{bw} CM in soybean oil by stomach tube, 5 days/week Duration: lifetime (the experiment was terminated once 75% mortality was observed in the positive control group, i.e. 135 weeks for males and 144 weeks for females) Uncertainty Factors: N/A | External human dose associated with an excess lifetime risk of 10 ⁻⁵ for combined liver cancers = 4.19E-05 mg/kg _{bw} -day | Rat PBPK model to determine daily internal doses of vinyl chloride liver metabolites + Multistage model to determine a POD + Human PBPK model to estimate external doses Oral SF = 10 ⁻⁵ /external human dose Given animal evidence of early-life sensitivity to vinyl chloride, a factor of 2 was applied to the oral SF of 2.4E-01 (mg/kg _{bw} -day) ⁻¹ for exposure during adulthood, to account for continuous lifetime exposure from birth (HC, 2013d). | Cancer (liver [hepatocellular angiosarcomas and carcinomas]) | CEPA: known human carcinogen (EC and HC, 2016) IARC: Group 1 carcinogenic to humans (IARC, 2012b) | HC, 2013d (based on Feron et al., 1981) |
| | Inhalation UR [US EPA (2000b): should not be used for exposures >10 mg/m ³] | 4.4E-03 (mg/m ³) ⁻¹ for continuous lifetime exposure during adulthood 8.8E-03 (mg/m ³) ⁻¹ for continuous lifetime exposure from birth | Study Type: chronic Species: Sprague-Dawley female rats Mode of Administration: inhalation (whole body exposure chambers) Exposure Regime: 0, 1, 5, 10, 25, 50, 100, 150, 200, 250, 500, 2500, 6000, or 10 000 ppm (0, 2.6, 12.8, 25.6, 63.9, 128, 256, 383, 511, 639, 1278, 6390, 15 340, 25 560 mg/ m ³) vinyl chloride, 4 hours/day, 5 days/week Duration: 52 weeks Uncertainty Factors: N/A A two-fold safety factor was applied to the adult value to account for continuous lifetime exposure from birth. | Based on the 95% upper confidence limit on excess cancer risk in female rats | PBPK model and linearized multistage model | Cancer (liver [angiosarcomas, angiomas, hepatomas, and neoplastic nodules]) | US EPA IRIS: Group A carcinogenic to humans (US EPA, 2000b) | US EPA, 2000b (based on Maltoni et al., 1981 and 1984) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|------------------------|---------------|-------------------------------------|---|--|---|---|--|---|
| Xylenes, mixed isomers | Oral TDI | 1.3E-02 mg/kg _{bw} -day | <p>Study Type: subchronic</p> <p>Species: male Wistar rats</p> <p>Mode of Administration: inhalation (whole body exposure chambers)</p> <p>Exposure Regime: a control group and 3 exposure groups: 1) m-xylene concentrations of 50 ppm (217 mg/m³) and 100 ppm (435 mg/m³), or 2) n-butyl alcohol 50 ppm (154 mg/m³) and 100 ppm (308 mg/m³), or 3) a 1:1 mixture of m-xylene and n-butyl alcohol (100 ppm [217 mg/m³ m-xylene + 154 mg/m³ of n-butyl alcohol] and 200 ppm [435 mg/m³ m-xylene + 308 mg/m³ of n-butyl alcohol]), 6 hours/day, 5 days/week</p> <p>Duration: 3 months</p> <p>Uncertainty Factors: 75 (10 for intraspecies variability, 2.5 for interspecies variability, and 3 for use of a subchronic study)</p> | NOAEL = 50 ppm (217 mg/m ³) m-xylene | <p>PBPK modeling to estimate internal rat blood concentration corresponding to NOAEL of 50 ppm = 0.138 mg/L + Conversion to external oral human dose using human PBPK model and assuming ingestion of 1.5 L drinking water/day</p> <p>NOAEL_{HEC} = 1.0 mg/kg_{bw}-day TDI = NOAEL_{HEC}/UF</p> | Neurotoxicity (impaired motor coordination) | <p>CEPA: Group IV unlikely to be carcinogenic to humans (EC and HC, 1993h)</p> <p>IARC: Group 3 not classifiable as to its carcinogenicity to humans (IARC, 1999a)</p> <p>US EPA IRIS: inadequate information to assess carcinogenic potential (US EPA, 2003b)</p> | HC, 2014b (based on Korsak et al., 1994) |
| | Inhalation TC | 1.0E-01 mg/m ³ | <p>Study Type: subchronic</p> <p>Species: male Wistar rats</p> <p>Mode of Administration: inhalation (whole body exposure chambers)</p> <p>Exposure Regime: a control group and 3 exposure groups: 1) m-xylene concentrations of 50 ppm (217 mg/m³) and 100 ppm (435 mg/m³), or 2) n-butyl alcohol 50 ppm (154 mg/m³) and 100 ppm (308 mg/m³), or 3) a 1:1 mixture of m-xylene and n-butyl alcohol (100 ppm [217 mg/m³ m-xylene + 154 mg/m³ of n-butyl alcohol] and 200 ppm [435 mg/m³ m-xylene + 308 mg/m³ of n-butyl alcohol]), 6 hours/day, 5 days/week</p> <p>Duration: 3 months</p> <p>Uncertainty Factors: 300 (10 for intraspecies human variability, 3 for interspecies variability, 3 for extrapolation from subchronic to chronic duration, and 3 for database uncertainties)</p> | NOAEL = 50 ppm (217 mg/m ³) | NOAEL adjusted for continuous exposure and difference in blood/gas partitioning in rats vs humans NOAEL _{HEC} = 39 mg/m ³ TC = NOAEL _{HEC} /UF | Neurotoxicity (impaired motor coordination) | <p>US EPA IRIS: inadequate information to assess carcinogenic potential (US EPA, 2003b)</p> | HC, 2024b (based on US EPA, 2003b [derived from Korsak et al., 1994]) |

| Substance | Type of TRV | TRV Value | Study Details | Threshold/ Non-threshold Endpoint | TRV Derivation Method | Critical Effect(s) | Carcinogenicity Classification | Source |
|-----------|---|---|---|---|---|--|--|--|
| Zinc | UL (HC) 0 to <6 mo 6 mo to <5 yrs 5 to <12 yrs 12 to <20 yrs ≥20 years | mg/kg _{BW} -day 4.9E-01 4.8E-01 5.1E-01 5.4E-01 5.7E-01 | <p>Yadrick et al., 1989 (adults)</p> <p>Study Type: epidemiological (prospective)</p> <p>Species: humans (adult women)</p> <p>Mode of Exposure: dietary supplements</p> <p>Exposure Concentrations: 10 mg/day (estimated dietary intake) + supplemental intake of 50 mg/day as zinc gluconate</p> <p>Duration: 10 weeks</p> <p>Uncertainty Factors: 1.5 (for intraspecies variability and extrapolation from a LOAEL to a NOAEL)</p> | LOAEL = 60 mg/day (adults) | <p>UL (IOM) = LOAEL/UF for intake of zinc from food, water, and supplements</p> <p>IOM adult ULs were adjusted to account for differences in HC's adult age group (HC, 2010a)</p> | Decrease in erythrocyte superoxide dismutase (ESOD) activity (sensitive indicator of copper status, reflecting copper utilization and the risk of copper deficiency) | <p>CEPA: see 2019 (draft) CMP assessment (ECCC and HC, 2019b)</p> <p>US EPA IRIS: Group D not classifiable as to human carcinogenicity (US EPA, 2005d)</p> | IOM, 2001 (based on Yadrick et al., 1989 and Walravens and Hambidge, 1976) |
| | | | <p>Walravens and Hambidge, 1976 (infants, children, and adolescents)</p> <p>Study Type: epidemiological (prospective)</p> <p>Species: humans (infants, 0–6 months)</p> <p>Mode of Exposure: dietary supplements</p> <p>Exposure Concentrations: control group: formula with 1.8 mg zinc/L; exposure group: formula with 1.8 mg zinc/L + supplement with 4 mg zinc/L (total of 5.8 mg zinc/L)</p> <p>Duration: 6 months</p> <p>Uncertainty Factors: 1 (because of a lack of evidence that formula intakes of 5.8 mg zinc/L result in infant toxicity)</p> | NOAEL = 5.8 mg/L (infants) | <p>NOAEL adjusted for estimated average human milk intake of 0.78 L/day</p> <p>NOAEL_{adj} = 4.5 mg/day</p> <p>Infant UL (IOM) = NOAEL_{adj}/UF = 4.5 mg/day (4 mg/day rounded down)</p> <p>IOM derived ULs for older infants, children, and adolescents based on the infant UL and relative body weight</p> <p>IOM ULs were adjusted to account for differences in HC's age groups (HC, 2010a)</p> | No effects of zinc on serum copper or cholesterol concentrations or other adverse effects were found | <p>CEPA: see 2019 (draft) CMP assessment (ECCC and HC, 2019b)</p> <p>US EPA IRIS: Group D not classifiable as to human carcinogenicity (US EPA, 2005d)</p> | IOM, 2001 (based on Yadrick et al., 1989 and Walravens and Hambidge, 1976) |

NOTES:

mg/kg_{BW}-day = milligrams per kilogram of body weight per day, (mg/kg_{BW}-day)⁻¹ = per milligram per kilogram of body weight per day, mg/m³ = milligrams per cubic metre, (mg/m³)⁻¹ = per milligram per cubic metre

N/A: not applicable

- ¹ The oral SF for arsenic (inorganic) is currently under review and may be subject to change in the future.
- ² Default adjusted ADAFs that HC recommends for federal contaminated site risk assessments of non-threshold carcinogens with a mutagenic mode of action are provided in HC (2013b).
- ³ HC recommends the oral TDI from EFSA (2014) as a provisional TRV for use in federal contaminated site risk assessments. Since Cr(III) can be present in various forms in the environment, EFSA (2014) should be consulted to confirm whether the provisional TRV is adequately representative. HC has not selected an inhalation TC for Cr(III). It is recommended that the TC for Cr(VI) be used as a provisional TRV.
- ⁴ HC has not derived a TRV for lead. Based on the available scientific literature, no threshold of effect could be established for the identified critical effect for lead (neurodevelopmental toxicity). HC (2013e) therefore recommended that lead be considered a non-threshold substance. The 95th percentile lower confidence limit of the benchmark dose (BMDL₀₁) associated with a population-wide 1 IQ point decrement from EFSA (2013) is recommended as a provisional TRV and is referred to as a risk specific dose in this document.
- ⁵ The unadjusted oral TDI of 2.0E-05 mg/kg_{BW}-day for PCBs (i.e., non dioxin-like) may be considered with the application of a congener specific approach and depending on the nature of the assessment, i.e., where there is a known point source of contamination that is confirmed analytically to be a mixture that is reasonably represented by Aroclor 1254.
- ⁶ PCDDs, PCDFs, and dioxin-like PCBs are assessed by converting their concentrations to units of 2,3,7,8-tetrachlorodibenzo-p-dioxin (TCDD) TEQs using TEFs. These TEFs are published in DeVito et al. (2024). The sum of the TEQs is then compared to the TDI for 2,3,7,8-TCDD.

TABLE 2: Recommended RPFs for Carcinogenic PAHs

| PAH | CAS No. | Benzo[a]Pyrene RPF ¹ |
|------------------------|----------|---------------------------------|
| Benzo[a]pyrene | 50-32-8 | 1 |
| Benzo[a]anthracene | 56-55-3 | 0.1 |
| Benzo[b]fluoranthene | 205-99-2 | 0.1 |
| Benzo[g,h,i]perylene | 191-24-2 | 0.01 |
| Benzo[j]fluoranthene | 205-82-3 | 0.1 |
| Benzo[k]fluoranthene | 207-08-9 | 0.1 |
| Chrysene | 218-01-9 | 0.01 |
| Dibenzo[a,h]anthracene | 53-70-3 | 1 |
| Indeno[1,2,3-cd]pyrene | 193-39-5 | 0.1 |

¹ The PAH RPFs in this table are those published in CCME (2010), and are recommended to evaluate the carcinogenic potential of PAH mixtures at federal contaminated sites.

TABLE 3: Provisional RPFs for Carcinogenic PAHs

| PAH | CAS No. | Provisional Benzo[a]pyrene RPF ¹ |
|----------------------------------|------------|---|
| Anthanthrene | 191-26-4 | 0.1 |
| Benzo[c]chrysene | 194-69-4 | 0.01 |
| Benzo[g]chrysene | 196-78-1 | 0.1 |
| Benzo[c]phenanthrene | 195-19-7 | 0.01 |
| Cyclopenta[c,d]pyrene | 27208-37-3 | 0.1 |
| Dibenzo[a,e]fluoranthene | 5385-75-1 | 1 |
| Dibenzo[a,e]pyrene | 192-65-4 | 1 |
| Dibenzo[a,h]pyrene | 189-64-0 | 1 |
| Dibenzo[a,i]pyrene | 189-55-9 | 1 |
| Dibenzo[a,l]pyrene | 191-30-0 | 100 |
| 9,10- Dimethylanthracene | 781-43-1 | 0.01 |
| 7,12- Dimethylbenzo[a]anthracene | 57-97-6 | 10 |
| 1,2- Dimethylbenzo[a]pyrene | 16757-85-0 | 1 |
| 1,6- Dimethylbenzo[a]pyrene | 16757-90-7 | 0.1 |
| 3,6- Dimethylbenzo[a]pyrene | 16757-91-8 | 1 |
| 4,5- Dimethylbenzo[a]pyrene | 16757-89-4 | 1 |
| 5,6- Dimethylchrysene | 3697-27-6 | 0.1 |
| 5,7- Dimethylchrysene | 52171-92-3 | 0.1 |
| 5,11- Dimethylchrysene | 14207-78-4 | 1 |
| 1,4- Dimethylphenanthrene | 22349-59-3 | 0.01 |
| 4,10- Dimethylphenanthrene | 23189-63-1 | 0.001 |
| 5- Ethylchrysene | 54986-62-8 | 0.1 |
| Fluoranthene | 206-44-0 | 0.001 |
| 7- Methylbenzo[a]anthracene | 2541-69-7 | 1 |
| 8- Methylbenzo[a]anthracene | 2381-31-9 | 1 |
| 9- Methylbenzo[a]anthracene | 2381-16-0 | 0.1 |
| 12- Methylbenzo[a]anthracene | 2422-79-9 | 0.1 |

| PAH | CAS No. | Provisional Benzo[a]pyrene RPF ¹ |
|---------------------------------|------------|---|
| 11- Methylbenzo[b]fluorene | 77969-74-5 | 0.01 |
| 1- Methylbenzo[a]pyrene | 40568-90-9 | 1 |
| 2- Methylbenzo[a]pyrene | 16757-82-7 | 1 |
| 3- Methylbenzo[a]pyrene | 16757-81-6 | 1 |
| 4- Methylbenzo[a]pyrene | 16757-83-8 | 1 |
| 5- Methylbenzo[a]pyrene | 31647-36-6 | 0.1 |
| 6- Methylbenzo[a]pyrene | 2381-39-7 | 0.1 |
| 11- Methylbenzo[a]pyrene | 16757-80-5 | 1 |
| 12- Methylbenzo[a]pyrene | 4514-19-6 | 1 |
| 5- Methylchrysene | 3697-24-3 | 1 |
| 6- Methylchrysene | 1705-85-7 | 0.1 |
| 2- Methylfluoranthene | 33543-31-6 | 0.001 |
| Phenanthrene | 85-01-8 | 0.001 |
| 2,9,10- Trimethylanthracene | 63018-94-0 | 0.01 |
| 2,3,9,10- Tetramethylanthracene | 66552-77-0 | 0.01 |

¹ The RPFs in this table are based on an analysis of available RPFs and scientific literature by EEI (2006), and may be used if these PAHs are measured at federal contaminated sites. The RPFs are relative to benzo[a]pyrene, provisional and based on limited data. RPFs based on more recent literature can be used with rationale.

TABLE 4: TEFs for PCDDs, PCDFs, and Dioxin-Like PCBs

| Substance | CAS No. | TEF ¹ |
|--|------------|------------------|
| Polychlorinated Dibenzo-p-dioxins | | |
| 2,3,7,8- Tetrachlorodibenzo-p-dioxin (TCDD) | 1746-01-6 | 1 |
| 1,2,3,7,8- Pentachlorodibenzo-p-dioxin (PeCDD) | 40321-76-4 | 0.4 |
| 1,2,3,4,7,8- Hexachlorodibenzo-p-dioxin (HxCDD) | 39227-28-6 | 0.09 |
| 1,2,3,6,7,8- Hexachlorodibenzo-p-dioxin (HxCDD) | 57653-85-7 | 0.07 |
| 1,2,3,7,8,9- Hexachlorodibenzo-p-dioxin (HxCDD) | 19408-74-3 | 0.05 |
| 1,2,3,4,6,7,8- Heptachlorodibenzo-p-dioxin (HpCDD) | 35822-46-9 | 0.05 |
| Octachlorodibenzo-p-dioxin (OCDD) | 3268-87-9 | 0.001 |
| Polychlorinated Dibenzofurans | | |
| 2,3,7,8- Tetrachlorodibenzofuran (TCDF) | 51207-31-9 | 0.07 |
| 1,2,3,7,8- Pentachlorodibenzofuran (PeCDF) | 57117-41-6 | 0.01 |
| 2,3,4,7,8- Pentachlorodibenzofuran (PeCDF) | 57117-31-4 | 0.1 |
| 1,2,3,4,7,8- Hexachlorodibenzofuran (HxCDF) | 70648-26-9 | 0.3 |
| 1,2,3,6,7,8- Hexachlorodibenzofuran (HxCDF) | 57117-44-9 | 0.09 |
| 1,2,3,7,8,9- Hexachlorodibenzofuran (HxCDF) | 72918-21-9 | 0.2 |
| 2,3,4,6,7,8- Hexachlorodibenzofuran (HxCDF) | 60851-34-5 | 0.1 |
| 1,2,3,4,6,7,8- Heptachlorodibenzofuran (HpCDF) | 67562-39-4 | 0.02 |
| 1,2,3,4,7,8,9- Heptachlorodibenzofuran (HpCDF) | 55673-89-7 | 0.1 |
| Octachlorodibenzofuran (OCDF) | 39001-02-0 | 0.002 |
| Non-ortho Substituted PCB Congeners | | |
| PCB 77 | 32598-13-3 | 0.0003 |
| PCB 81 | 70362-50-4 | 0.006 |

| Substance | CAS No. | TEF ¹ |
|---|------------|------------------|
| PCB 126 | 57465-28-8 | 0.05 |
| PCB 169 | 32774-16-6 | 0.005 |
| Mono-ortho Substituted PCB Congeners | | |
| PCB 105 | 32598-14-4 | 0.00003 |
| PCB 114 | 74472-37-0 | 0.00003 |
| PCB 118 | 31508-00-6 | 0.00003 |
| PCB 123 | 65510-44-3 | 0.00003 |
| PCB 156 | 38380-08-4 | 0.00003 |
| PCB 157 | 69782-90-7 | 0.00003 |
| PCB 167 | 52663-72-6 | 0.00003 |
| PCB 189 | 39635-31-9 | 0.00003 |

¹ Source: DeVito et al., 2024

TABLE 5: Recommended Dermal Relative Absorption Factors (RAF_{Derm})

| Substance | RAF _{Derm} ¹ | Substance | RAF _{Derm} ¹ |
|--------------------------------------|----------------------------------|------------------------|----------------------------------|
| Arsenic | 0.03 | n-Hexane ⁴ | 1 |
| Barium | 0.1 | Lead ⁵ | 0.006 |
| Benzene ² | 0.03 | Mercury ⁶ | 1 |
| Benzo[a]pyrene (B[a]P) ³ | 0.148 | Methylmercury | 0.06 |
| Beryllium | 0.1 | Nickel ⁷ | 0.09 |
| Cadmium | 0.01 | PAHs ³ | 0.148 |
| Carbon tetrachloride | 0.03 | PCBs | 0.14 |
| Chlorobenzene | 0.03 | PCDDs/PCDFs | 0.03 |
| Chromium, total | 0.1 | Selenium | 0.01 |
| Chromium, hexavalent | 0.1 | Tetrachloroethylene | 0.03 |
| Copper | 0.06 | Toluene | 0.03 |
| Dichlorobenzene, 1,2- (o-DCB) | 0.03 | Trichloroethylene | 0.03 |
| Dichlorobenzene, 1,4- (p-DCB) | 0.03 | Uranium | 0.1 |
| Dichloroethane, 1,2- | 0.03 | Vinyl chloride | 0.03 |
| Dichloroethylene, 1,1- | 0.03 | Xylenes, mixed isomers | 0.03 |
| Dichloromethane (methylene chloride) | 0.03 | Zinc | 0.1 |
| Ethylbenzene | 0.03 | | |

¹ RAF_{Derm} are those recommended by MOE (2011), unless otherwise noted.

² Unless otherwise indicated, the default value for volatile organic compounds (VOCs), including benzene, is 0.03 (MOE, 2011).

³ HC research on *in vitro* dermal absorption of B[a]P from commercial gardening soil spiked with 14C-B[a]P (Moody et al., 2007) identified a mean dermal absorption (total of receiver + skin depot) of 0.148 (14.8%) and is recommended as the dermal absorption of B[a]P from soil. Consistent with the MOE (2011) approach for other PAHs, the default RAF_{Derm} for all PAHs is the same as that for B[a]P (i.e., 0.148 or 14.8%).

⁴ No data regarding the relative dermal absorption of n-hexane were identified; therefore, an RAF_{Derm} of 1 is recommended, as per CCME (2011).

⁵ The dermal RAF for lead was determined by dividing 0.3% (absolute dermal absorption value [Moore et al., 1980]) by 50% (oral absorption of lead from food and water [US EPA 2007b]), i.e., 0.3% / 50% = 0.006 or 0.6%.

⁶ The RAF_{Derm} for mercury is based on the absolute dermal absorption (46.6%) in human skin (Moody et al., 2009), and is comparable to the range of oral absorption of mercuric chloride (HgCl₂) in water (30–40%) observed in male rats (Morcillo and Santamaria, 1995). Given the observed similarity in dermal and oral absorption of mercury, an RAF_{Derm} of 1 is recommended.

⁷ The RAF_{Derm} for nickel was determined by dividing 1.0% (absolute dermal absorption value [Moody et al., 2009]) by 11% (approximate oral bioavailability [Ishimatsu et al., 1995]), i.e., 1.0% / 11% = 0.09 or 9%.

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